

Idaho National Engineering Laboratory
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# BURYIT/ANALYZ: A Computer Package for Assessment of Radiological Risk of Low-Level Radioactive Waste Land Disposal 

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November 1984

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U.S. Nuclear Regulatory Commission

Washington, D.C. 20655
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National Technical Information Service
Springfield, Virginia 22161

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# BURYIT/ANALYZ: A COMPUTER PACKAGE FOR ASSESSMENT OF RADIOLOGICAL RISK OF LOW-LEVEL RADIOACTIVE WASTE LAND DISPOSAL 

James E. Fisher<br>Neil D. Cox<br>Corwin L. Atwood

Published November 1984

EG\&G Idaho, Inc.<br>Idaho Falls, Idaho 83415

## ABSTRACT

This report is a user's manual for a partially completed code for risk assessment of a low-level waste shallow-land burial site, to be used in the licensing of burial sites. This code is intended as a tool to be used for considering nuclide transport mechanisms, including atmospheric, groundwater, erosion, and infiltration to an underlying aquifer. It also calculates doses to individuals and the population through direct exposure, inhalation, and ingestion.

The methodology of the risk assessment is based primarily on the response surface method of uncertainty analysis. The parameters of a model for predicting dose commitment due to a release are treated as statistical variables in order to compute statistical distributions for various dose commitment contributions. The likelihood of a release is also accounted for by statistically evaluating the arithmetic product of the dose commitment distributions with the probability of release occurrence.

An example is given using the atmospheric transport pathway as modeled by a code called BURYIT. The framework for using other release pathways is described in this manual. Information on parameter uncertainties, reference site characteristics, and probabilities of release events is included.

## SUMMARY

The objective of this project is to provide the U.S. Nuclear Regulatory Commission (NRC) with a computer-implemented methodology for evaluating the public and operational risks due to radioactive releases from low-level waste. The scope of the risk methodology includes releases from a low-level burial site, a processing plant, and vehicles transporting the waste. The methodology is probabilistic in nature, allowing for the uncertainties in both the parameters of a computer model and the frequency of occurrence of release e ents. The methodology is intended to be used as a tool to assist in the licensing of low-level st al g'w-land burial sites.

The approach to the project was to base the probabilistic methodology upon a radionuclide transport modeling code developed for NRC by Science Applications, Inc. In this report, the code is called BURYIT. Numerous improvements to the code were made.

Because the atmospheric transport pathway was the first to be verified, it was the first to be treated probabilistically. Uncertainties in transport model and dose computation parameters were researched for use in the response surface method of uncertainty analysis. At the same time, an existing routine for automatically conducting response surface analysis (ANALYZ) was interfaced with BURYIT.

The result is a capability to generate uncertainty distributions for various dose commitment calculations, providing that a release has occurred. With these uncertainty distributions, it is possible to state a probability level for the event that a dose commitment would exceed any specified magnitude.

The next step in the methodology was to account for the frequency, or probability, of a release event. The risk of a release scenario is defined herein as the arithmetic product of dose commitment and probability of occurrence.

Since both the dose commitment and probability are subject to uncertainty, it is necessary to treat their preduct statistically. The result is a capability to generate an uncertainty distribution for the risk and, consequently, to make a probability statement that concerns risk exceeding a specified level. A way to graphically compare the risks from different scenarios is also provided.

Because of a reordering of funding priorities within NRC and consequent anticipated termination of this project after FY-1984, it was decided to produce a user's manual for the risk methodology in its current state. As funding becomes available, improvements are envisioned, to allow for easier input, more realistic modeling, and more output options.

## ACKNOWIEDGMENTS

This report could not have been written without the able assistance of S. J. Bengston, J. J. Einerson, and T. R. Meachum. Special thanks are due to
S. K. Rope for her dose factor research and tu K. L. Falconer for her research on reference site characteristics.

The authors are also grateful for the excellent conments from: P. E. Litteneker and S. K. Rope, Idaho Operations Office of the Department of Energy, and from B. C. Anderson and K. L. Smith, EG\&G Idaho, Inc.

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# BURYIT/ANALYZ: A COMPUTER PACKAGE FOR ASSESSMENT OF RADIOLOGICAL RISK OF LOW-LEVEL WASTE SHALLOW-LAND BUFIAL. 

\author{

1. INTRODUCTION
}

A computer program package is being evolved for calculating the radiological risk associated with shallow-land burial of low-level radioactive waste (LLW). The ultimate intended use of the package is as a tool for evaluating proposed or existing burial-site locations. The package includes two computer programs that are being adapted for the purpose of probabilisiic risk assessment. Risk assessment requires an estimate of the consequence of a release event, providing that the event has occurred, and an estimate of the likelihood, or probability, of occurrence of the event during some time interval. The arithmetic product of the consequence and the probability of occurrence is defined herein as the risk associated with the event.

The basic calculation of the radiological hazard (or dose commitment) due to a given release event, or scenario, is accomplished by the computer code BURYIT. The code ANAL YZ performs a statistical analysis of the results of BURYIT calculations for the scenario, and comptes the consequence and the risk distributions in terms of the population dose commitment.

> BURYIT ${ }^{1}$ was originally a deterministic code developed to calculate population doses resulting from nuclide transport from a burial site through one or more pathways. The pathways include atmospheric or groundwater transport of nuclides, occupational exposure during packaging, shipment, burial operations, or intruder (animal or human) contamination. The content of the waste material is categorized by nuclide amount and concentration. Thus, the code has the capability to select a nuclide inventory, which is representative of the type of waste expected, to calculate the migration of the inventory through the pathway of interest, and to calculate the radiological impact to the public. A separate release scenario ( e combination of patiways and release fractions) and nuclide inventory are chosen. At this time, the waste inventory categories and the release scenarios are the same as those described in Reference 1 , Volume 2 .


#### Abstract

The adaptation of BURYIT to a probabilistic risk assessment code involved the installation of a computer routine that added the capability of uncertainty analysis. An uncertainty analysis technique combines uncertainties in input parameters in a prescribed fashion in order to assign statistical limits to output responses. The driver program for BURYIT sets up a perturbation pattern, or design, for varying the inputs, runs BURYIT the required number of times with the inputs perturbed a cording to the design, and collects the responses (Je res) for each run on computer disk storage, for later use by ANALYZ.


ANALYZ was originally an uncertainty analysis code for calculating consequences of nuclear fuel thermal response during hypothetical accident scenarios. ${ }^{2}$ It read responses from nuclear fuel pin code calculations and used response surface methodology to calculate probability limits (or consequence levels) for nuclear fuel temperatures for the given hypothetical accident. The response surface method of uncertainty analysis has three steps. First, the code perturbation responses are fitted to a second-order polynomial equation, called the response surface equation, in terms of the variable input parameters. Next, second-order error propagation is used to determine the lower four statistical moments $a^{\prime}$ the response. Finally, these response moments are matched to an approximating probability density function so that the consequence of the calculated event can be determined at the specified probability value.

The adaptation of ANALYZ included minor modifications for compatibility with BURYIT responses and installation of a subprogram to perform the risk calculation.

A hypothetical risk calculation was done using the BURYIT/ANALYZ package for the atmospheric transport pathway. Fifteen variables were selected from Table 8 for demonstrating the probabilistic risk methodology.

Sections 2, 3, and 4 contain the theoretical description of the LLW risk methodology computer package. Included in Section 2 are the transport pathway models in BURYIT for the unsaturated soil column, the aquifer, the atmospheric, the soil erosion, and the direct radiation exposure pathways. Section 2 also discusses the population dose calculation subprogram. Section 3 contains an
explanation of the response surface method for uncertainty analysis. Section 4 gives the details of the risk calculation. Section 5 contains a description of input requirements. Section 6 contains the results of the atmospheric pathway methodology demonstration case and also describes the outputs generated by the computer package.

## 2. MODELS FOR NUCLIDE TRANSPORT PATHWAYS

In this section, the principal mathematical models for consequence analysis are discussed. There are seven computer subroutines: AQUIFER, ATMOS, DIRECT, DOSET, EROSIO, UNSAT and PUNC (see Table 1). The subroutine BURYIT directs the calls to these seven subroutines as required for each release scenario; hence, reference will be made to the BURYIT program hereafter. Table 2 gives the original sources of these subroutines.

Table 1. Principal subroutines used to model shallow-land burial consequences

| Subroutine | Function | Pathway |
| :--- | :--- | :--- |
| BURYIT | (Driver) |  |
| DOSET | Population and max- <br> imum individual doses <br> from radionuclides <br> released to the <br> biosphere | - |

There are some limitations of the BURYIT program that should be emphasized at the start of this discussion. Foremost is that only a single wind direction is modeled. This limitation affects both the ATMOS and EROSIO subroutines wherein multiple wind directions could have been allowed in principle. With a single wind direction, the population and the environs at risk are in a single sector of 22.5 degrees downwind.

Another limitation is that the infiltration of water into soil is modeled in a single vertical dimension. The groundwater may flow either upward or downward, but not laterally. The ultimate sink for downward flow is an unconfined squifer that uitimately becomes a source of water for human consumption.

Furthermore, there is no phenomenological basis for the consequence calculation for a number of postulated ru ease scenarios. These scenarios deal with the expo ure of a single person (usually) and are called h -rein "Other" scenarios. More will be said about these scenarios in Section 2.1.

As a wora of caution to those who must follow the details of the internal calculations, the reader is advised that the various parts of BURYIT utilize at least three different systems of units. However, efforts have been made to convert all required userinput data to the International System of Units (SI) with the only intended exception being temperature in degrees Celsius rather than Kelvin. Immediately upon being read in, the user-input is converted to the units appropriate for each algorithm, and annotations so indicate. All data files supplied with BURYIT ate in SI units.

### 2.1 Structure of BURYIT

In order to model a particular release scenario, BURYIT calls one or more of the seven principal models according to the pathway, or pathways, included in the scenario (see Table 1). These models include the three fundamental release mechanisms for radionuclides in disposed waste (Reference I).
2.1.1 Direct Radiation. Humans may be directly exposed to gamma radiation emanating from intact containers or from radionuclides in t'ie waste/soil

Table 2. Sources of principal subroutines used to model shallow-land burial consequences

Subroutine
Source

BURYIT Science Applications, Inc. (1981). ${ }^{1}$
AQUIFER W. V. Demier et al., GETOUT (Version GET005), Pacific Northwest Laboratory, PNL-2970 (1979) ${ }^{3}$

ATMOS J. R. Sagenderf et al., XOQ DOQ:Computer Program for the Meteorological Evaluation of Routine Effluent Releases at Nuclear Power St:*ions, NUREG/CR-2919, PNL-4380, September 1982.

DIRECT A. Foderaro, Photon Shielding Manual, Pennsylvania State University, 1976;5
T. Rockwell, III, Editor Reactor Shield Design Manwal, TID-7004, 1956. ${ }^{6}$

DOSET Science Applications, Inc., Based on NRC Regulatory Guide 1.109, 1977 and WASH-1400, 1975.

EROSIO N. D. Woodruff and F. H. Siddoway, WEROS, U.S. Department of Agriculture, 1965.7

PUNC Appendix A.
UNSAT B. Amirijafari and B. Cheney, HYDRO, Science Applications, Inc., Technical Report submitted to EG\&G Idaho, Inc., June 1979. Not published.
mixture. The dose resulting from this exposure is modeled with the subroutine DIRECT, designated in the program as Pathway 3.
2.1.2 Air. Air can transport radioactivity from exposed waste, either as vapor or as suspended solid particles. The atmospheric transport of nuclides is modeled with ATMOS (Pathway 2). The dose to humans from the intake of contaminated air and food and from cloud shine is modeled with DOSET (Pathway 4). BURYIT always calls DOSET following ATMOS in a modeling run. Therefore, the pathway-calling sequence in BURYIT is simply " 2 " for this scenario.

If the air-transport pathway also inclutes resuspended soil particles from the surface of the trenches, then the subroutine EROSIO (Pathway 5) is used to compute the airborne particulate concentration for ATMOS to use. BURYIT always calls ATMOS after EROSIO, and a call to DOSET then follows ATMOS. However, EROSIO must be preceded by the groundwater subroutine UNSAT (Pathway 9), which provides the nuclide concentration of the soil particles on the surface. The
pathway-calling sequence for transport of resuspended soil particles and vapor is, therefore, "95."
2.1.3 Water. Groundwater and surface water can act as transport agents for radionuclides; however, only groundwater transport is modeled using BURYIT. The infiltration of water into the top of a column of unsaturated soil is modeled with UNSAT (Pathway 9). The water percolates into a zone containing waste and dissolves radionuclides. Subsequently, the contaminated water may either return to the soil surface or continue percolating downward until an unconfined aquifer is reached. Evaporation at the surface leaves contaminated soil particles, transport of which is modeled by EROSIO (Pathway 5). Contaminated water entering the aquifer flows some distance befo e reaching a point of surface discharge or a well. Sat srated flow is modeled using AQUIFER (Pathway 1) DOSET, which computes the human dose due to ingestion of water, always follows AQUIFER. Thus, the prthway-calling sequence in BURYIT is " 91 " for in.iltration to an aquifer, and is "9." for infiltration coupled with air transport of contaminated soil. The calling sequence for the two processes together is "951."

The subroutines UNSAT and AQUIFER may be used to model a contaminated well or a contaminated river, but not both at the same time. In either case, the ingestion of nuclides is overestimated because the volumetric flow of the aquifer is not taken into account. It is assumed that the nuclides leaving the disposed waste enter a water supply source, which is at a user-specified distance from the center of the site. In the case of a well, a lower bound on water usage may be set at $7700 \mathrm{~m}^{3} / \mathrm{y}$ in order to account for low natural percolation in an arid, western site. Also, the number of people ingesting the contaminated water must depend on the type of water source being modeled
2.1.4 Puncture Wound. One release scenario in BURYIT pertains to a person suffering a puncture wound caused by a contaminated sharp object. A method for estimating the dose commitment, described in Appendix A, has been derived and installed in BURYIT. The pathway-calling sequence number for this scenario is simply " 6 ." The dose is calculated by the subroutine PUNC.
2.1.5 Other Scenarios. There are some release scenarios included within BURYIT that do not fall into the mechanisms described above. These include such events as theft of a usable item, package rupture during packaging operation, interim storage or transportation, and human or animal intrusion. A provision is made (Reference 1) for modeling these events as being, in an unexplained way, equivalent to the dose to a sir gle human resulting from a contaminated cloud t oving at a velocity not less than $2.2 \mathrm{~m} / \mathrm{s}(5 \mathrm{miles} / \mathrm{t})$. One supplies the dust cloud diameter and the wind velocity as input data. BURYIT then computes a pseudo-air concentration for input to DOSET which, in turn, calculates the dose to a single human. The pathway-calling sequence number foz this type of scenario is simply "4."

### 2.2 Subroutine DIRECT

The subroutine DIRECT was written to compute the external gamma whole body dose resulting from direct exposure to undispersed waste. One of three possible geometries is selected by the user and is input from TAPE20, Distances and up to five shielding materials are also input on TAPE20. The algorithms are based primarily on formulas from References 5 and 6. One difference is that the gamma-energy reatment is performed for 10 dis
crete energy groups. During verification activities, it was found that DIRECT compared favorably with the computer program ISOSHLD II, except for the nuclides cobalt- 60 and tritium. The results from DIRECT for these two nuclides were approximately one order of magnitude lower because DIRECT does not compute beta radiation.

### 2.3 DOSET and Associated Subroutines

This section applies to all pathways in which radionuclides are released to the environment. These are the pathways other than direct exposure and puncture wound.

The population and maximum individual doses are calculated in subroutine DOSET, and the results are cumulated in subroutine SUMDOS. The results are printed by OUTPUT, while PREDOS supplies concentration factors for food, milk, and meat.

The present version of DOSET utilizes two slightly different methods for computing dose commitments: (a) the original method described in Reference 1, and (b) a modified method. The ariginal method includes dose computations for eight organs and for three age groups. The modified method uses internal dose conversion factors (DCFs) that are a weighted sum of organ DCFs, resulting in effective whole body doses. The modified method computes the dose to adults only. The results from both methods are printed in order to facilitate a comparison of the two methods. In order to bypass one method or the order, minor rrogramming changes will be required.
2.3.1 Original DOSET. The dose is calculated for the total body and eight organs: bone, liver, kidney, gonad, lung, gastrointestinal tract, thyroid, and skin. Original DOSET involves two pathways for external exposure: nuclides in the air and nuclides on the ground. Original DOSET also involves two pathways for internal exposure: inhalation and ingestion. These latter two are age-dependent; thus, separate doses are computed for the child, teen, and adult age groups. This results in a total of eight pathway categories in Original DOSET, encompassing the organ and whole body dose commitments.

The methodology in Original DOSET is based partly on that in the NRC Regulatory Guide $1.109^{8}$ and partly on that in WASH-1400. ${ }^{9}$ DOSET differs in two principal respects: radionuclide release
is at a uniform rate over a specified time period rather than a continuing steady-state release, and the popu stion ingestion dose is based on the local production of contaminated foods (Reference 1, Volume 2, pp. 6-53, 54).

The Original DOSET was verified by updating the transfer constants and dose conversion factors using information more recent than that in Reference 1. In addition, manual calculations were used to verify the computational algorithms. This work is described in NUREG/CR-3210. ${ }^{10}$
2.3.2 Modified DOSET. There is some doubt that equal doses to organs of children and teens generate the same risks with respect to health effects as adult organ doses. As an alternative to age-specific doses, ICRP 30, Parts 1,2, and $3^{11}$ provides weighted or effective whole body DCFs for adults. These DCFs are based on the sum of organ doses that are weighted according to risk. This effective whole-body-dose methodology has been proposed by the NRC in a revision to 10 CFR Part 20, "Standards for Protection Against Radiation," and has been incorporated into DOSET. The DCFs are considered to be the best, currently available ones for risk assessments that stop at dose rather than estimating the health effects resulting from a unit dose. The effective whole body doses resulting from different scenarios should be easier to compare, because one dose value summarizes the dose to a number of organs.

### 2.4 ATMOS and Supporting Subroutines

The subroutine ATMOS and its supporting subroutines are based on recommended approximations to the time-integrated exposure resulting from airborne nuclides. ${ }^{12}$ The underlying theory is the Gaussian plume model, wherein advantage is taken of the fact that natural diffusion in the atmosphere is modeled well by a Gaussian distribution of airborne contaminants. As stated earlier, a single wind direction is permitted in the BURYIT package, although up to 10 wind speeds may be used (Reference 1). A plume-rise model that is said to be valid only for stable atmospheric conditions is installed in the current version. 13

[^0]plume rise
\[

$$
\begin{gathered}
\mathrm{H}(\mathrm{~L})=\mathrm{SH}+1.6\left(\mathrm{~F}_{\mathrm{b}}^{1 / 3} \times^{2 / 3}\right) / \mathrm{V}(\mathrm{~L}) \\
\mathrm{x} \leq 10000 \mathrm{~m}
\end{gathered}
$$
\]

final rise

$$
\begin{gather*}
\mathrm{H}(\mathrm{~L})=\mathrm{SH}+742.7\left(\mathrm{~F}_{\mathrm{b}}^{1 / 3}\right) / \mathrm{V}(\mathrm{~L}) \\
\mathrm{x}>10000 \mathrm{~m} \tag{1}
\end{gather*}
$$

where

$$
\left.\begin{array}{l}
\begin{array}{rl}
\mathrm{H}(\mathrm{~L})= & \begin{array}{l}
\text { plume height for each windspeed } \\
\text { index }(\mathrm{L}),(\mathrm{m})
\end{array} \\
\mathrm{SH}= & \text { stack height, }(\mathrm{m})
\end{array} \\
\begin{array}{rl}
\mathrm{F}_{\mathrm{b}}= & \text { buoyancy heat flux parameter, } \\
\left(\mathrm{m}^{4} / \mathrm{s}^{3}\right)
\end{array} \\
=
\end{array}\right\}
$$

This plume-rise model approximates observations downwind to perhaps 10000 m . The final-rise model is a constant beyond 10000 m .

Two comments may be beneficial to the user who wishes to study the internal programming. First, the dry deposition velocities are all set to $0.01 \mathrm{~m} / \mathrm{s}$ with the exception of the three noble gases $\mathrm{Kr}, \mathrm{Xe}$, and Rn , which are assigned values of zero. This is done in the absence of better information. Secondly, washout of the above three nuclides to the ground is assumed to occur during rainfall through solution. The relative effect on dose commitment induced by this assumption, as opposed to no ground deposition, has not been investigated.

### 2.5 UNSAT and Supporting Subroutines

UNSAT and its several supporting subroutines are used to model the infiltration of water in an
unsaturated coluinn of soil between the ground surface and an intersecting, unconfined aquifer. Lowlevel waste is buried in a layer of soil, the depth of which is user-specified. The underlying theory is Darcy's law with transient flow through multiple disparate soil layers and with hydraulic dispersion and ion exchange of nuclides with the soil allowed. The governing partial differential equation is given in Reference 1 (Volume 2, pp. 6-36). The finitedifference equation corresponding to the flow theory is consistent with published versions by Freeze and Cherry, Appendixes VI and VIII, p. 67. 14 The origin of the subroutine UNSAT is described in Reference 1, pp. 6-34.

A time-dependent boundary condition at the soil surface is used. Specifically, alternating wet and dry periods are allowed. Each wet period is modeled with the same constant rainfall rate, and each dry period is modeled with a constant evapotranspiration rate. Any other type of boundary condition, such as variable rainfall and evapotranspiration rates, would require modification of the program.

Each of the layers in the soil column may have distinct physical properties: hydraulic conductivity, density, thickness, and nuclide sorption characteristics. These properties are part of the input data as detailed in Section 5.2. However, hydraulic conductivity must be discussed in more detail. This property is read for the uppermost soil layer as a tabular function of the fractional water content (by volume). Upper and lower bounds on the applicability of the data are necessary and are established by water content input parameters WATL and WATH (Reference 1). (Also, a pressure head function is read in the same way.) For the second and subsequent soil layers, the hydraulic conductivity is computed by multiplying that of the first layer at the same moisture content by a constant (CONCOF) that is also supplied by the user for each soil layer. Thus, it is assumed th't the characteristic hydraulic conductivity-water content curves all have the same shape and are merely shifted, according to the value of CONCOF, relative to the curve for the first soil layer.

The nuclides dissolve in the groundwater contacting the waste in the specified burial (injection) layer. The nuclide concentration in the water leaving the burial lay $\%$ at each time step is assumed to be the solubility limit. A material balance on each nuclide is rationalized by computing the volume of waste in the injection layer using an emplacement efficiency
factor, EMEF, defined in NUREG/CR-1759, pp. 3-8, 10, 15 and in NUREG-0782, Appendixes E and F. ${ }^{16}$ This factor represents the ratio of volume of waste emplaced in the disposal cell to the total volume of available disposal space (soil plus waste volume). Recommended values for EMEF are 0.5 for random emplacement, and 0.75 for stacked emplacement (Reference 15).

The solubility limits of the nuclides should be elaborated on. This variable is called SOLFAC in the program and has units of curies per cubic metre of water. The user must input these values into the program in the same order as the nuclides in the waste inventory used (see Section 5.2.2.3). Suggested values for SOLFAC are listed in Table 3. Those values in groups I and II in the table are based on measurements at the Maxey Flats site (References 15 and 16). Values of the 23 nuclides in group III are guessed, pending other sources of information. Error factors (EF) are also listed such that one standard deviation uncertainty is given by (MEAN/EF; MEAN x EF). The EFs for the group I nuclides are derived from sample standard deviations. The EFs for groups II and III are subjective guesses that generally become larger as the perceived uncertainties of the estimates become larger. For instance, the group III error factors are generally orders of magnitude greater than those in group I.

The amount of each nuclide transported to the soil surface or to the underlying aquifer is computed for a unit area. In order to compute the total release of radioactivity, a separate trench area term, TRAR, is defined. This area term accounts for the desired area of infiltration and is a user-specified input having units of square metres (see Section 5.2.2.3). For example, the total radioactivity discharged to the aquifer is computed in the subroutine EXEC using TRAR. The term TRAR may represent the area of all trenches, one trench. or part of a trench as desired by the user.

### 2.6 Subroutine EROSIO

The EROSIO subroutine calculates the amount of radioactivity blown from the field surface of the site by the erosive action of the wind. The surface radioactivity is supplied by UNSAT, and it is assumed to be tr: asported to the surface via evaporation of water that has contacted the waste.

Table 3. SOLFAC values

Group $\quad$\begin{tabular}{l}
Element

 

Geometric Mean <br>
Solubility <br>
$\left(\mathrm{Ci} / \mathrm{m}^{3}\right)$

$\quad$

Error Factor <br>
(one standard deviation)
\end{tabular}

1 From Measurements, Reference 21

| ${ }^{3} \mathrm{H}$ | 0.17 | 10 |
| :--- | ---: | ---: |
| 60 Co | $5 \mathrm{E}-6$ | 10 |
| 90 Sr | $1 \mathrm{E}-4$ | 5 |
| ${ }^{137 \mathrm{Cc}}$ | $4 \mathrm{E}-6$ | 7 |
| ${ }^{238} \mathrm{Pu}$ | $9 \mathrm{E}-6$ | 7 |
| ${ }^{239} \mathrm{Pu}$ | $9 \mathrm{E}-6$ | 17 |
| ${ }^{24} \mathrm{Am}$ | $7 \mathrm{E}-7$ | 17 |
| 14 C | $3 \mathrm{E}-6$ | 13 |
| 238 U | $8 \mathrm{E}-9$ | 6 |
|  |  | 8 |

II Estimated in Reference 21 on basis of chemical similarity

| ${ }^{99} \mathrm{Tc}$ | 0.02 | 100 |
| :---: | :---: | :---: |
| ${ }^{129} 1$ | 0.02 | 100 |
| ${ }^{55} \mathrm{Fe}$ | 5E-6 | 25 |
| ${ }^{59} \mathrm{Ni}$ | 5E-6 | 25 |
| ${ }^{63} \mathrm{Ni}$ | 5E-6 | 25 |
| ${ }^{94} \mathrm{Nb}$ | $4 \mathrm{E}-6$ | 25 |
| ${ }^{135}$ Cs | 4E-6 | 7 |
| ${ }^{235}$ U | $8 \mathrm{E}-9$ | 8 |
| ${ }^{241} \mathrm{Pu}$ | 9E-6 | 17 |
| ${ }^{242} \mathrm{Pu}$ | $9 \mathrm{E}-6$ | 17 |
| ${ }^{237} \mathrm{~Np}$ | 9E-6 | 300 |
| ${ }^{243} \mathrm{Cm}$ | 9E-6 | 300 |
| ${ }^{244} \mathrm{Cm}$ | 9E-6 | 300 |
| ${ }^{243} \mathrm{Am}$ | $7 \mathrm{E}-7$ | 13 |

III Guesses based on nearest neighbor in periodic table

| 35 S | 0.02 | 1000 |
| :---: | :---: | :---: |
| ${ }^{51} \mathrm{Cr}$ | 0.02 | 1000 |
| ${ }^{54} \mathrm{Mn}$ | 0.02 | 1000 |
| ${ }^{58} \mathrm{Co}$ | 5E-6 | 5 |
| ${ }^{65} \mathrm{Zn}$ | $5 \mathrm{E}-6$ | 625 |
| ${ }^{95} \mathrm{Zr}$ | $4 \mathrm{E}-6$ | 625 |
| ${ }^{106} \mathrm{Ru}$ | 0.02 | 1000 |
| ${ }^{124} 5 \mathrm{Sb}$ | 0.02 | 1000 |
| ${ }^{125} 5$ Sb | 0.02 | 1000 |
| 125 | 0.02 | 100 |
| ${ }^{134}$ Cs | 4E-6 | 7 |
| 144 Ce | 8E-9 | 70 |
| ${ }^{152}$ Eu | 9E-6 | 90000 |
| ${ }^{154} \mathrm{Eu}$ | 9E-6 | 90000 |
| ${ }^{155} \mathrm{Eu}$ | $9 \mathrm{E}-6$ | 90000 |
| ${ }^{226} \mathrm{Ra}$ | IE-4 | 45 |
| ${ }^{230} \mathrm{Th}$ | 8E-9 | 70 |
| 232 Th | 8E-9 | 70 |
| ${ }^{240} \mathrm{Pu}$ | 9E-6 | 17 |
| ${ }^{242}$ Am | $7 \mathrm{E}-7$ | 13 |
| $2^{42} \mathrm{Cm}$ | $9 \mathrm{E}-6$ | 300 |

EROSIO then supplies ATMOS with activities of the nuclides for use in calculating air and ground concentrations of radioactivity.

The basis of EROSIO is the solution of the wind erosion equation, an empirical formulation of soil loss, E , in tons per acre per annum from a given agricultural field. This equation was developed by the U.S. Department of Agriculture to predict soil loss from the great plains region of the United States. ${ }^{17,7}$ The soil erosion is expressed in terms of five influential variables as
$E=f\left(I^{\prime}, K^{\prime}, C^{\prime}, L^{\prime}, V^{\prime}\right)$
where

$$
\begin{aligned}
& \mathrm{I}^{\prime}=\text { a soil erodibility index } \\
& \mathrm{K}^{\prime}=\text { a soil ridge roughness factor } \\
& \mathrm{C}^{\prime}=\text { a climatic factor } \\
& \mathrm{L}^{\prime}=\begin{array}{l}
\text { field length along the pr cvailing wind } \\
\text { erosion direction }
\end{array} \\
& \mathrm{V}^{\prime}=\begin{array}{l}
\text { equivalent quantity of vegetative } \\
\text { cover. }
\end{array}
\end{aligned}
$$

The computed soil loss, E (in tons/acre-year), is then multiplied, in consistent units, by the duration of the wind, the contaminated area, the fraction of soil lofted, and the nuclide concentrations in the soil. The nuclide concentrations are supplied by isotope and time period. In general, the UNSAT results consist of several blocks of time. A set of EROSIO/ATMOS/DOSET calculations is performed for each block, and the wind duration factor for each sequence of calculations corresponds to the UNSAT time block. The contaminated area term is the plane surface area of the trenches. This is because the unsaturated soil column model that calculates surface nuclide concentrations is a onedimensional model. Thus, the concentrations represent nuclide contamination for a unit surface area. The area which accounts for the total amount of surface contamination due to all the waste buried in the trenches is, therefore, that of the trench planes. The fraction of soil that is lofted, i.e., that remains suspended in the atmosphere, is largely subjective but is influenced by soil type. Estimates of the value range from a minimum of about $3 \%$ for clay particles to a maximum of about $40 \%$ for loam particles.

The computer solution to the wind erosion equation, which is done in the subroutine EROSIO, is addressed below. The solution follows the scheme outlined graphically in Reference 7.

Soil erodibility (I) is potential soil loss from a wide, unsheltered, isolated field with a bare, smooth, uncrusted surface. It is inversely related to the size of the aggregates in the surface soil and is tabulated as a function of percentage dry soil greater than 0.84 mm in diameter. Correction factors are then applied to soil erodibility to account for slope, ridge roughness, climate, dimensions and sheltering of the field, and vegetation. These correction factors are explained below.

Knoll erodibility $\left(\mathrm{I}_{\mathrm{s}}\right)$ is the correction factor which adjusts erodibility for windward slopes less than 152 m long. This factor varies with slope and can be as large as 7 for the top of a knoll with a $10 \%$ slope. The erosion rate for long windward slopes ( $>152 \mathrm{~m}$ ) is about the same as for level land. The Reference 7 data for knoll erodibility are tabulated in BURYIT for (a) the top of a knoll, and (b) for the portions of the windward slope where drag velocity and wind drag are the same as on top of the knoll (from about the upper third of the slope.)

Soil ridge roughness is a measure of the natural or artificial soil surface roughness other than caused by clods or vegetation. The roughness factor ( $\mathrm{K}^{\prime}$ ) is an empirical function of surface roughness ( k ) and varies between about 0.5 and 1. BURYIT contains tabular entries of the ridge roughness factor chart (Reference 3) for roughness values up to 0.25 m ( 10 in .). BURYIT also contains an equation for calculating the surface roug iness based on ridge height (h) and spacing ( s :
$\mathrm{k}=4 \mathrm{~h}^{2} / \mathrm{s}$.
The climatic factor correction includes the effects of mean annual temperature, precipitation, and prevailing wind conditions. For average annual soil loss calculations, the mean annual wind velocity data should be corrected to a standard height of 9 m . The climatic factor ( $\mathrm{C}^{\prime}$ ) is given by the empirical relation
$\mathrm{C}^{\prime}=34.483 \mathrm{~V}^{3} /(\mathrm{P}-\mathrm{E})^{2}$
where

[^1]P-E $=\underset{\text { tiveness ratio }^{18}}{\text { Thornthwaite's }}$ precipitation effec-

$$
=115(\mathrm{P} / \mathrm{T}-10)^{10 / 9}
$$

where

$$
\begin{aligned}
& \mathrm{P}=\text { mean annual precipitation }(\mathrm{mm}) \\
& \mathrm{T}=\text { mean annual temperature }\left({ }^{\circ} \mathrm{C}\right) .
\end{aligned}
$$

The equivalent field length correction factor ( $\mathbf{L}^{\prime}$ ) accounts for the unsheltered distance across the field along the prevailing wind erosion direction. The unsheltered travel distance for each component (j) of the erosion wind rose is:
$D_{j}=W \sec \left(A_{j}\right)$
where
$\mathrm{W}=$ field width $(\mathrm{m})$
$\mathrm{A}_{\mathrm{j}}=$ angle of deviation of wind rose vector from a direction perpendicular to the length of the field.

A geometric derivation of the above equation is given in Reference 1. However, in BURYIT, EROSIO considers only one wind rose direction vector, which should be in the direction of the prevailing wind. If barriers are modeled, the unsheltered travel distance is reduced by 10 times the barrier height. The result is then the equivalent field length (L) which is used to calculate $\mathrm{L}^{\prime}$.

The vegetative cover correction factor ( $\mathrm{V}^{\prime}$, includes the effects of quantity, type, and orientation of the vegetation. Reference 7 data, in tabular form, are included in BURYIT for modeling the three types of vegetative cover listed below:

1. Live or dead small grain crops in seedling and stooling stage, above the surface of the ground, for crops in 7.6 -cm-deep furrow (as created by a deep furrow drill) and on smooth ground.
2. Standing and flat anchored small grain stubble with any row width up to 0.25 m , including stover.
3. Standing and flat grain sorghum stubble of average stalk thickness, leafiness, and quantity of tops on the ground.

The solution of the wind erosion equation is done by a supervisory routine, COMPUT, which is called by EROSIO. EROSIO first converts the input variables (metric units) to English units and calculates the climatic factor correction. COMPUT then solves the wind erosion equation and returns the result to EROSIO, for calculation of the amount of radioactivity released to the atmosphere.

The wind erosion equation is solved in five steps. First, subroutine IPRIME determines the soil erodibility based on particle size, applies the windward knoll slope correction, and returns the result to COMPUT. The tabular lookup function for the soil erodibility and windward knoll slope tables (and ridge roughness correction and vegetative cover tables) is performed by a linear interpolation in INTER. The first intermediate result is
$\mathrm{E}_{1}=\mathrm{I}^{\prime}=\mathrm{II}_{\mathrm{s}}$.
Step 2 is the calculation of the soil ridge roughness correction factor in subroutine KPRIME. The second intermediate result is
$\mathrm{E}_{2}=\mathrm{I}^{\prime} \mathrm{K}^{\prime}$
The third step is the inclusion of the effects of local wind velocity and surface soil moisture, which gives
$E_{3}=I^{\prime} K^{\prime} C^{\prime}$
The fourth step accounts for the effect of field length ( $L^{\prime}$ )

$$
E_{4}=I^{\prime} K^{\prime} C^{\prime} f\left(L^{\prime}\right)
$$

$\mathrm{E}_{4}$ is not determined by a simple multiplication, because $\mathrm{E}_{2}, \mathrm{E}_{3}$, and $\mathrm{E}_{4}$ are all interrelated. Subroutine LPRIME calculates equivalent field length along the prevailing wind direction ( $\mathrm{L}^{\prime}$ ), and subroutine FLRM determines $\mathrm{E}_{4}$ based on $\mathrm{L}^{\prime}, \mathrm{E}_{2}$, and $E_{3}$. Calculation of equivalent vegetative cover is done in subroutine VEG. COMPUTE calls VEG separately to model each of the following types of vegetative cover:

1. Flat, anchored, small grain stubble
2. Live or dead small given crops in seedling and stooling stage
3. Flat grain sorghum stubble
4. Standing grain sorghum stubble.

The calls are made in the above order. It is important to specify only the desired cover type, because results of previous calls to VEG are overwritten, i.e., only the last specified type of vegetative cover is used for the correction factor. Subroutine FIN calculates the effect of the vegetative cover on soil erodibility as
$E_{5}=E=I^{\prime} K^{\prime} C^{\prime} f\left(L^{\prime}\right) f\left(V^{\prime}\right)$.
Again, $\mathrm{E}_{5}$ is not obtained by simple multiplication, because of interrelations among $\mathrm{E}_{4}, \mathrm{~V}$, and $\mathrm{E}_{5} . \mathrm{E}_{5}$ is the final, corrected soil loss result, E (ton/acreyear), which is returned to EROSIO.

After the soil loss is calculated, the value is compared to the maximum allowable erosion depth, which is set to 25.2 mm or 30 times the height of the smallest unmoved particle ( 0.84 mm ) (References 7 and 17). The equation which considers the erosion depth expresses the result as a maximum time for erosion at the calculated rate of soil loss
$t_{\text {max }}=\frac{(25.2)(\mathrm{e})(1-\mathrm{x})}{(0.224)(\mathrm{E})}$
where

$$
\begin{aligned}
t_{\max }= & \text { time which results in an erosion depth } \\
& \text { of } 25.2 \mathrm{~mm}(\mathrm{y}) \\
\mathrm{e}= & \text { soil density }\left(\mathrm{g} / \mathrm{cm}^{3}\right) \\
= & \text { fraction of aggregates greater than } \\
& 0.84 \mathrm{~mm} \\
0.224= & \text { units conversion factor (g-acre- } \\
& \left.\mathrm{mm} / \text { tons }-\mathrm{cm}^{3}\right) .
\end{aligned}
$$

The value of soil loss used in the calculation of airborne nuclide concentration is then the larger of the time block input from UNSAT and $\mathrm{t}_{\text {max }}$.

Once the corrected soil loss, E, is calculated, the next step is the computation of the radioactivity of the ith nuclide blown into the atmosphere. This may be expressed in the functional form as (in consistent units)
$\mathrm{C}_{\mathrm{i}}=(\mathrm{E})(\mathrm{A})(\mathrm{t})(\mathrm{Y})\left(\mathrm{X}_{\mathrm{i}}\right)$
where
$\mathrm{C}_{\mathrm{i}}=$ radioactivity to the atmosphere
$\mathrm{E}=$ computed total soil loss
$\mathrm{A}=$ contaminated area
$\mathrm{t}=$ duration of erosive wind
$\mathrm{Y}=$ percentage of soil loss into suspension
$\mathrm{X}_{\mathrm{i}}=$ radioactivity of ith nuclide per unit weight in the surface soil.

Two points in Equation (3) should be recognized. One, the area term should apply only to that surface that is subject to evaporation of water containing dissolved nuclides. UNSAT generates the contamination level per unit area of a soil column. Thus, the area to be used should be an appropriate trench area. Second, the percentage of soil loss into suspension, Y , is included because the computed soil loss, E, accounts for saltation and creep as well as suspension. Few measurements have been made so that the value of Y is uncertain. Estimates of the value range from no more than $40 \%$ to as low as $3 \%$ (Reference 1). For example, information from Chepil (Reference 17) (as cited in Soil Physics ${ }^{19}$ ) is shown in Table 4, where suspension of ciay particles is much less than loam particles and sand lies about in the middle.

### 2.7 AQUIFER and Supporting Subroutines

AQUIFER and supporting subroutines are usc 1 to model a release pathway for dissolved nuclides in water that has seeped through a waste disposal cell and entered an aquifer. The aquifer discharges at some user-specified distance into a body of water that is a supply for public consumption at a userspecified rate. The mathematical model is onedimensional, transient flow at a constant velocity, with axial dispersion and adsorption allowed (see Reference 1, pp. 6-7). Initially, the nuclide concentration is zero everywhere. A time-dependent boundary condition is used where the seepage column intersects the aquifer. The entraace of the nuclides is modeled as a uniform release during a specific time period (a band release). The subroutine sequentially selects these time periods. Radioactive decay up to the end of the release from the soil column is accounted for in the time-dependent boundary condition.

Table 4. Relative importance of saltation surface creep, and movement in suspension in the wind-erosion process (References 13 and 19)

|  | Soiil Removed <br> $(\%)$ |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
| Soil Type | Saltation |  | Surface Creep | Suspension |
| Scepter heavy clay | 72 |  | 25 | 3 |
| Haverhill loam | 55 | 7 | 38 |  |
| Hatton fine sandy loam | 55 | 13 | 32 |  |
| Fine drive sand | 68 | 16 | 16 |  |

The mathematical model has been solved analytically using Laplace transform techniques (Reference 3 ). The resulting solution was programmed and documented at Pacific Northwest Laboratory as the GET005 code. The present AQUIFER is a simplified version that does not account for daughter products and has a somewhat different input band description.

The analytical solution of the model in NUREG/CR-1963, pp. $6-8$ (Reference 1), is the same as that shown by D. H. Lester et al. ${ }^{20}$ under the same assumptions, although the latter reference should be consulted for the correct definitions of dimensionless variables. The supporting subroutines are SWD, FF10, and FERRNT. Together these subroutines compute the complimentary error function for large arguments. The result of these subroutines
was validated to 13 significant figures using the intrinsic function ERFC in the Control Data Corporation Fortran library.

The radioactivity discharged from the aquifer is merged with a user-specified flow of water that is ingested by people. An implicit assumption is that the flow of water includes the total flow discharged from the aquifer.

### 2.8 Subroutine PUNC

The dose from a puncture wound is calculated by subroutine PUNC, using the method described in Appendix A. In the output, this dose is printed separately from any dose resulting from direct radiation or from radionuclides in the environment.

## 3. RESPONSE SURFACE METHOD OF UNCERTAINTY ANALYSIS

### 3.1 Overview

The response surface method of uncertainty analysis is a statistical approach for calculating the consequence of a postulated event, or scenario, in the presence of a large number of influential variables. The methed is based on a systematic sampling of the true response surface that is then approximated by a polynomial equation in the input variables. The statistical moments of this polynomial response surface equation (RSE) are obtained using second-order error propagation. A Pearson distribution function, ${ }^{21}$ having the same statistical moments as the RSE, is used to determine the consequence of the scenario at a specified probability value. The response surface method is outlined in the following steps and explained in the discussion below.

1. Identify the influential input variables, and obtain the mean value, the standard error, and the probability ¿ensity function (PDF) for each variable.
2. Choose a statistical experimental design and assign each input variable to a column of the design matrix.
3. Perform the calculations using BURYIT. Make one run for each row of the design matrix, with the input variables perturbed according to their assigned factor levels for the row. The result of these calculations are the samples of the true response surface.
4. Fit a second-order polynomial RSE to the dose commitment response surface samples. Check the RSE for adequacy of fit and alter if needed.
5. Estimate the statistical moments of the RSE using second-order error propagation.
6. Match a probability density function (PDF) to the statistical moments of the RSE. The consequence of the scenario is then defined herein as the dose commitment which corresponds to a specified probability value (usually the $5 \%$ upper tail probability value) of the PDF.

The identification of influential input variables sometimes invol es a preliminary sensitivity study. The purpose of tee preliminary sensitivity study is to pare the number of variable input parameters to those which actually influence the response surface. For each candidatt variable, the mean or nominal value and the probability density should be known. The study is done with each of the candidate variables perturbed to its limiting value. These limiting values are usually taken to be three standard deviations above and telow the mean ( $\mu \pm 3 \sigma$ ); however, for some bounded distributions, notably the uniform probability density function, the extreme values occur at less than $\pm 3$ ofrom the mean. Two runs are done for each candidate, one for each limiting value. These runs are called the star points of the design. The responses for each run can be saved on computer disk storage for later use in estimating the quadratic terms of the RSE.

When the influential input variables have been identified, a statistical experimental design is chosen. A statistical experimental design is a pattern for perturbing the variable parameters, also called the factors of the design. An efficient design provides the information necessary to estimate the coefficients of all important terms of the RSE in as few computer runs as possible. Designs are classified according to the degree of confounding, or confusion of factors, caused by the pattern for perturbing the factors. The categories are actually in ascending order based on the resolution of terms of the RSE. Of interest are the following categories (Reference 6):

1. Resolution III designs: linear terms are free from confounding with other linear terms; but, are confounded with two-factor interactions and two-factor interactions with each other.
2. Resolution IV designs: linear terms are free from confounding with each other and with two-factor terms (also called interaction terms); but, two-factor terms are confounded with each other.
3. Resolution V designs: linear terms and two-factor terms are free from confounding with other linear or two-factor terms; but, two-factor terms are confounded with three-factor terms.

Resolution IV designs are well suited for calculation of second-order polynomial RSEs, because of the relative abundance of the linear and two-factor terms appearing in the actual equations. In general, each variable identified in the sensitivity analysis will have an associated linear term in the RSE. In addition, a few two-factor terms will appear, but usually the number is less than the number of linear terms. Resolution IV designs are efficient when the number of two-factor terms is not greater than the number of factors associated with a saturated design, i.e., a design containing the maximum number of factors for its size. The following discussion will clarify this concept.

Two types of designs are available in the BURYIT/ANALYZ package: fractional factorial and Plackett-Burman designs. Both types are of Resolution III in their basic forms. However, a Resolution IV design can be obtained for either by adding the foldover, which is simply the negative of the original design. This of course doubles the number of runs and, equivalently, the number of degrees of freedom available to estimate coefficients. The basic fractional factorial designs require the number of runs to be an integer power of two ( $\mathrm{n}=2 \mathrm{k}$ ), while the Plackett-Burman designs only require the number of runs to be an integer multiple of four $(\mathrm{n}=4 \mathrm{k})$. For both types, the maximum. number of factors is $\mathrm{n}-\mathrm{I}$. Thus, Plackett-Burman designs in certain circumstances are more efficient. Plackett-Burman designs of 20 and 24 runs are particularly useful in this respect. The disadvantage of the Plackett-Burman designs is the complexity of the confounding patterns. Algorithms are available for calculating confounding patterns for both fractional factorial and Plackett-Burman designs, but only the fractional factorial confounding pattern atgorithm is installed in BURYIT/ANAL YZ (Appendix B).

To illustrate confounding in fractional factorial designs, consider a 15 -factor, foldover design. Figure 1 is a 15 -factor, fractional factorial design including foldover and star point runs. The basic design requires 16 runs and 16 more are required for the foldover. The coefficients that can be estimated are: one constant term (also known as the grand mean), 15 linear terms (one for each factor), and 15 two-factor terms. The constant and the linear terms are free from confounding with twofactor interactions. However, there are 105 possible combinations of 15 factors taken two at a time, so the two-factor terms are in 15 groups each contain-
ing 7 confounded terms. Table 5 shows the confounding pattern for a 15 -factor, fractional factorial design with its foldover. Because of this confounding, the order of the input factors should be chosen so that each two-factor term suspected of being significant is in a group with only insignificant terms. If advance knowledge of significant twofactor terms is either not available or proves inadequate, a second calculation of the response polynomial can be made using a different sequence of input factors. Comparison of the coefficients of the two-factor terms of the two response polynomials should assist the isolation of significant two-factor terms.

BURYIT includes a driver program which automatically varies the input factors according to the design, makes the calculations for the response surface samples, and collects the results on computer disk storage. If a sensitivity analysis was done (Step 1), the results can be appended to the results of the linear analysis (Runs 1-32 of Figure 1). Thus, the sensitivity analysis results are the star points of the design (Runs 33-62), which are used for estimating the quadratic terms of the RSE.

The responsc surface samples are fitted to a second-order polynomial equation to fit the true response surface. If the equation fits well, the residuals, the errors in predicting the individual samples, are small and should appear to be random deviations. An equation is underfit if not enough terms are included to adequately model the response surface. Conversely, an equation is overfit if it contains so many terms that it tries to match the "random" residuals of the response. A good strategy ${ }^{22}$ witt guard against both underfit and overfit. ANYOLS, ${ }^{23}$ the code within ANALYZ that calculates the response surface equation, contains four regression strategies.

The first strategy uses the F-test (Reference 22) to determine which terms to include in the equation. The F-statistic is calculated for each term, to test whether the corresponding coefficient is zero. A large value of the F-statistic indicates that the coefficient is probably nonzero, and provides justification for including that term in the equation. To guard against overfit, terms in the model are reexamined at each step, and any with small F -statistics are removed from the model.

Another strategy uses the $\mathrm{C}_{\mathrm{p}}$ statistic 24,22 to estimate the mean squared error (MSE) of the fitting equation. $\mathrm{C}_{\mathrm{p}}$ measures the sum of the squared


Figure 1. Design matrix for a $2^{* *}(15-10)$ fractional factorial including foldover and quadratic runs.

Table 5. Confounding pattern for twofactor interactions for 15 -factor, fractional factorial design

|  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| $1-2$ | $1-3$ | $1-4$ | $1-5$ | $1-6$ |
| $3-11$ | $2-11$ | $2-12$ | $3-8$ | $2-8$ |
| $4-12$ | $4-13$ | $3-13$ | $4-9$ | $4-10$ |
| $6-8$ | $5-8$ | $5-9$ | $6-11$ | $5-11$ |
| $7-9$ | $7-10$ | $6-10$ | $7-12$ | $7-13$ |
| $10-15$ | $9-15$ | $8-15$ | $10-14$ | $9-14$ |
| $13-14$ | $12-14$ | $11-14$ | $13-15$ | $12-15$ |
| $1-7$ | $1-8$ | $1-9$ | $1-10$ | $1-11$ |
| $1-9$ | $2-6$ | $2-7$ | $2-15$ | $2-3$ |
| $2-9-10$ | $3-5$ | $3-15$ | $3-7$ | $4-14$ |
| $5-12$ | $4-15$ | $4-5$ | $4-6$ | $5-6$ |
| $6-13$ | $7-14$ | $6-14$ | $5-14$ | $7-15$ |
| $8-14$ | $9-13$ | $8-13$ | $8-12$ | $9-10$ |
| $11-15$ | $10-12$ | $10-11$ | $9-11$ | $12-13$ |
| $1-12$ | $1-13$ | $1-14$ | $1-15$ | $2-5$ |
| $2-4$ | $2-14$ | $2-13$ | $2-10$ | $3-6$ |
| $3-14$ | $3-4$ | $3-12$ | $3-9$ | $4-7$ |
| $5-7$ | $5-15$ | $4-11$ | $4-8$ | $8-11$ |
| $6-15$ | $6-7$ | $5-10$ | $5-13$ | $9-12$ |
| $8-10$ | $8-9$ | $6-9$ | $6-12$ | $10-13$ |
| $11-13$ | $11-12$ | $7-8$ | $7-11$ | $14-15$ |
|  |  |  |  |  |

biases (due to lack of fit) in the response equation at all sample points. This method depends on a good estimate of $\sigma_{b}{ }^{2}$ (variance due to "random" residuals in the sample data), and is, therefore, less desirable than $\mathrm{S}^{2}$ minimization or PRESS.

A third strategy for selecting terms to be included in the equation is minimization of $\mathrm{S}^{2}$, the mean residual sum of squares. The $\mathrm{S}^{2}$ is the estimate of $o_{\mathrm{b}}^{2}$ with the minimum variance if the residuals are truly random.
$S^{2}=\operatorname{RSS} /(\mathrm{N}-\mathrm{P})$
where
RSS $=$ the residual sum of squares, the sum of the squares of the residuals of the fitted polynomial at all sample points
$\mathrm{N}=$ the number of sample points
$\mathrm{P}=$ the number of terms in the equation.

Since adding terms to the polynomial decreases RSS and increases $P$, the minimum value of $S^{2}$ will occur at a small value of RSS, but before too many terms are included in the equation. Thus, the $\mathrm{S}^{2}$ minimization strategy guards against both overfit and underfit (Reference 23).

The fourth strategy is Prediction Error Sum of Squares (PRESS) ${ }^{25}$ In this method, all response samples, but one, are included in the fit, and the error in calculating the remaining point is measured. The sum of squares of these prediction errors, summed over all the sample points, is PRESS. Selection of terms for inclusion in the polynomial equation is based on minimizing PRESS. This strategy guards against both overfit and underfit and is the default-stepwise-regression method for ANALYZ.

Second-order error propagation is used to estimate the lower four moments of the response density under the assumption that the uncertainty variables are statistically independent. The computer program SOERP ${ }^{26}$ solves equations for the raw moments (mornents-about-the-origin) of the response distribution. These equations are based on a multivariable Taylor series expansion that is truncated after second-order terms and are, therefore, second-order error propagation equations. They use all moments of the input variables up to and including the eighth and produce estimates of the lower four moments of the fitted response surface equation. SOERP is incorporated as a subcode in ANALYZ.

The statistical moments of the response, calculated using SOERP, are then used to obtain a probability density function (PDF) which approximates the actual BURYIT dose commitment response density. The code PDFPLOT, ${ }^{27}$ within ANALYZ, matches the lower four response moments to those of a member of the Pearson family of distribution curves. The approximate response PDF is then used to estimate the consequence of the calculated scenario at the specified probability value of interest. Often the consequence is defined in terms of the spread of dose commitment that includes $90 \%$ of the response sample population. The bounds on the consequences of the calculated scenario thus correspond to the $5 \%$ lower and upper tail probability values, respectively. For dose commitment, the $5 \%$ upper tail probability value, which translates as the upper consequence bound, is the one of interest.

The relationship between the mean value and the spread of dose commitment response are such that some portion of the approximate response density could lie to the left of the origin. This situation occurred, in fact, in the methodology demonstration case (Section 6). To preclude the representation of negative dose commitments as physically meaningful, a one-sided truncation was included in the calculation of the Pearson censity approximation. The truncation population is formed by removing the portion of the density which lies to the left of the origin (defined as the point of truncation) and then including the removed fraction in the remaining density, to the right of the truncation point. The one-sided truncation incorporated into the PDFPLOT subcode is given by the following transformation. 28,29

If x is a random variable with density $\mathrm{f}(\mathrm{x})$ and cumulative distribution $\mathrm{F}(\mathrm{x})$, then the density of x truncated on the left at point a is given by
$g(x)=\frac{f(x)}{1-F(a)} I_{(a)}(x)$
where
$\mathrm{I}_{(\mathrm{a})}{ }^{(\mathrm{x})}$ is the indicator function;
$1=1$ for $a \leq x$
$1=0$ otherwise
$a \geq 0$.
The cumulative distribution of x truncated on the left at point a is:
$G(x)=\frac{F(x)-F(a)}{1-F(a)}{ }_{(a)}^{(x)}$.
The above transformation gives a conservative representation of the specified probability values (SPVs) that are used to estimate the dose commitment consequence. Quite simply, $\mathrm{G}(\mathrm{x}) \leq \mathrm{F}(\mathrm{x})$, and thus, $\mathrm{xG} \geq \mathrm{xF}$ for $\mathrm{G}=\mathrm{F}$. That is, the dose commitment $\times$ corresponding to a SPV defined by cumulative distribution G is never smaller than the x which corresponds to the same SPV defined by F.

Note that the truncation is performed on the values generated by the Pearson distribution. The truncation does not affect the response moments calculated by SOERP. Thus, the risk calculation (described in Section 4) uses the actual moments
calculated for the dose commitment response and not the truncated density. Truncation of the risk density, if necessary, is performed separately and follows the risk calculation.

Alternatively, the user has the option of fitting the logarithm of the response rather than the response itself. The fitting coefficients are computed by ANYOLS and passed to SOERP, which computes the first four moments and passes them to PDFPLOT. PDFPLOT then finds the corresponding Pearson distribution for $\log$ (response). This does not need to be truncated, because a logarithm can be positive or negative. Finally, PDFPLOT mathematically transforms the variable and plots the distribution for the original response. In situations when the final dose is roughly proportional to the important faciors, using logs only avoids the truncation problem, but may also provide a substantially improved fit in ANYOLS.

### 3.2 Sensitivity of Dose Commitment Computations

Once the response surface equation (RSE) has been obtained, the parametric sensitivity of the dose commitment computations may be studied. The term parametric sensitivity refers to the rate of change of a response with respect to an uncertainty factor, usually a parameter.

Two kinds of sensitivity coefficients are in common usage. For convenience here, these will be called the absolute sensitivity coefficient and the relative sensitivity coefficient. The absolute sensitivity coefficient refers to the change in response in its original units with respect to either a change in the parameter in its origina! units or a change in the parameter normalized with respect to its uncertainty standard deviation. The relative sensitivity coefficient refers to the fractional change of the response relative to its nominal value with respect to the fractional change of the parameter relative to its nominal value.

From the discussion in Section 3.1, it should be apparent that the linear coefficients in the response surface equation are the absolute sensitivity coefficients for the parameters. In mathematical terms, a linear term in a Taylor series expansion is the first partial derivative evaluated at the nominal (expansion) point
$\frac{\partial R}{\partial p}$
nom
where

$$
\begin{aligned}
& \mathrm{R}=\text { response } \\
& \mathrm{p}=\text { parameter. }
\end{aligned}
$$

This is precisely the definition of an absolute sensitivity coefficient. In the case of simple models, such coefficients may be computed directly. With more complex models and when a Resolution IV experimental design is used in evaluating the response surface coetficient, the linear terms are obtained without conlounding among themselves. In the methodology discribed herein, the absolute sensitivity coeffic ent is often expressed as rate of response change with respect to a parameter change of one standard deviation. Thus, if the parameter is normalized as $u=p / \sigma$, the absolute sensitivity coefficient will come out as

## $\frac{\partial R}{\partial i t}$ <br> $\overline{\partial_{11}}$ <br> nom

Such coefficients are produced by SOERP and are designated as LB( )s. The absolute sensitivity coefficients are useful for mapping the behavior of the response in the vicinity of the nominal point. In addition, the second-order terms of the RSE may be used to assess the magnitude and direction of deviations from linearity.

The relative sensitivity coefficients may also be derived from the linear terms of the RSE. If $\mathrm{R}_{\mathrm{o}}$ and $\mathrm{P}_{\mathrm{O}}$ are the values of the response and a parameter at the nominal point, the relative sensitivity coefficient is
$\left.\left(\frac{p}{R}\right)_{0} \frac{\partial R}{\partial p}\right|_{\text {nom }}=\left.\left(\frac{p}{\partial R}\right)_{0} \frac{\partial R}{\partial u}\right|_{\text {nom }}$
In other words, this sensitivity coefficient is the linear term in the RSE multiplied by the ratio $(\mathrm{p} / \mathrm{R})_{\mathrm{O}}$. The relative sensitivity coefficients are useful to those analysts who prefer to use percentage numbers to describe the perturbation behavior of the response.

Thus, the response surface uncertainty analysis methodology focuses on the calculation and use of the response surface equation. This second-order polynomial approximates the surface of the BURYIT code response over a small range of input parameter variation. The RSE forms the basis for the response density, which is a statistical representation of the likelihood of occurrence of a dose commitment, given a scenario and the statistical properties of the uncertainty varibbles. Section 4 describes the calculation of the risk of the dose commitment, given the consequence and the statistical properties of scenario occurrese probability.

## 4. CALCULATION OF RISK

The risk associated with an event, or scenario, is defined herein to be the product of the consequence of the event, providing that the event has occurred, and the probability of occurrence of the event within some convenient time period. That is,
$(\text { RISK })_{i}=(\text { Consequence })_{i} \cdot(\text { Probability })_{i}$.
The time period pertaining to the occurrence probability is often taken as one year; but, whatever time period is selected must be considered in the interpretation of computational results. The above definition of risk due to a particular scenario, i, arises as a term in the (probabilistic) expected consequence at a site
$E($ Consequence $)=\sum_{i=1}^{n}(\text { Probability })_{i}$
(Consequence) ${ }_{i}$

$$
\begin{equation*}
=\sum_{i=1}^{n}(\text { Risk })_{i} \tag{5}
\end{equation*}
$$

where the scenarios (events) are independent and mutually exclusive. Thus, each risk term represents a contribution to the expected consequence.

As discussca earler, the true consequence of a particular release scenario, (e.g.; the ith) is not known exactiy, but rather is represented by a probability density furction. At the same time, the true probability that a release scenario will take place is not known exactly and must be represented by a probability density function. As a result of these uncertainties, the risk associated with the ith release scenario must be represented by a probability density function, as must the ith consequence. The estimation of a probability density function for each risk term is discussed next. This is followed by a description of the estimation of the probability density function of the expected consequence, i.e., the sum of the risiss.

A direct application of the response surface method is used to determine the risk density, given the variable input parameters of scenario consequence and probability and their respective densities. Here the response surface equation is
obtained from the definition of risk in Equation (4). To obtain the cocfficients of the RSE, Equation (4) is expanded in a Taylor series about the mean values of the respective input variables. In the notation of Reference 26 ,

$$
\begin{align*}
Z\left(x_{1}, x_{2}\right)= & Z\left(v_{11}, v_{21}\right)+\frac{\partial h}{\partial x_{1}}\left(x_{1}-v_{11}\right) \\
& +\frac{\partial h}{\partial x_{2}}\left(x_{2}-v_{21}\right) \\
& +\frac{\partial^{2} h}{\partial x_{1} \partial x_{2}}\left(x_{1}-v_{11}\right)\left(x_{1}-v_{21}\right) \tag{6}
\end{align*}
$$

where

$$
\begin{aligned}
\mathrm{Z}= & h\left(\mathrm{x}_{1}, \mathrm{x}_{2}\right)=\mathrm{x}_{1} \cdot \mathrm{x}_{2} \\
v_{11}= & \begin{array}{l}
\text { expected value (mean value) of the } \\
\\
\\
\text { variable } \mathrm{x}_{1}
\end{array} \\
v_{21}= & \begin{array}{l}
\text { expected value (mean value) of the } \\
\\
\\
\text { variable } x_{2}
\end{array}
\end{aligned}
$$

The derivatives are

$$
\frac{\partial h}{\partial x_{1}}=x_{2}, \frac{\partial h}{\partial x_{2}}=x_{1}, \text { and } \frac{\partial^{2} h}{\partial x_{1} \partial x_{2}}=1
$$

Thus, the exact response surface equation consists of a constant (intercept) term, two linear terms, and one two-factor term.

$$
\begin{align*}
\mathrm{Z}=\mathrm{b}_{0}+ & \mathrm{b}_{1}\left(\mathrm{x}_{1}-v_{11}\right)+\mathrm{b}_{2}\left(\mathrm{x}_{2}-v_{21}\right) \\
& +\mathrm{b}_{12}\left(\mathrm{x}_{1}-v_{11}\right)\left(\mathrm{x}_{2}-v_{2}\right) \tag{7}
\end{align*}
$$

The coefficients of the RSE are:
$b_{0}=Z\left(v_{11}, v_{21}\right)$ (intercept coefficient)
$b_{1}=v_{21}$ and $b_{2}=v_{11}$ (linear coefficients)
$\mathrm{b}_{12}=1$ (two-fac or or cross-product coefficient).
Equation (7) is the KSE which is supplied to SOERP for calculating the risk dose commitment response moments, assuming statistically independent variables.

The absence quadratic terms in Equation (7) is fortunate. It was stated eariier that the equations solved by SOERP equations required the first eight central moments of the variable input parameters. Only the first four noments of scenario consequence dose commitment are available from SOERP. All the terms in the SOERP equations requiring input parameter moments 5 through 8 contain quadratic coefficients as well. All these terms vanish for the calculation of the risk moments. Thus, the first four moments of dose commitment are sufficient for the calculation.

In addition to the first four moments of the dose commitment, the risk calculation requires the first four moments of the scenario occurrence proiability PDF. Scenario occurrence probability is nodeled as a log-normal variate. Log-normal distributions are commonly described in terms of a median value and an error factor. The median value, $m$, is given by
$\mathrm{m}=\exp \left[\mu_{\mathrm{n}}\right]$,
where $\mu_{\mathrm{n}}$ is the mean of the underlying normal distribution of logarithms. The error factor, EF, often includes $90 \%$ of the population described by the distribution, and is given by
$\mathrm{EF}=\exp \left[1.645 o_{\mathrm{n}}\right]$,
where $\sigma_{\mathrm{n}}$ is the standard deviation of the underlying normal distribution of logarithms, and 1.645 is the 0.90 fractile of the two-sided, standard normal distribution. ne mean and variance of the distribution of scenaro : ccurrence probability are given in terms of the meen and variance of the underlying normal distribution of logarithms (Reference 29):
(4)
and

$$
o^{2}=\exp \left(2 \mu_{n}+o_{n}^{2}\right) \exp \left[\left(\sigma_{n}^{2}\right)-1\right] .
$$

Of course it is more useful to have the mean and variance of the distribution of scenario occurrence probability in terms of the given information. Thus,
$\mu_{1}=\exp \left[\ln m+\frac{1}{2}\left(\frac{\ln E F}{1.645}\right)^{2}\right]$
a d

$$
\sigma^{2}=\exp \left[2 \ln \mathrm{~m}+\left(\frac{\ln \mathrm{EF}}{1.645}\right)^{2}\right]\left[\exp \left(\frac{\ln \mathrm{EF}}{1.645}\right)^{2}-1\right] .
$$

The raw moments (moments-about-the-origin) are given by (Reference 29).

$$
v_{r}=\exp \left[r \mu_{n}+\frac{1}{2} r^{2} \sigma_{n}^{2}\right]
$$

where r dentes the order of the moment. SOERP requires the central moments of the variable input parameters. The first four central moments of a distribution are related to the raw moments as follows:

$$
\begin{aligned}
& \mu_{1}=0 \\
& \mu_{2}=v_{2}-v_{1}^{2} \equiv 0^{2} \\
& \mu_{3}=v_{3}-3 v_{2} v_{1}+2 v_{1}^{3} \\
& \mu_{4}=v_{4}-4 v_{3} v_{1}+6 v_{2} v_{1}^{2}-3 v_{1}^{4}
\end{aligned}
$$

Thus, given the statistical properties of the dose commitment and probability, along with the response surface equation which describes the risk in terms of the two inputs, SOERP estimates the lower four statistical moments of the risk. As in Section 3, these response moments are then used to obtain a PDF which approximates the actual risk density. PDFPLOT is used again to match the response moments to those of a Pearson distribution. The location of the $5 \%$ upper tail probability value of the Pearson distribution again corresponds to the dose commitment of interest. Therefore, the risk associated with a given scenario is the dose commitment value which includes $90 \%$ of the area unde the risk density curve. As in the previous section, a one-sided truncation is performed, if necessary, to avoid representation of negative dose values as physically significant.

If the $\log$ of the dose was originally fitted, computations similar to the above are used to get the moments, and then the PDF, of $\log$ (risk). This is then used to get the PDF for the risk itself.

Response surface methodology is, therefore, applied twice in the assessment of risk associated with a postulated event for a LLW disposal site. The first application involves the least-squares fitting of a RSE to the responses of BURYIT. The density
associated with this RSE is used to define the consequence of the scenario. The second application combines the consequence density with the density of the likelihood (probability) of the event to form a second response, the density of which is defined to be the risk of the scenario. The risk bounds calculated for the total population dose or the maximum individual dose for several scenarios may be plotted in bar graph form for comparison.

The estimation of the density of the expected consequence (sum of scenario risks) is now discussed, even though currently this is not included in the pro-
gram package. In accord with Equation (4), linear error propagation is used assuming that the scenario risks are statistically independent. It is also implicitly assumed that the consequence under no release is exactly zero. The means and first four central moments of the density for expected consequence are determined by the means and central moments of the risks, $\mu_{\mathrm{ij}}$. Again, as in Sections 3 and 4 , these moments may be fed to the subroutine PDFPLOT in order to obtain a density function for expected consequence (i.e., total risk). Just as before, upper and lower probability bounds on total risk may be computed.

## 5. INPUT AND OUTPUT

### 5.1 Analysis Flexibility

This section contains a brief discussion of the flexibility of , ie 3URYIT/ANALYZ package, followed by a a scription of the input requirements. A flexible con uter code or package is one which has the capability to perform only a portion of the entire sequence of calculations, so the results of the selected portion may be observed separately. Insight gained from the results of the selected calculation may then be used to modify the code inputs in order to obtain a more representative model for use in the remainder of the calculation sequence. An inflexible code, by contrast, may require execution of the entire sequence of calculations for each set of inputs. Inflexibility can result in wasted calculations and unnecessary computer costs. A flexible code is desirable for uncertainty analysis calculations, primarily because input parameter uncertainties may be imprecisely known, and advance knowledge of the effects of the uncertainties on the code responses may be sketchy. Flexibility options were incorporated in the BURYIT/ANALYZ package, as follows.

The BURYIT uncertainty analysis option includes the capability to do selected portions of an uncertainty analysis. The uncertainty analysis calculation, done by BURYIT, includes steps 1 through 3 of the response surface method, which was discussed in Section 3. The input variables are perturbed according to an experimental design, and the responses from each run of the design are collected on computer disk storage. Steps 4 through 6 of the response surface method involve obtaining the response surface polynomial equation and determining its statistical properties. ANALYZ performs these steps. The BURYIT flexibility options are:

1. Calculation of the experimental design and confounding pattern
2. Option 1 plus calculation of the scenario with all inputs set to nominal values
3. Option 1 plus calculation of the quadratic runs (star points) of the design
4. Complete uncertainty analysis calculation
5. Uncertainty analysis calculation of linear runs of the design, appending the results of quadratic runs previously obtained using Option 3.

Option 1 is useful for obtaining the confounding pattern for a fractional factorial design, and for determining the order of the input variables in the design. Option 2 can be used for checking for input correctness and determining whether the calculated dose commitment is a reasonable nominal value. Option 3 is used for a sensitivity study. The quadratic runs are made with each input variable perturbed singly to the practical limits of its density function (usually $3 a$ on either side of the nominal value). The results can be used to identify variables having little or no effect on the response. Option 5 is used in conjunction with Option 3, and includes the capability to call responses of selected variables from the results of the sensitivity study.

BURYIT was originally written as a deterministic program, i.e., with no variable input parameters. In this form, one or several scenarios can be calculated in the same run. This mode of calculation was retained in the present version and can be specified by merely omitting the uncertainty analysis control variable. The original version permitted interactive as well as batch operation. However, due to cost considerations and computer storage requirements associated with interactive execution of BURYIT, no effort was made to retain this capability. Thus, BURYIT in its present for n is for batch use only. The capability for calculating several scenarios in the same run was retained subject to the original restriction that all scenarios which call the unsaturated soil column model (UNSAT) use the same nuclide inventory. This stipulation is necessary because the nuclide retardation factors, used in the calculation of nuclide transport through the soil layers, depend on both the type of soil in each layer and the nuclide involved. Thus, the retardation factors, corresponding to the nuclide source list for the inventory specified for the scenario, are supplied in the geology data input. Therefore, the nuclide inventory must be correct for the geology data file being used.

ANALYZ includes features to improve its flexibility as well. The options available are the following:

1. Calculation of the statistical moments of scenario consequence, based on the response samples obtained using BURYIT and the densities of the input variables
2. Calculation of the statistical moments of scenario occurrence probability
3. Calculation of the statistical moments of scenario risk. This includes the calculations in Options 1 and 2, and the calculation of the product of the two results.

In all three options, the calculated moments are matched to a Pearson distribution in order to estimate the rasult (dose commitment or occurrence probability) which corresponds to a specified probability value. Additionally, the subcode PDFPLOT may be accessed directly in order to obtain an approximate probability density for a set of user-supplied statistical moments.

### 5.2 Input Requirements

Three types of input must be furnished to the BURYIT/ANALYZ package. They are: (a) the radiological data base, (b) the environmental characteristics of the location being modeled, and (c) the control information necessary to do the uncertainty analysis and the risk calculation. Items $a$ and $b$ are nearly the same as described in NUREG/CR-1963, Volume 1 (Reference 1). Changes to these items reflect the model improvements and corrections to coding errors as described in Section 2. Item c represents implementation of the response surface methodology for doing the uncertainty analysis and the risk calculation, described in Sections 3 and 4, respectively. Because of the amount and variety of input required, the information is divided among several disk files, each of which supplies a specific type of information. A summary of these files is contained in Table 6 and a detailed description of the function of each file is in the following paragraphs. The files which form the data base will not normally be changed by the user. Appendix C contains the scenario control file. Instructions for modeling the environmental characteristics of the site location under analysis were published by Science Applications, Inc., (Reference 1). However, for integrity of this document, a complete list of input instruc tions is included herein, which includes modeling changes and corrections that have been made to

## Table 6. Disk file allocation for BURYIT/ANALYZ

| File Name | Function |
| :---: | :---: |
| TAPE1 | Nuclide data |
| TAPE2 | Nuclide list |
| TAPE3 | Intercom terminal |
| TAPE4 | Uncertainty analysis responses |
| TAPE5 ${ }^{\text {a,b }}$ | Input-uncertainty analysis |
| TAPE6 | Printed output |
| TAPE7 | Scenario control and scenario descriptions |
| TAPE8 ${ }^{\text {b }}$ | Erosion data |
| TAPEs ${ }^{\text {b }}$ | Atmospheric data |
| TAPE10 ${ }^{\text {b }}$ | Geology and rainfall data |
| TAPE11 ${ }^{\text {b }}$ | Aquifer data |
| TAPE12 | Nuclide inventories |
| T/.PE14 ${ }^{\text {b }}$ | Input control and title |
| TAPE15 ${ }^{\text {b }}$ | Agriculture and population data |
| TAPE $20{ }^{\text {b }}$ | Shine exposure data |
| TAPE 21 | Internal BURYIT file for storage of intermediate dose results |
| TAPE44 | Sensitivity analysis input to BURYIT |
| TAPE45 | Sensitivity analysis results |
| TAPE97 | Scenario occurrence probability, input to ANALYZ. |
| TAPE98 | Input scenario risk file containing the results of risk calculations for previously done scenarios |
| TAPE99 | Output scenario risk file containing TAPE98 results plus results of current risk calculation |
| META | DISSPLA plot file |

[^2]BURYIT. Also contained in the following discussion are the instructions for using the uncertainty analysis and risk calculation options.
5.2.1 Computer Storage Files for BURYIT/ ANALYZ. TAPE1 is the nuclide data file, and is supplied with the BURYIT/ANALYZ package as part of the data base. The file contains dose equivalence factors, gamma exposure constants, and radioactive decay constants for each nuclide. The information for the nuclide data base was compiled by Science Application, Inc., and is documented in Reference 1.

TAPE2 is the nuclide index file, used for locating entries in TAPE1. It contains the nuclide names in the same sequence as the nuclide property entries in TAPE1.

TAPE3 is the intercom input file for BURYIT. It is used to specify the desired scenarios and to supply minimal information to the code if BURYIT is executed interactively. Although interactive capability is presently considered inactive, the installed coding pertinent to this feature was retained. The inputs supplied by the intercom terminal are included in the TAPE14 input for a batch job and are listed there.

TAPE4 is the uncertainty analysis response file, written by BURYIT for input to ANALYZ. Ir addition to the responses of the BURYIT runs, TAPE4 contains the control information for calculating the experimental design, the first eight central moments of the input variables, and pht title information.

TAPE5 is the primary input file. It contains the job control language and the uncertainty analysis control information. A detailed description of TAPES information is contained in the input requirements section, below.

TAPE6 is the printed output file. The contents depend on the options sp:cified, of course, but include echoes of input files, the results of BURYIT dose calculations, and uncertainty analysis results.

TAPE7 is the scenario control file. It contains, for each scenario, the control word that specifies the scenario description, the waste inventory, the pathway sequence, and the fraction of the waste specified to follow each pathway. It also contains the narratives describing the scenarios. Appendix C contains a listing of the contents of TAPE7.

TAPE8 (erosion data), TAPE9 (atmospheric data), TAPE10 (geology and rainfall data), TAPE11 (aquifer data), TAPE15 (agriculture and population data), and TAPE20 (shine exposure data) are specific to the site being modeled and are described in the input requirements section, below.

TAPE12 is the nuclide (waste) inventory file. It contains lists of isotopes and their amounts for six classifications of waste. These waste classifications are designated WS-1 through WS-6 and represent the following types of waste:

1. WS-I is highly activated light water reactor (LWR) components and is assumed to be stainless steel with cobalt-60 as the primary isotope.
2. WS-2 is LWR operational waste, assumed to be primarily comprised of low-level activity items such as filters, filter backwork, fitered phase separator decan: liquid, evaporator bottoms, demineralizer wastes, laundry wastes, and generai trash. This class contains 33 radionuclides.
3. WS-3 is LWR decommissioning waste, consisting primarily of cobalt-60, nickel-63, and iron-55.
4. WS-4 is assumed to result from decontamination of decommissioned facilities. It is similar in composition to LWR operational wastes but has a much higher specific activity. This class contains 33 radionuclides.
5. WS-5 consists of nonfuel cycle (institutional) waste and contains four radionuclides.
6. WS-6 is the waste inventory assumed to represent an average composition in a burial trench. The original WS-6 inventory supplied with BURYIT was based on information supplied by R. D. Smith. ${ }^{30}$ This class contains 44 radionuclides.

Appendix D lists the contents of the waste inventory file. The first line for each waste inventory contains the name of the inventory, the number of nuclides present, and the average density ( $\mathrm{kg} / \mathrm{m}^{3}$ ). The density is used for puncture-wound scenarios. The remaining lines for the inventory consist of a nuclide name and concentration $\left(\mathrm{Ci} / \mathrm{m}^{3}\right)$, one line per nuclide.

TAPE14 is the calculation control file for BURYIT. It contains the logic switches for specifying either a maximum individual dose or cumulative population dose, an in ercom or batch job, and the amount of printed output to be produced. It also contains the title card, and, for a batch job, the scenario number and pathwaydependent information. For an interactive job, the s. enario number and pathway-dependent input are supplied from the terminal (TAPE3). TAPE14 contents are completely described in the input requirements section, below.

TAPE21 is an internal scratch file, used by BURYIT to store intermediate results of the dose calculation.

TAPE44 and TAPE45 are used for storage and manipulation of BURYIT responses for sensitivity analysis runs. The responses are written to TAPE45 during the sensitivity analysis and read from TAPE44 during a subsequent uncertainty analysis. The quadratic runs which were done during the sensitivity analysis are not repeated. Instead, the results are read from TAPE44 and appended to the linear run responses which are written to TAPE4.

TAPE97 is the scenario probability file used by ANALYZ. The probabilities of occurrence for the scenarios are assumed to be log-normal variates, which are commonly described by a median value and an error factor. TAPE97 contains a median value and a $90 \%$ error factor for each scenario. Appendix E discusses these values and their derivation.

TAPE98 and TAPE99 are the scenario risk files used by ANALYZ. TAPE98 contains the results of the risk calculations (Equation 4) for previously done scenarios. TAPE99 contains the TAPE98 results, plus the result of the current scenario risk calculation.

The above files all fall within three categories. They are part of the data base furnished with BURYIT/ANALYZ, are internal files used for storage of intermediate results or for passing results between BURYIT and ANAI.YZ, or are output files containing the results of the calculations. In all cases, these files are normally not within the direct control of the user. The following files form the specific model for which the calculations are performed and are all user-supplied. Detailed instructions for creating the user-supplied inputs for BURYIT/ANALYZ are given in the following paragraphs.

### 5.2.2 Input Instructions

5.2.2.1 BURYIT Control and Title. TAPE14 contains control and title information for BURYIT. The control variables spicify interactive or batch mode, amount of printed output, and whether a maximum individual dose is calculated. Other input is specific to batch mode; interactive users supply this information from the terminal (TAPE3) in response to prompts by BURYIT.

## Card 1: IBATCH, IPRINT

## FREE FORMAT

IBATCH $\quad=0$, interactive job (interactive option is presently inactive)
IBATCH $=1$, batch job
IPRINT $\quad=0$, printed output consists of input file echoes and results summary
$=1$, printed output contains input file echoes, formatted input printout, intermediate results from UNSAT, AQUIFER, and/or EROSIO, and results summary
$=2$, printed output consists of the above plus detailed intermediate results from UNSAT.

Card 2: (ITITLE(I), $\mathrm{I}=1,8$ )
FORMAT(8A10)
ITITLE problem title which appears on printed output file and plots.

Cards 3-8 are for batch input only. Interactive users supply the following inputs in response to code prompts:
Card 3: ISCEN scenario code
FORMAT(A4).
Card 4: INVET waste inventory code
FORMAT(A9).
Cards $5-8$ are pathway-dependent and are determined by the scenario requested. Appendix C contains a listing of the scenario control file (TAPE7), which includes the list of pathways specified for each svenario. Each pathway for the scenario must include the appropriate cards, as follows:

## Paths 3.4

Card 5: CUMT, RF, EMEF, TIMWET, TIMCYC
FREE FORMAT
CUMT exposure time (h)
RF annual rainfall for ATMOS and EROSIO (m/y)
EMEF emplacement efficiency for UNSAT (waste volume/total volume)
TIMWET time for the rain cycle (h), for ATMOS and UNSAT
TIMCYC total time in one cycle of rain and dry (h)
Card 6: VOL volume of package $\left(\mathrm{m}^{3}\right)$
FREE FORMAT
Card 7: UD, DI
FREE FORMAT
UD wind velocity for equivalent contaminated cloud exposure model $(\mathrm{m} / \mathrm{s})$. Minimum value is $2.24 \mathrm{~m} / \mathrm{s}$ ( 5 miles $/ \mathrm{h}$ )

DI dust cloud diameter ( m )

## Paths 3.6

Card 5: CUMT, RF, EMEF, TIMWET, TIMCYC as above
'ard 6: VOL as above
Card 7: PDEPOS mass deposited in punc.ure wound ( $\mu \mathrm{g}$ )

## Paths 3-2, 3, 2

| Card 5: | CUMT, RF, EMEF, TIMWET, TIMCYC | as above |
| :--- | :--- | :--- |
| Card 6: | VOL | as above |


| Card 5: | CUMT, RF, EMEF, TIMWET, TIMCYC | as above |
| :--- | :--- | :--- |
| Card 6: | VOL | as above |
| Card 7: | CUMT, RF, EMEF |  |
|  | CUMT time span for unsaturated soil column calculation (y) |  |
|  | RF, EMEF, TIMWET, TIMCYC |  |

## Paths 951, 91

Card 5: CUMT, RF, EMEF, TIMWET, TIMCYC
CUMT time span for unsaturated soil column calculation (y)
RF, EMEF, TIMWET, TIMCYC

## Path 4

Card 6: CUMT, RF, EMEF, TIMWET, TIMCYC
CUMT exposure time (h)
RF, EMEF, TIMWET, TIMCYC as above
Card 7: UD, DI as above
Path 3-4951
Cards 5-7: as for Path 3-4
Card 8: CUMT, RF, EMEF, TIMWET, TIMCYC
CUMT time span for unsaturated soil column calculation (y)
RF, EMEF, TIMWET, TIMCYC as above
The following files contain information specific to the site being modeled. The SI (international metric) system of units is used with the exception of the mean annual temperature (MAT) where degrees Celsius should be used.
5.2.22 Aquifer Data. The aquifer data are input from TAPE11 using a free formatted read statement.

Aquifer Card: XZ, EI, VZ, FLOWR, (RNWV(I), I=1, NUMNUC)
$X Z \quad$ length of aquifer (m)
EI axial dispersion coefficient ( $\mathrm{m}^{2} / \mathrm{s}$ )
$\mathrm{VZ} \quad$ velocity of aquifer flow ( $\mathrm{m} / \mathrm{s}$ )
FLOWR water flow rate (L/y)

RNWV inverse equilibrium constant for each nuclide
NUMNUC number of nuclides in the inventory specified for the scenario. NUMNUC is supplied from TAPE12, however.
5.223 Geology and Rainfall Data. The information pertinent to the unsaturated soil column model is input from TAPE10. All the inputs are in free format.

## Card 1: JK, ND, IN

JK number of soil layers
ND number of entries in the hydraulic conductivity and pressure head tables
IN number of the soil layer initially containing the waste.
Card 2: DELW water content increment for the pressure head and hydraulic conductivity tables. Fractional water content is the independent variable for the two soil property tables, which are input on ( ards 5 and 6 , respectively. The fractional water content table is generated ir ternally in UNSAT. The first entry is 0 ; subsequent entries are equally spaced with increment DELW. The total number of entries is ND.

## Card 3: RAIN, DRY

RAIN rainfall rate during wet period $(\mathrm{m} / \mathrm{s})$
DRY evapotranspiration during dry period $(\mathrm{m} / \mathrm{s})$; DRY $<0$
Card 4: $\quad(\mathrm{DD}(\mathrm{I}), \mathrm{I}=\mathrm{I}, \mathrm{JK}+1)$
DD locations of soil layer boundaries $(\mathrm{m}), \mathrm{DD}(1)=0$
Card 5: $\quad(\mathrm{P}(\mathrm{I}), \mathrm{I}=1, \mathrm{ND})$ pressure head $(\mathrm{m})$. Pressure head is input as a tabular iunction of fractional water content. Fractional water content entries are generated internally within UNSAT, start with zero, and are equally spaced with increment DELW. Unfortunately, a minimum water content value exists at which the pressure head curve becomes asymptotic. This water content corresponds to the amount of moisture which is tound to the internal surfaces of the soil or rock complex due to the forces of molecular attraction (bound water). Pressure head loses physical significance below this asymptote; thus, tabular entries in this region of low water content represent artificial circumstances. The next to last entry in the pressure head table should correspond to saturation water content (zero pressure head). The final entry should have a positive value for P .

Card 6: $\quad(E(i), \mathrm{I}=1, \mathrm{ND})$ Hydraulic conductivity $(\mathrm{m} / \mathrm{s})$, input as a tabular function of fractional water content. Hydraulic conductivity, like pressure head, is asymptotic at some positive value of water content, and tabular entries in the region below the asymptote are without physical basis.

Card 7: $\quad(\mathrm{W}(1), \mathrm{I}=1, \mathrm{KK})$ Initial water content at each soil layer boundary, fraction by volume.
Card 8: HDRY, HWET, WATL, WATH
HDRY pressure head $(\mathrm{m})$ which corresponds to minimum water content

HWET pressure head $(\mathrm{m})$ which corresponds to satyration water content, normally zero.
WATL minimum water content (bound water), volume frection.
WATH maximum water content at soil saturation, volume fraction.
Cards 9-11 contain information specific to the soil layers. Input consists of JK sets of these three cards; however, some sets may be skipped. The sets are input in ascending layer order; skipped layers are assigned the characteristics of the next specified layer. Layer JK cannot be skipped.

Card 9: LYR layer number
Card 10: $\quad(\mathrm{XD}(\mathrm{I}, \mathrm{LYR}), \mathrm{I}=1$, NUMNUC $)$ distribution coefficients $(\mathrm{mL} / \mathrm{g})$ for each nuclide in the inventory list.

Card 11: CONCOF (LYR), DNSTY (LYR)
CONCOF conductivity factors for correction of hydraulic conductivity (Card 6) to the soil characteristics of a layer (dimensionless). CONCOF (1) is internally set to 1 .

DNSTY density of soil layer $\left(\mathrm{kg} / \mathrm{m}^{3}\right)$.
Card 12: (SOLFAC(I), I = 1, NUMNUC) nuclide solubilities $\left(\mathrm{Ci} / \mathrm{m}^{3}\right)$ for each nuclide in the inventory list (See Table 3).

Card 13: TRAR trench area subject to infiltration $\left(\mathrm{m}^{2}\right)$.
5.2.2.4 Erosion Data. The information required for the solution of the wind erosion equation is input from TAPE8. The entire group of cards is entered using free format.

Card 1: IK1, PAG84, IK3, MAT, ANGL, ANGWND, HTBR, FW, FL, R, CK5, CK8, CK11, CK13, SAIR

IK1 soil erodibility calculation control variable $=1$, calculate soil erodibility $\mathrm{I}^{\prime}$ for flat land $>1$, calculate I' and apply windward knoll slope correction to result as follows:
$=2$, correction is based on potential soil loss from top of knoll
$=3$, correction is based on potential soil loss from that portion of windward slope where drag velocity and windward drag are the same as on top of the knoll (from about the upper third of the knoll)

PAG84 percentage of soil in the top layer of the burial site that is greater than 0.84 mm diameter ( $1 \leq$ PAG84 $\leq 80$ )

IK 3 soil ridge roughness correction control variable
$=1$, ridge roughness is specified, and determines the ridge roughness factor $\mathrm{K}^{\prime}$.
$=2$, ridge height and spacing are specified. Calculate ridge roughness, then use roughness to determine $K^{\prime}$.
$=3, \mathrm{~K}^{\prime}=0.5$ (minimum value for ridge roughness factor)
$=4, \mathrm{~K}^{\prime}=1.0$ (maximum value for ridge roughness factor)
MAT mean annual temperature $\left({ }^{\circ} \mathrm{C}\right)$. MAT is a floating point variable
ANGL lower compass direction that parallels the long axis of the field $\left(0^{\circ}\right.$ is North, $90^{\circ}$ is East, 0 to 179).

ANGWND compass direction from which wind is blowing ( $0^{\circ}$ is North, $90^{\circ}$ is East, 0 to 359).

HTBR barrier height ( m )
FW field width (m) $0<$ FW < 1524 m
FL. field length ( m )
R vegetative cover $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$. Limits on R depend on cover type and are given in Table 7.

CK5 $\quad>0$, specifies anchored small grain stubble
CK8 $\quad>0$, specifies small grain crops in seedling and stooling stage
CK11 $>0$, specifies flat grain sorghum stubble
CK13 $>0$, specifies standing grain sorghum stubble
SAIR percentage of eroded soil which remains suspended (lofted soil).
The following cards are optional, and supply additional parameters as required by Card 1 .
Card 2: Include if $\mathrm{IK} 1>1$
KLSP $\quad$ nnoll slope in percent $(0 \leq$ KLSP $\leq 10)$. KLSP is a floating point variable
Card 3: Include if $\mathrm{IK} 3=1$
RDGRGH soil ridge roughness $(\mathrm{m}) .0 \leq R D G R G H \leq 0.254$
Card 4: Include if IK3 = 2
RDGHT soil ridge height ( m )
RDGSP soil ridge spacing (m)
Card 5: Include if CK5 $>0$
CK6 $\quad=0$, specifies standing small grain stubble
$=1$, specifies flat small grain stubble

## Table 7. Maximum input values for equivalent vegetative cover, $\mathbf{R}\left(\mathbf{k g} / \mathrm{m}^{\mathbf{2}}\right)$

$\xrightarrow[\text { Vegetation Type }]{\left.\begin{array}{c}\mathrm{R} \\ \text { (maximum) }\end{array}\right)}$

Small grain crops, seedling, and stooling

| In furrow | 1.3 |
| :--- | :--- |

$\begin{array}{ll}\text { On smooth ground } & 1.6\end{array}$
Anchored small grain stubble
Standing $\quad 3.1$
Flat 1.6

Grain sorghum stubble
Flat 1.4

Standing, $0.2 \mathrm{~m}(8 \mathrm{in}) \quad$.
Standing, $0.3 \mathrm{~m}(12 \mathrm{in}$.) 2.4
Standing, $0.4 \mathrm{~m}(16 \mathrm{in}) \quad$.
Standing, 0.5 m (20 in.) $\quad 3.6$

## Card 6: Include if CK8 $>0$

CK9 $\quad=0$, small grain crops on smooth ground
CK9 $\quad=1$, small grain crops in furrows
Card 7: Include if CK13 $>0$

HT height of standing grain sorghum stubble $(\mathrm{m})(0<\mathrm{HT} \leq 0.5 \mathrm{~m})$
5.2.2.5 Atmospheric Data. The atmospheric data are input from TAPE9. Included are stack (or release) height and energy release rate, for use in the Gaussian plume-rise model, and the wind frequency array, for use in the Gaussian plume-rise model and in the soil-erosion model. The wind frequency array consists of a wind speed vector ( $\mathrm{m} / \mathrm{s}$ ) and up to seven stability classes. Each frequency vector corresponds to the fraction of the total time that the atmospheric conditions fit a particular stability category. Each entry within the frequency vector corresponds to a value of wind speed and represents a fraction of the total time within the stability category at the corresponding wind speed. The total of the fractions of all the frequency vectors should be unity.

Card 1: $\quad \mathrm{SH}, \mathrm{SQ}$
FREE FORMAT

SH elease or stack height (m)

SQ stack energy release rate (W)

## Card 2: NU, KS, NS

## FREE FORMAT

NU number of entries in wind speed vector (maximum of 10)
KS optionally specify one wind speed for the ATMOS calculation
$=0$, use entire wind speed vector
$>0$, use wind speed entry KS , and use a wind frequency of 1 for that entry
NS number of stability categories (maximum of 7)

## Card 3: $\quad U(I), I=1, N U$

## FREE FORMAT

$\mathrm{U} \quad$ wind speed vector $(\mathrm{m} / \mathrm{s})$. Note: If an uncertainty or sensitivity analysis is to be performed with $K$ levels of wind $(K=3$ or 5$)$, then Card 3 must be input here $K$ times. Sce the discussion of the wind frequency array in Section 5.2.3.

Card 4: $\quad(\mathrm{F}(\mathrm{I}, \mathrm{J}), \mathrm{I}=1, \mathrm{NU}), \mathrm{J}=1, \mathrm{NS}$
fREE FORMAT
1 wind speed-stability class frequency array. Input consists of NS lines, NU entries per line. Note: As with Card 3, the array called "Card 4" must be i: put $K$ times here if an uncertainty or sensitivity analysis is performed with $K$ levels of wind.
5.2.2.6 Agricuiture and Population Data. The agriculture and population data are input from TAPE15.

Card 1: NR number of radial increments (maximum of 20)

## FREE FORMAT

Card 2: $\quad \mathrm{RM}(\mathrm{I}), \mathrm{I}=\mathrm{I}, \mathrm{NR}$

## FREE FORMAT

RM distance from the disposal trench to the center of the radial increment (m)
Card 3: BEEF, COWS, (FAGE(I), I = 1, 3), FCA, NCPY, PRODUC
FREE FORMAT
BEEF number of beef cattle per square ilc neter for all radial increments
COWS number of milk cows per square kilometer $f+r$ all radial increments
FAGE age group fraction breakdown by child, teen, nd adult, respectively, for all radial increments

FVA fraction of the total involved area available to plant leafy vegetables

NCPY number of crops per year
PRODUC food crop production ( $\mathrm{kg} / \mathrm{y}-\mathrm{km}^{2}$ )
Card 4: $\quad \operatorname{IPOP}(\mathrm{I}), \mathrm{I}=1, \mathrm{NR}$
FREE FORMAT
IPOP population in each radial increment
5.2.27 Direct Exposure Data. The direct exposure data are input from TAPE20.

Card 1: IST, RANGE, NSH, (MATRL(I), I = 1, NSH)
FREE FORMAT
IST direct source type as specified below
$=1$, point source
$=2$, line source
$=3$, volume source
RANGE distance from the direct source ( m )
NSH number of shielding materials around source container (minimum $=0$, maximum $=5$ )

MATRL composition of shielding materials
$=1$, aluminum
$=2$, iron
$=3$, lead
$=4$, ordinary concrete
$=5$, water
Card 2: $\quad$ THK(1), $1=1$, NSH
FREE FORMAT
THK thickness of shielding material I (m)

## Card 3: A

FREE FORMAT
If IST $=3, \mathrm{~A}=$ radius of the volume source $=$ half length of the volume source (m)

If IST $\quad=2, A=$ half the length of the line source $(\mathrm{m})$
If IST $\quad=1$, this card may be omitted.
5.2.28 Scenario Requirements Summary. Table 1-1 in Appendix I contains a list of scenarios available for BURYIT analysis. Included in the table are a description of the scenario, the waste inventory assumed for the site, and a list of the files required for the calculation.
5.2.3 Uncertainty Analysis Input. The uncertainty analysis input to BURYIT is on file INPUT. It includes the control variable that calls the uncertainty analysis option and the variables which specify the design, the input parameters and their uncertainties, and the responses.

## Card 1: IUNCRT, MAXI, IDERIV

## FREE FORMAT

IUNCRT $\quad=0$, no uncertainty analysis performed
$=1$, uncertainty analysis performed
MAXI $\quad=0$, cumulative population dose calculated
$=1$, maximum individual dose calculated
IDERIV $\quad=0$, no derivatives calculated
$=1$, derivatives calculated for certain responses and input parameters.
If no cards reside on file INPUT, all three variables are set to zero. If the uncertainty analysis option is chosen, and if MAXI $=1$, BURYIT is executed first under the nominal conditions to determine which individual receives the maximum dose. This same individual is then used for all the remaining runs, even though another individual may receive a larger dose on some of these runs. If the uncertainty analysis option is specified, then the uncertainty analysis variables are input using a Namelist READ statement, now described

Each line of namelist input begins in Column 2. The namelist name, IN, is on the first line and is preceded by a $\$$ ihe last line of namelist input is SEND.

SIN
IFL.AG flag for check runs
$=0$, catculate the design and confounding pattern (default)
$=1$, add the nominal BURYIT run
$=2$, do the uncertainty analysis specified by LTYPE and ISTART.
LTYPE type of analysis desired
$=1$, linear (default)
$=2$, linear plus foldover
$=3$, linear plus quadratic
$=4$, linear plus foldover plus quadratic
$=5$, linear plus quadratic, user-specified quadratic factor level
$=6$, linear plus foldover plus quadratic, user-specified quadratic factor level
LPB statistical experimental design flag
$=0$, fractional factorial design (default)
$=1$, Plackett-Burman design
LFAC(N) list of factors from Table 8 to be included in the design. Order is important because it determines the placement of the factors in the confounding patter 1.

LRES(N) List of responses from Table 9. LRES(1) means either the
$\operatorname{MRES}(\mathrm{N}) \quad$ cumulative population dose or the maximum individual dose, depending on the value of MAXI in the BURYIT control input (see Section 5.2.2).

FACTOR(J) Flag to mecify additive or multiplicative uncertainties for input factor J as listed in Table 8, i.e., $\mathrm{J}=1, \ldots, 100$.
$=0$, add tive
$=1$, multiplicative
$\mathrm{C}(1, \mathrm{~J}) \quad$ Uncertainty associated with factor J . Presently the I index is used if factor J has different uncertainties for each age group; i.e., $\mathrm{I}=1$ (child), 2 (teen), 3 (adult). I could also be used for other applications. Otherwise use $\mathrm{I}=1$.

If the uncertainty is additive, i.e., FACTOR(J) $=0$, then for most applications $C(I, J)$ should equal one standard deviation. The program multiplies $\mathrm{C}(\mathrm{I}, \mathrm{J})$ by $\pm 1$ or $\pm \mathrm{P}$, and adds this perturbation to the nominal value (mean), to produce a perturbed value for the factor. When LTYPE $\leq 4$, the program calculates P . When LTYPE $=5$ or 6 , the user specifies P , for example as 2 or 3 .

If the uncertainty is multiplicative, i.e., $\operatorname{FACTOR}(\mathrm{J})=1$, then $\mathrm{C}(1, \mathrm{~J})$ is the multiplier analogous to adding one or P standard deviations in the additive case. The program multiplies or divides the nominal value (median) by $\mathrm{C}(1, \mathrm{~J})$ or $[\mathrm{C}(1, \mathrm{~J})]^{\mathrm{P}}$, to produce a perturbed value for the factor. The perturbed value then has the same sign as the nominal value.

Note: When the uncertainties are additive, ANALYZ its the response as a linear or quadratic function of the factors. When any uncertainty is multiplicative, ANALYZ implicitly uses the logarithm of that factor, rather than the factor itself.

NOLD(J) A list of factors used in a previous sensitivity analysis, the responses for which are input on TAPE44. Some or all of the sensitivity study responses are assumed to be the quadratic factor level responses for the present uncertainty analysis. NOLD is used in conjunction with ISTART (below) and LFAC.

AMU( $1, J$ ) First eight central moments $(I=1,8)$ for Factor $J$. Default central moments are for a normal distribution.

ISTART Sensitivity analysis flag
$=0$, Perform a complete uncertainty analysis as specified by LTYPE

## Table 8. BURYIT uncertainty factor call list

## LFAC

Factor

## Meteorological Factors

1
2
3
4

5

6

7
8
9
10
11-29

## Consumption Factors

Amount of time specified for rainfall for UNSAT
Annual rainfall for ATMOS and EROSIO
Weather frequency array
Plume width crossrange standard error
Plume width vertical standard error
Nuclide deposition velocity
Stack (release) height
Stack energy release rate
Rainfall rate during wet period for UNSAT
Gaussian plume model lack of fit
Not assigned

Breathing rate for child/tuen/adult
Drinking water use per person, child/teen/adult
Fruit, vegetable, grain, and root crop consumption (maximum exposed individual) child/teen/adult

Leafy vegetable consumption (maximum exposed individual) child/teen/adult

Milk consumption (maximum exposed individual) child/teen/adult
Meat consumption (maximum exposed individual) child/teen/adult
Grain, root crop, etc., consumption (average individual) child/teen/adult
Milk consumption (average individual) child/teen/adult
Meat consumption (average individual) child/teen/adult
Vegetable consumption (average individual) child/teen/adult
Not assigned

Table 8. (continued)

| LFAC | Factor |
| :---: | :---: |
| Agriculte ral Factors |  |
| 50 | Leafy vegetable crop density |
| 51 | Areal grass density |
| 52 | Grass consumption rate by steer |
| 53 | Beef dressout fraction |
| 54 | Beef slaughter fraction |
| 55 | Cow weight (kg) |
| 56 | Milk production per cow (L/y) |
| 57 | Soil pool areal density ( $\mathrm{kg} / \mathrm{m}^{2}$ ) |
| 58 | Crop deposition fraction (except iodine) |
| 59-69 | Not assigned |
| Operations Factors |  |
| 70 | Total nuclide inventory ( $\mathrm{Ci} / \mathrm{m}^{3}$ ), $\Sigma$ ORI |
| 71 | Release fraction for Path 1 |
| 72 | Release fraction for Path 2 |
| 73 | Release fraction for Path 3 |
| 74 | Release fraction for Path 4 |
| 75 | Total human population |
| 76 | Direct exposure time (CUMT for DIRECT) |
| 77 | Distance from direct source (RANGE) |
| 78 | Thickness of Shield 1 |
| 79 | Thickness of Shield 2 |
| 80 | Thickness of Shield 3 |
| 81 | Thickness of Shield 4 |
| 82 | Thickness of Shield 5 |
| 83 | Thickness of Shield 6 |
| 84 | Waste deposited in puncture wound |
| 85-100 | Not assigned |

## Table 9. BURYIT uncertainty analysis response calls

| LRES | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MRES | Total Dose | Direct or Puncture Dose | Dose by Nuclide | Dose by Distance | Dose by Pathway | Dose by Organ | Dose by Age Group |
| 0 | TOTAL |  |  |  |  |  |  |
| 1 | None | Direct | Order in TAPE12 | Order in TAPE15 | Cloud <br> shine | Whole body | Child |
| 2 |  | Puncture wound |  |  | Ground shine | Bone | Teen |
| 3 |  |  |  |  | Inhalation | Liver | Adult |
| 4 |  |  |  |  | Resuspended particle inhalation | Kidney |  |
| 5 |  |  |  |  | Waler ingestion | Gonad |  |
| 6 |  |  |  |  | Leafy vegetation | Lung |  |
| 7 |  |  |  | - | Root crop, etc. | GI |  |
| 8 |  |  |  |  | Milk | Thyroid |  |
| 9 |  |  |  |  | Beef | Skin |  |

[^3]Example: A previous sensitivity analysis used:

```
$IN
LFAC(1) = 27, 7, 9, 1, 2, 8, 20, 12,
ISTART = 1,
```


## §END

The present uncertainty analysis is to use factors $7,8,9,27,2,12$

## SIN

$\operatorname{LFAC}(1)=7,8,9,27,2,12$, $\operatorname{NOLD}(1)=2,6,3,1,5,8$, ISTART $=2$,
.

## SEND

IPRINT Flag for printing additional design information, including design generators and one and two factor aliases (applicable to fractional factorial designs only)
$=0$, no (default)
$=1$, yes
SEND last line of namelist input.
If LTYPE > 4, user specifies quadratic factor levels, one entry per factor, immediately following \$END, FORMAT(10F5.0).

If the wind-frequency array is specified as a variable input factor $(\operatorname{LFAC}(\mathrm{J})=3)$, the associated uncertainty is not straightforward. Instead, the wind frequency varies in a complex fashion as atmospheric conditions become more or less stable. To accommodate the representation of uncertainty in such variables, a routine was incorporated into BURYIT which reads several sets of values for the entire array, and assigns the set to the array which represents the appropriate factor level for the present run. The sets of values are input in the following order:

1. nominal case
2. linear $(+1 \sigma)$ factor level
3. linear $(-1 o)$ factor level
4. positive quadratic factor level ( $\mathbf{P}$ )
5. Negative quadratic factor level.

The wind-frequency factor involves the perturbation of both the wind speed vector $(\mathrm{U})$ and the wind frequency array (F). For LTYPE $\leq 2$, Card 3 (Section 5.2.2.5) consists of three lines to input nominal and linear factor levels, and for LTYPE $\geq 3$, Card 3 consists of five lines to specify $U$ for nominal, linear, and quadratic factor levels. In a similar manner, Card 4 of Section 5.2.2.5 (the F array) consists of either 3 or 5 sets of NIS lines per set, to specify the wind frequency vectors for NS stability categories for nominal and linear factor levels (and quadratic factor levels for LTYPE $\geq 3$ ).

If the user desires variable input factors not presently included in Table 8, then installation is fairly simple. Assign an unused index number (up to 100) to the new variable input factor, and update the code at the appropriate location with

CALL DIALOT(ALP, JNDEX, INDEX, FDIALA, FDIALM)
VARIABLE $=($ VARIABLE + FDIALA $) *$ FDIALM
where
ALP $=$ current factor level

JNDEX $\quad=1$ for simple variable input factors. For factors which have different values and/or uncertainties according to age group or other characteristics, JNDEX is an index variable which is used to assign the appropriate uncertainty according to the index of the characteristic. For assigning uncertainties by age group, the updated coding becomes:

DO 30 JNDEX $=1,3$
CALL DIALOT(ALP, JNDEX, INDEX, FDIALA, FDIALM)
30 VARIABLE $($ JNDEX $)=($ VARIABLE $(J N D E X)+$ FDIALA $) *$ FDIALM

| for VARIABLE $(1)$ | $=$ child value |
| :---: | :--- |
| $\operatorname{VARIABLE}(2)$ | $=$ teen value |
| $\operatorname{VARIABLE}(3)$ | $=$ adult value |

INDEX $=$ assigned index number for the new variable input parameter

FDIALA = additive uncertainty amount

FDIALM $=$ multiplicative uncertainty factor

VARIABLE $=$ new variable input parameter to be used in uncertainty analysis.
5.2.4 ANALYZ Input. The main control variable for ANALYZ is on file INPUT. It determines the remainder of the input requirements, as follows:

Card 1: MBRANCH ANALYZ option control variable

MBRANCH $=1$, Perform uncertainty analysis using the responses from BURYIT. The required inputs are:

1. TAPE4, which was written by BURYIT, and contains the uncertainty analysis control variables and the BURY!T responses
2. ANALYZ uncertainty analysis control variables on file INPUT in Namelist PL(details below).

MBRANCH $=2$, Perform uncertainty analysis using responses from BURYIT, then do risk calculation for "cumulative population dose" or "maximum individual dose" response, as specified in BURYIT iaput. The required inputs are:

1. TAPE4 (as above)
2. Namelist PL on file INPUT (as above)
3. Scenario probability file (TAPE97)
4. TAPE98, which contains the results of risk calculations for previously done scenarios (optional).

In this case, TAPE99 should be saved, to use $\approx$ TAPE98 in the next run.
MBRANCH $=3$, Plot probability density function and cuniulative probability distribution for likelihood (or probability of occurrence) for specified scenario. Required inpu's are:

1. Scenario probability file (TAPE97)
2. Plot titles and scenario number on file INPUT (details below).

MBRANCH $=4$, Plot probability density function and cumulative probability distribution for a user-input density. Required input consists of plot titles and the lower four statistical moments of the desired density which are on INPUT (details below).

Card 2: LTNAME(I), $\mathrm{I}=1,11$. Input if $\mathrm{MBRANCH} \geq 3$.
FORMAT(8A10)
Plot title for PDF and CDF plots of scenario occurrence probability (MBRANCH $=3$ ) or user-input density (MBRANCH $=4$ ). For ANALYZ calculations using the responses from BURYIT, the plot titles are written by BURYIT onto TAPE4.

Card 3: $\operatorname{LYNAME}(1), I=1,4$. Input if $\operatorname{MBRANCH} \geq 3$.
FORMAT(8A10)
X-axis title for PDF and CDF plots. Cards 2 and 3 are either both included or both omitted.
Card 4: AMEAN, SIGMA2, RTBT1, BETA2, LPRINT. Input if MBRANCH $=4$.

FORMAT(4E15.5, I5)

AMEAN mean of user-input density

SIGMA2 variance
RTBT1 skewness coefficient, $\beta_{1}$
BETA2 kurtosis coefficient, e.g., 3 for the Gaussian distribution.
LPRINT print control variable
$\neq 0$, print complement of the cumulative distribution (CCD)
$=0$, bypass CCD printout.

## Card 5: ISCEN Input if MBRANCH $=3$.

## FORMAT(A4)

Scenario code specified for piot of occurrence probability PDF and CDF.
Card 6: $\quad$ Nảmelist PL optional input, MBRANCH $\leq 2$. This input block specifies the uncertainty analysis options for ANALYZ. Each line of namelist input begins in Column 2. The namelist name PL is on the first line and is preceded by $\$$.
\$PL
LISTR(I) List of responses for which uncertainty analysis is to be performed. Responses are specified by their numbers in the order in which they are designed in LRES and MRES (Section 5.2.3). If the risk calculation is requested. LISTR must include the "cumulative population dose" or "maximum individual dose", as appropriate ( $\operatorname{LRES}(\mathrm{I})=$ ). A maximum of 100 responses may be requested by LISTR.

LTRAN(1) Option to transform the independent variables (factors) by taking the exponential, logarithm, inverse, square, or square root prior to performing the regression analyst: in ANYOLS. The option was not installed in the original version of ANALYZ and has not been incorporated into the present version.

LOGFIT If LOGFIT $=0$, the response is fitted to a linear or quadratic function of the factors.

If LOGFIT $=0$, the natural logarithm of the response is used instead. The default is LOGFIT $=0$. It is useful to set LOGFIT $=0$ when the uncertainties on the factors are multiplicative and the response is approximately proportional to the product of the factors. In this case, the multiplicative uncertainties cause ANALYZ to use the logs of the factors rather than the factors themselves; and, the log of the response is approximately proportional to the sum of the logs of the factors so a good fit can be obtained. The distribution finally plotted is for the original response, not for its logarithm.

LPRINT Printout option for ANYOLS regressions, according to the following schedule:

|  | No Output | No Output | Summary |
| :--- | :--- | :---: | :---: |
| FINAL | Summary | 00 | 01 |
| $\underline{\text { MODEL }}$ | Summary plus regression <br> coefficients | 10 | 10 |
|  |  | 20 | 21 |

IPLT

NSPV

KKRIT

Plot control variable
$=0$, No plots (default)
$=1$, Up to six plots for current scenario of options requested by MBRANCH:
a. Residuals of response surface equation
b. Fitted values versus data
c. Probability density function (PDF) of response surface
d. Cumulative distribution function (CDF) of response surface
e. PDF of scenario risk
f. CDF of scenario risk
$=10$. Bar graph of scenario risk, at probability value specified by NSPV. If risk file (TAPE98) is present, include previous scenario risk values in bar graph.
$=11$. Same as $\operatorname{IPLT}=1$ plus IPLT $=10$.

Control the probability level of the risk plotted in the bar graph, when $\mathrm{IPLT} \geq 10$. The probabilities corresponding for values of NSPV are:

| NSPV | Probability | Upper Tail |
| :--- | :--- | :---: |
| 1 | 0.01 | 0.99 |
| 2 | 0.025 | 0.975 |
| 3 | 0.05 | 0.95 |
| 4 | 0.10 | 0.90 |
| 5 | 0.50 | 0.50 |
| 6 | 0.90 | 0.10 |
| 7 | 0.95 | 0.05 |
| 8 | 0.975 | 0.025 |
| 9 | 0.99 | 0.01 |

The default is NSPV $=7$, corresponding to a $95 \%$ upper bound on the risk.
Stepwise regression strategy option
$=1$, F-statistic
$=2, \mathrm{~S}^{2}$ minimization
\$END Last line of Namelist PL.
Card 7: $\operatorname{STAR}(\mathrm{I}), \mathrm{I}=1$, NFAC. Input, LTYPE $\geq 5$.
FORMAT(10F5.0)
S, ecify factor levels for the quadratic runs (star points) of the design.

## 6. UNCERTAINTY ANALYSIS EXAMPLE-ATMOSPHERIC PATHWAY

In this section, an example of a probabilistic risk assessment is presented in order to illustrate the breadth of analysis possible with the current status of BURYIT/ANALYZ. ANALYZ presently works on the atmospheric release pathway; that is, the subroutines ATMOS and DOSET, Pathway 2-4. The example uncertainty analysis deals with a hypothetical release at a waste processing or packaging facility and the resulting population dose commitment, release scenario P-4 with inventory WS-4. In what follows, the inputs for the BURYIT portion of the package will be discussed first. Then, the necessary inputs for the ANALYZ portion of the package will be discussed. (Note that Table 8 was rearranged subsequently to its present form.) Finally, the results of the uncertainty and risk analyses will be presented. The reader is cautioned that some of the BURYIT input was selected deliberately to magnify the computed release; therefore, there is no similitude to any real facility. The only purpose of this section is to demonstrate the capability for probabilistic risk assessment.

### 6.1 BURYIT Input

The release scenario, Code P-4 (Volume 1, Reference 1, pp. 4-22 and 7-18) was selected to represent the Pathway 2-4 with waste inventory WS-4. The scenario description is:
"Container wi'h volatile substance is ruptured during packaging. Volatile substance escapes to atmosphere."

The waste inventory is described as "LWR operational waste-high concentration" and has 33 nuclides with a total activity concentration of $32 \mathrm{Ci} / \mathrm{m}^{3}$. The scenario release fraction is 1.0 ; the volume of the package is $3 \mathrm{~m}^{3}$; and the release time is 200 hours. The cumulative population dose is calculated.

As discussed in section 5.2.2.1, input Card 6 for Path 2 also requires, in addition to the release time (CUMT), the annual rainfall in $m / y$ (RF), emplacement efficiency (EMEF), time for the rain cycle in hours (TIMWET), and total time in one cycle of rain and dry periods in hours (TIMCYC). The seven input cards entered on TAPE14 are thus:

| Card No. | Entry |
| :---: | :--- |
| 1 | 12 |
| 2 | Atmospheric Path Uncertainty <br> Analysis, Hanford Site: Maximum <br> Individual Dose (Hanford Site refers <br> roughly to population, rainfall, and <br> wind-frequency array.) |
| 3 | 0 |
| 4 | P-4 |
| 5 | WS-4 |
| 6 | 200 0.2 0.5 480 8760 |
| 7 | 3. |

In addition to the above cards, site-specific information must be available on TAPE9 and TAPE15. The former must contain stack height, energy release rate, and the wind speed-stability class array. The latter tape must contain agricultural and population data.

For this example problem, the stack (release) height is taken as 5 m ; and the stack energy release rate is taken as $29288000 \mathrm{~W}(7000000 \mathrm{cal} / \mathrm{s})$. Thus, it is assumed that the vaporizing material is somehow conveyed to an incinerator or furnace for combustion; however, the energy release rate is ctearly excessive for a realistic waste processing facility. Card No. 1 on TAPE9 appears as:

$$
5 \quad 29.288 \mathrm{E}+06
$$

In the example problem, six wind speed groups and seven stability classes are entered. The complete array is used for the calculations. Card No. 10 on TAPE9 appears as:

$$
\begin{array}{lll}
6 & 0 & 7 .
\end{array}
$$

Card No. 3 must contain the wind speed croups in meters per second; for the example problem:

$$
\begin{array}{llllll}
0.67 & 2.46 & 4.47 & 6.93 & 9.61 & 12.29 .
\end{array}
$$

Card No. 4 must contain the frequency for each of the six speed groups by wind stability class. Thus, for the example problem, the nominal array is:

## STABILITY CLASS

| $A$ | $: 017$ | $: 036$ |
| :--- | :--- | :--- |
| $B$ | 017 | $: 036$ |
| $C$ | $: 017$ | $: 036$ |
| $D$ | $: 049$ | $: 029$ |
| $E$ | $: 021$ | $: 023$ |
| $F$ | $: 051$ | $: 070$ |

TAPE15 contains site-specific agricultural and population data as described in section 5.2.2.6 Table 10 is a reproduction of the echo of these data for the example problem. Note that the population is concentrated mainly in the third and fourth of five radial incremental distances from the facility.

### 6.2 ANALYZ Input

This probabilistic risk assessment example was run using 15 uncertairty factors from Table 8 and 52 responses from Tabe 9. The discussion of the results will be limited to only one of the responses. The type of uncertainty analysis for BURYIT is the linear plus foldover plus quadratic, user-specified quadratic factor level; that is, LTYPE $=6$. The uncertainty variable with user-specified quadratic factor level is Wind Speed-Frequency Array, LFAC $(J)=3$. As discussed previously in section 5.2.3, a subroutine in BURYIT processes several complete sets of values for the purpose of uncertainty analysis.

ANALYZ sets up a statistical experimental design for the 15 factors. In this example, the design is the fractional factorial ( $\mathrm{LPB}=0$ ) that was shown previously in Figure 1. The perturbations of the numerical values of the factors away from the nominal value are expressed as number of standard deviations of the factor's distribution; i.e., the perturbations are $\pm 1$ or $\pm 3$ standard deviations. As shown in Figure 1, 62 perturbation runs of BURYIT are called by ANALYZ.

Figure 2 is a reproduction of the Namelist IN printed by the computer package. The first entries are FACTOR(J). It may be seen that of the 30 factors from Table 8, 19 have multiplicative uncertainties $(\operatorname{FACTOR}(\mathrm{J})=1)$, while the remainder have additive uncertainties ( $\operatorname{FACTOR}(\mathrm{J})=0)$. With multiplicative uncertainty, the logarithm (to the base e) is treated as a random variable. There is room for 100 entries in th's array. (The order of factors does not agree with that in Table 8, because Table 8 has been rearran red by the type of factor since the example was run.)

| .021 | .014 | $: 007$ | $: 005$ |
| :--- | :--- | :--- | :--- |
| $: 021$ | $: 014$ | $: 007$ | $: 005$ |
| $: 021$ | $: 014$ | $: 007$ | $: 005$ |
| $: 023$ | $: 017$ | $: 012$ | $: 009$ |
| $: 032$ | $: 045$ | $: 025$ | $: 013$ |
| $: 032$ | $: 045$ | $: 025$ | $: 013$ |
| $: 077$ | $: 041$ | $: 004$ | $: 000$ |

The next entries are the $\mathrm{C}(1, \mathrm{~J})$. These represent one stand ord deviation (standard error) of the distribution of each uncertainty factor. Provision is made for three age groups; thus, there is room for 300 entries. In the case of multiplicative uncertainties, a multiplication factor for the nominal value is used to yield the value at one standard deviation. For this example case, 13 of the 19 multiplicative uncertainties are assigned multiplication factors of 1.1 ( $10 \%$ uncertainty). This was an arbitrary choice made only for expediting the demonstration. All other uncertainties were estimated from the literature, as given in Appendix F.

The next two names in the list, $\operatorname{LRES}(\mathrm{N})$ and $\operatorname{MRES}(\mathrm{N})$, together designate the 52 responses, of which 33 are doses from specific nuclides. A response surface equation (RSE) is obtained for each of the responses. The RSE permits a sensitivity analysis; i.e., the rate of change of a response because of a change in factor level. A combined uncertainty analysis and probabilistic risk computation may be performed (based on the RSE) for any of the responses designated in the PL namelist in the name LISTR(I).

The next two names are LTYPE and LPB. Both of these were commented upon earlier in this section.

The next name, $\operatorname{AMU}(1, \mathrm{~J})$ pertains to the first eight central moments of the statistical distribution for the uncertainty variables. The default values are those for the Gaussian distribution, and these appear 100 times in Figure 2.

The next four names in Namelist IN are IPRINT, NOLD, IFLAG, and ISTART. Only IFLAG has a nonzero value, $\mathrm{IFLAG}=2$. The meaning of these names should be clear from the earlier namelist discussion.

The uncertainty analysis and probabilistic risk calculation options are specified in Namelist PL. Figure 3 shows the PL printout for the example. The name LISTR designates the responses to be treated, in this case, numbers $36,37,39,28$, and 1 .

Table 10. Input data wis TAPE15




[^4]$$
\dot{M O O} \underset{\sim}{\circ} \dot{0} \dot{\sim}
$$
$$
\dot{\operatorname{mig}} \dot{\circ} \dot{\sim} \dot{\sim} \dot{c}
$$
\[

$$
\begin{array}{ll}
m 00 \\
\therefore 00 \\
& 0
\end{array}
$$
\]

$$
\dot{m} \dot{0} \dot{0}
$$

$$
\dot{m o s} 0^{\circ+\pi}=\dot{0}
$$

$$
\dot{m i o s} \dot{\infty}^{-m}: \therefore \dot{0}
$$

muio mino nois

$$
\dot{m i n o ~} \dot{\sim}=\dot{m} \text { miso }
$$

$$
\begin{aligned}
& \text { LPES } \\
& \text { MRES } \\
& \text { LFAC } \\
& \text { ITYPE } \\
& \text { IPA }
\end{aligned}
$$

$$
\begin{aligned}
& \text { m゙் } \dot{\sim} \dot{\circ} \dot{\circ} \dot{\circ}
\end{aligned}
$$

$$
\begin{aligned}
& \text { Mus シ0 0. } \\
& \text { mos }{ }^{*+\infty} 00_{0}^{\circ}
\end{aligned}
$$

$$
\begin{aligned}
& \text { Moo ni: } \\
& \text { mo̊ } \\
& \text { mios }
\end{aligned}
$$

$$
\begin{aligned}
& \text { NI is!ouren } \quad \tau \text { गmil! }
\end{aligned}
$$

 $0000000000000000 \cup 0000000100000000000000000000000004$
$t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t+t$





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1．1．1． 0000072000000100000000000000000000000000000000000




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ベシั o்ősivi

ャロシ o்ō03u


SPI
LISTR
LTRAN

LPRINT
KKPIT
PREF
WSPY
IPIT
SEND

These numbers correspond to the order specified by the names LRES( N ) and MRES( N ) in Namelist IN. In particular, the responses are dose from cloud shine, dose from ground shine, resuspended particle inhalation dose, dose from nuclide plutonium-242, and cumulative population dose.

Continuing with Namelist PL, LTRAN is not used. LPRINT $=0$ means that no printout is desired from the regression subroutine ANYOLS. KKRIT $=4$ means that PRESS is the regression strategy option. Finally, the probability bounds for the uncertainty and risk distribution functions for all responses are specified by $\operatorname{PREF}(1)$; in this case, 0.05 and 0.95 .

### 6.3 Example Problem Results

6.3.1 Nominal Dose Computational Results. The nominal dose calculation is that using the "bestestimate" values of the parameters; i.e., the mean value when uncertainty is additive and the median value when the uncertainty is multiplicative. This is the "best estimate" of the population dose given the occurrence of the release of 96 Ci of inventory WS-4 (Appendix D). The computed results for the example problem are shown in Figure 4. (The effective whole body dose results were omitted to save space.) The cumulative population dose, which is also the population whole body dose, is 6930 person-rem with the largest contribution occurring in a radial increment centered at 40225 m from the point of release. A brief computation shows that $98 \%$ of the dose results from the three nuclides cesium-137, cesium-134, and plutonium-242. Furthermore, $97 \%$ of the dose is transported through the two pathways, resuspended particle inhalation and leafy vegetable inhalation.
6.3.2 Sensitivity Results. Lists of factor names and response names are printed by both the BURYIT and ANAL YZ subroutines (see Figure 5). Following the Namelist PL, which is printed by ANALYZ, the experimental design matrix appears. This is the same as Figure 1 in this case.

Following the design matrix, the confounding array appears. This provides the same information as Table 5, but in a different form. The confounding array is reproduced in Figure 6. The numbers across the top and down the side of the array represent the numbers of the factors in the same order as printed out in the list of factor names (see
previous paragraph). The entries in the confounding array represent the groups wherein the secondorder interactions are confounded (aliased). For example, the $(1,2)$ interaction is in group nine. Thus. the $(1,2)$ interaction coefficient cannot be estimated separately from any other interaction in group nine using only the present design matrix. (Additional periurbation runs or fewer factors would be necessary to do so.) This is the reason for the importance of the order in which the factors are entered by LFAC(N). As shown later, it is often possible to know that certain interaction terms must be zero because of the model structure.

The results printed next, but not reproduced here, include the levels of the factors in each perturbation run and the computed response. Then follows a tabulation of the actual response, the regression equation response (YHAT), and their difference (residual), also not reproduced here. The residuals are important for determining whether the response surface equation is adequately fitted. Plots of the residuals or YHAT versus response may also be obtained, e.g., as in Figures 7 and 8 .

A tabulation of the coefficients of the response surface equation follows in the ANAL.YZ printout. For example, Table 11 contains response surface coefficients for dose from cloud shine where the nominal (unperturbed) value is $8.34 \mathrm{E}-03 \mathrm{rem}$. This table shows linear coefficients of five factors. These are the absolute sensitivity coefficients, normalized with respect to standard deviation, for the factors: stack energy release rate [LB(2)], weather-frequency array [LB(3)], stack release height [LB(4)], fruit, vegetables, and grain consumption rate [LB(7)], and the lack-cl fit of the Gaussian plume model [LB(15)]. Such coefficients should always be subjected to critical examinations.

Factors 2, 3, 4, and 15 are "atmospheric" factors and clearly should affect the computed value of dose from cloud shine; and a study of the BURYIT model confirms this. On the other hand, Factor 7 is a "consumption" factor and does not enter the BURYIT model for cloud shine at all. This may be confirmed by the identical computed responses in perturbation Runs 47 and 48 where Factor 7 is set to $\pm 3$ (standard deviations) and all other factors are at the nominal levels (the two star points). Thus, the Factor 7 coefficient is spurious, and the reason for this is lack-of-fit by the quadratic RSE. Using either the defining relations printed by subroutine DESIGN or Table 5, it can be shown



Figure 5. Factors and responses.


Figure 6. Confou ading array.



[^5]Figure 8. Dose from cloud shine versus response surface equations.

Table 11. Coetficients of response surface equation for dose from cloud shine from the example problem

| Term Type |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Factor Number | Linear | Factor Number | Quadratic | Factor <br> Pair | Interaction |
| 1 | 0 | 1 | 0 | 1,2 | 0 |
| 2 | -5.04E-05 | 2 | 0 | 1,3 | 0 |
| 3 | $4.86 \mathrm{E}-04$ | 3 | -2.27E-05 | 1, 4 | 0 |
| 4 | $-2.64 \mathrm{E}-05$ | 4 | 0 | 1,5 | 0 |
| 5 | 0 | 5 | 0 | 1,6 | 0 |
| 6 | 0 | 6 | 0 | 1,7 | 0 |
| 7 | -4.50E-06 | 7 | 0 | 1,8 | 0 |
| 8 | 0 | 8 | 0 | 1,9 | 1.25E-04 |
| 9 | 0 | 9 | 0 | 1, 10 | -1.32E-05 |
| 10 | 0 | 10 | 0 | 1,11 | $-2.74 \mathrm{E}-05$ |
| 11 | 0 | 11 | 0 | 1,12 | 0 |
| 12 | 0 | 12 | 0 | 1,13 | 0 |
| 13 | 0 | 13 | 0 | 1,14 | 0 |
| 14 | 0 | 14 | 0 | 1,15 | 0 |
| 15 | 2.26E-04 | 15 | $3.02 \mathrm{E}-05$ | 2, 5 | 0 |

that LB(7) is aliased with the triple interaction (2, 3, 15). (Ways to do this are discussed in texts on experimental design.) Since LB(2), LB(3), and $\mathrm{LB}(15)$ are relatively large, their triple interaction is a reasonable identification of the source of lack-of-fit.

Table 12 also shows two quadratic coefficients, those being for Factors 3 and 15. In other words, the behavior of the response is somewhat nonlinear with variation in these two factors. At a perturbation of one standard deviation in Factor 3, the quadratic coefficient reduces the response change
by about $10 \%$ of that predicted by the linear coefficient. Conversely, a perturbation of one standard deviation in Factor 15 increases the response change by about $10 \%$ over that of a linear variation only.

Three interaction coefficients also appear in Table 11. As stated earlier, these coefficients are confounded (or aliased) and do not necessarity correspond to the factors indicated by the table. (As will now be seen, none of these coefficients corresponds correctly.) Additional critical examination is required in order to identify the interactions in, for example, a sensitivity study.


``` * * * * も * * *
```




```
* * * * * * * * *
```



ATMOSPHERIC DATH D A TEST CASE, HANFORD STTE: CUMULATIVE PQPUL. DOSE
SCENARIO 1Th, TVVENTNQY WS-4. DES * 36: DOSE FPOM CLOLD SHTNE (PERSON-PEMI

MF\&N-TNTEDCEOT * * * * * * * * * * * * * * * * * * * * * * . . . . 7549731F-95




THTR CENTRAL MDFENT $T$ FU3DLI $\ldots \ldots+\ldots * * * *-2278294 E-10$
CNEEETPTENT DF SKFHNESS SQUARED (RETA11 = . $1748735 E-01$
CDEEETCTENT OF SKENNESS (RTRT1I * * * * * * * $\quad-\quad 1322397 E+00$

ETIOTH CFNTRAL MOMENT $(M U 4 D L 1 * * * * * * * \quad .2973626 E-12$
PNEEET*TENT TF KURTOSTS (AETA2) ******** * $\quad$. $2101864 E+01$

A rule-of-thumb in examining an interaction from a response surface is that the coefficient probably does not correspond to the indicated factors if one factor or the other does not have a linear coefficient. Another way to say this is that a factor usually does not interact with others if it has no main effect. This rule, thus, suggests that not one of the three coefficients corresponds to the indicated factor pairs: $(1,9),(1,10)$, and (1, 11). In fact, the interactions probably result from the three factors with real linear coefficients.

Identification of a probable factor pair is facilitated by Table 5 or the confounding array in Figure 6 , from which the group containing the $(1,9)$ pair is repeated here.

$$
\begin{gathered}
1-9 \\
2-7 \\
3-15 \\
4-5 \\
6-14 \\
8-12 \\
10-11
\end{gathered}
$$

The only pair here that also has linear coefficients is $(3,15)$. Furthermore, a study of the cloud shine dose model in BURYIT shows that only the $(4,5)$ pair could possibly enter the calculation aside from (3, 15). Examination of the responses at the star points of Factors 4 and 5 reveals that changes in the response occur at the third significant figure, as might be expec. ed for insignificant linear coefficients. Thus, it is evident that the interaction coefficient printed as the ( 1,9 ) pair realty represents the (3, 15) pair. In a like manner, it may be determined that the $(1,10)$ pair in Table 11 is actually the $(2,15)$ pair; and the $(1,11)$ pair is actually the $(2,3)$ vair. This sort of analysis would be essential if - sparate statistical distribution functions for the ff tors were being used in the subsequent evaluawon of the uncertainty in the respr se and the risk. That is, it would be necessary to maich the proper distribution function to each factor.
6.3.3 Response Uncertainty Results. The results next printed out by the ANALYZ subroutine deal with the uncertainty in the response variables of interest, LISTR. For each, a table containing the moments of the statistical distribution function of the response is printed. (This information is computed by the subroutine SOERP.). Table 12 contains the moments for the dose from cloud shine response. Potentially interesting information in the
table or derivable from it includes the variance (3.10E-07) and the coefficient of variation (ratio of standard deviation to mean value expressed as a percentage, $66 \%$ ). The coefficient of skewness indicates, by a negative value, skewness to the ieft, and the coefficient of kurtosis (3.1) indicates a distribution function somewhat more "peaked" than a Gaussian function.

Perhaps the most important result in this area is the identification of the largest contributors to the uncertainty in the response. This is given in a table of tractional contributions to variance, and an abbreviated table for the dose from cloud shine is given in Table 13. The interpretation of this table is that $\sim 98 \%$ of the uncertainty in the computed dose from cloud shine arises from two factors only, numbers 3 and 15. Factor 3 is the wind-frequency factor, and Factor 15 is the lack-of-fit in the atmospheric transport model relative to field observations. If one wanted to reduce the uncertainty in this response, then one would first concentrate efforts on reducing the uncertainty in the wind-speed-frequency array. [Actually, one would probably not want to spend the effort to reduce the response uncertainty in this case, because the cloud shine dose is smaller than the total dose ( $6.93 \mathrm{E}+$ $03 \mathrm{rem})$ by a factor of eight million.]

Table 13. Fractional contributions to uncertainty in cloud shine dose

| Term |  | Fraction |
| :---: | ---: | ---: |
| LB(3) | 0.764 |  |
| LB(15) | 0.165 |  |
| CB(3,15) | 0.050 |  |
|  | sum $=0.979$ |  |

The central moments of the response are matched to one of those of a family of mathematical distributions called the Pearson family. This is done in the subroutine PDFPLOT. Certain parameters of the selected Pearson distribution are printed, and this is followed by a tabulation of the fitted density function and the comptementary cumutative distribution function (not reproduced here). Interpolations may be accomplished using the tabulation, and plots of the function may be obtained if desired.

The density function of dose from cloud shine appears in Figure 9. The two probability points on the plot are those specified in the input term PREF(I). Selected probability points are also tabulated, as shown in Table 14 for dose from cloud shine. The interpretation of the table is, for example: the prebubility of population dose from cloud shine exceeding $1.78 \mathrm{E}-03 \mathrm{rem}$ is 0.05 .
6.3.4 Risk Uncertainty Results. The next group of results in the printed output deals with the risk uncertainty. The probability of the occurrence of the release event and its error factor are input with the scenario number. For Scenario P-4, the probability a ssigned is 0.001 with an error factor of 5 . A response surface equation for the risk is computed, and the subroutine SOERP computes the properties of the statistical distribution of the risk. In the case of dose from cloud shine, Table 15, the mean risk is $1.36 \mathrm{E}-06$ rem and the standard deviation is $2.25 \mathrm{E}-06 \mathrm{rem}$. The resulting coefficient of variation is $166 \%$. The risk distribution is positively skewed, RTBTI $=6.8$, and is far removed from a Gaussian distribution, BETA2 $=150$, as may be
seen in Figure 10. The larger contribution to the risk uncertainty variance $(58 \%)$ is from the occurrence probability uncertainty. Selected probability values are listed in Table 16. From this tabulation, one can say that the probability of the risk from cloud shine exceeding $4.97 \mathrm{E}-06 \mathrm{rem}$ is 0.05 . When more than one scenario is run, bar charts comparing the total risks are a graphics option. In order to illustrate this option, a bar chart for a hypothetical set of scenarios is shown in Figure 11.

### 6.3.5 Sensitivity and Uncertainty for the Other

 Responses. The reader may recall that five responses were included in the call for risk evaluation, namely LISTR. In addition to dose from cloud shine, the responses treated include dose from ground shine, resuspended particle inhalation dose, dose from plutonium- 242 , and cumulative population dose. Since the tables and figures are siiailar to those already presented, they are not included in this report. Furthermore, the interpretation of the dose sensitivity analyses, the dose uncertainty analyses, and the risk evaluations would be done in the same way as described here.

Table 14. Complementary cumulative probability for dose from cloud shine

```
                    TABLE OF PROBABILITY POINTS
                    FOR SELECTED PROBABILITY VALUES
HEAN=.84429E-33 VARIANCE=.30962E-05
    .990 .4008123E-04
    .975 .9317651E-04
    .950 . . 1684541E-03
    .900 .2993977E-03
    .502. .8885119E-03
    .100 .1574571f-02
    .050 .1776752t-02
    .022 .1954732t-02
    .010 .2166160E-02
```

Table 15．Moments of risk due to cloud shine


```
**********
```



```
********** PRO******* PRDERTIES JF THE FUNCTIDN DF INDEPENDENT RANDOM VARIABLES
```



```
**もも**&も**
```




```
    ATMOSPHERIC PATH P R A TEST CASE, HANFORD SITE: MAXIMUM INDIVIDUAL DOSE
    RISK CALCULATIOY FOR RESPONSE S 36: LOSE FRON CLOUS SHINE (PERSON-REM)
```


MEAN-INTERCEPT**************************0*
MEAN................................................ . . . . . . . . $3362553 E-C 5$
SECOND MOMENT (EDEL2)........................ . . $5079043 E-11$

THIRD MOMENT (EDEL3)........................... . $7794355 E-16$
THIRD CENTRAL MOMENT (MU3DL)............. . . $7794355 E-16$
COEFFICIENT OF SKEWNESS SQUARED (BETA1) - .4636761E+C2
CJEFFICIEVT OF SKENNESS (RTBT1).......... . . $6809377 E+C 1$
FOURTH MOMENT (EDEL4)......................... . . $3861339 E-20$
FOURTH CENTRAL MOMENT (MU4DL)............ . . . $3861339 E-20$
COEFFILIEVT OF KURTOSIS (BETAL).......... . . $1496836 E+C 3$


Figure 10. Densitv function for risk due to cloud shine.

Table 16. Complementary cumulative probability for risik diue to cloud shine

$$
\begin{aligned}
& \text { MEAN= . } 23626 E-35 \text { VARIAMCE = . } 5079 C E-11 \\
& .997 \quad .2930283 E-06 \\
& .775 \quad .2965532 E-06 \\
& .952 \quad .3024290 E=06 \\
& \text {. } 9 \mathrm{CR} \quad .3141775 \mathrm{E}=06 \\
& .500 \quad .5811416 E-06 \\
& .100 \quad .3254435 E-C 5 \\
& .250 \quad .4965645 \mathrm{E}=05 \\
& .025 \quad .5959553 E-05 \\
& .010 \quad .9468322 E-25
\end{aligned}
$$



ATMOSPHERIC PATH UNCERTAINTY ANALYSIS, HANFORD SITE: MAXIMUM INDIVIDUAL DOSE

Figure 11. Bar chart illustration.

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## APPENDIX A

# EVALUATION OF POTENTIAL DOSE RECĖIVED BY CONTAMINATION RESULTING FROM A PUNCTURE WOUND 

S. K. Rope

October 1983

## APPENDIX A

## EVALUATION OF POTENTIAL DOSE RECEIVED BY CONTAMINATION RESULTING FROM A PIJNCTURE WOUND

BURYIT, a risk assessment code for low-level waste ${ }^{\text {A-1 }}$ contains 15 scenarios (out of 302 ) which involve a puncture wound of an individual. At present, the code correlates the potential dose received by contamination at the wound site to inhalation of air containing a certain concentration of nuclides. No directions or basis for this correlation are given in the code. Thus, a logical way of modeling the potential dose in the puncture wound scenarios is needed.

The system can be viewed as follows:


Once the contamination reaches the blood, the metabolism and dose can be described by the models which are used in the rest of the code. That is:

$$
\begin{equation*}
\text { DCF }_{p}=\frac{\text { DCF }_{\text {ing }}}{f_{1}} \tag{1}
\end{equation*}
$$

where

DCF $_{p}=$| Dose conversion factor describing |
| :--- |
| the dose per unit activity injected |
| into blood (Sv/Bq) |

DCF $_{\text {ing }}=$| Dose conversion factor describing |
| :--- |
| the dose per unit activity ingested |
| (Sv/Bq) |

$\mathrm{f}_{1}=$| Fraction of the activity ingested |
| :--- |
| which is assimilated into blood. |

DCF ing and $f_{1}$, values for the 44 nuclides in the BURYIT waste streams were obtained from ICRP Publication 30.A-2 Resulting DCF $\mathrm{p}_{\mathrm{p}}$ values are listed in Table A-I.

Thus, it remains to describe the deposition and translocation of contamination from the wound site. Much of the animal and human data availabte on radionuclide contamination of wounds relate to the transuranics, because of the hazard of these
long-lived nuclides when internal to the body. Several data summaries and case histories were studied. A-3, -4, -5, -6

The ICRP, in Publication 10A, discusses deposition in skeleton and excretion after contamination of wounds: A-7
"If one knew the retention function for plutonium in a wound, $\mathrm{W}_{5}(\mathrm{t})$, one could calculate its rate of movement from the site to blood, ${ }^{W}$ ' ${ }_{5}(t)$, on the assumption that all the material leaving the site of deposition went first to blood.
"The study of Hamilton et al., on the retention of plutonium at the site of intramuscular injection in rats showed that it left the site according to a power function of the form $\mathrm{W}_{\mathrm{s}}(\mathrm{t})=\mathrm{t}^{-\mathrm{b}}$, where $\mathrm{b}=0.07$ for $\mathrm{Pu}^{+3}, 0.21$ for $\mathrm{Pu}+4$, and 0.34 for $\mathrm{Pu}^{+6}$.
"Measurement of local retention as well as excretion of plutonium following contamination under, in, or on the skin of man gave widely varying results. One could conclude or calculate that $\mathrm{W}_{\mathrm{s}}(\mathrm{t})=\mathrm{t}^{-\mathrm{b}}$, where b varies from 0 to 0,4 . These variations are due to the chemical and physical nature of the plutonium and its site of deposition. The extreme variability makes it impossible to give a single formulation."

Thus, the retention function, $\mathbf{L}(\mathbf{t}$, for plutonium deposited at the wound site can be described by
$L(t)=\frac{W(t)}{W(t)}=t^{-b}$
where $t$ is in days and $b$ varies from 0 to 0.4 .
If $b$ is zero, none of the activity leaves the wound site. Under such cases, excision of the localized contamination is often performed (Reference $\mathrm{A}-6$ ). If $\mathrm{b}=0.4$, about $10 \%$ of the activity remains at the wound site after one year. The remaining $90 \%$ will have been translocated to blood. Plutonium is likely

Table A-1. DCF $_{p}$ values for evaluation of puncture wound scenarios ${ }^{\text {a }}$

| Nuclide | $\mathrm{DCF}_{\mathrm{p}}(\mathrm{Sv} / \mathrm{Bq})^{\mathrm{b}}$ | Nuclide | $\mathrm{DCF}_{\mathrm{p}}(\mathrm{Sv} / \mathrm{Bq})^{\mathrm{b}}$ |
| :---: | :---: | :---: | :---: |
| ${ }^{3} \mathrm{H}$ | $1.7 \times 10^{-11}$ | ${ }^{137} \mathrm{Cs}$ | $1.4 \times 10^{-8}$ |
| ${ }^{14} \mathrm{C}$ | $5.7 \times 10^{-10}$ | ${ }^{144} \mathrm{Ce}$ | $1.8 \times 10^{-5}$ |
| ${ }^{35} \mathrm{~S}$ | $1.8 \times 10^{-9}$ | ${ }^{152} \mathrm{Eu}$ | $1.6 \times 10^{-6}$ |
| ${ }_{51}^{51} \mathrm{Cr}$ | $3.6 \times 10^{-10}$ | ${ }_{154}^{154} \mathrm{Eu}$ | $2.4 \times 10^{-6}$ |
| ${ }^{54} \mathrm{Mn}$ | $7.3 \times 10^{-9}$ | ${ }^{155} \mathrm{Eu}$ | $3.7 \times 10^{-7}$ |
| ${ }^{55} \mathrm{Fe}$ | $1.6 \times 10^{-9}$ | ${ }^{226} \mathrm{Ra}$ | $1.8 \times 10^{-6}$ |
| ${ }^{58} \mathrm{Co}$ | $1.6 \times 10^{-8}$ | ${ }^{230} \mathrm{Th}$ | $7.5 \times 10^{-4}$ |
| ${ }^{60} \mathrm{Co}$ | $5.4 \times 10^{-8}$ | ${ }^{232} \mathrm{Th}$ | $3.7 \times 10^{-3}$ |
| ${ }^{59} \mathrm{Ni}$ | $1.1 \times 10^{-9}$ | 235 U | $3.4 \times 10^{-6}$ |
| ${ }^{63} \mathrm{Ni}$ | $3.0 \times 10^{-9}$ | ${ }^{238} \mathrm{U}$ | $3.1 \times 10^{-6}$ |
|  |  |  |  |
| ${ }^{90} \mathrm{Sr}$ | $3.2 \times 10^{-7}$ | ${ }^{238} \mathrm{Pu}$ | $1.5 \times 10^{-3}$ |
| ${ }^{94} \mathrm{Nb}$ | $1.4 \times 10^{-7}$ | ${ }^{239} \mathrm{Pu}$ | $1.6 \times 10^{-3}$ |
| ${ }^{95} \mathrm{Zr}$ | $4.6 \times 10^{-7}$ | ${ }^{240} \mathrm{Pu}$ | $1.6 \times 10^{-3}$ |
| ${ }^{99} \mathrm{Tc}$ | $4.3 \times 10^{-10}$ | ${ }_{241} \mathrm{Pu}$ | $2.5 \times 10^{-5}$ |
| ${ }^{106} \mathrm{Ru}$ | $1.2 \times 10^{-7}$ | ${ }^{242} \mathrm{Pu}$ | $1.5 \times 10^{-3}$ |
| 124 Sb | $2.5 \times 10^{-7}$ | ${ }^{241}$ Am | $1.2 \times 10^{-3}$ |
| ${ }^{125} \mathrm{Sb}$ | $7.0 \times 10^{-8}$ | ${ }^{242} \mathrm{Am}$ | $1.1 \times 10^{-3}$ |
| 1251 | $1.0 \times 10^{-8}$ | ${ }^{243} \mathrm{Am}$ | $1.2 \times 10^{-3}$ |
| 1291 | $7.4 \times 10^{-8}$ | ${ }^{242} \mathrm{Cm}$ | $3.6 \times 10^{-5}$ |
| ${ }^{134} \mathrm{Cs}$ | $2.0 \times 10^{-8}$ | ${ }^{243} \mathrm{Cm}$ | $7.8 \times 10^{-4}$ |
| ${ }^{135} \mathrm{Cs}$ | $1.9 \times 10^{-9}$ | ${ }^{244} \mathrm{Cm}$ | $6.0 \times 10^{-4}$ |

a. When ICRP gives more than one solubility class, the most conservative $D C F / f_{1}$ value is shown.
b. $\quad 1 \mathrm{~Sv} / \mathrm{Bq}=3.7 \times 10^{12} \mathrm{rem} / \mathrm{Cl}$.
to be one of the least mobile of elements in terms of translocation from the wound site to the blocd Its low mobility in intra- and intercellular fluids is evident by its very low uptake potential from the gut to blood (Reference A-2).

Because of these factors, and because of the uncertainties in formulating, a priori, the outcome of a given accident, it is proposed to assume that all activity deposited at the wound site enters the bloodstream within a year.

It is necessary to estimate the quantity of radioctive material which could be deposited at the wound site. Johnson and Lawrence (Reference A-6) report initial contamination of wounds with from 0.4 to 45 nCi ( 0.004 to $0.7 \mu \mathrm{~g}$ plutonium-239). Hammond and Putzier (Reference A-3) report cases
in which up to $2.5 \mu \mathrm{~g}$ plutonium have been deposited in wounds. Excision of the contamination often follows.

It is proposed to assume that $1 \mu \mathrm{~g}$ of material could be deposited at the wound site and subsequently be transfocated to the bloodstream. This parameter may be varied as part of the sensitivity analysis.

The dose from the puncture wound can thus be described by

$$
D=\sum_{i=1}^{n} \frac{C_{i} * W(1) * D C F}{e} \times 10^{9}
$$

where
$\mathrm{D}=$ dose resulting from puncture wound (rem)
$\mathrm{n}=$ total number of nuclides in the waste.
$\mathrm{C}_{\mathrm{i}}=$ concentration of nuclide i in waste container ( $\mathrm{Cl} / \mathrm{m}^{3}$ )
$W(1)=$ quantity of material deposited at the wound site ( $\mu \mathrm{g}$ )
$\mathrm{DCF}_{\mathrm{pi}}=$ blood-to-dose conversion factor for nuclide i (Sv/Bq)
$\mathrm{Q} \quad=$ density of waste $\left(\mathrm{kg} / \mathrm{m}^{3}\right)$
DCF $_{\mathrm{pi}}$ values are given in Table A-1. $\mathrm{C}_{\mathrm{i}}$ is defined in the BURYIT waste streams. A waste density for the different waste sireanis must be defined.

## References

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A-6. L. J. Johnson and J. N. P. Lawrence, "Plutonium Contaminated Wound Experience and Assay Techniques at the Los Alamos Scientific Laboratory," Health Physics, 27, 1974, pp. 55-59.

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## APPENDIX B

## CONFOUNDING PATTERN ALGORITHM FOR FRACTIONAL FACTORIAL DESIGNS

## APPENDIX B <br> CONFOUNDING PATTERN ALGORITHM FOR FRACTIONAL FACTORIAL DESIGNS

Figure 3-1 contains that part of subroutine DESIGN that computes confounding patterns for fractional factorial experimental
designs. The pattern aids one in identifying real alias terms in a response surface equation, as illustrated in the text.

S3O IFILTYPE,NE, 3. A A ) , LTYPE, NE, 4) GO TO 700 COTIONAL INPUT: READ IN PUADRATIE FACTORS INSTEAD TF USING

602 द.j弓ᄅी. RAD $=$ INRUN*** $+2 * N R U N * N F A C+$ NRUN $+* 0.5$ ALP:( (NRUN-RAD) $1-2.1) * * 3$ NF $B C 2=2 * N F A C$ 00610 I=1,NFAC?
610
OJ 620 I, J, NFAC IF(IREAD, EO, I) ALP:STAR(J)

620 NRUN =NRUN+2 *NFAC+1 630 DO ${ }^{630}$ RUN, Ji~N.
C
700 IF $\{\mathrm{LPR}, \mathrm{EQ}, 11 \mathrm{GO}=0.775$


735 KRITE $(5,706)$ NFA:bYरJN
 TOB URITE $(6,709)(\mathrm{I}, \mathrm{I}=1$, NFAC
7 L 9 FORNATI

```
        NFACP2=NFAC4?
```

711 FORMAFIZ

IFILYPE,GE; ${ }^{31}$ ALP
4WRITE 6,720$)^{\text {ALP }}$

C URITE DESIGN GENERATORS

Figure B-1. Subroutine Design


Figure B-1. (continued)


Figure B-1. (continued)


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| :---: |
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-4inninnminn

SUBROUTINE DESIGN


## APPENDIX C <br> SCENARIO CONTROL FILE TAPE7

## APPENDIX C

## SCENARIO CONTROL FILE TAPE7

This appendix contains a printout or. Table C-1 of the scenario control file as it appears on TAPE 7 The scenarios are in the same order as in Table I-1, Appendix I. The scenario list is identical to that in NUREG/CR-1963 ${ }^{\text {C-1 }}$ eve. 1 though i is probable that some are no longer applicable; for example, drums will be required to contain $i=$ more than a small amount of liquid.

Each scenario corresponds to three lines. The first line contains the scenatio code in columns 1-4, taken from Reference C-1. Solumn 8 contains the number
of release pathways in .ee scenario. Columns 14-25 contain a pair of numbers. The first of the pair is the calling sequence for the release pathway; e.g., 951 means UNSAT, EROSIO, AQUIFR, and ATMOS. (Recall that ATMCS is always calied by EROSIO.) The second numt r of thr yair is the fractional release of nuclides to the pathway. A.aditional pairs of numbers for calling ser, uence and release fraction appear in successive groups of 12 columns until the total number of pathvid given in column one is reached. The second and third lines for the scenario contain a brief $\mathrm{v}^{-}$rhal description of the scenario.

## Reference

C-1. D. Lester et al., System Analysis of Shallow-Land Burial, NUREG/CR-1963, Vol. 1-3, March 1981.

Table C-1. Scenario control file


Table C-1. (continued)


Table C-1. (continued)
 ER $-1,1 O N^{1}$ UR WASHINGG ${ }^{1}$ OUFTOI OF BACKFILL. INADEQUATE BACKFILL DEPTH.


INTRUSION BY SCAVENGEKS OIGGING FOR ARTIFACTS. REMJVAL OF CONTAMINATED 1TEAS.

F-5 OF ${ }^{1}$ THE BUǨIAL ${ }^{3}$ SiTE AE AS A PASTURE FOR DJMESTIC ANIMALS.
F-O $I N L S I O N$ BY ANIMALS. OE-O4MALS GECOME COVTAMINATED.

BURIAL SITE.
EUNĊJVERING OF ${ }^{3}$ THL B BURIED WASTE BY EARTHZUAKE.
PIGHLY ACTIXATED ${ }^{\frac{1}{2}}{ }^{1}$ R OE COMO ONENTS ARE MISHANDLED OURING PACKAGING INTJ
A STLILDED CASK, CAUSING WOKKEKS E EXPOSURE.
CHRONLC ${ }^{2}$ DIRECT RADI ${ }^{3}$ OTAON TO WORKERS ENGAGED IN PAGKAGING HIGHLY
ACTVAFED LWK COMPONEATS INTO SHIELDEDGASKS.
P-3, ID 1
${ }_{5-4}{ }^{5}$ PLED.
 P-5 2

FIR ERUPTS DURING ${ }^{2}$ PACKAGING OF ${ }^{2}$ COEMBUSTIBLE WASTES. FIRE IS ALLOWEO
${ }_{P \rightarrow 7}^{T O}$ TRN OUT. 3 1.0E-012 2 2.0E-02.951 1.OE-01
FIKE ERUPTS UURING PACKAGING OF COMBUSTIBLE WASTES. FIRE IS QUENEHED WITH WATER. $3 \quad 1.0 \mathrm{E}+002$ 1.0E-03 951 , $1.0 \mathrm{E}-01$

P-9 ${ }^{3}$. ${ }^{3} \quad 1.0 E+00{ }^{2}$ 1.0E-03 $951 \quad 1.0 E-01$
EXPLOSION DURING PACKAGING OF SC:ID WASTES.

STACK UURING PACKGGINGIPROCESSING OF WASTES.

CHRJNIG UISCHAKGE TO ATMOSPHERE OF RADIONUCLIDES DURING INCINERATIJN JF
DIS:HARGE OF RADIONUCLIDEES THKOUGH OFF-GAS STACI. WITH FAILED FILTERS DURING PACKAG ING/PROCESSING OF WASTES.


THE PACNAGE C JNTAINING WASTES INADEQUATELY DECONTAMINATED PRIOD TO
$K=15$ ASE TO SHI 3 PMENT: $0 E+00 \quad 4 \quad 1.0 E \sim 02$
IREADAATEDICONTAKINITED USABLE ITEMS ARE REMOVED FRIAM WASTES DURING
PACKAGING OR PROCESSING.
P-17
MRORESSING。

Table C-1. (continued)


Table C-1. (continued)


```
SOXES SR CARTONJ FILLED WITH SOLIDS OR LO THE BVERPACK
A TRANSPQRT VEHICLE IS ABAND ONED OR DESTRDYED.OEGRING TRANSIT. LIQUID
SUBSTANCE IS SPILLED FROM DAMAGED CONTAINERS JNTUING TRANSITA ROAUWAY,
```



```
A TKANSPORT VEHICLE IS DAMAGED OR DESTROYED DURING TRANSIT. VOLATILE
SUBSTANCE IS SPILLED FKUM DAMAGED CONTALNERS ONTO THE ROADEAY.
-14 Mo 951 -OE-01
ATRANSPORT, YEHACLE IS OAMAGED OR DESTROYED DURING TRANSIT. SOLID TR
YIOJ IU HASTES SPTLLED ON THE ROADWAY ARE FLODDED BY RAINFALL.
A TRANSPORT VEHICLE IS DAMAGEO DR DESTROYED IN TRANSIT. SOLID OR LIQUIO
WASTE SPILLED ON ROAD, DISPERSED BY HIGH VELDCITY WINDS.
IRKADIATEDICJNTAMINATED ITENS ARE REMOVED FROM WASTE SCATTERED AS A
RESJLT OF TRANSPORT VEHICLE DAMAGE OR DESTRUCTION.
T-17PRER IS INNURED OEY COO COAMINATEE \$ SHARP JBJECT PROTRUDING FROM
RUPTUKED CONTALNER UURING POST-ACEIDENT CLEANUY OF THE ROADWAY.
```


## APPENDIX D <br> NUCLIDE (WASTE) INVENTORY FILE TAPE12

## APPENDIXD

## NUCLIDE (WASTE) INVENTORY F'L: TAPE12

This appendix contains a listing in Table D-1 of the nuclide (waste) inventory file on TAPE12. The inventories designated as WS-1 through WS-6 are the same as those published in NUREG/CR-1963, Volume $1^{\text {D-1 }}$.

The number appearing immediately on the right side of the inventory designator is the number of
nuclides in the inventory. The number next on the right is the density of the waste, in $\mathrm{kg} / \mathrm{m}^{3}$. This density is used only for scenarios involving puncture wounds. On the right side of each nuclide, the specific activity is given, in $\mathrm{Ci} / \mathrm{m}^{3}$.

## Reference

D-1. D. Lester et al., System Analysis of Shallow-Land Burial, NUREG/CR-1963, Vol. 1-3, March 1981.

Table D-1. Listing of the nuclide (waste) inventory file.


APPENDIX E
PROBABILITY ASSIGNMENTS CONTAINED IN TAPE97

## APPENDIX E

## PROBABILITY ASSIGNMENTS CONTAINED IN TAPE97

Tentatively, probabilities or frequencies for the events leading to radioactive releases have been assigned. These are tabulated in Table E-1 where the material is categorized according to the type of event appearing in the first column. For each event type, the information source, upon which the probability or frequency is based or deduced, is given. One may observe that in eight cases, the information source is subjective judgment. All probabilities and frequencies are, of course, subject to revision.

Also, there is tabulated an error factor (EF) for each probability or frequency that is used to set an upper or lower bound according to whether the EF is used as a multiplier or divisor. Note that the use of the EF implies that the logarithm of probability is a random variable. Where data were available, the bounds thus computed are estimates of $90 \%$ (double-sided) confidence bounds (assuming the Gaussian distribution of the logarithm of probability). In most cases, however, the EF is a subjective judgment of what would be reasonable and conservative for $90 \%$ confidence bounds. (In this context, conservative means that the computed upper bound radioactive release is larger than that resulting when data are available. For example, the vehicle accident EF of 12 yields an upper bound of about 24 accidents per year for a single reference site as opposed to an expected number of about 2.)

The rationale of some of the assignments is now discussed in more depth. For example, chronic events always exist during the operational phase and, therefore, are assigned unit probability. Typical of such events is direct radiation.

During the postburial administrative period and/or the postburial unrestricted use period, some events are considered inevitable; thus, each is assigned unit probability. Such events include surface water intrusion, scavenger intrusion, farming intrusion, and pasturage intrusion. The former of
these always occurs to some extent while no effort is made to prevent the latter three during the unrestricted-use period. Subsidence is considered inevitabie during the administrative control period for unsolidified waste (a conservative assumption). However, the probability for washout during a 500 -year unrestricted-use period is assigned 0.11 in order to account for a crash of a sinall aircraft that causes damage to one or more trench covers.

With respect to events with subjective evaluations, those considered highly unlikely because of licensing requirements were assigned frequencies on the order of $10^{-6}$. Those considered unlikely were assigned frequencies on the order of $10^{-4}$. Washout of solidified waste during the administrative control period was assigned the same frequency as an earthquake opening a trench, since the latter event can initiate the former one. The removal of a contaminated item from the facility was assigned a frequency of $1.0 /$ year based on the assumption that one person could be sufficiently tempted each year despite warnings and work rules. The error factor for this theft event allows for the participation of up to 10 persons. For the joint frequency of a theft and a truck accident, a reduction in theft frequency to 0.1 is made because usable items will not be uncovered in most accidents.

The frequency of a puncture wound is taken as that for nonfatal lost time accidents in municipal refuse collection and disposal as reported to the National Safety Council for the years 1970 and 1971 (F.eferences E-2 and E-3). The frequency of fire on board a truck during the arrival phase is based on the occurrence of one similar event during the operations at 22 Atomic Energy Commission jurisdictions during 25 years (Reference E-4). The remaining frequencies with cited references are either derived in a similar fashion or are taken directly from the references. In the latter case, the frequencies are judged to be reasonable. All the above information is entered in TAPE97.

Table E-1. Probability estimates for scenarios

| Type of Event | Activity Phase | Scenario Description | Probability or Frequency Estimate | Error <br> Factor | Reference Used to Derive |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Chronic | Arrival | A-10, -11 | I (for each phase) | - | By defini ion |
|  | On-Site | B-2, -11, -12 |  |  |  |
|  | Trench | C.7, 8 |  |  |  |
|  | Backfill | D-2, -11 |  |  |  |
|  | Packaging | P.2, -10, -11, -12 |  |  |  |
|  | Interim | S-10, -11 |  |  |  |
|  | Transportation | T-1, -2 |  |  |  |
| Surface Water Intrusion | Maintenance | E-2 | 1.0 (for phase) | - | E-11, Inevitable |
|  | Unrestricted Use | F-2 | 1.0 (for phase) | - | E-11, Inevitable |
| Scavenger Intrusion | Maintenance | E. 3 | 1E-4/y (unlikely) | 30 | E-11, Subjective |
|  | Unrestricted Use | F-3 | 1.0 (for phase) | - | E-11, Inevitable |
| Farming Intrusion | Unrestricted Use | F-4 | 1.0 (for phase) | - | E-11, Inevitable |
| Pasturage Intrusion | Unrestricted Use | F-5 | 1* (for phase) | - | E-11, thevitable |
| Washout or Subsidence |  |  |  |  |  |
| Unsolidified | Maintenance | E-1 | 1.0/Phase | 10 | E-10, Subjective |
| Solidified | Maintenance | E-1 | 1E-3/y | 10 | E-10, Subjective |
|  | Unrestricted Use | F-1 | 0.11 (for 500 years) | 1.1 | Attachment I |
| Earthquake Opens Trench | Unrestricted Use | F 8 | 1E.3/y | 10 | E-7, Attachment 3 |
| Animat Intrusion | Open Trench | C. 6 | 1.0/each phase | - | frevitable |
|  | Maintenance | L.4 |  |  |  |
|  | Unrestricted Use Period | F-6 |  |  |  |
| Explosion Aboard Truck | Arrival | A-7, -8 | IE-6/y (highly unlikely) | 100 | Subjective |
| Ruptured Liner | On-Site | B-1 | 3E-6/y (highly | 65 | E-1 |
|  | Backfill | D. 1 | unlikely) |  |  |
| Explosion in Waste | On-Site | B-9, -10 | 1E-6/y (highly | 100 | Subjective |
|  | Backfill | D-9, -10 | unlikely) |  |  |
|  | Packaging | P-8, -9 |  |  |  |
|  | Interim Storage | S-7, -8 |  |  |  |
|  | Transportation | T-10, 11 |  |  |  |
| Long Term Flooding | Unrestricted Use | F-7 | 1E-4/y (ualikely) | 30 | Subjective, Site Dependent. However, see Attachment 2. |
| Removal of ltem from Facility | Arrival | A-9 | 1.0/y | 10 | Subjective fone per |
|  | On-Site | B-14 |  |  | year) |
|  | Trench | C. 5 |  |  |  |
|  | Packaging | P. 16 |  |  |  |
|  | Interim | S-9 |  |  |  |
|  | Transporiation | T. 3 |  |  |  |
| Theft After Truck Accident | Transportation | T-16 | 7E-5/truck-y (O.1) (7E-4) | 120 | E-2, -5, Subjective |
| Puncture Wound | Arrival | A-4, T-7 | 0.2/person-y | 2 | E-2, -3 |
|  | Packaging | P-17 |  |  |  |
|  | Interim Storage | S-4 |  |  |  |
|  | Shipping | T-7 |  |  |  |
| Truck Accident and Puncture Wound (400 Mile Trip) | Transportation | T. 17 | 1.4E-4/truck-y | 20 | E-2, -5 |
|  |  |  |  |  |  |
|  |  |  |  |  |  |

Table E-1. (continued)

| Fire Aboard Truck | Arrival | A-5, -6 | 0.002/y | 10 | E-4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Ruptured Drum | Arrival On-Site Backfill <br> Packaging Interim Storage Transportation | $\begin{gathered} \text { A-1,-2,-3 } \\ \text { B-3, -4,-5,-6 } \\ \text { D-3, -4,-5,-6 } \\ \text { P-3, }-4,-5 \\ \text { S-1, }-2,-3 \\ \text { T-4, }-5,-6 \end{gathered}$ | 0.001/drum-y | 5 | E. 1 |
| Fire in Waste | On-Site <br> Trench <br> Backfill <br> Packaging Interim | $\begin{aligned} & \text { B-7, }-8 \\ & \text { C-1, }-2 \\ & \text { D-7, }-8 \\ & \text { P-6, }-7 \\ & \text { S-5, -6 } \end{aligned}$ | 0.4/y | 2 | E-9 |
| Trench Flooded | Open Trench | C. 3 | $0.1 / \mathrm{y}$ (site dependent) | 4 | E-9 |
| Dispersal by High Wind | Open Trench | C-4 | 1E-3/y | 10 | Attachment 4 |
| Vehicle Accident | Transportation | T-12, -13, -14 | 7E-4/truck-y (1.7E-6/mile) $x$ ( 400 mile) | 12 | E. 5 |
| Vehicle Acvident and High Wind | Transportation | T-15 | 7E-7/truck-y | 120 | E-5, Attachments 4 and 5. |
| Fire Aboard Truck | Transportation | T-8, -9 | 0.0002 /truck-y <br> ( 400 mile trip) | 10 | E-1 |
| Vehicle Contaminated (One Way Trip) | On-Site | B-13 | 0.05/truck-y | 5 | Subjective |
| Container Contaminated | Packaging | P-15 | 2E-3/container-y | 25 | E-6 |
| Cask Mishandled | Packaging | P-1 | 1E-3/cask-y | 10 | E-6, -1 |
| Off-Gas Release, Processing | Packaging | P. 13 | 6E-5/y | 10 | E. 8 |
| Off-Gas Release, Incineration | Packaging | P. 14 | 6E-5/y | 10 | E-8 |

## ATTACHMENT 1

## Aircraft Jazard

A crast: of an aircraft into a low-level waste (LLW) disposal site is a potential initiator of radionuclide releases during the 500 -year unrestricted-use period. For exampie, a crash of a sufficiently large airplane, such as a commercial airliner, could cause a crater that penetrates into the buried waste below the surface. The crash itself could eject radionuclides into the atmosphere. Furthermore, the LLW would be open to water intrusion if the damage to the trench caps were not repaired. As another example, the crash of a small, general aviation airplane could scar one or more trench caps and initiate water erosion. Eventually, rainwater would enter the disposed LLW. The purpose of this section is to estimate the frequency of such an initiating event so that subsequent evaluation of the radiological hazard may be made.

During the period 1970-77, certificated and supplemental air carriers experienced 342 accidents. ${ }^{\text {E }} 12$ This number excludes foreign carrier accidents in the United States (U.S.) and all military accidents. The average annual number of accidents is 42.75 and the sample standard deviation is 10.4 . Many of these accidents occur in the immediate vicinity of an airport. An LLW site should not be sited near an airport in order to minimize the hazard. We estimate the fraction of accidents in the vicinity of an airport using data on "Notable Aircraft Disasters."E-13 In the period 1953-80, 75 such
events occurred. From the brief narrative given, 27 are apparently close to an airport; that is, about $36 \%$. This factor is used to adjust the average annual number of accidents to 27.4 away from airports, with a corresponding change in standard deviation to 6.7. We now use the total U.S. land area ( $7.7 \mathrm{E}+12 \mathrm{~m}^{2}$ ) to estimate an areal crash frequency of $3.6 \mathrm{E}-12\left(\mathrm{y} / \mathrm{m}^{2}\right)$. A typical LLW site would occupy about $6 \mathrm{E}+05 \mathrm{~m}^{2}$. We use the Poisson probability function now to estimate the probability of at least one crash of a large aj craft onto an LLW site in the 500 -year unrestricted-use period:
$\operatorname{Pr}($ crash of large ship $)=1-\mathrm{e}^{-(3.6 \mathrm{E}-12)(6 \mathrm{E}+05)(500)}$

$$
=1.1 \mathrm{E}-03
$$

with the sample standard deviation being 0.3E-03.
During the period 1970-77, general aviation experienced 35012 accidents (Reference E-12). Using the same reduction factor for air carriers to estimate those accidents away from airports yields 22408. The average annual number of such accidents is 2800 and the sample standard deviation is 128.

Following the same procedure as above yields:
estimated areal crash frequency $-364 \mathrm{E}-12\left(\mathrm{y} / \mathrm{m}^{2}\right)$
$\operatorname{Pr}$ (at least one crash of small ship) $=0.11$ (in 500 -year period). Standard deviation $=0.005$.

## ATTACHMENT 2

## Hurricane Hazard

Most people in the United States (U.S.) are familiar to some extent with storms known as hurricanes. Born at sea, these storms build up into huge heat engines and cover large areas, typically having a diameter of 370 km ( 200 miles). When these storms come inland, their high winds and heavy rainfall begin to dissipate; but not always before flooding and wind damage occur. Such a storm would be apt to produce serious damage to a lowlevel waste disposal site in its unrestricted-use period. The heavy rainfall and possible attendant flooding could wash out the trench caps and open the disposed waste to continuous water percolation and wind erosion. Some concern for the hurricane hazard is called for.

In the 81 -year period 1900 to 1980, 22 major hurricanes have entered the U.S. through the Atlantic
and Gulf coastlines. E-14 (Such storms in the Pacific Ocean are called typhoons and apparently do not enter the contiguous U.S.) We judge that severe flooding could occur as far inland as 185 km ( 100 miles). The Atlantic and Gulf coastlines comprise 5.9 Mm ( 3700 miles). Assuming each hurricane had a diameter of 370 km ( 200 miles) when it crossed the coastline, the entire coastline has been exposed 1.2 times during 81 years. We assume the occurrence of major hurricanes follows the Poisson distribution function so that the Poisson rate is estimated as ( $22 / 81$ year) or $0.27 / \mathrm{y}$. E-15 This yields a probability of one or more major hurricanes in one year of about 0.24 . The expected number of hurricanes during the unrestricted-use period ( 500 years) is about 136 . These would expose the entire coastline on the average 7.3 times ( $136 \times 370 \mathrm{~km} / 5.9 \mathrm{Mm}$ ).

## ATTACHMENT 3

## Earthquake Hazard

An earthquake, with enough magnitude to rupture the ground surface, would expose the buried low-level waste (LLW) to water and wind intrusion during the unrestricted-use period. Radionuclides would then be released to water and air pathways. Such an earthquake would inave an intensity of VII or greater on the Mercalli Intensity Scale. E-16 The acceleration of the ground would be $200 \mathrm{~cm} / \mathrm{s}^{2}$ $(0.2 \mathrm{~g})$ or $\operatorname{larger}^{\mathrm{E}-17}$ and would have to attain Richter magnitude of 5.5 or larger (Reference E-16). Such earthquakes are not uncommon in certain areas of the U.S., including Alaska, and Canada. Sixty-one earthquakes with an intensity of VIII or larger, have occurred in the two nations during the 334 -year period of 1638 to 1971, inclusive (Reference E-16). Since the unrestricteduse period of an LLW disposal site is considered to be 500 years, there must be some concern about preventing earthquake-induced releases.

Perhaps the best way to minimize the effect of an earthquake is selecting the site judiciously. The tectonic plate patterns assist this selection. According to the Reference E-16, page $15,95 \%$ of the total seis nic energy of earthquakes around the world is released in the vicinity of ocean ridges and plate subduction zones (where plates are moving toward the interior of the earth).

Major earthquakes have occurred within the contiguous U.S. at points not apparently connected to a subduction zone. These were probably caused by tectonic plate movement, but the mechanisms are as yet unknown. These areas may be identified through historical records. One active region is along the western slopes of the Rocky Mountains from Arizona into Montana, the Intermountain Seismic Belt. There are relatively seismically quiet (aseismic) regions on either side of this belt; for example, the Snake River Plain, southeastern Utah, and parts of southern Arizona. There are two identifiable active zones underlying Nevada. E-17 Both seem to connect with the active zones underlying California, and one seems to connect with the Intermountain Seismic Belt (Reference E-16, p. 175).

Continuing with historical records, major earthquakes have occuried in Missouri near the southern
tip of Illinois (New Madrid and Charleston, Missouri). A major earthquake occurred at Charleston, South Carolina (Mercalli X). This writer concludes that the Missouri earthquakes may be tectonicaily related to the one at Charleston, South Carolina.

Four major earthquakes have occurred around the St. Lawrence River during the period 1663 to 1941. Two major earthquakes that may be tectonically related occurred in Massachusetts (1638 and 1755). In addition, two major earthquakes in New York and three others in Canada may also be related.

The figure in Reference E-16, p. 175, is a seismic risk map of the U.S. that was prepared for the Applied Technology Council in 1976 and 1977. This map shows four other areas of moderate seismic activity, accelerations up to $100 \mathrm{~cm} / \mathrm{s}^{2}$. The map also shows vast areas of no significant seismic risk which, in turn, translate to little or no radiological hazard due to an earthquake in the area of an LLW burial site. An eyeball estimate of the fractional area of the contiguous U.S. with little seismic risk is $60 \%$. However, all of this land may not be suitable because of other constraints, such as potential volcanic activity.

The exposure of LLW, due to earthquakes, could conceivably result in a significant radiological hazard. The risk assessment of LLW disposal, therefore, requires an estimate of the probability of at least a Mercalli intensity of VII earthquake at a disposal site. Two independent sources of information are used to derive frequency estimates, one based on statistical modeling and the other on reported insurance premiums.

The TERA Corporation has developed a method for estimating the frequency of an earthquake with a specific acceleration or larger. E-18,-17 The method is based on seismic measurements at a specific site. The occurrences of scismic events are modeled with a statistical distribution function, from which the desired frequency estimate is extracted. TERA studied the Idaho National Engineering Laboratory (INEL) on the Snake River Plain and predicted the return period of earthquake with peak acceleration of 0.2 g (Reference E-17). The predicted period was 2000 years with a 10 upper limit of 2800 years and a $1 \sigma$ lower limit of

900 years. These numbers translate to frequencies of $0.5 \times 10^{-3} / \mathrm{y}, 0.36 \times 10^{-3} / \mathrm{y}$, and $1.1 \times 10^{-3 /} \mathrm{y}$, respectively.

TERA also demonstrated that method on two hypothetical cases said to be typical for low and high seismicity regions (Reference E-18). For the low seismicity region, similar to parts of the eastern U.S., the estimated frequency of the 0.2 g earthquake was about $10^{-4} / \mathrm{y}$. For the region of high seismicity, similar to Central American or southern Alaska, the estimated frequency was about $10^{-2 / y}$.

In Reference E-16 (p. 167), Bolt gives some insurance premium information dating from 1973. In all cases, there was a deductible amount before coverage took effect. For purposes of estimating earthquake frequency here, this deductible is assumed to represent damage from an earthquake with Mercalli intensity rating of VI or less. The premium for the insurance was assumed to contain a component for profit and overhead. Let this fraction be denoted by, f. The remainder of the premium, $\mathbf{P}$, was assumed to represent an even proposition with respect to coverage, C , so that a return period, R , could be estimated by
$R=C /(1-f) P$.
This calculation was done with $f$ taken as 0.1 and 0.2 for aseismic regions in New Zealand, United

Kingdom, France, Canada, and Australia. The computed frequencies, $1 / \mathrm{R}$, ranged from $0.4 \times 10^{-3}$ to $1.2 \times 10^{-3} / \mathrm{y}$. Considering the assumptions made, this range seems reasonable. Insurance premium information is also given for two regions of high seismicity: highest zone in California and Newfcundland, Canada. The estimates here ranged from $1.6 \times 10^{-3}$ to $2.7 \times 10^{-3} / \mathrm{y}$. These are at least a factor of 2 greater than the low seismicity frequencies.

Based on the above considerations, it is possible to assign a frequency to the occurrence of a major earthquake for the purposes of probabilistic risk assessment of LL.W burial. For the aseismic regions, the frequency may be taken as $10^{-3}$ (return period $=1000$ years). An error factor, EF , may be taken as 10 for one standard deviation. (The error factor is defined such that the upper bound of the estimate, $\mathbf{E}$, is 10 E , while the lower bound is $\mathrm{E} / 10$.) The resulting bounds more than cover the uncertainty in Reference E-18 and the results using insurance premiums (returns period bounds: 100 years, 10,000 years).

Should the site to be assessed be one of high seismicity, an expected frequency of $10^{-2}$ is recommended, again with an error factor of 10 . (Thus, the return period bounds are 10 years, 1000 years.)

## ATTACHMENT 4

## High Wind Hazard

When one examines the information on wind hazards, ${ }^{-19,-20,-21}$ one quickly becomes aware that relatively few tornadoes occur in the western states but that fast straight winds are frequent. The converse is generally true for those states east of the Rocky Mountains. Given this observation, the generic probability of damage due to high wind velocity, applicable to the contiguous United States (U.S.), should be the sum of probabilities for straight winds and tornadoes. Assuming that
$161 \mathrm{~km} / \mathrm{h}(100 \mathrm{mile} / \mathrm{h})$ is a threshold velocity for wind-induced damage to buildings and trucks, the references, especially E-19, show that the return periods cluster in the order-of-magnitude of 1000. A few return periods are an order-of-magnitude lower, and a few are an order-of-magnitude higher. An error factor of 100 encompasses all 19 sites in the U.S. covered in Reference E-19. These observations lead to a generic high wind frequency of $0.001 / y$ with an error factor of 100 to represent two standard deviations. (The error factor for one standard deviation 1 10.) The lognormal distribution function is assumed.

## ATTACHMENT 5

## Truck Accident Probability

Motor transportation accident rates for 1977 are published in Reference E-3, p. 58. The data for three types of vehicles we consider representative
of those that carry low-level waste (LLW) are reproduced here (Table E-2).

From these numbers, others may be derived as in Table E-3.

## Table E-2. Accident rates for intercity carriers

| Type of Vehicle | Vehicle Miles (thousands) | Accident Rate Per Million Miles | Number of Vehicles |
| :---: | :---: | :---: | :---: |
| Common carrier | 1350643 | 3.07 | 9408 |
| Private carrier | 270321 | 4.34 | 3920 |
| Contract carrier | 33248 | 4.63 | 410 |
| Combined | 1654212 | 3.31 | 13738 |

Table E-3. Truck accident probabilities

Probability of Accident

| Type of Vehicle | Per 1000 Miles | Per Vehicle Day |
| :---: | :---: | :---: |
| Common carrier | 0.0031 | $0.9 \mathrm{E}-03$ |
| Private carrier | 0.0043 | $0.8 \mathrm{E}-03$ |
| Contract carrier | 0.0046 | 3.1E-03 |
| Combined | 0.0033 | 1.1 E-03 |

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## APPENDIX F <br> PARAMETER AND UNCERTAINTY VALUES FOR FACTORS IN TABLE 8

## APPENDIX F

## PARAMETER AND UNCERTAINTY VALUES FOR FACTORS IN TABLE 8

This appendix (Table F-1) provides 29 parameter values and uncertainty standard deviations for factors listed in Table 8. Thirteen of the 29 factors have
arbitrarily been assigned standard deviations equal to $10 \%$ of the nominal values. This was an expediency to enable an early test of the program package.

Table F-1. Uncertainty factor nominal values and statistical properties in the example problem


Table F-1. (continued)


## References

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# APPENDIX G <br> UNCERTAINTIES FOR SOME FACTORS NOT INCLUDED IN TABLE 8 

## APPENDIX G

## UNCERTAINTIES FOR SOME FACTORS NOT INCLUDED IN TAPLE 8

The 1 ocedure to follow in order to add uncertainty factors other than thua in Table 8 to an analysis has been described i the text of this manual. The purpose of this appendix is t. provide
uncertainty information gleaned from the literature for certain other factors. It is hoped this information, presented in Table G-1, will be useful; but no claim is made for coverage of all important factors.

Table G-1. Suggested ur certaindies for factors not in Table 8

| Computer Subroutine | Parameter | Standard Deviation |  | $\begin{aligned} & \text { Distribution } \\ & \text { Type } \\ & \hline \end{aligned}$ | Source |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Percent of Mean | Error Factor |  |  |
| DOSET | Resuspension factor, DUM when $K K=4$. | - | 6 | Lognormal | G-1 |
| PREDOS | Food concentration factor, PLIV | - | 10 | Lognormal | G-2 |
|  | Milk concentration factor, PFM | - | 2 | Lognormal | G-2 |
|  | Meat concentration factor, PFF | - | 2 | Lognormal | G-2 |
|  | Deposition velocity, VDI | 80 | 1.8 | Lognormal | G-3 |
|  | Dose factors, DOSFAC | - | 5 | Lognormal | G-4 |
| BURYIT | Waste curie content, INVET | 30 | - | Normal | G-1 |
| AQUIFER | Inverse equilibrium constant | - | - | - | G-5 |
|  | (Inverse retardation factor) RNWV | - | 4 | Lognormal | G-6 |
|  | Aquifer water velocity, VZ | - | 4 |  |  |
|  | Axial dispersion coefficient, EI | 10 | - | Normal | G-8 |
| UNSAT | Sorption distribution coef- <br> ficient, XD | - | 4 | Lognormal | G-6 |
|  | Soil density, DNSTY |  | - | Normal | G-9 |
|  | Hydraulic conductivity, E | 70 | - | Normal | G-10 |
|  | Water content, W | 9 | - | Normal | G-9 |
| ATMOS | Wind speed, U | 45 | - | Normal | G-11 |
|  | Rainout coefficient, RC | - | 10 | Normal |  |
|  | Plume rise lack-of-fit, BUOY | 18 | - | Normal | G-12 |
| EROSIO | Percentage of eroded soil which remains suspended, SAIR | 100 | - | Normal | G-1 |
| NOT | Porosity | 10 | - | Normal | G-10 |
| USED | Cation exchange capacity | 35 | - | Normal | G-8 |
| DIRECTLY |  |  |  |  |  |

## References

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G-11. F. A. Gifford, "Turbulent Diffusion-Typing Schemes: A Review," Nuclear Safety, 17, 1, January through February 1976, pp. 68-77.

G-12. G. Briggs, "Consequences of Effluent Release," Nuclear Safety, 12, 1, January through February 1971.

## APPENDIX H REFERENCE SITES

## APPENDIX H

## REFERENCE SITES

Five geographic regions were selected as being representative of those in the United States where low-level waste may be disposed. These five regions are designated as northeast, southeast, midwest, northwest, and southwest. Tabies H-1 through H-6 provide most of the information needed to use the predictive program BURYIT for estimating the dose commitments resulting from a low-level shallowland burial site. Much of the information was gleaned from the sources listed in the references for this appendix. However, the reader should recognize that some data were not available, so that subjective judgments were required. Thus, the results from BURYIT using these tabulations should 1 used for relative comparisons.

Table H-1 contains input for the EROSIO subroutine that is pla il on TAPE8. The dimensions of th aite correspo d to those for the disposal of 1 miltion $\mathrm{m}^{3}$ of wavt.. The percentage of soil greater in sise than 0.84 mm and the percentage of soil lofted that remains suspended are rough estimate. The mean annual temperature of the northeast vite is less than $10^{\circ} \mathrm{C}$, but BURYIT cannot har dle such a number. Hence, the MAT at the northeast site was set arbitrarily at $10.1^{\circ} \mathrm{C}$. At all sites, the field angle was set arbitrarily perpendicular to the prevailing wind di ection.

Table H-2 contains wind data for TAPE9. Only the average annual wind speed is given for each site.

The stability class frequency numbers should be considered as preliminary. The release height, SH, and the stack energy release rate, SQ, should be selected by the modeler to suit his/her needs.

Table H-3 contains geology and rainfall data for input on TAPE10. Note that the actual number of distinct soil layers is given rather than the divisions needed for the computer model (variable JK). The user must select JK in accordance with the accuracy desired from the solution of the differential equation. This selection also affects the soil layer containing the waste, variable IN . The entries for moisture contents at layer boundaries pertain to the actual soil layers. The soil density entries pertain to the actual soil layers. All other data in Table H-3 should be clear from Section 5.2.2.3.

Table H-4 contains aquifer data for input on TAPE11. The variable FLOWR pertains to the dilution water into which the aquifer empties. All the other variables should be clear.

Table H-5 contains input data needed for TAPE14. Section 5.2.2.1 defines the variables.

Table H-6 contains the agriculture and population input data for TAPE15. The variables FVA and PRODUC should be regarded as rough estimates. The population values, IPOP, pertain to a single compass segment of 22.5 degrees.

Table H-1. Reference site data needed for BURYIT/ANALYZ-erosion data, TAPE8

|  |  |  | Reference Site |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Input Parameter | Description | Units | Northeast | Southeast | Midwest | Northwest | Southwest |
| IKI | Erodibility control | - | 3 | 3 | 3 | 3 | 3 |
| PAG84 | Percent of soil greater than 0.84 mm | \% | 20 | 40 | 10 | 25 | 25 |
| IK3 | Roughness control | - | 2 | 2 | 2 | 2 | 2 |
| MAT | Mean annual temperature | ${ }^{\circ} \mathrm{C}$ | 10.1 | 18 | 11 | 12 | 19 |
| ANGL | Field angle | Degrees | 157.5 | 112.5 | 90 | 22.5 | 135 |
| ANGWND | Direction from which wad blows | Degrees | 247.5 | 202.5 | 180 | 292.5 | 225 |
| HTBR | Barrier height | m | 0 | 0 | 0 | 0 | 0 |

Table H-1. (continued)

|  |  |  | Reference Site |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Description | Units | Northeast | Southeast | Midwest | Northwest | Southwest |
| FW | Field width | m | 550 | 550 | 550 | 550 | 550 |
| FL | Field length | m | 917 | 917 | 917 | 917 | 917 |
| R | Vegetative cover | $\mathrm{kg} / \mathrm{m}^{-2}$ | 1.6 | 1.6 | 1.6 | 1.6 | 1.6 |
| CK5 | Control | - | 1 | 1 | 1 | 1 | 1 |
| $\mathrm{Cr}_{8}$ | Control | - | 0 | 0 | 0 | 0 | 0 |
| CKII | Control | - | 0 | 0 | 0 | 0 | 0 |
| CK13 | Control | - | 0 | 0 | 0 | 0 | 0 |
| SAIR | Percent soil lofted that remains suspended | \% | 20 | 20 | 20 | 20 | 20 |
| KLSP | Percent knoll slope | \% | 3. | 3. | 3. | 3. | 3. |
| RDGHT | Ridge height | m | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 |
| RDGSP | Ridge spacing | m | 10.7 | 10.7 | 10.7 | 10.7 | 10.7 |

Table H-2. Reference site data needed for BURYIT/ANALYZ-wind data, TAPE9

| Input Parameter | Description | Units | Reference Site |  |  |  | Southwest |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Northeast | Southeast | Alidwest | Northwest |  |
| SH | Release height-modeler's choice | - | 0 | 0 | 0 | 0 | 0 |
| SQ | Stack energy release rate-modeler's choice | W | 0 | 0 | 0 | 0 | 0 |
| NU | Number of entries in wind speed vector ( 10 max) | - | 1 | 1 | 1 | 1 | 1 |
| KS | Option for one wind speed or all | - | 0 | 0 | 0 | 0 | 0 |
| NS | Number of stability categories (7 max) | - | 7 | 7 | 7 | 7 | 7 |
| U | Wind speed vector | $\mathrm{m} / \mathrm{s}$ | 4.6 | 3.4 | 4.7 | 4.5 | 3.8 |
| F | Wind speed-stability class frequency | - |  |  |  |  |  |
|  | (A) |  | 0 | 0.212 | 0.04 | 0.0 | 0 |
|  | (B) |  | 0 | 0.118 | 0.02 | 0.30 |  |
|  | (C) |  | 0.33 | 0.171 | 0.07 | 0.14 | 0.33 |
|  | (D) |  | 0.34 | 0.284 | 0.33 | 0.0 | 0.34 |
|  | (E) |  | 0.33 | 0.172 | 0.54 | 0.32 | 0.33 |
|  | (F) |  | 0 | 0.035 | 0 | 0.24 | 0 |
|  | (G) |  | 0 | 0.008 | 0 | 0.0 | 0 |

Table H-3. Reference site data needed for BURYIT/ANALYZ-geology and rainfall data, TAPE10

| Input Parameter | Description |  | Reference Site |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Units | Northeast | Southeast | Midwest | Northwest | Southwest |
| ND | Actual number of distinct soil layers at site | - | 1 | 3 | 2 | 6 | 2 |
|  | Number of entries in the hydraulic conductivity and pressure head arrays $(\max =52)$ | - | 32 | 26 | 32 | 33 | 10 |
|  | Number of the actual soil layer initially containing the waste | - | 1 | 2 | 2 | 1 | 1 |
| DELW | Water content increment | Volume |  |  |  |  |  |
|  |  | Fraction | 0.0078 | 0.02 | 0.012 | 0.009 | 0.05 |
| RAIN | Rainfall rate during wet period | m/s | $9.4 \mathrm{E}-08$ | $1.3 \mathrm{E}-07$ | 8.9E-08 | 1.6E-08 | 5.1E-08 |
| DRY | Evapotranspiration during dry period | m/s | -0.18E-7 | -0.19E-7 | -0.19E-7 | -0.21E-8 | -0.16E-7 |
| DD | Actual soil layer boundaries, starting with 0 | m |  |  |  |  |  |
|  | DD (1) |  | 0 | 0 | 0 | 0 | 0 |
|  | (2) |  | 12 | 1.2 | 3.06 | 14.94 | 10 |
|  | (3) |  |  | 7.3 | 12 | 37.00 | 250 |
|  | (4) |  |  | 13.7 |  | 41.00 |  |
|  | (5) |  |  |  |  | 43.00 |  |
|  | (6) |  |  |  |  | 47.00 |  |
|  | (7) |  |  |  |  | 60.05 |  |
| P | Pressure head array (ND entries) | m |  |  |  |  |  |
|  | P (1) |  | -1.97 | $-8.16$ | -1.97 | -1.48 | -408 |
|  | (2) |  | -1.97 | -7.65 | -1.97 | -1.48 | -188 |
|  | (3) |  | -1.97 | -7.14 | -1.97 | -1.48 | $-24$ |
|  | (4) |  | -1.97 | -6.63 | -1.97 | -1.48 | -7 |
|  | (5) |  | -1.97 | -6.12 | -1.97 | $-1.48$ | - 3 |
|  | (6) |  | -1.97 | -5.61 | -1.97 | -1.48 | -1.6 |
|  | (7) |  | -1.97 | -5.10 | -1.97 | -1.48 | -0.9 |
|  | (8) |  | -1.97 | -4.59 | -1.97 | $-1.0$ | -0.02 |
|  | (9) |  | -1.97 | -4.08 | -1.97 | -0.84 | 0.0 |
|  | (10) |  | -1.97 | -3.57 | -1.97 | -0.71 | 0.1 |
|  | (11) |  | -1.97 | -3.06 | -1.97 | -0.61 |  |
|  | (12) |  | -1.97 | -2.55 | -1.97 | -0.55 |  |
|  | (13) |  | -1.97 | -2.04 | -1.97 | -0.49 |  |
|  | (14) |  | -1.73 | -1.53 | -1.73 | -0.45 |  |
|  | (15) |  | -1.57 | -1.07 | -1.57 | -0.42 |  |
|  | (16) |  | -1.43 | -0.90 | -1.43 | -0.39 |  |
|  | (17) |  | -1.33 | -0.80 | -1.33 | -0.37 |  |
|  | (18) |  | -1.23 | -0.74 | $-1.23$ | -0.35 |  |
|  | P (19) |  | 4.16 | $-0.63$ | $-1.16$ | -0.33 |  |
|  | (20) |  | -1.09 | -0.63 | -1.09 | -0.31 |  |
|  | (21) |  | $-1.04$ | -0.58 | $-1.04$ | -0.30 |  |
|  | (22) |  | -1.00 | -0.53 | -1.1 | -0.28 |  |
|  | (23) |  | -0.96 | -0.35 | -0.96 | -0.27 |  |
|  | (24) |  | -0.92 | -0.18 | -0.92 | -0.264 |  |
|  | (25) |  | -0.89 | 0.0 | $-0.89$ | -0.258 |  |
|  | (26) |  | -0.87 | 0.01 | -0.87 | -0.246 |  |
|  | (27) |  | -0.85 |  | -0.85 | -0.233 |  |
|  | (28) |  | -0.82 |  | -0.83 | -0.221 |  |
|  | (29) |  | -0.79 |  | -0.79 | -0.209 |  |
|  | (30) |  | -0.73 |  | -0.73 | -0.203 |  |
|  | (31) |  | 0.0 |  | 0.0 | -0.0625 |  |
|  | (32) |  | 0.1 |  | 0.1 | 0.0 |  |
|  | (33) |  |  |  |  | 0.1 |  |

Table H-3. (continued)
$\longrightarrow$

| Input Parameter | Description | Units | Northeast | Southeast | Midwest | Northwest | Southwest |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| E | Hydraulic conductivity array | m/s |  |  |  |  |  |
|  | (ND entries) |  |  |  |  |  |  |
|  | E (1) |  | 8.8E-12 | 4.5E-8 | $2.6 \mathrm{E}-7$ | 2.2E-10 | 3E-14 |
|  | (2) |  | $8.8 \mathrm{E}-12$ | 4. SE-8 | 2.6E-7 | 2. $2 \mathrm{E}-10$ | $2 \mathrm{E}-13$ |
|  | (3) |  | 8.8E-12 | 4. 5E-8 | 2.6 E-7 | 2.2E-10 | 3.3E-11 |
|  | (4) |  | 8.8E-12 | 4.5E-8 | $2.6 \mathrm{E}-7$ | 2.2E-10 | 6.6E-10 |
|  | (5) |  | 8.8E-12 | 4.5E-8 | 2.6E-7 | 2.2E-10 | $5.5 \mathrm{E}-9$ |
|  | (6) |  | 8.8E-12 | 4.5E-8 | 2.6E-7 | 2.2E-10 | $2.9 \mathrm{E}-8$ |
|  | (7) |  | 8.8E-12 | 4.5E-8 | 2.6E-7 | 2.2E-10 | 1.1E-7 |
|  | (8) |  | 8.8E-12 | 4. 5 E-8 | 2.6E 7 | 1.3E-9 | $3.4 \mathrm{E}-7$ |
|  | (9) |  | 8.8E-12 | 4.5E-8 | 2.6E-7 | 3.1E-9 | $9.2 \mathrm{E}-7$ |
|  | (10) |  | 8.8E-12 | $4.5 \mathrm{E}-8$ | $2.6 \mathrm{E}-7$ | 6.6E-9 | 9.2E-7 |
|  | (1) |  | 8.8E-12 | 4.5E-8 | $2.6 \mathrm{E}-7$ | 1.3E-8 |  |
|  | (12) |  | 8.8E-12 | 4.5E-8 | $2.6 \mathrm{E}-7$ | 2.1E-8 |  |
|  | (13) |  | 8.8E-12 | 4.5E-8 | $2.6 \mathrm{E}-7$ | $3.7 \mathrm{E}-8$ |  |
|  | (14) |  | 1.2E-11 | 4.5E-8 | 3.6E-7 | $5.6 \mathrm{E}-8$ |  |
|  | (15) |  | 1.7E-11 | $4.6 \mathrm{E}-8$ | 4.9E-7 | $7.8 \mathrm{E}-8$ |  |
|  | (16) |  | 2.2E-11 | $5.8 \mathrm{E}-8$ | 6.5E-7 | $1.1 \mathrm{E}-7$ |  |
|  | (17) |  | 3.0E-11 | $7.9 \mathrm{E}-8$ | 8.7E-7 | $1.4 \mathrm{E}-7$ |  |
|  | (18) |  | $3.8 \mathrm{E}-11$ | $1.1 \mathrm{E}-7$ | 1.1E-6 | $1.8 \mathrm{E}-7$ |  |
|  | (19) |  | $5.0 \mathrm{E}-11$ | 1.5E-7 | 1.5E-6 | 2.2E-7 |  |
|  | (20) |  | 6.2E-11 | $2.0 \mathrm{E}-7$ | 1.8E-6 | $3.0 \mathrm{E}-7$ |  |
|  | (21) |  | $6.9 \mathrm{E}-11$ | $2.9 \mathrm{E}-7$ | $2.0 \mathrm{E}-6$ | 4.0E-7 |  |
|  | (22) |  | $8.0 \mathrm{E}-11$ | 4.1E.7 | $2.6 \mathrm{E} \cdot 6$ | $4.8 \mathrm{E}-7$ |  |
|  | (23) |  | 9.9E-11 | 4.1E-7 | 3.9E.6 6 | 6.5E-7 |  |
|  | (24) (25) |  | 1.4E-10 | 4. 1E.7 | 4.0E-6 | $7.3 \mathrm{E}-7$ |  |
|  | (26) |  | 1.5E-10 | 4.1E-7 | 4.5E-6 | 9.1E-1 |  |
|  | E (27) |  | 1.8E-10 |  | 5. $2 \mathrm{E}-6$ | 1.2E-6 |  |
|  | (28) |  | $2.0 \mathbf{E}-10$ |  | 5.8E-6 | 1.9E-6 |  |
|  | (29) |  | 2.2E. 10 |  | 6. $4 \mathrm{E}-6$ | 1.9E-6 |  |
|  | (30) |  | $2.4 \mathrm{E}-10$ |  | 7. 1E-6 | 2.4E-6 |  |
|  | (31) |  | $2.6 \mathrm{E}-10$ |  | 7.6E-6 | 2.6E-6 |  |
|  | (32) |  | $2.6 \mathrm{E} \cdot 10$ |  | $7.6 \mathrm{E}-6$ | 2.6E-6 |  |
|  | (39) |  |  |  |  | $2.6 \mathrm{E}-6$ |  |
| W | Initial water content at each layer boundary | Fraction by |  |  |  |  |  |
|  |  |  | . 0936 | 0.300 | 0.144 | 0.054 | 0.089 |
|  | (2) |  | 0.234 | 0.300 | 0.236 | 0.102 | 0.089 |
|  | (3) |  |  | 0.300 | 0.360 | 0.185 | 0.40 |
|  | (4) |  |  | 0.420 |  | 0.200 |  |
|  | (5) |  |  |  |  | 0.212 |  |
|  | (6) |  |  |  |  | $\begin{aligned} & 0.223 \\ & 0.270 \end{aligned}$ |  |
|  |  |  |  |  |  |  |  |
| HDRY | Pressure head which corresponds to | m | -1.97 | -1.27 | . 1.97 | -1.48 | -34.2 |
|  | WATL water content |  |  |  |  |  |  |
| HWET | Pressure head which corresponds to | m | 0 | 0 | 0 | 0 | 0 |
|  | WATH water content |  |  |  |  |  |  |
| WATL | Minimum water content (bound water) | Volume |  |  |  |  |  |
|  |  | Fraction | 0.0936 | 0.270 | 0.144 | 0.054 | 0.089 |
| WATH | Maximum water content at soil saturation | Volume | 0.234 | 0.420 | 0.360 | 0.270 | 0.40 |
|  | Maximum waier coment at soil saluration | Fraction |  |  |  |  |  |

Table H-3. (continued)


Table H-3. (continued)


Table H-4. Reference site data needed for BURYIT/ANALYZ-aquifer data, TAPE11

|  | Description | Northeast | Reference Site |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter |  |  | Southeast | Midwest | Northwest | Southwest |
| XZ | Length of aquifer (m) | 500 | 500 (wells) | 1250 | 2000 | 30 (wells) |
| El | Axial dispersion coefficient ( $\mathrm{m}^{2} / \mathrm{s}$ ) | IE.9 | 1E-9 | 1E-9 | 1E-9 | 1E-9 |
| vz | Velocity of aquifer ( $\mathrm{m} / \mathrm{s}$ ) | 6E-09 | 4E-08 | 2E-08 | 3E-07 | 3E. 6 |
| FLOWR | Flow rate (L/y) | $3.1 \mathrm{E}+09$ | $8.9 \mathrm{E}+09$ | $2.7 \mathrm{E}+09$ | $2.5 \mathrm{E}+09$ | 7700 tone nerson) |

RNWV Inverse equilibrium constant-1/K ${ }^{\text {a }}$

| (1) | ${ }^{3} \mathrm{H}^{*}$ |
| :---: | :---: |
|  | 14 C |
|  |  |
| (4) | ${ }^{51} \mathrm{C}$ |
| (5) |  |
|  |  |
| (7) |  |
| (8) <br> (9) |  |
|  | ${ }^{59}$ |
| (10) | ${ }^{63} \mathrm{~N}$ |
| (11) |  |
| (12) | ${ }^{90} \mathrm{~S}$ |
| (2) |  |
| (14) |  |
| (15) | 99 |
| (16) |  |
|  | ${ }^{124} \mathrm{~S}$ |
| (18) |  |
| (19) | 125 |
| ) | 129 |
|  |  |
| (22) |  |
|  |  |
|  |  |


| 1 | 1 | 1 | 1 | 1 |
| :--- | :--- | :--- | :--- | :--- |
| $1.0 \mathrm{E}-01$ | $1.0 \mathrm{E}-01$ | $1.0 \mathrm{E}-01$ | $2.9 \mathrm{E}-03$ | $1.0 \mathrm{E}-01$ |
| $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-03$ | $2.9 \mathrm{E}-3$ |
| $2.9 \mathrm{E}-03$ | $2.9 \mathrm{E}-2$ | $2.9 \mathrm{E}-2$ | $2.9 \mathrm{E}-02$ | $2.9 \mathrm{E}-2$ |
| $2.9 \mathrm{E}-03$ | $2.9 \mathrm{E}-2$ | $2.9 \mathrm{E}-2$ | $2.9 \mathrm{E}-02$ | $2.9 \mathrm{E}-2$ |
| $2.0 \mathrm{E}-04$ | $4.0 \mathrm{E}-04$ | $4.0 \mathrm{E}-04$ | $2.0 \mathrm{E}-03$ | $7.8 \mathrm{E}-4$ |
| $2.9 \mathrm{E}-04$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-03$ | $1.2 \mathrm{E}-3$ |
| $2.9 \mathrm{E}-04$ | $6.0 \mathrm{E}-04$ | $6.0 \mathrm{E}-04$ | $2.9 \mathrm{E}-03$ | $1.2 \mathrm{E}-3$ |
| $3.0 \mathrm{E}-04$ | $6.0 \mathrm{E}-04$ | $6.0 \mathrm{E}-04$ | $2.9 \mathrm{E}-03$ | $1.2 \mathrm{E}-3$ |
| $3.0 \mathrm{E}-04$ | $6.0 \mathrm{E}-04$ | $6.0 \mathrm{E}-04$ | $2.9 \mathrm{E}-03$ | $1.2 \mathrm{E}-3$ |
| $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-03$ | $2.9 \mathrm{E}-3$ |
| $1.5 \mathrm{E}-02$ | $2.8 \mathrm{E}-02$ | $2.8 \mathrm{E}-02$ | $1.2 \mathrm{E}-01$ | $5.6 \mathrm{E}-2$ |
| $1.0 \mathrm{E}-04$ | $2.0 \mathrm{E}-04$ | $2.0 \mathrm{E}-04$ | $2.9 \mathrm{E}-03$ | $4.7 \mathrm{E}-4$ |
| $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-03$ | $2.9 \mathrm{E}-3$ |
| $2.0 \mathrm{E}-01$ | $2.5 \mathrm{E}-01$ | $2.5 \mathrm{E}-01$ | $7.7 \mathrm{E}-01$ | $3.3 \mathrm{E}-1$ |
| $1 . \mathrm{E}-3$ | $1 . \mathrm{E}-3$ | $1 . \mathrm{E}-3$ | $1.0 \mathrm{E}-03$ | $1 . \mathrm{E}-3$ |
| $1 . \mathrm{E}-1$ | $1 . \mathrm{E}-1$ | $1 . \mathrm{E}-1$ | $1.0 \mathrm{E}-01$ | $1 . \mathrm{E}-1$ |
| $1 . \mathrm{E}-1$ | $1 . \mathrm{E}-1$ | $1 . \mathrm{E}-1$ | $1.0 \mathrm{E}-01$ | $1 . \mathrm{E}-1$ |
| $2 . \mathrm{E}-01$ | $7 . \mathrm{E}-01$ | $7 . \mathrm{E}-01$ | $7 . \mathrm{E}-01$ | $2 . \mathrm{E}-1$ |
| $2 . \mathrm{E}-01$ | $7 . \mathrm{E}-01$ | $7 . \mathrm{E}-01$ | $7 . \mathrm{E}-01$ | $2 . \mathrm{E}-1$ |
| $1.4 \mathrm{E}-03$ | $1.4 \mathrm{E}-2$ | $2.9 \mathrm{E}-03$ | $1.4 \mathrm{E}-02$ | $5.8 \mathrm{E}-3$ |
| $1.4 \mathrm{E}-03$ | $1.4 \mathrm{E}-2$ | $2.9 \mathrm{E}-03$ | $1.4 \mathrm{E}-02$ | $5.8 \mathrm{E}-3$ |
| $1.4 \mathrm{E}-03$ | $1.4 \mathrm{E}-2$ | $2.9 \mathrm{E}-03$ | $1.4 \mathrm{E}-02$ | $5.8 \mathrm{E}-3$ |
| $2.0 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-3$ | $2.9 \mathrm{E}-03$ | $2.9 \mathrm{E}-3$ |

Table H-4. (continued)


Table H-5. Reference site data needed for BURYIT/ANALYZ-BURYIT control, TAPE14

| Input <br> Parameter | Description | Reference Site |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northeast | Southeast | Midwest | Northwest | Southwest |
| RF | Annual Rainfall (m/y) | 1.04 | 1.20 | 0.77 | 0.16 | 0.11 |
| EMEF | Emplacement efficiency in trench ( $\mathrm{m}^{3} / \mathrm{m}^{3}$ ) | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| TIMWET | Time for the rain cycle (h) | 3072 | 2664 | 2400 | 720 | 600 |
| TIMCYC | Total time in one cycle of rain and dry periods (h) | 8760 | 8760 | 8760 | 8760 | 8760 |
| VOL | Volume of package ( $\mathrm{m}^{3}$ ) | - a | -a | - ${ }^{\text {a }}$ | -a | _a |
| UD | Wind velocity of hypothetical cloud ( $\mathrm{m} / \mathrm{s}$ ) | _b | -b | _b | -b | -b |
| DI | Diameter of hypothetical cloud (m) | - ${ }^{\text {a }}$ | - ${ }^{\text {a }}$ | - ${ }^{\text {a }}$ | -a | -3 |

a. Modeler's choice.
b. Moderler's choice $>2.24$.

Table H-6. Reference site data needed for BURYIT/ANALYZ-agriculture and population data, TAPE15

| Input Parameter | Description | Reference Site |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Northeast | Southeast | Midwest | Northwest | Southwest |
| NR | Number of radial increments | 6 | 6 | 6 | 6 | 6 |
| RM | Distance to center of radial increments (m) (1) | 4020 | 4020 | 4020 | 4020 | 4020 |
|  | (2) | 12070 | 12070 | 12070 | 12070 | 12070 |
|  | (3) | 24140 | 24140 | 24140 | 24140 | 24140 |
|  | (4) | 40.230 | 40230 | 40.230 | 40230 | 40230 |
|  | (5) | 56330 | 56.310 | 56330 | 56330 | 56330 |
|  | (6) | 72420 | 72420 | 72420 | 12420 | 72420 |
| BEEF | Number of beef cattle per unit area ( km$)^{-2}$ | 1 | 4 | 9 | 3 | 1 |
| cows | Numser of milk cows per unit area (km) ${ }^{-2}$ | 9 | 1 | 2 | 2 | 0.05 |
| FAGE | Age proup fraction breakdown - |  |  |  |  |  |
|  | (1) | 0.2 | 0.24 | 0.2 |  |  |
|  | (2) | 0.12 | 0.16 | 0.12 0.68 | 0.16 0.62 | 0.12 0.68 |
|  | (3) | 0.68 |  | 0.68 |  |  |

Table H-6. (continued)

| Reference Site |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Description | Northeast | Southeast | Midwest | Northwest | Southwest |
| FVA | Fraction of area planted to leafy vegetables | 0.5 | 0.3 | 0.65 | 0.5 | 0.1 |
| NCPY | Number of crops/y | 1 | 1 | 1 | 1 | I |
| PRODUC | Food crop production $\mathrm{kg} /\left(\mathrm{y}-\mathrm{km}^{2}\right)$ | $14 \mathrm{E}+05$ | $9 E+05$ | $13 \mathrm{E}+05$ | $1 E+05$ | $23 \mathrm{E}+05$ |
| IPOP | Population in each radial increment (within 22.5 degrees) |  |  |  |  |  |
|  | (1) | 210 | 130 | 190 | 0 | 2 |
|  | (2) | 1280 | 510 | 310 | 1 | 18 |
|  | (3) | 4600 | 2250 | 1740 | 365 | 60 |
|  | (4) | 7600 | 7810 | 6510 | 4060 | 142 |
|  | (5) | 34790 | 12710 | 7620 | 3980 | 214 |
|  | (6) | 05 500 | 6560 | 22440 | 6980 | 285 |

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## APPENDIX I

## SCENARIO LIST

## APPEND!X I

## SCENARIO LISY

The scenarios available for BURYIT analysis are shown in Table 1-1. Included in this table is a description of the scenario and the files required for
the calculation. Each scenario may be used in conjunction with any one of the six inventories defined in Appendix D.

Table 1-1. Required files for scenario input
Scenario TAPE11 TAPE10 TAPE8
Code Aquifer
Aeology
A-1 Erosion

Table 1-1. (continued)

| Scenario TAPE11 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Code |
| B-3 |
| B-4 |

Table 1-1. (continued)

| Scenario Code | TAPE11 Aquifer | TAPE10 Geology | TAPE8 <br> Erosion | TAPE9 Atmospheric | TAPE1s <br> Agricult: $=$ | TAPE20 <br> Direct | Scenario Description |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| C. 7 | 0 | 0 | 0 | 0 | 0 | 1 | Chronic direct radiation to workers engaged in the activities in the vicinity of uncovered wastes. |
| C-8 | 0 | 0 | 0 | 1 | 1 | 0 | Chronic escape of radionuclides to atmosphere from the uncovered wastes. |
| D-1 | 1 | I | 1 | 1 | 1 | 1 | L. ner containing highly activated L.WR components is accidently ruptured during burial or backfill operation. Wastes are spilled from the liner. |
| D. 2 | 0 | 0 | 0 | 0 | 0 | 1 | Chronic direct radiation to workers engaged in burying the liner, containing highly activated LWR components. |
| D. 3 | 0 | 0 | 0 | 0 | 0 | 1 | Drum with liquid waste containers is ruptured during burial or backfill operation. Liquid is spilled into backfill. |
| D-4 | 1 | 1 | 1 | 1 | 1 | 0 | Drum with liquid waste containers is ruptured during burial or backfill operation. Liquid is spilled into trench. |
| D. 5 | 0 | 0 | 0 | 1 | 1 | 0 | Drum containing volatile substance is ruptured during burial or backfill operation. Volatile substance escapes to atmosphere. |
| D-6 | 0 | 0 | 0 | 1 | 1 | 0 | Drum, carton, or bux containing solid wastes is ruptured during burial and backfill operation. |
| D. 7 | 0 | 0 | 0 | I | 1 | 0 | Fire erupts in the trench containing burnable cartons. hoves or lonse hundles during hurial and hackfill operations. Fire is allowed to burn out. |
| D-8 | 1 | 1 | 1 | 1 | 1 | 0 | Fire erupts in the trench containing burnable cartons, boxes, or loose bundles during burial and backfill operations. Fire is quenched with water. |
| D. 9 | 1 | 1 | 1 | 1 | 1 | 1 | Explosion ia the trench containing drums or boxes with volatile substances or liquid containers during backfill operations. |
| D-10 | 1 | 1 | 1 | 1 | 1 | 1 | Explosion in the trench containing drums, boxes cartons, or loose bundles (in solid state) during burial and backfill operations. |
| D-11 | 0 | 0 | 0 | 0 | 0 | 1 | Chronic direct radiation to workers engaged in burial and backfill operations. |
| E-1 | 1 | 1 | 1 | 1 | 1 | 0 | Erosion or washout of backfill. Inadequate backfill depth. |
| E. 2 | 1 | 1 | 0 | 0 | 1 | 0 | Intrusion of surface water. Water seepage to water table through buried wastes. |
| E-3 | 0 | 0 | 0 | 0 | 1 | 1 | Intrusion by scavengers (site worker/outside person) Removal of contaminated items. |
| E-4 | 0 | 0 | 0 | 0 | 1 | 0 | Intrusion by animals (rats, rabbits, etc.). Animals become contaminated. |
| F-1 | 1 | 1 | 1 | 1 | 1 | 0 | Erosion or washout of backfill. Inadequate backfill depth. |
| F-2 | 1 | 1 | 0 | 0 | 1 | 0 | Intrusion of surface water. Water seepage to water table through buried wastes. |

## Table 1.1. (continued)

| Scenario Code | TAPEII <br> Aquifer | TAPE10 <br> Geology | TAPE8 <br> Erosion | rAPE9 <br> Atmospheric | TAPE 15 Agriculture | TAPE 20 Direct | Sceitaitu Sextiprioum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| F. 3 | 0 | 0 | 0 | 0 | I | 1 | Intrusions by scavengers digging for artifacts. Removal of contaminated items. |
| F-4 | 0 | 0 | 0 | 0 | I | 0 | Farming of the burial site for crops. |
| F. 5 | 0 | 0 | 0 | 0 | 1 | 0 | Use of the burial site as a pasture for domestic animals. |
| F-6 | 0 | 0 | 0 | 0 | 1 | 0 | Intrusion by animals. Animals become contaminated. |
| F-7 | 1 | 1 | 1 | 1 | 1 | 0 | Long-term flooding of the burial site. |
| F-8 | 0 | 0 | 0 | 0 | 0 | 1 | Uncovering of the buried waste by earthquake. |
| P. 1 | 0 | 0 | 0 | 0 | 0 | 1 | Highly activated IWR components are mishandled during packaging into a shielded cask, causing direct radiation exposure to workers. |
| P. 2 | 0 | 0 | 0 | 0 | 0 | 1 | Chronic direct radiation to workers engaged in packaging highly activated LWR components into shielded casks. |
| P. 3 | 0 | 0 | 0 | 0 | 0 | 1 | Liquid waste containers are ruptured during packaging. Liquid is spilled. |
| P. 4 | 0 | 0 | 0 | 1 | 1 | 0 | Container with volatile substance is ruptured during packaging. Volatile substances escape to atmosphere. |
| P. 5 | 1 | 1 | 1 | 1 | I | 0 | Solid wastes are spilled and dispersed during packing. |
| P-6 | 0 | 0 | 0 | 1 | 1 | 1 | Fire erupts during packaging of combustible wastes. Fire is allowed to hiren out |
| P. 7 | 1 | 1 | 1 | 1 | I | 1 | Fire erupts during packaging of combustible wastes. Fire is quenched with water. |
| P.8 | 1 | I | 1 | 1 | 1 | 1 | Explosion during packaging of volatile substances or liquid. |
| P-9 | 1 | 1 | 1 | 1 | 1 | 1 | Explosion during packaging of solid wastes. |
| P-10 | 0 | ) | 0 | 0 | 0 | 1 | Chronic direct radiation to workers engaged in packaging of wastes or processing. |
| P-11 | 0 | 0 | 0 | 1 | I | 0 | Chronic discharge of radionuclides to atmosphere from facility off-gas stack during packaging/processing of wastes. |
| P. 12 | 0 | 0 | 0 | 1 | 1 | 0 | Chronic discharge of radionuclides to atmosphere during incineration of wastes |
| P. 13 | 0 | 0 | 0 | 1 | I | 0 | Discharge of radioneclides through off-gas stack with failed filters during packaging/processing of wastes. |
| P-14 | 0 | 0 | 0 | 1 | 1 | 0 | Discharge of radionuclides through off-gas sywem with failed filters during waste incineration. |
| P.15 | 0 | 0 | 0 | 0 | 1 | 1 | The package containing wastes inadequately decon: taminated prior to release to shipment. |
| P. 16 | 0 | 0 | 0 | 0 | 1 | 1 | Irradiated/contaminated usable items are removed from wastes during packaging of processing. |
| P-17 | 0 | 0 | 0 | 0 | 0 | 1 | Worker is injured by contaminated sharp object during packaging or processing. |

Table 1-1. (continued)

| Scenario Code | TAPEII Aquifer | TAPEIO Geology | TAPE <br> Erosion | TAPE9 <br> Atmospheric | TAPE 15 <br> Agriculture | $\begin{aligned} & \text { TAPE } 20 \\ & \text { Direct } \end{aligned}$ | Scenario Description |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| S-1 | 0 | 0 | 0 | 0 | 1 | 1 | A ruptured container with liquid substance causes spill to contaminate the storage or handling area. |
| \$-2 | 0 | 0 | 0 | 0 | 1 | 1 | A ruptured co atainer with volatile substance causes release to contaminate the handling or storage area. |
| S. 3 | 0 | 0 | 0 | 0 | 1 | 1 | A ruptured drum, catton, or box containing solids causes release to contaminate the handling or storage area. |
| S. 4 | 0 | 0 | 0 | 0 | 0 | 1 | Worker is injured by contaminated sharp object protruding from ruptured drum, carton, or box during interim bandling or storage. |
| S-5 | 0 | 0 | 0 | I | 1 | 1 | Fire erupts in the handling or storage area containing combustible cartons, boxes, or loose bundles. Fire is allowed to burn out. |
| S-6 | 1 | 1 | 1 | 1 | 1 | 1 | Fire erupts in the handling of storage area containing combustible cartons, boses, or loose bundies. Fire is quenched with water. |
| 5.7 | 0 | 0 | 0 | I | 1 | 1 | Explosion in the handling or storage area containing drums or boxes with volatile substances of liquid containers. |
| 8-8 | 0 | 0 | 0 | 1 | 1 | 1 | Explosion in the handling of storage area containing drums, boxes, or cartons filled with solids or loose bundles. |
| 5-9 | 0 | 0 | 0 | $\theta$ | 1 | I | Irradiated/contaminated usable items are removed from wastes. |
| 5-10 | 0 | 0 | 0 | 0 | 0 | 1 | Chronic direct radiation to workers engaged in the handling and storage of drums, boxes, cartons, and loose bundles. |
| S. 11 | 0 | 0 | 0 | 1 | 1 | 0 | Chronic escape of radionueliden to atmosphere during the handling and storage of drums, bokes, cartons, and toose bundies. |
| T. 1 | 0 | 0 | 0 | 0 | 0 | 1 | Chronic direst radiation to workers engaged in the loading of drums, boxes cartons, and loose bundles on transport vehicles. |
| T. 2 | 0 | 6 | 0 | 1 | 1 | 0 | Chronic escape of radionuclides to atmosphere during the inspection of drums, boses, cartons, and loose bundies prior to loadiag on transport vehicte |
| T-3 | 0 | 0 | 0 | 0 | 1 | 1 | Irradiated/contaminated usable items are removed from waste during loading on transport vehicle. |
| T.4 | 0 | 0 | 0 | 0 | I | 1 | A ruptured drum with liqsid substance causes spill to comaminate the vehicle or the averpack interior. |
| T-5 | 0 | 0 | 0 | 0 | 1 | 1 | A ruptured drum with volatile substance causes frlease to contaminate the vehide or the overpack interior. |
| T-6 | 0 | 0 | 0 | 0 | 1 | 1 | A ruptured drum, carton, of bos containing solids eauses reltase to contaminate the vehiche or averpack inferior. |
| 7.9 | 0 | 0 | 0 | 0 | 0 | 1 | Worker is injured by consaminated sharp object pro truding from ruptured drum, carton, of bos doring receiving inspection. |

Table 1.1. (continued)

| Scenario Code | TAPE1I Aquifer | TAPE10 Geology | TAPE 8 Erosion | TAPE9 <br> Atmospheric | TAPE15 <br> Agriculture | TAPE 20 Direct | Scenario Description |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| T-8 | 0 | 0 | 0 | 1 | I | 1 | Fire erupts in the transport vehicle of in the overpack containing combustible cartons, boxes, or loose bundles along transportation reute. Fire is allowed to burn out. |
| T-9 | 1 | 1 | 1 | I | 1 | 1 | Fire erupts in the transport vehicle or in the overpack containing combustible cartons, bokes, or loose bundles along transportation route. Fire is quenched with water. |
| T-10 | 0 | 0 | 0 | 1 | 1 | 1 | Explusion in the transpont vehicle or in the overpack containing drums or boxes with volatile substances of liquid containers. |
| T-11 | 0 | 0 | 0 | 1 | 1 | 1 | Explosion in the transport vehicle or in the overpack containing drums, boxes, or cartons filied with solids or loose bundles. |
| T-12 | 1 | 1 | 1 | 1 | 1 | 1 | A transport vehicle is abandoned or destroyed during transit. Liquid subscance is spilled from damaged containers onto the roadway. |
| T. 13 | 0 | 0 | 0 | 1 | 1 | 6 | A transport vehicle is damaged or destroyed during transit. Volarile substance is spilled from damaged containers ante the roadway. |
| T. 14 | 1 | 1 | 1 | 1 | 1 | 0 | A transport vehicle is damaged or destroyed during f ansih. Solid or liquid wastes spilled on the roadway are flooded by rainfall. |
| T. 45 | 0 | 0 | 0 | 1 | 1 | 1 | A transport vehicle is damaged or destroyed during fransit. Solid waste is spilled on the roadway. High velceify wind causes lifling and dispersal of those radionoclides which are attached to dul, light powders. loose papers, boards, ete. The materials are disper sed over the raadway and neighboring countryside. |
| 1. 16 | 0 | 0 | 0 | 0 | 1 | t | Irradiated/contaminated items are removed from waste scattered as a result of transport vehicle Sanh seg of destruction. |
| T-47 | 0 | 0 | 0 | 0 | 1 | 1 | A worker is injured by contaminated sharp abject protruding from ruptured waste consainer during pontaceldent cleanup of the roadway. |



This report is a user's manual for a partially complefed code for risk assessment of a low-level waste shailow-land burial site, to be used $\operatorname{h}$ the liceniing of burial sites. This code is intended as a tool to be used for consid/ring nuclideyransport mechanisms, including atmospheric, groundwater, erosion, fnd infiltration te an underlying aquifer. It also calculates doses to individuals and the population through direct exposure, inhalation, and ingestion.

The methodology of the risk assessment is base primarily on the response surface method of uncertainty analysis. The parameters of a model for predicting dove commitment due to a release are treated as statispcal variables in order to compute statistical distributions for various dose curnm ment contributions. The likelifiegd of a release is also accounted for by statistically evaluating the arithmetic produs: of the dose commitment distributions with te probability of release occurrence.

An example is given using the atmospheric /ratsport pathway as modeled by a code called BURYIT. The framework for using of her release pathways is described in this manual. Information on parameter uncery inties, reference site characteristics, and probabilities of release events is included



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[^0]:    A "transitional" plume rise model is used.

[^1]:    $\mathrm{V}=$ mean annual wind velocity for a particular geographic location (miles per hour)

[^2]:    a. User-created file, input to ANALY:
    b. User-created file, input to BURYIT.

[^3]:    $=1$, Determine the experimentai design specified by LTYPE, but do only the quadratic runs (sensitivity analysis runs) and write the responses to TAPE45
    $=2$, Calculate the design specified by L fYPE, but use the responses from the sensitivity analysis runs for the quadratic portion of the design. Attach the sensitivity analysis responses as TAPE44, call in the desired responses with NOLD(J), where $J$ is the sensitivity analysis factor sequence number, and $\operatorname{NOLD}(\mathrm{J})$ is listed in the order of the present (complete uncertainty analysis) factors LFAC(J) as given in Table 8. The following example demonstrates the use of NOLD(J):

[^4]:    SIN

    ## FACTDR

[^5]:    ATMOSPHERIC PATH P R A TEST CASE, HRNFORD SITE: CUMULATIVE POPUL. DOSE SCENARIO 136, INVENTORY WS-4.

