

Updated Baseline Human Health Risk Assessment

March 2011

United Nuclear Corporation Church Rock Tailings Site Church Rock, New Mexico



United Nuclear Corporation Gallup, New Mexico

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List of Acronyms and Abbreviations

ACL	alternate concentration limit
ALARA	as low as reasonably achievable
ARARs	applicable or relevant and appropriate requirements
CERCLA	Comprehensive Environmental Response, Compensation, and Liability Act
COPC	chemical of potential concern
COC	chemical of concern
CRUMP	Church Rock Uranium Mining Project
CSF	cancer slope factor
EPA	U.S. Environmental Protection Agency
EPC	exposure point concentration
FS	feasibility study
ft/yr	feet per year
gpm	gallons per minute
GWPS	NRC Source Materials License groundwater protection standards
HEAST	Health Effects Assessment Summary Tables
HHRA	human health risk assessment
HEAST	Health Effects Assessment Summary Tables
HI	hazard index
HQ	hazard quotient
IRIS	Integrated Risk Information System
IC	institutional control
IUR	inhalation unit risk
L/day	Liters per day
L/m ³	Liters per meters cubed
MCL	federal primary maximum contaminant level
m ³ /hr	meters cubed per hour
m ³ /day	meters cubed per day
mg/L	milligrams per liter
mg/kg-d	milligrams per kilogram per day
MNA	monitored natural attenuation
MOU	Memorandum of Understanding
NA	natural attenuation
NCP	National Contingency Plan
NMED	New Mexico Environment Department



List of Acronyms and Abbreviations

NNEPA	Navajo Nation Environmental Protection Agency
NPL	National Priorities List
NRC	U.S. Nuclear Regulatory Commission
pCi/L	picocuries per liter
POC	point of compliance
POE	point of exposure
PLSS	public land survey system
PRG	preliminary remediation goals
RAGS	Risk Assessment Guidelines for Superfund
RAO	remedial action objective
RfD	reference dose
RME	reasonable maximum exposure
ROD	Record of Decision
SF	slope factor
SFS	supplemental feasibility study
SWSFS	site-wide supplemental feasibility study
SMCL	federal secondary maximum contaminant level
TDS	total dissolved solids
TTHMs	total trihalomethanes
TI	technical impracticability
UCL95	upper confidence limit on the mean at the 95% confidence level
UMTRCA	Uranium Mill Tailings Radiation Control Act
μg/L	micrograms per liter



Section 1 Introduction

On behalf of United Nuclear Corporation (UNC), Chester Engineers has prepared this updated baseline human health risk assessment (HHRA) for UNC's Church Rock Mill and Tailing s Site (hereinafter the Site or Church Rock Site) near Gallup, New Mexico. The previous baseline risk assessment, called the Public Health Assessment (PHA), was completed as Chapter 4 of the original Feasibility Study (EPA, 1988b), to assess the potential hazards to public h ealth associated with the release or threat of release of hazardous substances from the Site. Much of the content in this updated HHRA has been de veloped addressing speci fic sections of the Environmental Protection Agency's (EPA's) comment letter of September 2, 2010 (EPA, 2010a) and in accordance with the appr oach presented in a conference call held on November 1, 2010, with EPA, the United States Nuclear Re gulatory Commission (NRC), the New Mexico Environment Department (NMED), and the Navajo Nation Environmental Protection Agency (NNEPA).

1.1 Site Location, Description, and Background

1.1.1 Site Location

The Church Rock Site is loca ted approximately 17 miles northeast of Church Rock, McKinley County, New Mexico (see Figure 1). The Site com prises two Sections (i.e., as described using the Public Land Survey System [PLSS]) owned by UNC: Section 2 of New Mexico Township 16 North, Range 16 West (known hereinafter as Section 2) and Section 36 of New Mexico Township 17 North Range 16 W est (known hereinafter as Section 36). Section 2 contains the former uranium ore processing mill facilities and a byproduct material (tailings) disposal site (hereinafter tailings impoundments or tailings site), which cover about 25 and 100 acres, respectively. Section 36 adjoins the southern border of the Navajo Reservation.

The area surrounding the Site is sparsely populated and the primary land use is grazing for sheep, cattle, and horses. Two underground uranium mines were for merly operated in the vicinity of the Site. UNC operated the form er Northeast Church Rock (NECR) mine, which is located to the northwest and ad jacent to the Site, and Qu ivira (formerly Kerr-McGee) operated a mine to the north of the Site (Figure 2).

1.1.2 Site Operational History

The UNC uranium mill was operated from 1977 to 1982. Uranium ore was processed at the facility using a com bination of crushing, grin ding, and acid-leach solven t extraction methods. The milling operation produced an acidic slurry of ground rock and fluid (tailings) that was pumped into the tailings impoundments, into which an estimated 3.5 million tons of tailings were disposed. Details of the Site operational histor y have been summarized in EPA (2008), N.A. Water Systems (2008d), and Chester Engineers (2011).

From approximately 1969 to 1986, large volumes of groundwater were pumped from the nearby NECR and Quivira mines to dewater the underg round workings. This m ine water was



discharged to the local arroyo (known as Pipe line Arroyo), which runs through the Site. A portion of the mine discharge water infiltrated into the subsurface and significantly saturated the near-surface alluvium and Zone 1 and Zone 3 of the Upper Gallup Sandstone Form ation. As designated in the Record of Decision (ROD; EPA, 1988c), this infiltrated water represents the background groundwater conditions for the Site (i.e., the groundwater present prior to tailings disposal which is known alterna tively as (1) post-m ining, pre-tailings water, (2) anthropogenic background groundwater, or (3) background groundwater).

Acidic tailing liquids from the tailings impoundments seeped into the Southwest Alluvium and the Zone 1 and Zone 3 bedrock units of the Upper Gallup Sandst one Formation. The tailingsimpacted groundwater m ay have relatively low (acidic) pH and elevated concentrations of certain heavy metals, radionuclides, sulfate, and total dissolved solids (TDS).

1.1.3 Site Regulatory and Remediation History

EPA listed the Site on the National Priorities Li st (NPL) of Superfund sites in September 1983 and conducted a Site Remedial Investigation (RI) and Feasibility Study (FS) from 1984 through 1988. The RI report (EPA, 1988a) concluded that tailings seepage had contaminated the background water in the Southwest Alluvium, Zone 1, and Zone 3.

A Public Health Assessment (PHA) was completed as Chapter 4 of the FS (EPA, 1988b), to assess the potential hazards to public health associated with the release or threat of release of hazardous substances from the Site. Following the EPA's original PHA, this risk assessm ent addresses each hydrostratigraph ic unit separately. The ROD (EPA, 1988c) indicates that although there was no exposure at that time to local residents from ingestion of groundwater in domestic and livestock wells within four miles of the site, EPA concluded that adverse health or environmental hazards could result in the future if no action was taken to prevent exposure to groundwater contaminants found at the Site. These conclusions were based on the assum ed ingestion, of non-potable background and im pacted well waters, having constituent concentrations measured during the 1985 RI sampling in Sections 1, 3, 36, and a few locations in Section 2. However, groundwater use beneath Sections 2 and 36 will be permanently precluded by the Ur anium Mill Tailing s Radiation Control Act (UMTRCA, discussed more be low); groundwater use in Section 1 (Z one 1) is precluded by lim ited saturated thickness and nonpotable quality; and groundwater use in Section 3 (Southwest Alluvium) is precluded by nonpotable quality.

In the PHA, EPA indicated their analysis was conservative because dilution, dispersion, and natural attenuation were expected to reduce concentrations, from those assumed, if seepage continued to migrate downgradient from the Site (EPA, 1988b).

Under a 1988 Memorandum of Understanding (MOU) between EPA and the NRC (53 Fed. Reg. 37887 [September 28, 1988]), NRC is designated the elead federal agency responsible for regulating the reclam ation and closure activities completed at the tailings impoundments pursuant to the NRC's Source Materials Lice nse SUA-1475 (License) and the Uranium Mill Tailings Radiation Control Act (UM TRCA) of 1978, 42 U.S.C. §7901 et seq. Under the MOU,



the NRC-regulated reclamation and source con trol actions are subject to EPA monitoring and review to ensure that such actions will allow attainment of the CERCLA requirements outside of the tailings impoundments. NRC assumes responsibility within the licensed area (within Section 2, containing the tailings dis posal area); EPA assu mes responsibility for groundw ater offsite (outside of Section 2).

The remedy selected for the Site by EPA in the 1988 ROD was gr oundwater extraction and evaporation, along with groundwater monitoring. The ROD required groundwater extraction in the Southwest Alluvium, Zone 1, and Zone 3 hydros tratigraphic units to limit further migration of tailings-impacted water. Once the reclam ation and remediation activities are complete, the tailings disposal impoundments and associated property will be transferred to the U.S. Department of Energy (DOE) for long-term control and surveillance monitoring.

Between 1999 and 2001, extraction well operation in Zone 1 was perm anently stopped as declining groundwater levels reduced extraction efficiency to rates beneath the minimum needed for pumping as defined by the S ource Materials License; it was temporarily stopped in Zone 3 because pumping was found to inadvertently accelerate contaminant transport away from the Site towards the northwest; and a natur al attenuation test initiated in the Southwest Alluvium has continued to the present. St arting in 2005, extraction well pumping in Zone 3 was done under a revised pumping regime. UNC continues to revise and im prove upon the Zone 3 rem edial system and has recently installed the first injection well just downgradient of the northern edg e of impact in Zone 3, to serve as a hy draulic barrier, and the injected water will be amended with alkalinity in order to stabilize the tailings-impacted water in situ. Pum ping in Zone 3 continues to be characterized by very small, and diminishing, well yields. This results in lim ited ability to hydraulically control the groundwater flow that is governed by the dipping bedrock slope which controls the elevation heads.

There is a long history of re medy performance monitoring at the S ite. The groundwater performance monitoring plan was originally described by the Corre ctive Action Plan (UNC, 1989a), Remedial Design Report (Canonie Environmental Services Corp. [Canonie], 1989a) and Remedial Action Plan (UNC, 1989b). The program has been modified over time, as described in the annual reports (Canonie, 1989b, 1990, 1991, 1992, 1993 and 1995; Sm ith Technology, 1995 and 1996; Rust, 1997; Earth Tech, 1998, 1999, 2000, 2002a and 2002b; USFilter, 2004a; N.A. Water Systems, 2004, 2005, 2007a, 2008a; Chester E ngineers, 2009, 2010, and 2011), to adjust the monitoring requirements as the corrective action has progressed. The compliance monitoring program is required under Condition No. 30 of the NRC Source Materials License. Figure 2 is a Site map that shows the location of the performance monitoring wells, the decommissioned and temporarily idled extraction wells, the evaporation on ponds, and the reclaim editailings areas. Figure 2 also shows the Remedial Action Target Area for each hydrostratigraphic unit, where the impacts of tailings seep age were originally identified and corrective action was implemented (EPA, 1988c) (although the target areas shown for Zone 1 were refined on the basis of pH and isoconcentration mapping (Canonie, 1989a)).



Three CERCLA Five-Year Reviews have been com pleted to date; the most recent was issued in September 2008 (EPA, 2008). The third Five-Year Review concluded that the Site remedy is, ". . . curren tly considered protective of hum an health and the environment because there is no evidence that there is exposure;" however, ". . . there rem ains the question as to the long-term protectiveness of the Site ground-water operable rem edy" (EPA, 2008, from the related Memorandum of Approval in the front of the report). For this reason, EPA has asked UNC to complete a Site-Wide Supplemental Feasibility Study (SWSFS) to review and develop potential remedial alternatives (including alternatives to restrict exposure to contam inated groundwater). EPA has also required that UNC prepare an update d baseline human health risk assessment (i.e., this HHRA document) to support the SWSFS remedy re-evaluation.

1.2 Scope of Risk Assessment

1.2.1 Site-Specific Risk Assessment Objectives

The objective of a hum an health risk assessment is to evaluate the likelihood of adverse effects occurring in human populations potentially exposed to contaminants released in the environment. As described in the EPA's Septem ber 2, 2010, comment letter (EPA, 2010a), the site-specific objectives for preparing an updated HHRA for the UNC Church Rock Mill and Tailings Site are the following:

- 1. Update the risk estimates for the Site using current risk assessm ent methods and information;
 - Comment 18 indicates that the "his toric assessment may no longer provide adequate assessment of the risk under current Site conditions."
 - Comment 18 indicates that the "risk assessment needs to be updated to address the carcinogenic and non-carcinogenic risk posed by the COCs [Chemicals of Concern], including both radiologic and non-radiologic C OCs. The updated assessment should include relevant Risk Assessment Guidance for Superfund (RAGS) revisions, applicable exposure pathways (e.g., derm al (RAGS E) and inhalation (RAGS F), and current toxicological information for each COC."
- 2. Support the reassessment of remediation levels;
 - Comment 17 states that "Part of the SW SFS is to reassess existing or baseline remediation standards or levels set forth in EPA's 1988 ROD and potential changes to those remediation levels that may be necessary to ensure the protectiveness of the remedy. Protection of hum an health should be discussed in terms of cancer and non-cancer risk associated with exposure to ground water."
- 3. Provide a basis for comparing remedial alternatives;
 - Comment 17 states that "Knowing the ri sk posed by ground water exposure will assist in the evaluation of alternatives with respect to demonstrating the potential



for achieving the RAOs [Re medial Action Objectives] and ground water protection standards established at the Site; protection of human health; long-term and short-term effectiveness; and reduction in toxicity, mobility, and volume."

- 4. Identify Point of Com pliance (POC) and Po int of Exposure (POE) concentrations in accordance with NRC requirements;
 - Comment 18 indicates that the exposur e assessment component of the risk assessment should "determine the maximum permissible levels of COCs at the POC that are protective of human health and the environment at the POE."

The principal focus of the HHRA is the assum ed future exposure to groundwater contam inants residing in all three of the hydr ostratigraphic units. Because the hydrogeologic characteristics, COPCs, and remedial alternatives for each of the units are distinct, the risks of potential future exposure to groundwater in each of the units have been evaluated separately.

1.2.2 Risk Assessment Approach

This updated HHRA re port has been prepared in the format of a ba seline risk assessment in accordance with current EPA guidelines for ris k assessment including relevant revisions to RAGS and statistical procedures (e.g., EPA 1989, 1991a, 2001b, 2004, 20 07, 2009). According to EPA (1989), HHRAs comprise the following four principal steps:

- 1. Data Collection and Evaluation gathering and analyzing the Site data relevant to the human health evaluation and identifying the substances that are the focus of the risk assessment process (i.e., the Chemicals of Potential Concern [COPCs]). For the Church Rock Site, much of this work had been previously completed and approved by EPA; however, additional screening was conducted to select the final list of contam inants for inclusion in the risk analysis.
- 2. Exposure Assessment estimating the magnitude of the actual and/or potential hum an exposures, the frequency and duration of the ese exposures and the pathways by which humans are potentially exposed. The results of this assessment are quantitative pathways specific intakes for exposures to individual substances. For the Church Rock Site, the exposure assessment focused on a potential future exposure to groundwater outside UNC-owned property (Sections 2 and 36), because there is no current exposure and no potential future exposure to groundwater contaminants on UNC-owned property.
- 3. Toxicity Assessment The toxicity assessment component considers (1) the types of adverse health effects associated with chemical exposures, (2) the relationship between the magnitude of exposure and adverse effects, and (3) related uncertainties such as the weight of evidence of a particular chemical's carcinogenicity in humans. For the Church Rock Site, the toxicity assessment was based on existing toxicity inform ation available from EPA and other sources, including in formation regarding carcinogenic and non-carcinogenic effects associated with radionuclide and non-radionuclide COPCs.



4. Risk Characterization - The risk characteri zation step summarizes and combines outputs of the exposure and toxicity assessments to characterize risk in a set of quantitative and qualitative statements. For the Church Rock Site, this includes the consideration of risks associated with background groundwater COPC concentrations and the uncertainties associated with the risk assessment process.

1.3 Report Organization

The rest of this HHRA report is organized in a m anner that generally corresponds to the risk assessment steps described in Section 1.2.2:

Section 2	Identification of Chemicals of Potential Concern
Section 3	Exposure Assessment
Section 4	Toxicity Assessment
Section 5	Risk Characterization
Section 6	Uncertainty Analysis
Section 7	Risk Assessment Summary
Section 8	References

Risk assessment calculations have been prepared following the Risk Assessm ent Guidance for Superfund: Volume I: Hum an Health Evaluation Manual (Part D, Standardized Planning, Reporting, and Review of Superfund Risk Assessments; EPA, 2001b) and are documented using the RAGS Tables 1 to 10 format (i.e., the EPA RAGS Part D risk assessment summary tables) in Appendix A. Additionally, site-specific reference documents related to exposure assessment are provided on CD in Appendices B through D.



Identification of Chemicals of Potential Concern

2.1 Overview of Site-Specific Data Evaluation Considerations

The principal focus of the Churc h Rock Site HHRA is the assum ption of potential future exposure to seepage-impacted groundwater contaminants in each of the three hydrostratigraph ic units (i.e., Southwest Alluvium, Zone 1, and Zone 3) at locations outside Section 2. However, groundwater use in Section 1 (Z one 1) is precluded by lim ited saturated thickness and non-potable quality, and groundwater use in Section 3 (Southwest Alluvium) is precluded by non-potable quality (discussed further below).

Because the hydrogeologic characteristics, contaminants, and remedial alternatives for each of the units are distinct, the risks of potential future exposure to groundwater in each of the units have been evaluated separately and COPCs have been selected for each unit. The COPCs are screened to derive the sets of COCs in Section 7 (Risk Assessment Summary).

2.2 Process to Select Chemicals of Potential Concern

A COPC selection process was used to focus the quantitative risk assessment on chemicals and radionuclides that pose the greatest risks to hum an health. Tabulated risk screening levels were used to evaluate groundwater performance monitoring data to select risk assessment COPCs for each of the hydrostratigraphic units.

The groundwater performance monitoring program was established at the Site on the basis of the Site Remedial Investigation/Feasibility Study (RI/FS) activities and NRC License conditions and has been ongoing, with approved modifications, for 22 years. Modifications have been made to the program over time, including the elimination of certain monitoring parameters that were no longer considered relevant to the remedy implementation (e.g., a set of trace metals plus iron); these historical m onitoring parameters were not considered as potential COPCs for the quantitative risk estimates because (1) the data were not current and (2) EPA and NRC concurred with their deletion from the monitoring program. In contrast, the individual isotopes of uranium are of interest for risk calculations, but histo rically have not been part of the perform ance monitoring program. In this situation, simplifying assumptions were made so that exposure to the uranium isotopes could be evaluated in the HHRA and to re duce the chance that potential human health risk is underestim ated. Ho wever, it should be ac knowledged that these assumptions could result in an overestimation of risk because the actual uraniu m isotope ects of these issues are considered in the discussion of activities are not known. The eff uncertainties presented in Section 6.

For the Church Rock Site HHRA, it was necessary to identif y a subset of the histor ic groundwater performance monitoring data that were representative of seepage-im pacted groundwater under current Site cond itions, had adequate data quality, and for which exposure point concentrations (EPCs) could be statistically estimated. This evaluation was performed by UNC at EPA's request in 2008 (N.A. W ater Systems, 2008c) for each of the hydrostratig raphic



units, and was subsequently approved by EPA. Comparable statistical estimates of background constituent concentrations were also made in a separate analys is (N.A. Water System s, 2008b). The reports of these analyses are provided in Appendices A and B (on CD). Additional screening was subsequently conducted to select the risk assessm ent COPCs for inclusion in the Church Rock Site HHRA; this process is described below.

The following steps were used to select the ri sk assessment COPCs for each of the stratigraphic units.

- 1. Monitoring period selection, m onitoring well selection, and statistical data evaluation. This data reduction step was completed in a 2008 report by UNC (N .A. Water Systems, 2008c).
 - a. An eight-quarter period (July 2006 through April 2008 inclusive) of performance monitoring data was selected as representative of current conditions. This period comprised the most recent eight quarters of sampling available at the tim e of the calculations. The performance monitoring program COPCs included 23 analytes (Table 1 of this report).
 - b. Monitoring data from the eight-quart er period were rev iewed in each hydrostratigraphic unit to identify the wells in which tailings seepage-impacted groundwater was present. Groundwater th at was representative of tailings seepage-impacted groundwater was di scriminated from non-im pacted (i.e., background) groundwater based on water quality including indicator param eter concentrations (e.g., pH or bicarbonate c oncentrations). The wells selected for each hydrostratigraphic unit are shown in Table 2 of this report.
 - c. Statistical analyses were perform ed, using ProUCL (EPA, 2007) to calculate the 95th percentile upper confidence lim it of the mean (UCL95) for the perform ance monitoring COPCs (Table 1) for tailings seepage-im pacted groundwater in each of the three hydrostrat igraphic units (N.A. Water Systems, 2008c). Many of the calculated UCL95 are utilized as EPCs for COPCs in the exposure assessment step of the HHRA. Statis tical analyses were not com pleted for historical monitoring parameters comprising several trace metals and iron (Table 1). These historical monitoring parameters were e liminated from further consideration in the quantitative risk assessment. These parameters are characterized by very low concentrations and toxicities, and their potential effect with respect to ris k calculation is addressed in the uncertainty analysis in Section 6.
- 2. Monitoring COPCs representing common ions (i.e., sulfate, chloride, nitrate-nitrogen), general chemistry parameters (i.e., tota 1 dissolved solids), and groups of multiple parameters (i.e., gross alpha, total radium) were eliminated from further consideration in the quantitative risk as sessment calculations. The potential effect of these parameters with respect to risk calculation is addressed in the uncertainty analysis in Section 6.



- 3. Monitoring COPCs with no detected values during the eight quart er monitoring period were eliminated from further consideration in the quantitative risk assessment. The remaining monitoring COPCs were retained for further evaluation in the quantitative risk assessment and the EPC was estimated or assigned using other means.
- 4. Estimation of uranium isotope con centrations total uran ium mass concentrations are reported for groundwater samples, but there is no site-specific information regarding the abundance of the individual isotopes that ca n be used to evaluate the radioactive carcinogenic risk of uranium at the Site. Therefore, the activities of the three major uranium isotopes (U-234, U-235, and U- 238) in groundwater within each hydrostratigraphic unit were estimated from the total uranium mass concentrations using the assumption that the isotopes were present in proportion to their natural abundance.
- 5. Screening process Consistent with EPA ri sk assessment guidance (e.g., EPA, 2001b) a screening process was employed to further focus the COPC select ion for quantitative evaluation in the risk assessment for each of the three hy drostratigraphic units. The screening process varied according to the contaminant type, as follows:
 - a. Non-radionuclides
 - The non-radiologic, non-carcinoge nic analytes with m aximum concentrations exceeding 0.1 tim es the EPA Risk Screening Level (RSL) for the tapwater ingestion plus inhalation exposure pathways (i.e., a target hazard quotient (HQ) of 0. 1) were retained as COPCs in the quantitative risk assessment. The use of a target HQ of 0.1 as a screening level accounts for the potential exposure to m ultiple constituents.
 - The non-radiologic carcinogenic analytes with maximum concentrations exceeding the EPA RSLs, com piled on the November 2010 EPA Risk Screening Level Table (EPA, 2010b) for the tapwater ingestion and inhalation exposure pathways (set at a target risk level of 1E-06), were retained as COPCs in the quantitative risk assessment.
 - All non-radiologic, non-carcinogenic an alytes that were retained as COPCs for the tapwater ingestion ex posure pathway were also retained for the derm al exposure pathway. This represents a conservative measure, because EPA guidance for the derm al-water exposure pathway requires only those chem icals which contribute a significant dose relative to the oral exposure pathway (EPA, 2004) be retained.
 - Non-radiologic analytes that were r etained for the tapwater inhalation pathway were restricted to volat ile organic compounds (i.e., organic compounds with a Henry's Law Constant greater than 1E-05 atm



 m^3 /mole and a molecular weight of less than 200 g/mole (EPA, 2010b)). The only Site analyte that met these criteria was chloroform.

- Uranium has chemical toxicity as well as radiotoxicity and was retained as a non-carcinogenic COPC in the quant itative risk assessment in those hydrostratigraphic units where the maximum concentrations exceeded 0.1 times the RSL for the tapwater ingestion and derm al exposure pathways (i.e., a target HQ of 0.1).
- b. Radionuclides
 - All detected radiologic analytes were reta ined as COPCs in the quantitative risk assessment. In addition, the uranium isotopes for which activities were estimated from total uranium mass concentrations and were retained in as COPCs.
 - The skin is generally an effect ive barrier against absorption of radionuclides and the dermal absorption exposure pathway is considered very minor with respect to other e xposure routes, such as ingestion. Therefore, radionuclides were not retained as C OPCs under the dermal exposure pathway.
 - The only radiologic analytes that were retained for the tapwater inhalation pathway in the risk assessment were those for which EPA has included the inhalation pathway in the calculation of its radionuclide Preliminary Remediation Goals (PRGs). The only Site analyte that m et this criterion was radiu m-226, for which the d ecay product is the gas radon-222.

A summary of the data evalu ated and the COPCs selected for each hy drostratigraphic unit is provided in Section 2.4.

2.3 Relationship of Background COPC Concentrations to COPC Screening Process

The chemistry of the background groundwater in each of the hydrostratig raphic units is well known and is not considered suitable for use as a drinking water source. Sulfate and TDS, which are non-hazardous constituents, have exceeded New Mexico groundwater qu ality standards in the background water in each of the hydrostratigraphic units (Chester Engineers, 2011). Because the sulfate concentrations are controlled by geochemical equilibrium with gypsum (or anhydrite) and calcite, they are irreducible in the background water. Exceedances of these and other COPCs have been documented in the background water from each hydrostratigraphic unit.

Statistical estimates of upper confidence limits (UCL95s) for the population m eans of the background chemical concentrations have be en calculated (N.A. W ater Systems, 2008b) and approved by EPA. Background concentrations of som e constituents equal or exceed concentrations in seepage-im pacted groundwater. W here these background concentrations



exceed applicable or relevant and appropriate requirements ARARs, background concentrations may be selected as cleanup levels. Additionally, a statistical evaluation of background water quality by NRC led to their recomm endation that manganese, sulfate, and TDS should not be regulated site constituents and they s hould not be used as bases f or corrective action (NRC, 1996). Consistent with EPA guidance regarding b ackground concentrations (EPA, 2002), COPCs in seepage-impacted groundwater that are present in background groundwater have been included in the quantitative risk assessm ent calculations of the seepage e-impacted groundwater. The resulting non-carcinogenic hazard and carcinogenic risk estimates (i.e., estimates which include background risk) may accurately quantify the total hazard and risk of exposure to groundwater, but may overestimate the risk associated w ith seepage-impacted groundwater. Therefore, background concentrations have been considered qualitatively and, in some cases, quantitatively in the risk characterization with respect to risk drivers and should be c onsidered in any future reassessment of Site rem ediation levels. A di scussion regarding uncertainties in the HHRA is presented in Section 6.

2.4 Summary of Selected Chemicals of Potential Concern

The wells included in the COPC selection evaluation for each hydrostratigraphic unit are listed in Table 2 and their locations are shown on Figure 2. For Z one 1 and Zone 3, wells that were within Section 2 were excluded from the calculations because Section 2 encompasses the tailings disposal area, which will eventually be transferred to the U.S. Department of Energy (DOE) for long-term care. This transfer will effective ely eliminate the potential exposure to groundwater within this area. Furthermore, the high levels of seepage impact evident in Zone 1 and Zone 3 wells proximal to the tailings disposal cells are not expected to migrate and occupy areas outside of Section 2. This judgm ent is based on analys is of water-quality monitoring results from the past 22 years, and the conclusion the at the tailings cells are no longer a source of measureable quantities of seepage fluid (USFilter, 2004b).

However, it was necess any to include data from seepage-impacted wells in Section 2 for the Southwest Alluvium dataset due to statistical limitations that were encountered when using only samples from outside Section 2.

The following subsections summ arize the results of the EPC statistical analyses (N.A. Water Systems, 2008c) and the subsequent COPC screening process for each of the hydro stratigraphic units.

2.4.1 Southwest Alluvium COPCs

The HHRA COPCs for the Southwest Alluvium were selected as follows:

1. Seven monitoring parameters were eliminated from further consideration in the H HRA because they were not detected: bery llium, cadmium, lead, m olybdenum, nickel, vanadium, and lead-210. Two parameters (cobalt and selenium) that were each detected only one time during the monitoring period were retained for further screening.



2. The remaining monitoring parameters were screened by comparing the m aximum detected concentration agains t the toxicity screening values defined in Section 2.2. Aluminum and seleniu m were elim inated as risk assess ment COPCs because their maximum concentrations were below the corresponding toxicity screening values. The results of the screening process are shown in Table 2.1 in Appendix A.

2.4.2 Zone 1 COPCs

The HHRA COPCs for Zone 1 were selected as follows:

- Five monitoring parameters were eliminated from further consideration in the HHRA because they were not detected: b eryllium, cadmium, lead, molybdenum, and lead-210. Two parameters (selenium and vanadium) that were each d etected only one time during the monitoring period were retained for further screening.
- 2. The remaining monitoring parameters were screened by comparing the m aximum detected concentration agains t the toxicity screening values defined in Section 2.2. Aluminum, nickel, selenium, and uranium were eliminated as risk ass essment COPCs because their maximum concentrations were below the corresponding toxicity screening values. However, the uranium isotopes (for which ac tivities are estimated from the uranium mass concentration) were retained for evaluation as radiologic carcinogens. The results of the screening process are shown in EPA risk assessment Table 2.1 in Appendix A.

2.4.3 Zone 3 COPCs

The HHRA COPCs for Zone 3 were selected as follows:

- 1. One monitoring parameter, lead, was eliminated from further consideration in the HHRA because it was not detected. Two parameters (selenium and vanadium) that were each detected only one time during the monitoring period were retained for further screening.
- 2. The remaining monitoring parameters were screened by comparing the m aximum detected concentration agains t the toxicity screening values defined in Section 2.2. Selenium was eliminated as risk ass essment COPC because its m aximum concentration was below the corresponding toxicity screening value.



Exposure Assessment

3.1 Characterization of Exposure Setting

The exposure assessment process is used to que antify the type and mean agnitude of the total exposure by potential receptors to COPCs across exposure pathway combinations. This section evaluates and documents the sources, receptors, exposure pathways, and exposure duration and frequency to quantify the human exposure to the Site risk assessment COPCs.

In their Septem ber 2, 2010 comm ents letter (EPA, 2010a), EPA stated that the updated risk assessment should include the following exposure assessment considerations:

- Information related to recepto r population, expected land use, and grou nd water use in the vicinity of the Site (Comment 18).
- Relevant RAGS revisions, applicable expos ure pathways (e.g., derm al (RAGS E) and inhalation (RAGS F)), and current toxico logical information for each COC (Comment 18).
- A paragraph regarding the expos ure routes and pathways [pre sented in SWSFS Part I], including potential exposur e through the inhala tion pathway associated with the evaporation ponds.

Each of the exposure assessment-related issues is addressed within this section of the report and the toxicological information is provided in Section 4.

3.1.1 Physical Setting

The Church Rock Site is loca ted approximately 17 miles northeast of Church Rock, McKinley County, New Mexico. The local climate is arid, with approximately 10.6 inches of precipitation per year. The principal surface water feature in the vicinity of the Site is the Pipeline Arroyo (and Pipeline Canyon), which runs through the Site from northeast to southwest. Surface water flows in the arroyo seasonally and alluvium is present along the feature and its floodplain.

The Site is situated on alluvial valley fill, sandstone, and shale of Cretaceous age at the southern margin of the San Juan Basin. T he stratigraphic units of interest at the Site are the three uppermost water-bearing units (i.e., hydrostratigraphic units) as follows (EPA, 2008):

- 1. Alluvium, which is located along Pipe line Arroyo, has a m aximum thickness of approximately 150 ft and a maximum width of 4,000 ft;
- 2. Zone 3 (the upperm ost stratigraphic unit of the Upper Gallup Sandstone, having a thickness of 70 to 90 ft in the area of the tailings impoundments); and
- 3. Zone 1 (the lowest stratigr aphic unit of the Upper Gallup Sandstone, having a thickness of 80 to 90 ft in the area of the tailings impoundments).



There are three genetic classes of groundwater present in the vicin ity of the Church Rock Site: (1) pre-mining water, (2) post-m ining, pre-tailings water, and (3) tai lings-seepage impacted groundwater. Pre-m ining water is natural water that is present only in the U pper Gallup Formation north of UNC-owned property; this water does not underlie the Site and is not a focus of the HHRA. From approximately 1969 to 1986, large quantities of groundwater were pumped from the nearby NECR and Quivira mines to dewater the underground workings, and discharged to the Pipeline Arroyo. A portion of the mine discharge infiltrated into the alluvium and then into the Zone 3 and Zone 1 bedrock units. T his water is referred to as the post-m ining, pre-tailings water in the ROD which designated it as the background groundwater at the Site.

Seepage from the tailings, which were depos ited in the impoundments beginning in 1977, subsequently impacted the Site background water. Im pact from the tailings seepage has been observed in the alluvium to the west and so uthwest of the tailings impoundments (i.e., the Southwest Alluvium) and, because they are in contact with both the alluvium and the tailings in the vicinity of the impoundments, in Zone 3 and Zone 1 to the nort heast and east of the impoundments (Figure 3).

3.1.2 Potentially Exposed Populations

Information related to the poten tial receptor population, current and expected land use, and groundwater use near the Church Rock Site has been used to identify potentially exposed populations and to develop and sele ct exposure scenarios. Land use information is available in UNC's Annual Land Use Reports, which are prepared and submitted to NRC as a condition of UNC's NRC Source Materials License. The Annual Land Use Report for 2009 is provided in Appendix D (on CD).

Land use in the vicinity of the Site has not changed significantly in more than 30 years. The area surrounding the Site is sp arsely populated and the primary land use is grazing for sheep, cattle, and horses. The 2009 Land Use Report indicates that there are a total of thirty-six hom e sites and eight wells within approximately two miles of the former mill site. Two of the wells listed in the 2009 Land Use Report are abandoned and two ar e used as monitoring wells. Only two of the wells are identified as having domestic use (including the Church Rock Site water supply well (the mill well, which is very d eep and open to the W estwater Canyon Formation) and the Circle Wash Well (an alluvium well south of the Puerco River). Three wells, including the Circle Wash Well, the Friendship W ell (14T-586) and Well 15K-303 are used f or livestock watering, but due to their locations relative to hydraulic gradient s, the first two wells cannot be impacted by seepage from the Chur ch Rock Site. The Friendship well taps the Lo wer Gallup Formation. Well 15K-303, located more than two m iles to the northeast of the mill Site, is the only local well known to tap the Upper Gallup For mation and is used for livestock watering; however, it is too distant to be im pacted by seepage from the Church Rock Site, and the results of sampling (King, 2007) indicate the water has not been impacted by tailings seepage and it is unsuitable for human consumption. No residents have private wells for dom estic water supply and many haul their own water from known (although often unregulated) sources for dom estic supply and livestock watering. King (2007) cites the results of a 1999 survey by the Church



Rock Uranium Mining Project (C RUMP) which i ndicated that m ore than 80 percent of the nearby Churchrock Chapter residents haul water even when connected to a public water supply system. King (2007) also cites CR UMP groundwater monitoring data which indicate that the Friendship Well (Well 14T-586) was abandoned in 2003.

There is no current human exposure to groundwat er at the S ite (EPA, 2008) except during the quarterly groundwater sampling conducted by UNC personnel, and no potential future exposure to groundwater contaminants on UNC-owned property, because no groundwater supply wells drawing on any of the three hydrostratigraphic units will be allowed on UNC property, and the same restriction will apply on ce this property is turne d over to the DOE f or long-term surveillance monitoring.

Current potential effects on the ecology are mainly from the disc harge of pumped water from Zone 3, and purged water from quarterly groundwater sampling, into the evaporation ponds on the South Cell. Illegally grazing stock have very rarely consumed water here but Site access is restricted according to the NRC License and key parts of the Site fencing have recently been physically strengthened, which has further decreased the rate of incursions.

Considering land ownership patterns and lim ited water availability, alternate future land use is unlikely, with the possible exception of additional mining-related activities such as in-situ leach mining. The assumed, potentially exposed populations to COPCs in groundwater, in the future residential exposure scenarios evaluated in this report, are those indi viduals that would use groundwater for dom estic purposes from hypothetical wells overlying the seepage-im pacted groundwater in locations just outside Section 2 (for the Southwest Alluvium and Zone 1) and just north of Section 36 (for Zone 3).

3.2 Identification of Exposure Pathways

Potential exposure pathways are id entified to estim ate the doses of contaminants to which populations may be exposed. The following four elements are necessary to identify a potentially complete exposure: (1) a source and m echanism of release of contaminants to the environment, (2) an environmental transport m edium or m echanism of transfer of contaminants among environmental media, (3) a point of potential contact of humans to the contam inated medium, and (4) an identified route of exposure. An id entified pathway indicates that the potential for exposure exists; it does not imply that exposures do or may actually occur.

3.2.1 Sources and Receiving Media

The source of the COPCs in the environm ent is the tailings im poundments and the only environmental medium impacted by the tailings is groundwater. Acidic tailings liqu ids seeped from the impoundments into the groundwater in the Southwest Alluvium, Zone 1, and Zone 3. The affected groundwater has relatively low (acidic) pH and elevat ed concentrations of certain heavy metals, radionuclides, sulfate, and total dissolved solids TDS. There is no local discharge of groundwater to surface water. Downdip to the north, the deep regional continuation of the Gallup Formation is in ferred to leak upward and discharge to the San Juan River, which is



approximately 150 m iles from the Site (St one, 1981; Raymondi and Conrad, 1983) and far beyond the potential reach of Site impacted gro undwater). Source control and on-site surface reclamation activities conducted within Section 2 under the direction of the NRC (pursuant to the facility's NRC Source Materials License) have eliminated the potential COPC releases from the tailings impoundments to the atmosphere. Potential radionuclide emissions from the evaporation ponds are monitored within Section 2 under the direction of the NRC, pursuant to the facility's NRC Source Materials License, and are at acceptable levels.

3.2.2 Fate and Transport in Release Media

Seepage-impacted groundwater has slowly migrated from the tailings impoundments in each of three underlying affected hydrostrat igraphic units (Southwest Alluvium, Zone 1, a nd Zone 3). The groundwater in the alluvi um flows to the southwest beneath Pipeline Arroyo. The groundwater in both Zone 1 and 3 flows prim arily in a north-northeasterly direction, following the direction of the bedrock dip. Therefore, tailings-seepage impact has been ob served in the alluvium to the west and southwest of the ta ilings impoundments (i.e., the Southwest Alluvium) and in Zone 3 and Zone 1 to the northeast and east of the impoundments. The extent of seepage-impacted groundwater m igration in October 2010 is shown in Figure 3 (Chester Engineers, 2011). Historically, the directi ons of groundwater flow in Zone 3 and Zone 1 were to the northeast and east, respectively, due to groundwater mounding in the overlying alluvium. These earlier flow directions are also reflected in the historical distribution of seepage impacts within these hydrostratigraphic units.

Currently, seepage impacts in the Southwest Alluvium extend beyond Section 2 into Sections 3 and 10, and seepage impacts in Zone 1 extend into Section 1. Seepage impacts in Zone 3 have been observed beyond Section 2 in Section 36, but have not migrated beyond the UNC property boundary to the north. Investigations are underway to explore the transport of seepage-impacted water beyond Section 36; for this risk assessm ent, it is p resumed that seepage-impacted water could potentially migrate beyond the Section 36 boundary (so as to over-estimate possible risks). A portion of the Zone 3 seepage-im pacted groundwater extends off the property into Section 1 (Figure 3); however, NRC has determ ined that this area is not a pot ential point-of-exposure (POE) because of limited and declining saturation (NRC, 1999).

Downgradient seepage-impacted water m igration has been and is expected to continue to be limited by remediation activities and natural attenuation (Chester Engineers, 2011). In Zone 3, source control (neutralizing and later dewatering of the North Cell), neutralization of the seepage by natural attenuation, and active rem ediation have limited the migration of seepage impact. In Zone 1, COPC concentrations are attenuated by reactions between the bedrock m atrix, and the tailings fluids. The natural system is succes sfully attenuating the seepage im pacts by th e processes of neutralization, pr ecipitation, and adsorption. In Zone 1 and the Southwest Alluvium, natural geochem ical attenuation has re duced COPC concentrations s uch that no hazardous constituents exceed Site standard s (except for lead-210) ou tside the UNC property boundary within seepage-impacted water. An unexpectedly larg e number of detections and exceedances of lead-210 occurred in the laboratory analytical results for the samples collected in



October 2010 (further discussed in Chester Engin eers, 2011). Off-site impacted groundwater in the Southwest Alluvium has quality that is equal to or better than the off-site background water quality; both types of groundwater are unsuitable for human consumption.

There is likely to be insufficient volumes of water available in Zones 1 and 3 for use as a potable water sources. In the ROD, EP A stated the following regarding the inacc essibility and unsuitability of Zone 1 for water supply we lls: "EPA studies indi cate that the physical characteristics of Zone 1 are su ch that sufficient quantities of water coul d not be pumped from the sandstone to support volumes required for domestic or livestock purposes. Therefore, Zone 1 would not be a good candidate for locating a dom estic or livestock well even if there were no impacts from tailings seepage" (R OD, EPA 1988c, Appendix H [Re sponsiveness Summary], Response to Comment 9 in Section 2, p. 4). Recent studies related to the operation of the Zone 3 remediation pumping system indicate that the potential use of Zone 3 groundwater as a potable water source is also limited. Twenty-two years of remedial pumping have resulted in significant dewatering of Zone 3. The saturated thickness measured in Zone 3 wells has declined by 69 percent on average since the third quarter of 1989. Additionally, most of the Zone 3 pumping wells have reduced yields controlled by the foll owing physical factors: (1) encrustation along the wellbore of iron oxyhydroxides, carbonates, and/or gypsum; (2) precipitation of am orphous aluminosilicates (e.g., well EPA 14); (3) alteration of feldspar to clays within the bedrock matrix; and (4) reduced saturated thickn esses (Chester Engineers, 2010). At some time in the future, there will be a balance between the tendency for the irreducible elevation head (caused by the bedrock dip) to promote the continuing northward migration of seepage-impacted water and the tendency for the seepage-induced permeability reductions (due to factor 3, listed above) to hold the groundwater in place. Moreover, the quantity of acidic tailings seepage water is fixed while the availability of alkalinity in Zone 3 is unlimited in comparison. It is inevitable that the tailings seepage will be fully neutralized by reactions with carbonate minerals in the Gallup Form ation. The exact tim ing and location of the m aximum extent of seepage-impacted wate r in Zone 3cannot be predicted with precision (e.g., Chester Engineers, 2011); however, it is reasonable to estimate that it would probably be on the order of hundreds of feet rath er than thousands of feet. Therefore, setting a hypothetical groundwater expos ure point immediately to the north of the Section 36 boundary represents the m aximum potential exposure to Z one seepage-impacted groundwater.

3.2.3 Exposure Points and Exposure Routes

This section identifies the potential exposure points and exposure routes that m ake up the potential exposure pathways. As previously stated, groundwater is the only seep age-impacted medium, there is no current exposure to seepage-impacted groundwater (EPA, 2008) and there is no potential for future human exposure to groundwater in the property owned by UNC (Sections 2 and 36). Therefore, potential future exposur e to seepage-impacted groundwater could occur only at exposure points outside Sections 2 and 36. Because each of the three hydrostratigraph ic units have different water chemistries, likelihoods of exposure, and rem edial alternatives, the



following potential future groundwater exposur e points at the UNC property boundary are selected for evaluation in the HHRA:

- Southwest Alluvium a hypothetical future we ll located just west of the UNC property boundary in Section 3;
- Zone 1 a hypothetical future well located just east of the UNC property boundary in Section 1; and
- Zone 3 a hypothetical future well located just north of the UNC property boundary in Section 36.

It is unlikely that actual human exposure to seepage-impacted water using a dom estic well will occur at any of the three hypot hetical exposure points because both the background water and impacted water in each of the hydrostratigraph ic units are not suitab le for use as a prim ary drinking water source (e.g., due to sulfate and other chemicals that affect potability). Nonetheless, a hypothetical future residential la nd-use scenario represents the only conceivably possible exposure pathway to grou ndwater, given the current and an ticipated future land use. The residential RME exposure scenario assumes that residents would cons truct residences and live adjacent to the UNC property boundary near the tailings impoundments for up to 30 years, and that residents would use seepage-im pacted groundwater for all domestic water needs. Risk calculations based on this scenario will provide maximum estimates of the risk.

To assess the potential exposure of a hypothetical future resident, three exposure pathways were selected for evaluation:

- Ingestion of groundwater as the drinking water source;
- Direct dermal contact with groundwater through bathing; and
- Inhalation of volatile compounds in groundwat er during showering exposure and other domestic tapwater uses.

A thirty year exposure period was evaluated, consistent E PA risk assessment guidance. A residential adult (aged 7-30) and a young child (age d 1 to 6) were select ed as the potentially exposed populations for non-carcinogenic C OPCs. A com bined child/adult receptor r was selected as the potentially exposed population for the carcinogenic COPCs , including radionuclides. EPA dermal risk assessment guidance (EPA, 2004) indicates that the age-adjusted child/adult receptor typically is the most sensitive receptor for cancer endpoints and the child typically is the most sensitive receptor for non-cancer endpoints. For non-carcinogenic compounds, groundwater exposure rates for children ar e higher than for adults because the ratio of intake rate to body weight is higher. For non-radionuclide carcinogens, the com bined child/adult exposure scenario is conservative primarily because the combined intake for the child and adult (a larger intake than either a child or adult intake alone) is averaged over the lifetime of the receptor because of the assumption that cancer may develop even after actual exposure has ceased.



A second potential exposure scenario was consid ered: the hypothetical future secondary hum an exposure from consumption of m eat or milk (i.e., food pathway) from liv estock watered with groundwater that has been impacted by tailings seepage. This second scenario was excluded for the following reasons: (1) expo sure would be insignificant co mpared to hypothetical use of impacted water as a domestic water supply; (2) the land use survey indicates it is not a current or anticipated future exposure pathway because there are no livestock watering wells that are currently, or anticipated to be, i mpacted by seepage, and (3) there is significant t uncertainty related to exposure assumptions for this hypothetical exposure scenario (e.g., percentage of local consumption of local m eat/milk products sourced from seepage-impacted area, likelihood that livestock would consume impacted water, and bioaccumulation factors). This determination was made with EPA during a teleconference on November 1, 2010.

3.3 Quantification of Exposure

3.3.1 Exposure Point Concentrations

Statistical analyses were perform ed using Pr oUCL (EPA, 2007) to ve rify the statistical distribution of COPC concentrations in groundwater in each of the hydrostratigraphic units and to estimate an EPC for each COPC. The statis tical analyses are described in the EPA-approved submittal (N.A. Water Systems, 2008c), which is attached as Appendix B (on CD).

The EPC is the concentration term used in the exposure equations and represents the average COPC concentration that is c ontacted over the exposure period. Because of the uncertainty associated with estim ating the true averag e COPC concentrations, the 95 percent upper confidence limit of the arithmetic mean (UCL95) is us ed to rep resent this variable when calculating the reasonable maximum exposure (RME) for selected exposure scenarios. As part of the analysis, the statistic cal distribution (i.e., norm al, log-normal, or non-param etric) represented by the data was determined so that the proper r statistical test could be applied to calculate the appropriate UCL95. The UCL95 provides reasonable c onfidence that the true Site average concentration will not be underestimated.

The EPCs calculated for the exposure scenarios for each hydrostratigraphic unit are shown on Table 3.1.RME and Table 3.2.RME in Appendix A. As described in Section 2.2, total uranium mass concentrations are analyzed and reported for r Church Rock Site groundwater sam ples, but there is no site-specific data regarding the a bundance of the individual uranium isotopes to evaluate the radioactive carcinogenic risk of uranium at the Site. Therefore, the EPCs shown for the three major uranium isotopes (U-234, U- 235, and U-238) in groundwater (pCi/L) wer e estimated from the to tal uranium mass concentrations (mg/L) using the assum ption that the isotopes were present in proporti on to their natural abundance. The calculations are shown on supplemental Table 3.A.RME in Appendix A.

3.3.2 Estimation of Non-Radiological Chemical Intakes

Environmental medium-specific exposure algorithms were developed for each of the identified exposure route/pathways to estimate COPC intake of non-radiological COPCs by receptors (e.g.,



adult and young child residents) in potentia lly exposed populations. The exposure to radiological COPCs was assessed using similar algorithms and assumptions, but the total intake was calculated for the entire exposure period. Expos ure to radiological COPCs is discussed in Section 3.3.3.

For each identified pathway, an RME scenario was assessed in which the exposure factors used are both average and upper-bound (90th to 95th percentile distribution) point estimates. The RME scenario is intended to represent the highest exposure that is reasonably expected to occur at a site (EPA, 1989).

The exposure factors and exposure algorithms to estimate intake of Site-related contam inants through the groundwater ingestion and derm al exposure pathways for potential future adult and young child residents are listed in Tables 4.1.R ME and 4.2.RME in Appendix A. Details regarding the modeled intake methodologies for the derm al absorption exposure pathway are provided in supplemental Tables 7.A.RME and 7.B.RME in Appendix A. Details regarding the modeled intake methodology for the inhalation e xposure pathway to non-radionuclides s are provided in supplemental Tables 7.C.RME and 7.D.RME in Appendix A. T he exposure calculation methods are described in the following subsections.

3.3.2.1 Ingestion of Groundwater

A principal assumption of the future residential exposure scenario associated with the Church Rock Site was that groundwater would be used as the only water s upply for all domestic needs. The chronic daily intake (CDI, mg/kg-d) of non-radiological, non-carcinogenic COPCs in groundwater due to ingestion was calculated by the following equations:

$$CDI = \frac{CW \times IR_{W} \times EF \times ED}{BW \times ATn}$$

where:

CDI = chronic daily intake of contaminants in groundwater (mg/kg-day) CW = COPC concentration in groundwater (mg/L) IR_w = ingestion rate of water (L/day) EF = exposure frequency (days/year) ED = exposure duration (years) BW = body weight (kg) ATn = averaging time – non-carcinogenic effects (days).

The RME ingestion r ates for drinking water were assumed to be two liters per day (2 L/day; EPA, 1989) for adults and 1 L/day for child ren (California EPA, 1994) over an exposure frequency of 350 days/ year for exposure durations of 6 and 24 years for a child and adult, respectively. The averaging times used for non-carcinogenic COPC effects were 2,190 days (6 years) for a child and 8,760 days (24 years) for an adult.



For carcinogenic exposure under the com bined child/adult exposure scenario, an age-adjusted ingestion rate and corresponding intake equation was used:

$$CDI = \frac{CW \times IR_{\scriptscriptstyle Wadj} \times EF}{ATc}$$

And

$$IR_{\textit{Wadj}} = \frac{EDc \times IRWc}{BWc} + \frac{EDa \times IRWa}{BWa}$$

where:

 $\begin{array}{l} \text{CDI} = \text{chronic daily intake of contaminants in groundwater (mg/kg-day)} \\ \text{CW} = \text{COPC concentration in groundwater (mg/L)} \\ \text{IR}_{\text{Wadj}} = \text{age-adjusted ingestion rate (1.09 L-year/kg-day)} \\ \text{IR}_{\text{Wa}} = \text{adult ingestion rate (2 L/day)} \\ \text{IR}_{\text{Wc}} = \text{child ingestion rate (1 L/day)} \\ \text{EF} = \text{exposure frequency (350 days/year)} \\ \text{EDc} = \text{child exposure duration (6 years)} \\ \text{EDa} = \text{adult exposure duration (24 years)} \\ \text{BWc} = \text{body weight (15 kg)} \\ \text{BWa} = \text{body weight (70 kg)} \\ \text{ATc} = \text{averaging time - carcinogenic effects (25,550 days [70 years])}. \end{array}$

3.3.2.2 Dermal Contact With Groundwater

The dermal contact exposure pathway accounts for daily exposure to water while bathing. With respect to non-radionuclide COPCs, the dermal exposure pathway is important for many organic contaminants and som e inorganic contam inants. Under EPA guidance for the derm al-water exposure pathway, only those chemicals which cont ribute a significant dose relative to the oral exposure pathway (i.e., more than 10% of the dose from drinking water ingestion) are required to be carried through the risk assessm ent (EPA, 2004). However, as a conservative m easure, dermal exposure was evaluated for each of the non-radionu clide COPCs (i.e., those selected for the ingestion exposure pathway) in each of the hydrostratigraphic units.

The dermally absorbed dose (DAD) serves as the intake rate for non-radiologic COPCs in seepage impacted groundwater due to derm al absorption while bathing. DAD is calculated by the following formula:

$$DAD = \frac{DA_{event} \times EV \times ED \times EF \times SA}{BW \times ATn}$$

where:

DAD = dermally absorbed dose (mg/kg-day)



- DA_{event} = absorbed dos e per event (mg/cm2-event) The calculation of DAevent is based on chemical-specific parameters and equations presented in EPA (2004) and shown in Table 4.1 in Appendix A and calculated in Tables 7.A.RME and 7.B.RME i n Appendix A.
- SA = surface area of exposed skin (cm2)
- EV = event frequency (events/day)
- EF = exposure frequency (days/yr)
- ED = exposure duration (years)

BW = body weight (kg)

ATn = averaging time – noncarcinogenic effects (days).

Default RME exposure parameter values were obtained from EPA dermal guidance (EPA, 2004). Body surface area values are 18,000 cm2 for adults and 6,600 cm2 for a child. Bathing tim e values are assumed to be 0.58 hours for adults and one hour for a child per event and the event frequency was assumed to be one event (i.e., bath or shower) per day. The averaging times used for non-carcinogenic COPC effects were 2,190 days (6 years) for a child and 8,760 days (24 years) for an adult. For the combined child/adult risk exposure scenario for carcinogens, an age-adjusted DAD was calculated as follows:

$$DAD = \left(\frac{DA_{event-c} \times EV \times EDa \times EF \times SAc}{BWc} + \frac{DA_{event-a} \times EV \times EDc \times EF \times SAa}{BWa}\right) \times \frac{1}{ATc}$$

where:

DAD = dermally absorbed dose (mg/kg-day)

- $DA_{event-c} = Absorbed dose per event-child (mg/cm2-event) The calculation of DAevent is based$ on chemical-specific parameters and equations presented in EPA (2004).
- $DA_{event-a} = absorbed dose per event-adult (mg/cm2-event) The calculation of DAevent is based$ on chemical-specific parameters and equations presented in EPA (2004).
- SAc = surface area of exposed skin child (6,600 cm2)
- SAa = surface area of exposed skin adult (18,000 cm2)
- EV = event frequency (1 events/day)
- EF = exposure frequency (350 days/yr)
- EDc = exposure duration (6 years)
- EDa = exposure duration (24 years)
- BWc = body weight child (15 kg)
- BWa = body weight adult (70 kg)
- ATc = averaging time carcinogenic effects (25,550 days).



3.3.2.3 Inhalation of Groundwater

The inhalation of volatile organic com pounds (VOCs) in groundwater while bathing was included as a potential exposure pathway under the future resident scenario. An exposure model developed by Foster and Chrostowski (1987) was used to estim ate VOC exposure point concentrations in bathroom air during and after bathing (see supplemental risk assessment Tables 7.C.RME [for adult exposure] and 7.D.RME [for child exposure] in Appendix A). C hloroform, which is detected in groundwater at low concentrations, is the only non-radionuclide COPC that is sufficiently volatile to be included in the assessment of this pathway. The following equation was used to calculate the non-radionuclide COPC exposure concentration:

$$EC = \frac{CA \times EV \times ET \times EF \times ED}{CF \times AT}$$

where:

EC = Average Daily Exposure Concentration (mg/m³)

CA = chemical concentration in air (mg/m³) derived using Foster and Chrostowski shower model (1987)

EV = event frequency (1 event/day)

EF = exposure frequency (350 days/year)

ED = exposure duration (6 years child, 24 years adult)

ET = exposure event time (0.58 hours/event adult, 1 hour/event child)

AT = averaging time (2,190 days child, 8,760 days adult)

CF = conversion factor 24 hrs/day.

For the combined child/adult risk exposure scen ario for carcinogens, an age-adjusted EC was calculated as follows:

$$EC_{adj} = \left(\frac{CA_a \times EVa \times ETa \times EDa \times EF}{CF \times AT}\right) + \left(\frac{CA_c \times EVc \times ETc \times EDc \times EF}{CF \times AT}\right)$$

where:

 EC_{adj} = age-adjusted average daily exposure concentration (mg/m³) CAa = chemical concentration in air – adult (mg/m³) (using Foster and Chrostowski, 1987) CAc = chemical concentration in air – child (mg/m³) (using Foster and Chrostowski, 1987) EVa = event frequency - adult (1 event/day) EVc = event frequency – child (1 event/day) EF = exposure frequency (350 days/year) EDa = exposure duration – adult (6 years) EDc = exposure duration – child (24 years adult) ETa = exposure event period - adult (0.58 hours/event) ETc = exposure event period – child (1 hour/event) AT = averaging time (25,550 days)



CF = conversion factor 24 hrs/day.

3.3.3 Estimation of Radiological Intakes

Generally, standardized default exposure equations for radionuclides are similar to those for non-radionuclides (EPA, 1991a). However, according to EPA (1991a), there are three principal differences:

- The equations utilize input qua ntities of activity (e.g., pCi/L) rathe r than mass (mg/L) because health effects due to radionuclide effects are directly related to the amount, type and energy of the radiation deposited in specific body tissues and organs;
- Radionuclide exposure equations consider only the carcinogenic effects of radionuclides; and
- Radionuclide exposure equations use cancer slope factors that are best estim ates (i.e., median or 50th percentile values) of the age-averag ed, lifetime excess total can cer risk per unit intake of a radionuclide. Radionuclide slope factors given in the EPA's He alth Effects Assessment Summary Tables (HEAST, <u>http://www.epa.gov/radiation/heast;</u> EPA, 2001a) are calculated for individuals us ing a non-threshold, linear dose-response model that accounts for radionuclide absorp tion into and distribution throughout the body and also accounts for the ag e, sex and we ight of an in dividual. The model then averages the risk over a lifetim e exposure (i.e., 70 years). Consequently, radionuclide slope factors are not expressed as a function of body weight and time and do not require adjustments for gastrointestinal absorption or lung transfer efficiencies.

For the Church Rock Site HHRA, radionuclide COPC exposure was evaluated using the same potential future residential re ceptors as those eval uated for chem ical COPCs. The exposure pathways evaluated included groundwater inges tion and inhalation of groundwater related to domestic water uses. The groundwater inhalati on exposure pathway has been included for one COPC, radium-226, which has the radioactive gas radon-222 (Rn-222) as its decay product. This is consistent with EPA's inclusion of the einhalation pathway in it s tapwater radionuclide PRG calculation for radium-226 (http://epa-prgs.ornl.gov/radionuclides/).

Dermal uptake is typically not an important exposure route for radionuclides and it has not been evaluated in this risk assessment. The skin is generally an effective barrier against absorption of radionuclides, which have small permeability constants (EPA, 1989), and the derm al absorption exposure pathway is considered very minor with respect to ot her exposure routes, such as ingestion and inhalation.

Other exposure pathways that are often considered in risk assessments with radionuclide COPCs include (1) inhalation of airborne particulates, and (2) external ga mma radiation. As previously described, these exposure pathways also do not apply to the Church Rock Site, because (1) source control and on-site surface reclamation activities were conducted within Section 2 under the direction of the NRC, pursuant to the f acility's NRC Source Materials License and (2) the operation of the evaporation ponds and the m onitoring of potential radionuclide emissions are



conducted within Section 2 under the direction of Source Materials License.

the NRC, pursuant to the facility's NRC

Ingestion exposure to radionuclides was assessed using the sam e equations presented in the previous sections for chem ical contaminants, except that the body we ight and averaging time terms are omitted for reasons described previously. The result of these calculations is an estimate of total intake over the exposure duration, expressed in terms of activity (i.e., pCi/L), instead of a body weight norm alized chronic daily intake (e .g., mg/kg-day). As with the carcinogenic chemical risk assessment, combined adult/child exposure assessment was completed for each of the selected exposure pathways for radionuclide COPCs.

3.3.3.1 Ingestion of Groundwater

The intake of radiological COPCs in groundwater due to i ngestion was calculated for the child/adult receptor by the following formulas:

$$Intake = CW_{R} \times IR_{Wadj} \times EF \times ED$$

and

$$IR_{Wadj} = (IR_{Wc} \times EDc + IR_{Wa} \times EDa) \times \frac{1}{ED}$$

where:

Intake = intake of COPCs in groundwater (pCi) CW_R = radionuclide activity in groundwater (pCi/L) IR_{Wadj} = age-adjusted ingestion rate (1.8 L/day) IR_{Wa} = adult ingestion rate (2 L/day) IR_{Wc} = child ingestion rate (1 L/day) EF = exposure frequency (350 days/year) ED = exposure duration (30 years) EDc = child exposure duration (6 years) EDa = adult exposure duration (24 years).

3.3.3.2 Inhalation of Groundwater

Inhalation exposure to radium -226 in groundwater was assessed using a different method than that used for the chemical COPC chloroform. The Andelman volatilization factor K (Andelman, 1990), which is used by EPA in t he development of its radionuclide PRGs and referenced in RAGS Part B (EPA, 1991a), was applied to calculate the fraction of radionuclide COPCs that are transferred from groundwater to the air through all uses of hous ehold water (e.g., showering, laundering, dishwashing). Residents are assumed to potentially be exposed to indoor air in their homes for 24 hours a day resulting in an exposure tim e ratio of 1. This is meant to be protective of sensitive populations who stay at home (i.e., young children or the elderly). Although radium



is not considered to be volatile, radon has a Henry's Law constant approximately 30 times that of chloroform. Andel man's K (0.0005) is a unitl ess constant, but it is comm only given with a conversion factor of 1000 L/m³ that is used so the resulting air concentration is expressed in units of pCi/m³. The Andelman K represents an average transfer efficiency of 50 percent, in that half of the concentration of each chemical in water will be transferred into air by all water uses (e.g., a 10 pCi/L concentration in water will result in a 5 pC i/m³ concentration in air). EPA (1991a) indicates that the Andelman K is based primarily on experimental data from the volatilization of radon from water and cites assumptions used in the development of Andelman K to include (1) the volume of water used in a residence for a family of four is 720 L/day (approxim ately 190 gallons/day), (2) the volume of the dwelling is 150,000 L, and (3) the air exchange rate is 0.25 m³/hr.

The intake of the radiological COPC radium -226 in groundwater through the inhalation pathway was calculated for the child/adult receptor by the following formulas (also see Table 7.E.RME in Appendix A):

$$Intake = CW_R \times IRA_{adj} \times EF \times ED \times ET \times K \times \frac{1}{CF}$$

and

$$IF_{adj} = (IRAc \times EDc + IRAa \times EDa) \times \frac{1}{ED}$$

Intake = intake of contaminants in groundwater (pCi) CW_R = radionuclide activity in groundwater (pCi/L) IRA_{adj} = age-adjusted inhalation rate (18 m³/day) IRA_a = adult inhalation rate (20 L/day) IRA_c = child inhalation rate (10 L/day) EF = exposure frequency (350 days/year) ED = exposure duration (30 years) ET = exposure time (24 hours/day) EDc = child exposure duration (6 years) EDa = adult exposure duration (24 years) CF = conversion factor 24 hrs/day K = volatilization factor of Andelman (0.5 L/m³)



Section 4 Toxicity Assessment

The toxicity assessment component of the HHRA considers (1) the types of adverse health effects associated with chem ical exposures, (2) the relations hip between the m agnitude of exposure and adverse effects, and (3) related uncertainties such as the weight of evidence of a particular chemical's carcinogenicity in hum ans. For the Church Rock Site, the toxicity assessment was based e ntirely on existing toxicity information available from EPA and other sources, including information regarding carcin ogenic and non-carcinogenic effects associated with radionuclide and non-radionuclide COPCs.

Two categories of toxicological effects, cance r and non-cancer causing health effects, were evaluated in the toxicity assessm ent for each identified COPC for the Church Rock Site. Toxicity values are used to quantify the probability of observing cancer and non-cancer effects in a potentially exposed population. Several types of toxicity values are used to express a COPC's dose-response-effect relationship:

- Oral Reference Dose (RfD) an RfD, expre potential non-carcinogenic effects through radiological contaminants. Oral RfDs are exposure to COPCs (i.e., absorbed RfD).
 ssed in mg/kg-day, is used for estimating ingestion exposure, prim arily from nonalso used to calculate RfDs for derm al
- Inhalation Reference Concentration (R fC) RfCs are expressed in m g/m³ and used for estimating potential non-carcinogenic effects though inhalation exposure, primarily from non-radiological contaminants.
- Oral and inhalation Slope Factor (SF) an SF is us ed for estim ating potential carcinogenic effects. An oral SF for non-radiological effects is expressed in the units of (mg/kg-day)⁻¹. The oral SFs are also used to calcu late slope factors for derm al exposure to COPCs (i.e., absorbed cancer SF). An oral or inhalation SF for radiological COPCs is expressed in the units of (pCi)⁻¹ (i.e., risk/pCi) for radiological contaminants.
- Inhalation Unit Risk (IUR) an IUR is used for estimating potential carcinogenic effects for non-radionuclides and is expressed in the units of $(\mu g/m^3)^{-1}$.

In their September 2, 2010 comment letter (EPA, 2010a), EPA requested that the risk assessment include updated toxicity factors. For the Chur ch Rock Site HHRA, all toxicity f actors were updated in accordance with EPA's Superfund program hierarchy of human health toxicity values (EPA, 2003), which is as follows:

- Tier 1 EPA's IRIS IRIS is the first tier of the recommended hierarchy as the generally preferred source of human health toxicity values. IRIS generally contains RfDs, RfCs, SFs, and IUR values that have gone through a peer review and EPA consensus review process.
- Tier 2 EP A's Provisional Peer R eviewed Toxicity Values (PPRTVs) The Office of Research and Developm ent/National Center for Environm ental Assessment/Superfund



Health Risk Technical Support Center (STSC) develops PPRTVs on a chemical specific basis when requested by EPA's Superfund program.

• Tier 3 - Other Toxicity Values – Tier 3 includes additional EPA and non-EPA sources of toxicity information. Priority should be given to those sources of information that are the most current, the basis for which is transparent and publicly available, and which have been peer reviewed.

Information on the non-carcinogenic toxicity factors and effects for the COPCs is listed in Tables 5.1 and 5.2 and carcinogenic toxicity eff ects for the C OPCs (both non-radiological and radiological) are listed in Ta bles 6.1 and 6.2 in Appendix A. Sections 4.1 and 4.2 below summarize the development of toxicity information related to non-carcinogens and carcinogens, respectively.

4.1 Toxicity Information for Non-Carcinogens

Chronic RfD or RfCs are estimates of the daily exposure (intake) a human population (including any sensitive subpopulation) that is unlikely to cause an increased incidence of adverse health effects during a lifetime of exposure (i.e., a chronic exposure). Chronic RfD or RfC values are specifically developed to be protective for long-term exposure to a constituent. Subchronic RfDs and RfCs, which are u sed to evaluate the h ighest average daily exp osure over shorter tim e periods (i.e., between 2 weeks and 7 years) that will not cause adverse health effects, have not been used in the Church Rock Site HHRA, because there are no short-term exposure pathways evaluated.

As discussed during the November 1, 2010, teleconference, the November 2010 EPA RSL Table (EPA, 2010b) was the prim ary source of m ost COPC toxicity values used within the Church Rock Site HHRA. The toxicity values used as "defaults" in the RSL table are selected by EPA in a manner consistent with their 2003 guidance re garding hierarchy of hum an health toxicity values (<u>http://www.epa.gov/region9/superfund/prg</u>). The original sources of the tox icity information used on the RSL Table, as well as other sources used for ce rtain chemical COPCs, are noted on Tables 5.1 and 5.2 in Appendix A. Uranium is the only radionuclide COPC which has been determined to have chemical toxicity comparable to its radiotoxicity and for which an RfD has been included on the EPA RSL table to evaluate chemical toxicity. For the purposes of this HHRA, both effects (chemical toxicity and radiogenic cancer risk) are considered.

EPA guidance (EPA, 2004) was utili zed to calculate derm al toxicity f actors for the derm al contact with groundwater pathway. Derm al toxicity factors (RfDs) are based on absorbed dose (i.e., the amount of a chem ical absorbed through the skin), while oral RfDs are based on an administered dose (i.e., the am ount of a chem ical ingested). Therefore, to obtain dermal RfDs from oral R fDs, the oral RfD is ad justed using an oral ab sorption efficiency values (ABS_{GI}), which represents the fraction of an orally adm inistered dose of chemical that was absorbed by the gastrointestinal tract (i.e., the absorbed dose) in the critical study that was the basis of the oral RfD. The m agnitude of toxicity factor adjust ment is inversely proporti onal to the absorption



fraction in the critical study (i.e., if the absorp tion efficiency value is high, the absorbed dose approaches the administered dose, resulting in little difference in a dermal RfD derived from an oral RfD). As absorption efficiency in the cr itical study decreases, the difference between the absorbed dose and administered dose increases, and the dermal RfD decreases with respect to the oral RfD. The ABS_{GI} values used to calculate the dermal RfDs were identified in EPA guidance (EPA, 2004 and EPA, 2010b). The ABS_{GI} value selected for vanadium was that for "vanadium and compounds" in the EPA RSL Table (ABS_{GI} = 1; EPA, 2010b), rather than the value in the EPA dermal risk assessment guidance (ABS_{GI} = 0.026, EPA, 2004) because this value e corresponded with the selected oral RfD.

Certain inorganic groundwater contaminants at the Site do not have toxici ty values established (e.g., sulfate, chloride). These contaminants have not been included in the quantitative risk assessment calculations, but are addressed qualitatively in the uncertainty analysis portion of the risk assessment (Chapter 6).

4.2 Toxicity Information for Carcinogens

Cancer risks are expressed as the incremental probability of an individual developing cancer over a lifetime as a result of exposure to the potenti al carcinogen (i.e., ex cess individual lifetim e cancer risk). The SF and IUR toxicity values that define a quantitative relationship between dose and response are used to convert the estim ated daily intakes averaged over a human lifetime of exposure directly to an increm ental risk of an individual deve loping cancer. This approach assumes that the dose-response relationship is a linear relationship in the low-dose portion of the dose-response curve.

The Church Rock Site risk assessment consider s cancer risk associated with both exposures to chemicals and radionuclides. For r exposure to chem icals, slope factors typically represent an upper bound estim ate or 95 th percent confidence lim it value that has been obtained from extrapolation from laboratory experiments (EPA, 1989). Cancer slope f actors for radionuclides are central tendency es timates (i.e., median or 50 th percentile values) of the ag e-averaged increased lifetime cancer risk that are based on epidemiological studies of radiogenic cancers in humans (EPA 1989).

As discussed with EPA in the November 1, 2010, teleconference, the November 2010 EPA RSL Table (EPA, 2010b) was used as the prim ary source of most chemical COPC slope factors and IURs. As with the non-carcinogeni c toxicity values, the slope factors used in the RSL table are selected by EPA in a manner consistent with their 2003 guidance regarding hierarchy of hum an health toxicity values (RSL Table User's Guide, <u>http://www.epa.gov/reg3hwmd/risk/human/rb-concentration_table/usersguide.htm).</u> The original sources of the toxicity inform ation used on the RSL Table, as well as other sources used for certain chemical COPCs are noted on Tables 6.1 and 6.2 in Appendix A. For the derm al groundwater exposure pathway, Table 6.1 shows the application of the oral absorption efficiency values (ABS _{GI}) to calculate the absorbed cancer slope factor; however, no adjustm ents to the oral SFs were necessary because the ABS _{GI} value



was one (i.e., 100 percent) for each of the evaluated carcinogenic COPCs (EPA, 2004; EP A 2010b).

EPA has classified all radionuc lides as known hum an carcinogens (Class A carcinogens) based on their property of em itting ionizing radiation in epidemiological studies of radiogenic cancers in humans. W ith the exception of uranium (s ee Section 4.1), the chem ical toxicity of radionuclide COPCs is low in com parison to the carcinogenic risk; therefore chemical risk was not considered further in the risk assessment for radionuclides other than uranium.

Radionuclide cancer slope factors (i.e., risk coefficients for total cancer morbidity) were obtained from the EPA's HEAST (EPA, 2001a). The e 2001 update of the HEAST for radionuclides incorporates slope factor values based on Fe deral Guidance Report No. 13 (EPA 1999), which incorporates state-of-the-art models and m ethods that take into account age and gender dependence of radionuclide intake, m etabolism, dosimetry, radiogenic cancer risk, and competing risks. The EPA docum ent "User's Guide: Radionuc lide Carcinogenicity" (http://www.epa.gov/radiation/heast/) describes the derivation of the slope f actors in the Radionuclide Table, information about the table, and contact information.

The User's Guide indicates that selected rad ionuclides and radioactive decay chain products are designated with the suffix "+D" (e.g., U-238+D, Ra-226+D) to indicate that cancer risk estimates for these radionuclides include the contributions from their short-lived decay products, assuming equal activity concentrations (i.e., secular equilibrium) with the principal or parent nuclide in the environment. The use of the "+D" designation can be important because some decay products can be more toxic than the parent isotope. Fo r example, Ra-226, which has a half-life of 1600 years, decays to radon-222 (Rn-222, a noble gas) by alpha particle em ission. However, Rn-222 and its daughters emit three additional alpha particles and two beta particles through the principle decay modes with a total half-life of less than four days. The User's Guide indicates that in the absence of site-specific data re garding secular equilibrium, the "+D" values for radionuclides should be used. For the Church Rock Site Risk Assessment, the cancer slope factors used for the following radionuclides had the "+D" designati on: uranium-235+D, uranium-238+D, radium-226+D, radium-228+D, and lead-210+D (see Tables 6.1 and 6.2 in Appendix A).



Section 5 Risk Characterization

The risk characterization step of the HHRA pr ocess summarizes and com bines outputs of the exposure and toxicity assessments to characterize risk in quantitative and qualitative statements. For the Church Rock Site HHRA, COPC intakes for RME exposure scenarios are combined with the toxicity values to estimate health hazards and cancer risks, as described in Sections 5.1 and 5.2. The RME exposure represents the highest exposure that is reasonably expected to occur at a site (EPA, 1989) and, therefore, the resulting risk estimate is conservative (i.e., well above the average case).

The only environm ental medium affected by the tailings im poundments is groundwater. Because there is no current expo sure to seepage-impacted groundwater, the HHRA does not evaluate past or current expos ures. Following the EPA's origin al PHA (EPA, 1988b), this risk assessment addresses each hydrostr atigraphic unit separately. R isk and hazard estim ates are presented for each o f the potential futu re exposure scenarios and are grouped by hydrostratigraphic unit (Sections 5.3 to 5.5) and the EPA RAGS Part D risk assessment summary tables (EPA, 2001) are provided in Appendix A.

The risk and hazard estimates for exposure to groundwater from each of the hydrostratigraphic units are based on the assum ption that future re sidents will install wells and m eet all the ir domestic water needs with seepage-im pacted groundwater from one of the three affected hydrostratigraphic units. The ri sk and hazard estim ates for the exposure scenario s within a selected hydrostratigraphic unit are considered to be additive because COPC exposure could occur independently through multiple scenarios in a residential setting (i.e., a receptor would likely be exposed both through groundwater inge stion and through derm al contact with groundwater during b athing). However, the risk and hazard estimates between hydrostratigraphic units are exclusive of one another because total daily groundwater exposure is limited to a receptor living at one location and using one groundwater source.

5.1 Method for Evaluating Non-Cancer Hazard

Risk estimates for non-cancer effects are calculated using the in take or exposure concentration calculated in the exposure assessment (Section 3) and toxicity benc hmarks (i.e., RfDs or RfCs) that represent intake levels below which it is unlikely that a receptor will experience adverse health effects following a chronic exposure. The non-carcinogenic toxicity, or hazard, to an individual for a specific COPC, is the average daily intake divided by the RfD or RfC. This ratio is known as the hazard quotient (HQ):

$$HQ = \frac{CDI}{RfD}$$
 (for ingestion exposure)

Or



$$HQ = \frac{EC}{RfC}$$
 (for inhalation exposure)

where: HQ = hazard quotient CDI = chronic daily intake (mg/kg-day or mg/m³) EC = Exposure concentration (mg/m³) RfD = Reference Dose (mg/kg-day) RfC = Reference Dose (mg/m³)

The sum of the HQs for each COPC evaluated w ithin an exposure scenario is k nown as the hazard index (HI). If the HI is les s than one, it is generally assumed that no adverse im pact to human health will occur as a res ult of the de fined exposure scenario. EPA conservatively assumes dose and effect to be additive for non-carcinogenic effects; if the HI exceeds one for the exposure scenario, this provides an indication that the exposed receptor may be subject to an adverse health impact (EPA, 1989). However, EPA (1989) has also noted that adding all HQ or HI values m ay overestimate hazards, becaus e the assumption of additivity is probably appropriate only for those chem icals having the sa me toxicological effect. Furtherm ore, the potential for adverse health effect do es not necessarily increase linearly as an RfD is approach ed or exceeded because RfDs do not have equal accuracy or precision and are not based on the same severity of toxic effects.

Therefore, if the HI is greater than one as a consequence of summing several hazard quotients of similar value, it is app ropriate to further evaluate the haz ard calculations by segr egating the compounds by target organ, toxico logical effect, or toxic m echanism, and to derive separate hazard indices for each group (EPA 1989). If any se gregated HI exceeds the target hazard level, this may indicate a potential adverse health impact. However, if all of the segregated HIs are less than the target hazard level, non-cancer health effects are unlikely to result from exposure to the COPCs included in the HI. In Sections 5. 3 to 5.5, non-cancer health hazards based on chemical contaminants are presented for CO PCs for each receptor p opulation and for each exposure pathway and then summarized across all media and exposure pathways. Target organs for non-cancer health hazards are indicated, and orga n-specific health hazards are also presented as applicable.

5.2 Method for Evaluating Cancer Risk

Risk estimates for cancer effects are expressed as an increased probability of contracting cancer (i.e., excess cancer risk). The EPA "target" acceptable excess cancer risk range (e.g., EPA, 1991a) is 1E-06 to 1E-04 (i.e., 1 in 1,000,000 to 1 in 10,000). Cancer risks above 1E-04 are generally considered u nacceptable and warrant so me form of remedial action. The risk calculation equation for non-radiological contaminants is as follows:

 $Risk = CDI \times SF$



or

 $Risk = EC \times IUR$

where:

Risk = probability (unitless) of an individual developing cancer over a lifetime CDI = chronic daily intake averaged over 70 years (mg/kg-day) EC = cxposure concentration (μ g/m³) SF = slope factor, expressed in (mg/kg-day)⁻¹ IUR = inhalation unit risk value, expressed in (μ g/m³)⁻¹.

The risk calculation equation for radiological contaminants is as follows:

 $Risk = Intake \times SF$

where:

Risk = a unitless probability (e.g., 1E-06) of an individual developing cancer over a lifetime Intake = total lifetime intake above background (pCi) $SF = slope factor, expressed in (pCi)^{-1}$.

The excess cancer ris k from exposure to multip le carcinogens is assum ed to be additiv e. Therefore, the total can cer risk from radiol ogical and non-radiological COPCs is calculated by summing the individual cancer risks for all COPC s across all exposure m edia and pathways. However, EPA (1989) also recommends that, due to differences in the way cancer toxicity values for radiological and chem ical risk are develope d, risk estimates for radiological and che mical risk should be tabulated separately in the final baseline risk assessment. In Sections 5.3 to 5.5, carcinogenic risks based on chemical and radiological contaminants are presented for COPCs for each receptor population and for each exposure pathway and then summ arized across all media and exposure pathways.

5.3 Risk Characterization Results for the Southwest Alluvium

Potential future residents may be exposed to COPCs in Southwest Alluvium groundwater if they install a w ell in Section 3, west of the Section 2 boundary, and use seepage-im pacted groundwater for drinking water and other domestic uses. Calculated hazards and risks associated with potential future exposure to seepage-impacted groundwater in the S outhwest Alluvium are summarized in Tables 7.1.RME through 7.3.RME, 8.1.RME, and 9.1.RME through 9.3.RM E (Appendix A). For the Southwest Alluvium , potential future non-carcinogenic hazards are associated primarily with the groundwater i ngestion exposure pathway and excess carcinogenic risks are associated primarily with the ingestion and inhalation pathways.

5.3.1 Ingestion of Groundwater - Southwest Alluvium

The non-cancer chem ical (i.e., non -radiological) hazard calculations are presented in Tables 7.1.RME and 7.2.RME for potential future adult a nd child residents, respectively. The child



resident is a sensitive subpopul ation and all COPC haza rd quotients for the child resident receptors exceed those for adult resident receptors.

The HI for ingestion of Southwest Alluvium groundwater by a future child resident is 12.9 and for a future adult resident is 5.5. For the child, the HQs for manganese (7.5), uranium (2.7), and cobalt (2.1) exceed one. For the ad ult, the HQs for manganese (3.2) and uranium (1.2) exceed one. For both potential future residents, m anganese accounts for m ore than half the hazard quotient associated with groundwater ingestion, and the sum of HQs for cobalt, m anganese, and uranium represent over 95 percent of the health hazard, although the cobalt HQ is less than one (0.92) for the adult receptor.

The chemical (i.e., non-radiological) and radiological cancer risk associated with the ingestion of groundwater from the Southwest Alluvium was evaluated for a combined RME child/adult exposure. The results of the evaluation are summarized in Table 7.3.RME (Appendix A). Arsenic and chloroform are the only non-radio nuclide, carcinogenic COPCs evaluated for the Southwest Alluvium. The results on Table 7.3.RME show that the chemical cancer risk is 5.9E-05, which is within the EPA acceptable risk range of 1E-04 to 1E-06. Arsenic accounts for more than 95 percent of the risk. The data used for the risk assessment show that average arsenic concentrations in impacted water (EPC UCL95, 0.00256 mg/L) are below the arsenic m aximum contaminant level (MCL, 0.010 m g/L) and are similar to average background water concentrations (UCL95, 0.00116 mg/L). The range of detected concentrations was the same for both impacted and background water and the freque ncy of non-detected re sults for arsenic was 86.5 percent in impacted water and 93.1 percent in background water.

The radiological cancer risk associated with the ingestion of groundwat er from the Southwest Alluvium is 1.5E-04, which exceeds the EPA accep table risk range of 1E -04 to 1E-06 (Table 8.1.RME in Appendix A). None of the individual radionuclides exceeds the EPA acceptable risk range. Three radionuclides have individual risks greater than 1E-05: uranium-234 (5.8E-05), uranium-238+D (7.0E-05), and radium 228+D (1.7E-05). The sum of the risks associated with these three radionuclides represents more than 96 percent of the total radionuclide risk.

The risk associated with radionuclides due to ingestion of backgr ound groundwater in the Southwest Alluvium is likely to be slightly lo wer (i.e., near the high end of the acceptab le risk range), based on the UCL95 concentrations of radionuclides reported in background groundwater (Appendix C).

5.3.2 Dermal Contact with Groundwater - Southwest Alluvium

The non-carcinogenic health effects of der mal contact with Southwest Alluvium groundwater through bathing was evaluated for potential future adult and child residents. The HI for derm al contact with groundwater by an adult is 0.43 and the HI for derm al contact with groundwater by a child is 1.3. Manganese contributes over 95 percent of the hazard for each receptor and the manganese HQ (1.2) exceeds one for the future child receptor.



The non-radiological cancer risk associated with the dermal contact with groundwater from the Southwest Alluvium was evaluated for a combin ed RME child/adult exposure. The results shown in Table 7.3.RME (Appendix A) indicate that the excess cancer risk is 4.7E-07, which is lower than the EPA target accep table range of 1E-04 to 1E-06. As described in Section 3.3.3, dermal absorption is typically not an im portant exposure route for radionuclides and it has not been evaluated for radionuclides within this risk assessment.

5.3.3 Inhalation of Groundwater - Southwest Alluvium

The non-carcinogenic health effects of inhalation exposure to groundwater from the Southwest Alluvium through bathing were evaluated for pot ential future adult a nd child residents. Chloroform is the only volatile COPC that was evaluated in this scenario. The results, which are provided in Table 7.1.RME and 7.2.RME in Appendix A, show that the chloroform HQ (and the pathway HI) for the groundwater inhalation exposure e pathway is below one for both an adult (0.0017) and a child (0.0041).

The non-radiological cancer risk associated with the inhalation exposure to groundwater from the Southwest Alluvium was evaluated for a combined RME child/adult exposure. The only volatile COPC evaluated in this scenario was chloroform and the results are provided in Table 7.3.RME (Appendix A). The cancer risk is 2.1E-06, which is within the EPA acceptable risk range of 1E-04 to 1E-06.

The radiological cancer risk associated with the inhalation exposure to groundwater through domestic tapwater use (bathing, dishwashing, laundry, etc.) fr om the Southwest Alluvium was evaluated for a combined RME child/adult exposure. Radium-226 is the only radiological COPC for which this exposure pathway is considered complete. The results presented in Table 8.1.RME (Appendix A) indicate th at the excess cancer risk for radium-226 is 2.9E-04, which exceeds the EPA acceptable risk range of 1E- 04 to 1E-06. However, the radiu m-226 EPC concentration in seepage-im pacted groundwater (UCL95, 0.267 pC i/L) is approximately three times lower than the background groundwater concentration in the Southwest Alluvium (UCL95, 0.798 pCi/L). Therefore, the risk due to inha lation exposure to seepage-im pacted Southwest Alluvium groundwater is within the range of background risk.

5.3.4 Hazard and Risk Summary - Southwest Alluvium

The total non-carcinogenic hazard and risks f or future resident exposure to seepage-im pacted groundwater in the Southwest Alluvium have been calculated. The HIs, based on RME exposure factors for future adult and child receptors and summed across all media and exposure pathways, are shown in Tables 9.1.RME and 9.2.RME (Appendix A). The total HI for an adult is 6.0 and for a child is 14.2, resulting m ostly from the ingestion exposure pathway and from the COPCs manganese, cobalt, and uranium.

Several segregated total HIs exceed one for target organs or toxicological effect. The HIs based on central nervous system effects are 8.7 for the child and 3.6 for the adult, due to the ingestion of manganese in groundwater. The segregated HIs for kidney toxi city are 2.7 for the child and



1.2 for the adult, due to uranium . The segregat ed HI for thyroid toxicity is 2.1 for the child, based on the cobalt con centration. Hazard indices for other specific organs or targ ets are less than one.

Manganese is the most significant contributor to the total hazard in the Southwest Alluvium. Manganese is a tra ce nutrient; the Adequate In take (i.e., for those nutrien ts that have not yet received enough scientific study to merit setting of an official Recommended Dietary Allowance, the AI represents an amount that appears to sustain good health) for manganese identified by the Institute of Medicine (2001) ranges from 1.2 mg/day for a child to 2.3 mg/day for an adult m ale. The RfD used in the HHRA for manganese was obtained from the EPA RSL table (EPA, 2010b) and was based on a modification of the IRIS RfD (0.14 mg/kg-day), which includes manganese from all sources, including diet. The RSL Table User's Guide (http://www. epa.gov/region9/superfund/prg/) indicates that the author of the IRIS assessment for manganese recommended that the dietary contribution from the normal United States diet (an upper lim it of 5 mg/day) be subtracted when evaluating non-fo od (e.g., drinking water or soil) exposures to manganese, leading to a RfD of 0.071 mg/kg-day for non-food ite ms. A m odifying factor of three is also applied when calculating risks asso ciated with non-food sources due to a num ber of uncertainties that are discussed in the IRIS file for manganese, leading to a RfD of 0.024 mg/kgday.

The non-carcinogenic hazard associated with potential future residential exposure to background groundwater in the Southwest Alluvium would be lower than the impacted water, but would also exceed one for both the adult and child receptors. The hazard for ingestion of cobalt in seepage-impacted groundwater is based on one detected result out of 96 sa mples. Cobalt is m ore frequently detected in background groundwater in the Southwest Alluvium, at a slightly higher concentration (UCL95, 0.0121 m g/L) than the m aximum detected concentration in seepage-impacted groundwater (0.01 m g/L) that was us ed as the EPC. Background m anganese and uranium concentrations would al so contribute to a background HI greater than 1 for both adult and child receptors.

The total radiological and chem ical carcinogenic risk, sum med across all m edia and exposure pathways, and based on RME exposure factors for the future child/adult receptor, is shown in Table 9.3.RME (Appendix A). The total risk for a combined adult/child receptor is 5.0E-04, resulting mostly from the ingestion and inhalation of radionuclide COPCs. The total radiological and chemical carcinogenic risk in background gr oundwater would also exceed the EPA target risk range, prim arily due to inhalation expose ure to radium -226, which has a background concentration (UCL95, 0.798 pCi/L) that is appr oximately three times the EPC (UCL95, 0.267 pCi/L) used for seepage-im pacted groundwater in the Southwest Alluvium . Radium -228 and uranium would also contribute to background risk.

The selection of COCs from the COPCs, for all hydrostratigraphic units, is discussed in Section 7.2.



5.4 Risk Characterization Results for Zone 1

Potential future residents may be exposed to COPCs in Zone 1 groundwater if they install a well in Section 1, east of the Section 2 boundary, and use seepage-impacted groundwater for drinking water and other dom estic uses. Calculated haza rds and risks associated with potential future exposure to seepage impacted groundwater in Zone 1 are summarized Tables 7.4.R ME through 7.6.RME, 8.2.RME, and 9.4.RME through 9.6.RME in Appendix A. For Zone 1, potential future non-carcinogenic hazards are asso ciated primarily with the ingestion pathway and excess carcinogenic risks are associated primarily with the ingestion pathways.

5.4.1 Ingestion of Groundwater - Zone 1

The non-cancer chemical (i.e., non-radiological) hazards are summarized in Tables 7.4.RME and 7.5.RME for potential future adult and child residents, respectively. The total HI f or groundwater ingestion by a future child resident is 19.9 and the HI for a fu ture adult resident is 8.5. For the child resident, the H Qs for cobalt (11.9), manganese (5.2), and vanadium (2.6) exceed one. For the adult resident, the HQs for cobalt (5.1), manganese (2.2), and vanadium (1.1) also exceed one. For both the adult and child receptors, cobalt accounts for more than half the hazard quotient, and the sum of HQs for the cobalt, manganese, and vanadium represent over 98 percent of the health hazard. Ho wever, because the Zon e 1 background manganese UCL95 concentration (2.519 mg/L) exceeds the manganese EPC concentration (UCL95, 1.95 m g/L) in seepage-impacted groundwater in Zone 1 (N .A. Water Systems, 2008b, 2008c), the background groundwater HI would also exceed one.

The chemical (i.e., non-radiological) and radiological cancer risk associated with the ingestion of groundwater from Zone 1 was evaluated for an RME combined child/adult exposure. Arsenic and chloroform are the only non-radionuclide, car cinogenic COPCs evaluated for Zone 1. The results provided in Table 7.6.RME (Appendix A) indicate that the chemical cancer risk is 3.3E - 05, which is within the EPA acceptable risk range of 1E-04 to 1E-06. Arsenic accounts for more than 99 percent of the calculated excess risk for r the ingestion exposure pathway. The arsenic EPC concentration for impacted water (UCL 95, 0.00145 mg/L) and the UCL95 for background water (0.00117 m g/L) are approxim ately equal and below the MCL (0.010 mg/L). The maximum result detected in Zone 1 background groundwater (0.003 m g/L) (N.A. Water Systems , 2008b, 2008c).

The radiological cancer risk associated with the ingestion of groundwater from Zone 1 is 5.3E-05, which is within the EPA acceptable risk range of 1E-04 to 1E-06 (Table 8.2.RME in Appendix A). The only radionuclide with an excess risk exceeding 1 E-05 is radium 228+D (4.1E-05). The average radium 228 activity (UCL95, 2.946 pCi/L) in Zone 1 background groundwater exceeds the radium-228 activity in seepage-impacted water (EPC UCL95, 2.046 pCi/L); therefore, the risks associated with ing estion of background water would exceed those for impacted water.



5.4.2 Dermal Contact with Groundwater - Zone 1

The non-carcinogenic health effects of dermal contact with Zone 1 groundwater through bathing was evaluated for potential future adult and child residen ts. The hazard index (HI) for der mal contact with groundwater for an adult is 0.32 and for a child is 0.95. Manganese contributes 90 percent of the hazard for each receptor. These results are below the EP A target level of one for this pathway which indicates that adverse health effects are unlikely to be due to chronic derm al exposure. Furthermore, as described in Sec tion 5.4.1, average Zone 1 background m anganese concentration exceeds the EPC ma nganese concentration in im pacted water; therefore non-carcinogenic hazard associated with the potential dermal exposure to manganese in background water exceeds that of the Zone 1 impacted water.

The non-radiological cancer risk associated with the dermal contact with groundwater from Zone 1 was evaluated for a com bined RME child/adult exposure. The results, provided in Table 7.6.RME in Appendix A, show that the calculated cancer risk is 2.1E-07, which is lower than EPA acceptable risk range of 1E-04 to 1E-06.

5.4.3 Inhalation of Groundwater - Zone 1

The non-carcinogenic health effects of inhala tion exposure to Zone 1 groundwater through bathing was evaluated f or potential future adult and child residents. The only volatile COPC evaluated in this scenario was chloroform. The results, provided in Table 7.4.RME and 7.5.RME (Appendix A), show that the HI for the inhalation exposure pathway for an adult is 0.0003 and a child is 0.0008. These results are well below the EPA target level of one, which indicates that adverse health effects are unlikely as a result of chronic exposure to chloroform under the Zone 1 groundwater inhalation exposure scenario.

The non-radiological cancer risk associated with the inhalation exposure to groundwater from Zone 1 was evaluated for a com bined RME child/adult exposure. As for the non-carcinogenic risk evaluation, the only volatile COPC evaluated in this scenario was chloroform. The results are provided in Table 7.6.RME (Appendix A). The cancer risks are 4.2E-07, which is below the EPA acceptable risk range of 1E-04 to 1E-06.

The radiological cancer risk associated with the inhalation exposure to Zone 1 groundwater through domestic tapwater use (e.g., bathing, di shwashing, laundry, etc.) was evaluated for a combined RME child/adult exposure. Radium-226 is the only radiological COPC for which this exposure pathway is considered com plete. The results are provided in Table 8.2.RME (Appendix A). The calculated cancer risk for ra dium-226 is 1.3E-03, which exceeds the EPA acceptable risk range of 1E-04 to 1E-06. However, the radium-226 EPC concentration in Zone 1 seepage-impacted groundwater (1.213 pCi/L, based on the UCL95) is lower than the Zone 1 background groundwater radium-226 UCL95 (1.314 pCi/L). Therefore, the radiological cancer risk for seepage-impacted water is less than that of background water.



5.4.4 Hazard and Risk Summary - Zone 1

The total non-carcinogenic hazard and risks f or future resident exposure to seepage-im pacted groundwater in Zone 1 have been calculated. The HIs, summed across all m edia and exposure pathways and based on RME exposure factors for future adult and child receptors, are shown in Tables 9.4.RME and 9.5.RME (Appendix A). The total HI for an adult is 8.9 and for a child is 20.9, resulting mostly from the in gestion exposure pathway and from the COPCs cobalt, manganese, and vanadium.

Several segregated total HIs exceed one for target organs or toxicological effect. The HIs based on thyroid effects are 11.9 for the child and 5.1 for the adult, du e mostly to the ingestion of cobalt in groundwater. The segreg ated HIs for the central nervous system are 6.1 for the child and 2.5 for the adult, due mostly to ingestion of manganese. The total HIs for the m etabolic system (as indicated by decreased h air cystine) are 2.6 for the child and 1.1 for the adult, due to the ingestion of vanadium. Hazard indices for other specific organs or targets are less than one.

The non-carcinogenic h azard associated with b ackground groundwater in Zone 1 would also exceed one for both the adult and child recep tors. Although cobalt concentrations are lower in background groundwater than in Zone 1 seepag e-impacted groundwater and vanadium is not detected in Zone 1 background groundwater, th e Zone 1 ba ckground manganese concentration (UCL95, 2.519 m g/L) exceeds th e manganese EPC concentration (UCL95, 1.95 m g/L) in seepage-impacted groundwater (N.A. Water Systems, 2008b, 2008c). Furthermore, certain other potential COPCs that could contribute to non-carcinogenic hazard (e.g., molybdenum, cadmium and lead) have been infrequently detected in background groundwater, but were not detected in Zone 1 seepage-im pacted groundwater during the selected HHRA monitoring period (N.A. Water Systems, 2008b, 2008c).

The total radiological and chem ical carcinogenic risk, sum med across all m edia and exposure pathways and based on RME exposure factors f or future adult and child receptors, is shown in Table 9.6.RME (Appendix A). Th e total risk for a combined adult/child receptor is 1.4E-03, resulting mostly from the inhalation of the radionuclide COPC radium-226. As was the case for Southwest Alluvium, total radiological and ch emical carcinogenic risk in Zone 1 background groundwater would also exceed the EPA target risk range, primarily because the Zone 1 background groundwater radium-226 concentration (UCL95, 1.314 pCi/L) exceeds the Zone 1 seepage-impacted groundwater radium-226 EPC (UCL95, 1.213 pCi/L). The selection of COCs from the COPCs, for all hydrostratigraphic units, is discussed in Section 7.2.

5.5 Risk Characterization Results for Zone 3

Potential future residents m ay be exposed to COPCs in Zone 3 groundw ater if the tailings seepage migrates beyond the northern Section 36 boundary and future residents were to install a well and use seepage-im pacted groundwater as a drinking water and dom estic source. Calculated hazards and risks asso ciated with potential future exposure to seepage e impacted groundwater in Zone 3 are summarized Tables 7.7.RME through 7.9.RME, 8.3.RME, and 9.7.RME through 9.9.RME (Appendix A). For Zone 3, potential future non-carcinogenic



hazards are associated p rimarily with the inge stion pathway and excess carcinogenic risks are associated primarily with the ingestion and inhalation pathways.

5.5.1 Ingestion of Groundwater - Zone 3

The non-cancer chemical (i.e., non-radiological) hazards are summarized in Tables 7.7.RME and 7.8.RME for potential future adult t and child residents, respectively. The HI for groundwater ingestion by a future child resident is 236 and the HI for a future adult resident is 101. For the child, the HQs for aluminum (2.5), a rsenic (87.8), cadmium (8.0), cobalt (93.5), m anganese (29.0), molybdenum (9.4), nickel (1.6), and vanadium (2.3) exceed the E PA target level of one. For the adult, the hazard quotients for aluminum (1.1), arsenic (37.6), cadmium (3.4), cobalt (40.1), manganese (12.4), molybdenum (4.0), and vanadium (1.0) equal or exceed one. For both potential future residents, the sum of the HQs for arsenic, cobalt, and m anganese accounts for almost 90 percent of the health hazard.

Two non-carcinogenic Zone 3 COPC s (molybdenum and uranium) are present in Zone 3 background groundwater at con centrations that exceed Zone 3 seepage-im pacted groundwater concentrations. The UCL95 molybdenum concentration in background water is 17.43 mg/L and the EPC UCL95 concentration in impacted water r is 0.739 m g/L. The UCL95 ur anium mass concentration in background water is 0.107 m g/L and the EPC (UCL95) in im pacted water is 0.0431mg/L. For com parison, the HI for inge stion of Zone 3 background groundwater was calculated using the background concentrations of the COPCs selected for the seepage-impacted-water hazard calculation (see supplemental Tables 7.F.RME and 7.G.RME in Appendix A). The background HI calculation does not include hazard as sociated with constituents that have been detected in background groundwater but are not present in seepage-impacted groundwater at sufficient concentration to be evaluated as COPCs in the HHRA (e.g., lead). The results showed that the HI for ingestion of the background groundwater was higher than that f or seepage-impacted groundwater for both a potential adult re ceptor (125) and a potential child receptor (292).

The risk associated with background groundwater ingestion is related primarily to molybdenum; individual HQs for background Zone 3 ground water that exceed one for a future child receptor r are arsenic (37.3), cadm ium (1.4), cobalt (18.7), m anganese (9.2), molybdenum (223), and uranium (2.3). The individual HQs for backgr ound Zone 3 groundwater that exceed one for an adult receptor are arsenic (16.0), cobalt (8.0), manganese (3.9), and molybdenum (95.5).

The chemical (i.e., non-radiological) and radiological cancer risk associated with the ingestion of Zone 3 groundwater was evaluated for a combined RME child/adult exposure. The results of the evaluation are summarized in Table 7.9.RME (Appendix A). Arsenic and chlorof orm are the only non-radionuclide, carcinogenic CO PCs evaluated for Zone 3. The results show that the chemical cancer risk is 9.2E-03, which exceeds the EPA acceptable risk range of 1E-04 to 1E-06. Arsenic accounts for almost 100 percent of the non -radionuclide cancer risk. Arsenic is als o present in Zone 3 background groundwater at a UCL95 concentration of 0.175 mg/L, which is approximately 42 percent of the Zone 3 seep age-impacted groundwater EPC (UCL95, 0.412



mg/L). The chemical cancer risk associated with ingestion of Z one 3 background groundwater for a combined RME child/adult exposure is 3 .9E-03, which also exceeds the EP A acceptable risk range of 1E-04 to 1E-06 (see supplemental Table 7.H.RME in Appendix A).

The radiological cancer risk asso ciated with the ingestion of seepage-im pacted Zone 3 groundwater for a combined RME child/adult is 5.3E-04, which exceeds the EPA target range of 1E-04 to 1E-06. The largest excess cancer risks are associated with the following individual radionuclides: radium -228+D (3.5E-04), uran ium-234 (2.0E-05), uranium -238+D (2.4E-05), radium 226+D (8.1E-05) and lead-210+D (5.5E-05) . The radiological can cer risk associated with the ingestion of Zone 3 background groundwater is 2.8E-04, which also exceeds the EPA target range of 1E-04 to 1E-06 (see supplemental Table 8.A.RME in Appendix A). The elevated risk in background is related primarily to uranium-234, uranium-238+D, radium-226+D, radium-228+D, and lead-210+D. The Zone 3 backgroun d uranium groundwater concentration (UCL95, 0.107 mg/L) is approximately 2.5 times higher than the uranium concentration in seepageimpacted groundwater (EPC UCL95, 0.0431 mg/L). The thorium -230 activity in background groundwater (UCL95, 1.426 pCi/L) also exceed s, by approximately 5.5 times, the thorium-230 EPC (UCL95, 0.259 pCi/L) in seepage-im pacted groundwater. Radium -226, radium-228 and lead-210 activities (U CL95) in background groundwater are 25 to 70 percent of the corresponding activities (EPC, UCL95) in seepage-impacted Zone 3 groundwater.

5.5.2 Dermal Contact with Groundwater - Zone 3

The non-carcinogenic health effects of dermal contact with Zone 3 groundwater used as tapwater for bathing were evaluated for potential future adult and child residents. The HQ for derm al contact with groundwater for an adult is 2.6 and the HQ for a child is 7.8. The individual HQs for dermal hazards associated with Zone 3 groundwater that exceed one for the potential child receptor are cadmium (1.1) and manganese (4.8). The only individual CO PC HQ that exceeds one for dermal hazards associated with Zone 3 groundwater for the potential adult receptor is manganese (1.6).

The non-radiological cancer risk associated with the dermal contact with groundwater from Zone 3 was evaluated for a com bined RME child/adult e xposure. The results are provided in Table 7.3.RME in Appendix A. The cancer risks are 5.3E-05, which is with in than EPA acceptable risk range of 1E-04 to 1E-06.

5.5.3 Inhalation of Groundwater - Zone 3

The non-carcinogenic health effects of inhala tion exposure to Zone 3 groundwater through bathing was evaluated f or potential future adult and child residents. The only volatile COPC evaluated in this scenario was chloroform. The results provided in Table 7.7.RME and 7.8.RME in Appendix A show that the HI for the inhala tion exposure pathway for an adult is 0.0016 and a child is 0.0040. These r esults are well below the EPA target level of one which indicates that adverse health effects are unlikely follo wing a chronic exposure under the groundwater inhalation exposure scenario.



The non-radiological cancer risk associated with the inhalation exposure to groundwater from Zone 3 was evaluated for a com bined RME child/adult exposure. As for the non-carcinogenic risk evaluation, the only volatile COPC evaluated in this scenario was chloroform. The results are provided in Table 7.9.RME (Appendix A). The cancer risks are 2.0E-06, which is within the EPA acceptable risk range of 1E-04 to 1E-06.

The radiological cancer risk associated with the inhalation exposure to Zone 3 groundwater through domestic tapwater use (e.g., bathing, di shwashing, laundry, etc.) was evaluated for a combined RME child/adult exposure. Radium-226 is the only radiological COPC for which this exposure pathway is complete. The results are provided in Tabl e 8.3.RME (Appendix A). The cancer risks for radium-226 are 1.2E-02, which exceeds the EPA acceptable risk range of 1E-04 to 1E-06.

However, Table 8.A.RME (Appendix A) indicates that the risk associated with radium -226 inhalation of Zone 3 background groundwater is 5.5E-03, which also exceed s the EPA acceptable risk rang e. The elevated backgrou nd risk is the resu lt of a Zone 3 background radium-226 activity (UCL95) of 4.996 pCi/L.

5.5.4 Hazard and Risk Summary - Zone 3

The total non-carcinogenic hazard and risks for future reside ntial exposure to seepage-impacted groundwater in the Zone 3 have been calcul ated. The HIs summ ed across all m edia and exposure pathways, and based on RME exposure fact ors for future adult and child receptors, are shown in Tables 9.7.RME and 9.8.RME (Appendix A). The total HI for an adult is 104 and for a child is 244, resulting mostly from the ingestion exposure pathway and from the COPCs cobalt, arsenic, and manganese; however, the HIs for several individual COPCs exceed one. As can be determined from the description of the individual pathways in the preceding sections, the sum of non-carcinogenic hazard associated with back ground groundwater in Z one 3 across all m edia would also significantly exceed one for both the adult and child receptors.

Several segregated total HIs exceed or equal one for target organs or toxico logical effect, as follows:

- The segregated HIs based on thyr oid effects are 94.2 for the child and 40.3 for the adult, due mostly to the ingestion of cobalt in groundwater.
- The segregated HIs for skin to xicity are 88.4 for the child and 37.8 for the adult, due primarily to arsenic ingestion.
- The segregated HIs for central nerv ous system effects are 36.3 for the child and 15.1 for the adult due primarily to manganese ingestion.
- The segregated HIs for kidney effects are 19.5 for the child and 8.3 for the adult, due to molybdenum, cadmium and uranium. As described above, molybdenum and uranium are present in Zone 3 background groundwater at concentrations that exceed Zone 3 seepage-impacted groundwater concentrations.



- The segregated HIs for the m etabolic system (as indicated by decreased hair cystine) are 2.3 for the child and 1.0 for the adult, due to the ingestion of vanadium.
- The segregated HI for reduced body and organ weights is 1.6 for the child receptor due to nickel.

Segregated hazard indices for liver and gastrointestinal effects for both the adult and the child are less than one. Additionally, the hazard index for reduced b ody and organ weights is less than one for the adult receptor.

The total radiological and chem ical carcinogenic risk, sum med across all m edia and exposure pathways, and based on RME exposure factors for future adult and child receptors are shown in Table 9.9.RME (Appendix A). The total risk fo r a combined adult/child receptor is 2.2E-02 resulting mostly from the inhalation of the radionuclide COPC ra dium-226. However, the risk associated with ingestion of arsenic and radium-228 also individually exceed the EPA target risk range. As was the case for the Sout hwest Alluvium and Zone 1, total radiological and chem ical carcinogenic risk in Zone 3 b ackground groundwater would also exceed the EPA target risk range, primarily due to inhalati on exposure to radium -226. The selection of COCs from the COPCs, for all hydrostratigraphic units, is discussed in Section 7.2.



Section 6 Uncertainty Analysis

This section describes the uncertainties identified in the Church Rock Site HHRA. Uncerta inty may relate to the variability of the available data or the variability in an estimate of a missing value for a param eter of interest (i.e., an EPC or exposure factor). This uncertainty analysis comprises a mostly qualitative assessment of a ssumptions used in the risk assessment and a discussion of whether the quantitative risk as sessment process m ay have overestim ated or underestimated the hazard and risk levels. T he analysis presented in this section describes uncertainties related to the following:

- Exposure point concentrations and data;
- Exposure assessment;
- Toxicity assessment; and
- Risk characterization.

6.1 Exposure Point Concentrations and Data Uncertainties

The risk assessment process relies on the estim ation of UCL95 EPCs in the exposure m edia to estimate risks and hazards. There is som e bias associated with using an EPC to represen t a receptor's exposure. There is also uncertainty associated with the sampled medium, including the numbers and locations of the sampling points (i.e., sampling bias), and possible measurement errors related to the sample collection and analysis. Our understanding of risk should include the context that allowable concentrations of constituents in public drinking water supplies (MCLs) sometimes exceed the risk range that EPA says must be used for baselin e risk assessments (e.g., for arsenic).

The use of tailings-impacted groundwater concentrations representative of current conditions (e.g., current concentrations of impacted water in Section 36) is considered to be conservative for evaluation of future exposure scen arios, because future concentrations are anticipated to be attenuated or reduced by the natural processes (primarily adsorption and precipitation) that are known to occur in impacted groundwater at the Site. Furthermore, because it was necessary t o include data from seepage-impacted wells in Section 2 for the Southwest Alluvium , the associated risks will be overestimated.

Regarding sampling bias, the assumption that a pot ential future receptor would be exposed only to impacted groundwater (in the is case, tailings-impacted groundwater) would typically be considered conservative and likely to result in an overestimate of risk and hazard. In the case of the Church Rock Site, this uncertainty is not so important because the risk and hazard associated with background groundwater is similar to that of impacted water in many locations (see Section 5).



The uncertainty related with accurate sam ple collection and analysis of COPC concentrations is likely to be relatively low and could result in an overestimate or underestimate of the hazard and risk. However, the statistical process used to calculate the EPC (typically UCL95) is intended to minimize the chance that the average concentration is underestimated; therefore, it is likely that the result of the EPC estimation process results in an overestimate of risk and hazard. Additionally, the risk and/or hazard estimates are likely overestimated for each of the hydrostratigraphic units because certain monitoring parameters retained as COPCs were detected at very low frequency (i.e., le ss than 10 percent) in imacted groundwater. These COPCs include: cobalt in the Southwest Alluvium; vanadium in Zones 1 and 3 (vanadium has been detected only once in Zone 1 impacted water); and lead-210 and thorium-230 in Zone 3.

There is also some uncertainty related to missing data. The individual isotopes of uranium are of interest for radionuclide cancer risk calculations, but historically have not been part of the performance monitoring program. Therefore, a simplifying assumption was made that uranium isotope concentrations in groundw ater are proportional to their natural abundance. Using this assumption improved the radiological cancer risk estimate relative to excluding the isotopes; however, because isotope activities in grou ndwater may not be pr oportional to natural concentrations, the calculated ra diological risk due to uranium could be underestim ated. For example, Rhodes et al. (2006) indicate that th e activity ratio of uranium -234 to uranium-238 in groundwater increases with time due to the alpha particle recoil effect. A ratio higher than one-to-one would result in a lower calculated risk because the water cancer slope factor used for uranium-238+D is approxim ately 1.2 tim es higher than slope factor for uranium-234 (note however, that the U-234 slope factor is approximately 1.1 times the water cancer slope factor for U-238 without decay products).

There is low uncertainty associated with the missing data related to the elimination of monitoring parameters that were no longer considered relevant to the rem edy implementation (i.e., a set of trace metals plus iron). These p arameters were not included in the ri sk and hazard estim ates because they were no t analytes during the eight-quarter period (July 20 06 through April 2008 inclusive) of perfor mance monitoring data select ed as representative of current conditions. These missing data are likely to represent a sm all underestimation of the hazard and risk; inclusion of a larger numb er of contaminants in the risk assessment is not likely to affect site decisions if the add itional contaminants do not contribute significantly to the total risk. These parameters were eliminated from the monitoring program with EPA and NRC concurrence.

6.2 Exposure Assessment Uncertainties

Uncertainties associated with the exposure assessment for the Church Rock Site HHRA include (1) land use assumptions, (2) exposure factors, and (3) exposure models. The land use assumptions represent the most significant exposure assessment uncertainty, b ecause the assumptions made about future exposure scenarios may result in either an overestimation or an underestimation of hazards and risk s. In this case, the principal land use assumptions were as follows:



- Future residents would live adjacent to the Site and use seepage-im pacted groundwater from one of the three hydrostratigraphic unit for drinking water.
- Other than the future use of i mpacted groundwater by future residents, land use would remain generally unchanged.

The likelihood that future res idents will live a djacent to the Site and use impacted water for domestic use is believed to be very low; therefore the calculated risks and hazards represent an overestimation. The background water quality in the hydrostratigraphic units of interest in the vicinity of the Site is poor due to high concentrations of sulfate, chloride, TDS, metals, and radionuclides, and is not considered suitable for use as a drinking water source. Furtherm ore, there is unlikely to be sufficient saturated thickness available in Zone 1 for use as a potable water source. In the ROD, EPA stated that "the physical characteristics of Zone 1 are such th at sufficient quantities of water could not be pu mped from the sandst one to support volum es required for domestic or livestock purposes" a nd "Zone 1 would not be a good candidate for locating a domestic or livestock well even if there were no impacts from tailings seepage" (EPA, 1988c).

The assumption that future land use would rem ain generally unchanged is believed to be realistic; however, should future land use include exposures additional to the residential exposure scenario, the current risk and hazard estim ates could be underestimated. Regarding exposure factors, an RME exposure was evaluated for the Church Rock Site HHRA and EPA default values were used for param eters such as grou ndwater ingestion rates, exposure duration, and event frequency. Because the RME represents the highest exposure that is reason ably expected to occur at a site (EPA, 1989), the resulting risk estim ate is conservative (i.e., well above the average case and tends to overestim ate exposure). However, it is possible that some of these exposure factors represent underestimates when considering the local population and, therefore, result in an underestimate of risk and hazard:

- A 30-year exposure duration m ay be low with respect to a loca 1 Navajo resident population because the Navajo resident population may be more likely to rem ain in one area than the general population (although it is also likely that there would be insufficient impacted water to use as a domestic supply for 30 years in Zones 1 and 3).
- A drinking water ingestion rate of two liters per day may be low with r espect to a local population residing in a semi-arid environment.
- The assumption of a 350-day exposure freque ncy could be slightly low for the local population, but is bounded at 365 days, so the minimal.

Conversely, it is possible that exposure factor s related to the groundwater inhalation scenario represent overestimates when considering the local population and, therefore, result in an overestimate of risk and hazard. For exam ple, 2000 census infor mation posted on the Navajo Churchrock and Pinedale Chapters indicates that approximately 39 pe rcent of the Chapter



residents lack indoor plum bing and 38 percent l ack kitchen facilities. Neither of these circumstances eliminate the potential exposure to groundwater contam inants through the inhalation pathway, however, the use of the Fo ster and Chrostowski shower m odel and the Andelman volatilization factor may overestimate potential risk for these potential receptors.

With respect to exposure m odels, there are two factors that contribute to uncertainty associate d with the model utilized to calculate the exposure point concentration for radium -226 under the groundwater inhalation exposure pathway and to calculate the corresponding risk. This exposure pathway represents a significant percentage of the total risk calculated for each of the three hydrostratigraphic units and the model utilizes protective a ssumptions where measured radium concentrations are not available. However, the assumption that, in a domestic water supply, radium-226 becomes sufficiently airborne and i nhaled to justify the use of the radium -226+D cancer slope factor instead of the slope f actor for its gaseous decay product radon-222+D, appears to be conservative. The radium -226 cancer slope factor is approxim ately 650 times higher than the radium-226+D cancer slope factor, and consequently, the calculated risks would be proportionally lower. Secondly, the Andelman (1990) volatilization factor, which is used by EPA in the development of its radionuclide PRGs, appears very conservative for the non-volatile radium-226 because it assumes that half of the contaminant is transferred from the water through all domestic uses. At least on e other source suggests that this tr ansfer factor is too high for radon. In a United States Geological Survey Water-Resources Investigation Report, Lindsey and Ator (1996) indicate that well water radon concentrations equal to 10,000 pCi/L typically release approximately 1 pCi/L of radon to the air; this equa ls a transfer factor that is 20 percent of the Andelman factor.

The uncertainty associated with the use of the Foster and Chrostowski Model f or inhalation exposure to chloroform is low because the chloroform concentrations are low when compared to other COPCs, resulting in lower risk and hazard. Sim ilarly, the hazard and risk through the dermal exposure pathway for any of the hydrostra tigraphic units represents a small proportion of the ingestion pathway (up to approximately 10 percent in the Southwest Alluvium).

6.3 Toxicity Assessment Uncertainties

The principal COPCs in groundwater at the Church Rock Site include radionuclides and m etals including manganese, uranium, arsenic, coba lt, and m olybdenum. The toxicity assessment portion of the Church Rock Site HHRA relied on existing toxicity information available from EPA and other sources, including information regarding carcinogenic and non-carcinogenic effects associated with radionuclide and non-radionuclide COPCs. The toxicity information is presented on Tables 5.1, 5.2, 6.1, and 6.2 (Appendix A).

The uncertainty as sociated with to xicity information varies depending on the COPC. EPA methodologies for both cancer and non-cancer toxicity evaluation (i.e., development of RfDs and cancer slope factors) are in tentionally designed to be protective. In many cases, data are extrapolated from animals to sensitive humans by the application of uncertainty factors to an estimated No Observed Adverse Effect Levels (NOAELs) or Lowest Observed Adverse Effect



Levels (LOAELs) for non-cancer effects. It is likely in many cases that uncertainty factors overestimate the magnitude of differences that may exist between hu man and anim als, and among humans. However, the extent to which toxi city values may overestimate toxic potency is not clear, and it is possi ble that the toxicity values for some compounds may not be adequately protective (e.g., where multiple COPCs are present or when the toxic ity studies did not detect a sensitive adverse effect).

In addition, the derivation of cancer slope factors often involves linear extrapolation of effects at high doses to potential effects at lower doses commonly seen in environmental exposure settings. It is likely that the assumption of linearity is conservative and yields slope factors that are unlikely to lead to underestim ation of risks for most carcinogens. It is also possible that the dose-response curves for individual carcinogens vary from the linear extrapolation in a way that leads to an underestimate of risk.

There is generally less uncertainty associated with carcinogenic risk from radionuclides than with non-radionuclides because of the m ethod in which the tox icity numbers are typ ically developed. For exposure to chemicals, slope factors typically represent an upper bound estimate or 95th percent confidence limit value that has been obtained from extrapolation from laboratory experiments (EPA, 1989). Cancer slope factors for radionuclides are central tendency estimates (i.e., median or 50th percentile values) of the age-averaged increased lifetime cancer risk that are based on epidemiological studies of radiogenic cancers in humans.

There is significant uncertainty associated with the calculation of absorbed RfDs for the derm al exposure pathway, primarily in that the EPA guidance (2004) recommends the use of a 100 percent ABS_{GI} value for inorganics that have not otherwise been determined. This assumption may contribute to an underestim ation of hazard or risk for those inorganics that are actually poorly absorbed in the gastroin testinal tract. As the ABS _{GI} value decreases, the contribution of the dermal pathway to overall risk increases, relative to the ingestion pathway.

Some of the COPCs evaluated in the HHRA, including manganese, uranium, and vanadium have special considerations associated with the toxicity values cited on the RSL table that m ay affect the uncertainty of the risk asses sment. Manganese is a trace nutrient that has both a physiologically-required intake level and an intake level that is considered toxic. The "Adequate Intake" for manganese established by the National Research Council ranges from 1.2 mg/day for a child to 2.3 m g/day for an adult m ale (Institute of Medicine, 2001). The m anganese RfD obtained from the EPA RSL table (0.024 mg/kg-d) incorporates an adjustment to the manganese RfD of 0.14 m g/kg-d listed on IRIS http://www.epa.gov/iris/subst/0373.htm. The RSL Table User's Guide (http://www.epa.gov/region9/superfund/prg/) indicates that the modifications were in accordance with a recommendation by the author of the IRIS assessment to subtract a normal United States dietary contribution (an upper limit of 5 mg/day) when evaluating non-food (e.g., drinking water or soil) exposures . The RSL Table User's Guide indicates that IRIS further recommends using a m odifying factor of 3 wh en calculating risks as sociated with non-food sources, leading to the RfD of 0.024 mg/kg-day. The results of the risk calculations indicate that manganese is the most significant contributor to the total hazard in the Southwest Alluvium and



also contributes significant hazard in Zones 1 and 3. These hazards may be overestimated due to the uncertainty, or conservatism , in the ad justed toxicity value for non-food m anganese exposure.

The RfD us ed for uranium non-carcinogenic hazard was for "uranium - soluble salts," which appears appropriate for a drinking water exposure in which the potential receptor ingests salts dissolved in the drinking water. However, the use of this RfD may ove restimate the calculated hazards slightly. IRIS references a 1949 study by Maynard and Hodge which ind icates that soluble uranium salts are more toxic than the insoluble uranium salts (IRIS, <u>http://www.epa.gov/iris/subst/0421.htm</u>). Groundwater m onitoring concentrations are reported for total uranium at the Church R ock Site (i.e., based on an unfiltered sample), rather than dissolved uranium (i.e., based on a filtered sample). Therefore, the reported concentrations could include insoluble uranium as well as soluble u ranium. The overall affect of this uncertainty would be expected to be low. Non-carcinog enic uranium hazard was significant only in the Southwest Alluvium.

Vanadium HI values are above one for ingestion of impacted groundwater from Zones 1 and 3. EPA derived the oral R fD toxicity value for vanadium in the RSL Table from the IRIS RfD for vanadium pentoxide by factoring out the molecular weight of the oxide ion. It is not known whether this adjustment imparts any uncertainty into the risk calculations.

Toxicity values are not available for all COPCs. Therefore, health risks and hazard s cannot be quantitatively assessed for all contam inants and the total risk or hazard for the site may be underestimated in such circum stances. At the Church Rock Site, m onitoring COPCs representing common ions (i.e., su lfate, chloride, nitrate [m easured as nitrate-n itrogen]) were eliminated from further consideration in the quantitative risk assessment calculations because with the exception of a nitrate RfD for for mula ingestion by infants (0 to 3 m onths old), they do not have associated toxicity values. The nitrat e RfD is 1.6 mg/kg-d for early clinical signs of methemoglobinemia for infants of age 0 to 3 months old. The exclusion of these param eters results in some uncertainty in the risk assessment as described below.

UCL95 chloride concentrations in seepage- impacted water at the Site exceed UCL95 background concentrations in each of the hydr ostratigraphic units. However, the UCL95 chloride concentrations in seepage impacted water are below the secondary m aximum contaminant level (SMCL, 250 m g/l) in each unit, and the maximum detected concentrations exceed the SMCL only in the Southwest Allu vium seepage-impacted water (374 mg/L, see (Appendix B). SMCLs are established by EPA only as guidelines to assist public water systems in managing their drinking water for aesthetic considerations (e .g., taste, co lor and odor) and EPA does not consider these contaminants to present a risk to hum an health at the S MCL. The chloride SMCL is established for a salty taste associated with chloride. Because the UCL95 concentrations in seepage impacted water are similar to or below the SMCL, chloride has little or no effect on uncertainty associated with the risk assessment.



Sulfate is a component of the tailings seepage fluid, and it is present in impacted water in the three hydrostratigraphic units with UCL95 values ranging from approximately 2,800 to 4,000 mg/L. However, sulfate is also presen t in the background gr oundwater and sulfate concentrations in the three hydr ostratigraphic units are fixed by geochemical equilibrium with gypsum (or anhydrite) and calcite and are therefore unchangeable. Sulfate has an EPA SMCL of 250 mg/L based on a bitter, salty, or m edicinal taste and higher concen trations have been associated with laxative effects. EPA proposed in 1994, but did not finalize, an MCL for sulfate of 500 mg/L based on the potential laxative or diar rheal health effects of sulfate in drinking water (59 Fed. Reg. 65578-65604). The New Me xico Water Quality Control Commission domestic water-supply standard for sulfate is 600 m g/L (New Mexico Adm inistrative Code 20.6.2.3103). However, there is conflicting info rmation regarding these potential effects: ATSDR (2010) cites two studies the at have statements regarding the "laxative effect" with drinking water, wherein one study concluded that a laxativ e effect resulted from water containing 1 g sulfate/L, while a second study, using drinking water containing 1.2 g/L, found no "laxative effect." ATS DR (2010) also consider ed the effects of inge sting groundwater sulfate concentrations in Wyoming that are similar to those at the Church Rock Site. They state that drinking water from a well with sulfate concentrations of $3,640,000 \ \mu g/L$ ($3,640 \ m g/L$) would result in daily intakes of 7.3 g/day for an adul t and 3.6 g/day for a s mall child and that both of these daily intake rate s would ex ceed the in take rate asso ciated with the gastrointestinal discomfort (2 g/day). Therefore, it is likely the at a potential future receptor would become ill from ingestion of the se epage-impacted water (or background water) from the vicinity of the Church Rock Site. However, if a potential receptor is cap able of ingesting water with a sulfate concentration that is ten times the taste-based SMCL, the effect on the hazard and risk estimates is almost certainly the opposite: due to the acute gast rointestinal effects, water with sim ilar elevated concentrations of sulfate would not be drinkable for su fficient time to experience the ingestion-related chronic hazard and risk effects of the COPCs present in the water. In this situation, the calculated hazard and risk estimates would likely be substantial overestimates and actual risks would be related only to the inhalation and dermal exposures.

Nitrate has a Federal MCL of 10 mg/L (as nitrate-nitrogen). Infants below the age of six months who drink water containing nitrate in excess of the MCL are at particular risk of nitrate tox icity, which causes methemoglobinemia and can be fatal. According to the Ohio EPA (2005), healthy adults and older children can consume higher levels of nitrate than infan ts because of their fully developed digestive system s. Ohio EPA recommends that (1) wom en who are pregnant or nursing consult with their physicians about limiting nitrate consumption, and (2) people with medical conditions that m ay make them more susceptible to methemoglobinemia, such as reduced stomach acidity, should also consult their physicians. The Idaho Departm ent of Environmental Quality (2001) summarized research on nitrate effects on human health, stating that studies have im plicated nitrate exposure as a possible risk factor associated with non-Hodgkin's lymphoma, gastric cancer, hypertension, thyroid disorder and birth defects, but that it is difficult to dem onstrate a link between nitrate in drinking water and can cer or birth defects. This is due, in part, to the widely variable exposure to nitrates from other sources. The National



Research Council (1995) stated that nitrate and nitrite have been tested for carcinogenicity in laboratory animals, and epidemiologic studies of human cancer rates am ong populations with high nitrate or nitrite exposure c oncentrations have been performed. The results have generally indicated that nitrate and nitrite are not carcinogenic in laboratory animals when administered in the absence of nitros atable amines; but when n itrite and nitros atable amines are administered together, carcinogenic nitrosamines can be formed in the stomach and lead to various tumors.

HEAST cancer slope factors for radionuclides we re obtained from the EPA Radionuclide PRGs website (http://epa-prgs.ornl.gov/radionuclides/). Certain radionuclide COPCs (U-235, U-238, Ra-226, Ra-228 and Pb-210) are designated with the suffix "+D" (e.g., U-238+D, Ra-226+D) to indicate that cancer risk estimates for these radionuclides in clude the contributions from their short-lived decay products, assuming equal activit y concentrations (i. e., secular equilibrium) with the principal or parent nuclide in the environm ent. The PRG Users Guide indicates that using the "+D" design ation can be important because some decay products can be m ore toxic than the parent isotop e. EPA states that in the absence of site-specific data regarding secular equilibrium, the "+D" values for radionuclides s hould be used; therefore, for the Church Rock Site HHRA, the cancer slope factors used for r the following radionuclides had the "+D" designation: uranium -235+D, uranium-238+D, radium-226+D, radium-228+D, and lead-210+D(see Tables 6.1 and 6.2 in Appendix A).

There have been no site -specific determinations of secular equilibrium. Uncertainty associated with the use of these +D designa tions varies with COPC. For both of the radium isotopes, the difference in slope factor between the principal nuclide and the +D value is less than one percent. However, for uranium-235, the difference is approximately three percent; for uranium -238, the difference is 36 percent; and for lead-210, the difference is 44 percent. Therefore, the uncertainty related to the assumption of secular equilibrium is greater for uranium than radium and is in the range of three to 44 percent.

6.4 Risk Characterization Uncertainties

The risk characterization step of the HHRA pr ocess summarizes and com bines outputs of the exposure and toxicity assessments to characterize risk in quantitative and qualitative statements. There is uncertainty in this process related to both non-carcinogenic and carcinogenic risk. For the non-carcinogenic hazard, there is uncertainty re lated to the significance of HQs greater than one and the process of summing individual HQs f or multiple COPCs and ac ross multiple exposure pathways. Because there is conservatis m built into most toxicity numbers (i.e., RfDs) through uncertainty factors and m odifying factors (e.g., the range of com bined uncertainty and modifying factors shown in Table 5.1 is 3 to 1000), there is inherent conservatism in the HQ value calculated by dividing the estimated intake (which in turn may be conservative due to the use of conservative exposure factors) by the toxicity number.

Furthermore, EPA (1989) indicates that the pote ntial for an adverse health effect does not necessarily increase linearly as an RfD is appro ached (i.e., an HQ of 1) or exceeded because RfDs do not have equal accuracy or precision and are not based on the same severity of toxic



effects. This presents uncertainty w hen calculating the HI, which is the sum of the individual COPC HQs evaluated within an exposure scenar io or across exposure scenarios. HIs are calculated in this HHRA first using the assumption that all the hazards are additive, and secondly by toxic effect or targ et organ. These HI calcu lations are important uncertainties in the r isk characterization. The assumption that the risks are additive likely (but not always) would result in an overestimation of the hazard. It is unknow n whether COPC interact ions are synergistic, antagonistic, or additive or whether the severity of effects used to develop the RfDs are comparable. These uncertainties may be re duced by sum ming the COPCs by toxic effect or target organ; however, the hazard is likely still overestimated due to conservatism in the RfD development and RfDs that are based on toxic e ffects that are difficult to categorize (e.g., the reported effect for vanadium is "decreased hair cystine").

Another major source of uncertainty in the Church Rock HHRA, as well as for future risk management decisions, is the risk as sociated with background COPC concentrations. There are three principal situations under which unacceptable background hazard or risk is encountered at the Church Rock Site:

- 1. Background COPC concentrations higher than s eepage-impacted waters In this case, a hazard or risk driver is present at higher concentrations in background water than in seepage-impacted waters. An ex ample of this is the groundwater inhalation exposure pathway for the Southwest Alluvium, where radium-226 concentrations represent more than half the total risk, a nd where radium-226 concentrations in background water are three times higher than in seepage impacted water.
- 2. Background COPC concentrations lower than s eepage-impacted waters In this case, a hazard or risk driver is present at lower concentrations in background water than in seepage-impacted waters, but the risk associated with background water exceeds the EPA acceptable range of 1E-04 to 1E-06. An exam ple of this is the groun dwater ingestion exposure pathway for Zone 3, where arseni c concentrations in im pacted water are approximately 2.4 times the concentration in background water, but the non-radiological carcinogenic risk associated with both background and seepage-impacted water concentrations exceed the EPA acceptable range.
- 3. Different COPC distributions in seepage-im pacted and background waters In this case, the COPC distribution in background and seep age-impacted waters are different, but the risk or hazard associated with both waters exceed the applicable EPA acceptab le range. An example of this is the non-carcinogenic groundwater ingestion exposure pathway for Zone 3, where the HI associated with seepage-impacted water for a child is estimated to be 236, due primarily to cobalt (93.5), arsenic (87.8), and manganese (29.0); whereas the background water ingestion exposure HI for a child is estim ated to be 292 and is primarily due to molybdenum (222), arsenic (37.3), and cobalt (18.7).

Clearly, some of the hazards and risks that ar e calculated for the seepage-im pacted water are overestimates because background hazards or risks either ex ceed or represent a large portion of



the calculated hazard o r risk. The variab ility of the back ground conditions results in som e uncertainty in describing and quantifying the incremental risk due to seepage im pacts, and will ultimately affect risk management decisions. For example, EPA guidance (200 2) describes situations where COPCs with concentrations below natural or anthropogenic background concentrations are excluded from further consideration in the establishment of remediation goals. However, the application of this m ethod to select individual COCs for each o f the th ree hydrostratigraphic units may appear somewhat subjective when considering the background risks as a whole.



Section 7 Risk Assessment Summary

The objective of a hum an health risk assessment is to evaluate the likelihood of adverse effects occurring in human populations potentially exposed to contaminants released in the environment. This section summarizes the results of the Church Rock Site HHRA, including the risk calculations, the evaluation of uncertainty and the selection of chemicals of concern (COCs). The site-specific objectives for updating the Church Rock Site HHRA are the following:

- 1. Update the risk estimates for the Site using current risk assessm ent methods and information;
- 2. Support the reassessment of remediation levels;
- 3. Provide a basis for comparing remedial alternatives; and
- 4. Identify Point of Com pliance (POC) and Po int of Exposure (POE) concentrations in accordance with NRC requirements.

The following sections describe how the results of the HHRA have met, or can be used to meet, these objectives.

7.1 Hazard and Risk Summary

There is no current human exposure to groundwat er at the S ite (EPA, 2008) except during the quarterly groundwater sampling conducted by UNC personnel, and no potential future exposure to groundwater contaminants on UNC-owned property, because no groundwater supply wells drawing on any of the three hydros tratigraphic units will be allowed on UNC property, and the same restriction will apply once the license is transferred to the DOE for long-term surveillance monitoring.

Therefore, the focus of the Church Rock Site HHRA was the potential future exposure to seepage-impacted groundwater contaminants in one of the three hydrost ratigraphic units (i. e., Southwest Alluvium, Zone 1, and Zone 3) at locations outside Section 2. A residential tapwater (i.e., groundwater) exposure scenario was selected. Because the hyd rogeologic characteristics, contaminants, and remedial alternatives for each of the un its are distinct, the risks of potential future exposure to groundwater at the following locations have been evaluated separately:

- Southwest Alluvium a hypothetical future well located adjacent to the UNC property boundary in Section 3;
- Zone 1 a hypothetical future well located adjacent to the UNC property boundary in Section 1; and
- Zone 3 a hypothetical future well located to the north o f and adjacent to the UNC property boundary in Section 36.



The residential tapwater exposur e scenario assumes that residents would construct residences and live adjacent to the UNC property bound ary near the tailings impoundments for up to 30 years, and that residents would use s eepage-impacted groundwater for all domestic water needs. To assess the potential exposure of a hypothetical future resident, three exposure pathways were selected for evaluation:

- Ingestion of groundwater as the drinking water source;
- Direct dermal contact with groundwater through bathing; and
- Inhalation of volatile compounds in groundwater through showering exposure and, for radionuclides, through other domestic tapwater uses.

The non-carcinogenic hazard and carcinogenic risk to potential RME receptors were evaluated for each of the exposure pathways. A future re sident adult and a future resident young child (aged 1 to 6) were selected as the potentially exposed populations for non-carcinogenic COPCs. A future resident combined child/adult recept or was selected as the potentially exposed population for the carcinogenic COPCs, including radionuclides.

The results of the risk assessment calculations are provided in the R AGS Part D tables in Appendix A and are summarized for the child and combined child/adult receptors as follows:

Hydrostratigraphic Unit	Exposure pathway	Total Non- carcinogenic Hazard Index (Child)	Chemical Carcinogenic Risk (Child/Adult)	Radionuclide Carcinogenic Risk (Child/Adult)	Total Carcinogenic Risk (Child/Adult)
Southwest Alluvium	Ingestion	12.9	5.9E-05	1.5E-04	2.1E-04
Southwest Alluvium	Dermal	1.3	4.7E-07	N/A	4.7E-07
Southwest Alluvium	Inhalation	0.0041	2.1E-06	2.9E-04	2.9E-04
Southwest Alluvium	Total	14.2	6.2E-05	4.4E-04 5.0	E-04
Zone 1	Ingestion	19.9	3.3E-05	5.3E-05	8.6E-05
Zone 1	Dermal	0.95	2.1E-07	N/A	2.1E-07
Zone 1	Inhalation	0.0008	4.2E-07	1.3E-03	1.3E-03
Zone 1	Total	20.9	3.4E-05	1.4E-03 1.4	E-03
Zone 3	Ingestion	236	9.2E-03	5.1E-04 9.7	E-03
Zone 3	Dermal	7.8	5.3E-05	N/A	5.3E-05
Zone 3	Inhalation	0.0040	2.0E-06	1.2E-02	1.2E-02



Hydrostratigraphic Unit	Exposure pathway	Total Non- carcinogenic Hazard Index (Child)	Chemical Carcinogenic Risk (Child/Adult)	Radionuclide Carcinogenic Risk (Child/Adult)	Total Carcinogenic Risk (Child/Adult)
Zone 3	Total	244	9.3E-03	1.3E-02 2.2	E-02

Notes:

N/A = Not applicable, radionuclides were not retained as COPCs under the dermal exposure pathway

Italics indicate that the hazard or risk shown for seepage-impacted groundwater is within background hazard or risk.

The summary table shows that there is significant to tal non-carcinogenic hazard and total ris k associated with a hypo thetical residential exposure scenario in each of the hydrostratigraphic units and that the highest hazard and risks are associated with Zone 3 groundwater. These calculations may accurately reflect the combined risk of exposure to seepage-impacted and non-seepage-impacted (i.e., background) groundwater, but likely overestimate the risk associated with seepage-impacted groundwater due the non-seepage-impacted COPC concentrations.

Total non-carcinogenic HI values exceed one in each of the hydrostratigraphic units. The ingestion exposure pathway is the most im portant for non-car cinogenic hazard, where total hazards exceed one for each of the hydrostratigraphic units. For the dermal exposure pathway, hazards exceed one for the Southwest Alluvium and Zone 3. The inhalation exposure pathway is not important with respect to non-carcinogenic hazard for any of the hydrostratigraphic units, because the only volatile non-rad iologic compound present in seepage-impacted groundwater is chloroform at very low concentrations, and only at locations immediately adjacent the tailing s impoundment. The hazard associated with the ingestion exposure scenario in background groundwater for Zone 3 exceeds that of the seepage-impacted water.

HIs segregated by target organ also exceed one w ithin each of the hydrostratigrap hic units as follows:

- For the Southwest Alluvium The HIs based on central nervous system effects are 8.7 for the child and 3.6 for the adult, due to th e ingestion of, and derm al contact with, manganese in groundwater. Segregated HIs for kidney effects are 2.7 for the future resident child and 1.2 for the future resident adult (due mostly to uranium ingestion) and the HI for thyroid effects is 2.1 for the child (due mostly to cobalt in gestion). Hazard indices for other specific organs or targets are less than one.
- For Zone 1 Several segregated total HIs ex ceed one for target organs or toxicological effect. The HIs based on thyroid effects are 11.9 for the child and 5.1 for the adult (due mostly to the ingestion of cobalt in groundw ater). Total HIs for the central nervous system are 6.1 for the child and 2.5 for the adult (due to manganese), and the total HIs for the metabolic system are 2.6 for the child an d 1.1 for the adult (due to vanadium). Hazard indices for other specific organs or targets are less than one.



• For Zone 3 – The segregated HIs based on thyroid effects are 94.2 for the child and 40.3 for the adult, due m ostly to the ingestion of cobalt in groundwater. Segregated HIs for skin toxicity are 88.4 for the child and 37.8 for r the adult. Segregated HIs for the cen tral nervous system are 36.3 for the child and 15.1 for the adult, and the segregated HIs for the kidney are 19.5 for the child and 8.3 fo r the adult. T he segregated HIs for the metabolic system (as indicated by decreased hair cystine) are 2.3 for the child and 1.0 for the adult, due to the vanadium ingestion. Hazard indices for the li ver, for both the adult and the child, are less than one. Additionally, the hazard index for gastroin testinal system effects and reduced body and organ weights is less than one for the adult receptor.

The summary table shows that total cancer ris k exceeds EPA's target risk range of 1E-04 to 1E-06 for each of the hydrostrat igraphic units. The radionuclide carcinogenic risk associated with inhalation exposure pathways for the Southwest Alluvium and Zone 1, and the radionuclide risk associated with the ingestion pathway for r Zone 1, are below background risks for these pathways within these units. In each case, radium-226 and/or radium -228 activities in background water exceed those in seepage-impacted water.

There is significant uncertainty associated with the risk assessment results, primarily with respect to the following factors:

- Background COPC concentrations. Some of the hazards and risks that are calculated for the seepage-impacted water are overestimates because background hazards or risks either exceed or represent a large portion of the calculated hazard or risk.
- Background water quality is not considered su itable for use as a prim ary drinking water source (e.g., due to sulfate and other chemicals that affect potability);
- Toxicity numbers (particularly for non-radionu clides) are typically con servative due to the incorporation of uncertainty factors and modifying factors. Furthermore, summations of total hazards and total risks may or may not be appropriate.
- Inhalation risks may be overestimated for the following reasons:
 - The model for inhalation risk to radi um-226 may not be appropriate because radium-226 is not volatile. The evalua tion of exposure to the radium-226 decay product radon m ight be m ore appropriate, but m easurements of radon in groundwater are not available. Furt hermore, the use of the Andelm an volatilization factor m ay be overly conservative for radium -226; a U.S. Geological Survey report (Lindsey and Ator, 1996) in dicates that the typical transfer of radon from well wa ter to re sidential air is 20 percent of that represented by the Andelman factor.
 - Inhalation exposure may be lower than estim ated, because many local residents don't have running water in their hom es and the models used to approxim ate RME intake may be inappropriate. However, some local residents may also haul



water from local wells and exposure factor s for this potential exposure scenario have not been identified.

- Hazards and risks may be underestimated based on usage of the following exposure factors:
 - A 30-year exposure duration m ay be low with respect to a local resident population;
 - A drinking water ingestion rate of two liters per day may be low with respect to a local population residing in a semi-arid environment;
 - The assumption of a 350-day exposure frequency could be slightly low for the local population, but is bounded at 365 days.
- There is likely to be in sufficient water available in Zone 1 for use as a potable w ater source for the exposure duration evaluated in the HHRA.
- Downgradient seepage impacts have been, and are expected to continue to be, limited by natural attenuation in all three hydrostratigraphic units.
- Assumptions that certain radionuclide decay products are at secular equilib rium with their parent nuclides.
- Assumption that uranium isotopes are present in proportion to natural abundance.

In summary, for the RME individ ual that meets assumptions made in this as sessment with the established uncertainties, there is a potential f or human health risk that exceeds the criteria established by the EPA for remedial action to be conducted.

7.2 Support the Reassessment of Remediation Levels and Provide a Basis for Comparing Remedial Alternatives

The results of this risk assessment, together with the background risk information and data from the three hydrostratigraphic units at the Site, can be used to support the reas sessment of remediation levels within the SWSFS and provide a basis for comparing remedial alternatives.

Consistent with EPA r isk assessment guidance regarding background concentrations (EPA, 2002), COPCs that are present in both im pacted and background groundwater have been carried through the quantitative risk assessment calculations of the seepage-impacted groundwater. The resulting non-carcinogenic hazard and carcinogenic risk estimates (i.e., estimates which include background risk) may accurately quantify the total hazard and risk of exposure to groundwater, but may overestimate the excess risk associated with seepage-impacted groundwater due to the hazards or risk asso ciated with background concentrations. T herefore, background concentrations should be considered in any futu re reassessment of Site rem ediation cleanup levels. Considerations may include the following:



- Where background concentrations exceed ARARs, background concentrations may be selected as remediation levels.
- Where background COPC conce ntrations exceed COPC concentrations in seepage impacted water, COPCs may be eliminated from further consideration as COCs.
- Where background water hazards or risks exceed EP A target levels, it would be m ore effective and appropriate to im plement remediation alternatives that restrict exposure to contaminated groundwater as compared to the existing groundwater remedy.

COCs were identified using a two step process. First, COCs were identified using the following criteria:

- Those COPCs which contribute at least 1E-06 cancer risk to an exposure scenario (i.e., total risk) that exceeds EPA's target risk range of 1E-04 to 1E-06; or
- Those COPCs contributing an HQ of at least 0.1 to an HI (for a segregated total HI) of 1 for non-cancer effects.

COCs that meet these criteria are shown on Ta bles 10.1.RME to 10.9.RME. The second step was to compare COC concentrations for each of the hydrostratigraphic un its against the corresponding background concentrations an d background risks. The following table summarizes the selected COCs and the rationale used to select them.

Hydrostratigraphic Unit	COCs Identified in "Table 10s"	Non-carcinogen or Carcinogen	Selected as COC	COC Selection Rationale
Southwest Alluvium	Arsenic	Carcinogen	No	Similar to background concentrations; below MCL
Southwest Alluvium	Cobalt	Non-carcinogen	No	One detected result in impacted water; background concentrations higher than impacted water concentrations
Southwest Alluvium	Manganese	Non-carcinogen	Yes	HI = 8.7 (Child)
Southwest Alluvium	Uranium	Non-carcinogen	Yes	HI = 2.7 (Child)
Southwest Alluvium	Chloroform	Carcinogen	Yes	Risk = 1.7E-06 – Ingestion and dermal Risk = 2.1E-06 – Inhalation
Southwest Alluvium	Uranium isotopes	Carcinogen	Yes	Risk > 1E-04



Hydrostratigraphic Unit	COCs Identified in "Table 10s"	Non-carcinogen or Carcinogen	Selected as COC	COC Selection Rationale
Southwest Alluvium	Radium-226	Carcinogen	No	Background concentrations higher than impacted
Southwest Alluvium	Radium-228	Carcinogen	No	Background concentrations higher than impacted
Zone 1	Cobalt	Non-carcinogen	Yes	HI = 11.9 (Child)
Zone 1	Manganese	Non-carcinogen	No	Background concentrations higher than impacted
Zone 1	Vanadium	Non-carcinogen	No	Hazard based on only one historical detection in seepage impacted water
Zone 1	Arsenic	Carcinogen	No	Similar to background concentrations; below MCL
Zone 1	Radium-226	Carcinogen	No	Background concentrations higher than impacted water concentrations
Zone 1	Radium-228	Carcinogen	No	Background concentrations higher than impacted water concentrations
Zone 1	Thorium-230	Carcinogen	No	Risk = 1.1E-06, within background radiological risk
Zone 3	Aluminum	Non-carcinogen	Yes	HI = 2.5 (Child)
Zone 3	Arsenic	Carcinogen and Non-carcinogen	Yes	HI = 88.4 (Child) Risk 9.3E-03
Zone 3	Cadmium	Non-carcinogen	Yes	HI = 9.1 (Child)
Zone 3	Cobalt	Non-carcinogen	Yes	HI = 94.2 (Child)
Zone 3	Manganese	Non-carcinogen	Yes	HI = 33.8 (Child)
Zone 3	Molybdenum	Non-carcinogen	No	Background concentrations higher than impacted water concentrations
Zone 3	Nickel	Non-carcinogen	Yes	HI = 1.6 (Child)



Hydrostratigraphic Unit	COCs Identified in "Table 10s"	Non-carcinogen or Carcinogen	Selected as COC	COC Selection Rationale
Zone 3	Vanadium	Non-carcinogen	Yes	HI = 2.3 (Child)
Zone 3	Uranium	Non-carcinogen	No	Background concentrations higher than impacted water concentrations
Zone 3	Chloroform	Carcinogen	Yes	Risk = 1.6E-06 – Ingestion and dermal Risk = 2.0E-06 – Inhalation
Zone 3	Uranium Isotopes	Carcinogens	No	Background concentrations higher than impacted water concentrations
Zone 3	Radium-226	Carcinogen	Yes	Risk = 8.5E-05 – Ingestion Risk 1.2E-02 Inhalation
Zone 3	Radium-228	Carcinogen	Yes	Risk = 3.5E-04 – Ingestion
Zone 3	Lead-210	Carcinogen	Yes	Risk = 5.5E-05 – Ingestion

These COCs may require consideration for rem edial action in the SWSFS. The SWSFS, which is currently underway, will cons ider the complicated nature of overlapp ing human health risks and hazards associated with seep age-impacted and background water. This information will be used to support any future CERLCA decision- making regarding rem edy modification and, if necessary and appropriate, provide a basis for pot entially waiving ARARs due to TI, consistent with the NCP and EPA TI waiver guidance.

7.3 Identify Point of Compliance (POC) and Po int of Exposure (POE) Concentrations in Accordance with NRC requirements

The results of this risk assessment, together with the background risk information and data from the three hydrostratigraphic units at the Site, can be used outside of the CERCLA program to support the identification of N RC Point of Compliance (POC) and Point of Exposure (POE) concentrations (and associated risks) in accordance with NRC requirements.

The elements of the ex posure assessment for the NRC are estab lished in NRC guidance for review of reclamation plans (NRC, 2003, known as NUREG 1620), according to which:



- The POC is the location at which the groundwat er is monitored to determine compliance with the groundwater protection standards. POCs have been established for each of the hydrostratigraphic units in the UNC's NRC Source Materials License.
- The POE is any location where people, wildlife, or other species could reasonably be exposed to hazardous constituents from groundwater contaminated by uranium m ill tailings. The POE is genera lly located at the downgradient edge of land that will be transferred to eith er the Federal g overnment or the state for long-term institutional control. For the Church Rock Site, the la nd to be transferred for administration by DOE includes Section 2, which com prises the tail ings disposal site, and may also include Section 36, which is currently owned by UNC and adjoins Section 2 to the north.

With the exception of the previously m entioned, unexpected high fr equency of lead-210 concentrations in October 2010, the Church Ro ck Site meets all NRC License ground water protection standards at the POCs for the Sout hwest Alluvium and, upon approval of the ACL application for nickel and chloroform in Zone 1, will m eet all the N RC License com pliance standards in Zone 1. (T he first uranium exceedance (0.312 mg/L) in So uthwest Alluvium POC well 509-D during October 2010 is an exception which UNC will monitor.)

The results of the risk assessm ent indicate that the there could potentially be unacceptable risk associated with using Zone 3 groundwater for residential use. Attenuation will occur in the Zone 3 plume, but it would likely be insufficient to reduce hazards and risks to potential receptors at a hypothetical POE that is established at the nor thern border of Section 36. NRC's NUREG 1620 (2003; section 4.3.3.2 Exposure Assessm ent) indicates that, using either em pirical data or calculations, maximum permissible levels of COCs at a given POC should be determ ined and those levels should be protective of human health and the environment at the P OE. UNC is presently evaluating the feasibility of conducting numerical groundwater modeling of Zone 3. The results of this task are pending and it is premature to define maximum permissible COC concentrations at a given POC. UNC will work with NRC to determine the most appropriate approach to these POC-POE issues.

The Zone 1 risk assessment evaluation in this report provides additional documentation to NRC that Zone 1 risks associated with chloroform and nickel are low at the proposed POE, and can be used in support of UNC's submitted ACL application (N.A. Water Systems, 2008d).



- Andelman, J.B.,1990, Total Exposu re to Organic Chem icals in Potable Water: Chapter 20 in Cantor, K.P., Christm an, R.F., Ram, N.M.(editors), Significance and Treatment of Volatile Organic Compounds in Water Supplies; Lewis Publishers. pp. 485-504.
- California EPA, 1994, Preliminary Endangerment Assessment Guidance Manual (PEA); Department of Toxic Substances Control, Sacramento, California.
- Canonie Environmental Services Corp., 1995, EPA Remedial Action and NRC Gr ound Water Corrective Action, Five-Year Review (1989-1994). January.
- Canonie Environmental Services Corp., 1993, Ground Water Corrective Action, Annual Review 1993, Church Rock Site, Gallup, New Mexico. December.
- Canonie Environmental Services Corp., 1992, Ground Water Corrective Action, Annual Review 1992, Church Rock Site, Gallup, New Mexico. December.
- Canonie Environmental Services Corp., 1991, Ground Water Corrective Action, Annual Review 1991, Church Rock Site, Gallup, New Mexico. December.
- Canonie Environmental Services Corp., 1990, Ground Water Corrective Action, Annual Review 1990, Church Rock Site, Gallup, New Mexico. December.
- Canonie Environmental Services Corp., 1989a, Re medial Design Report, Church Rock Site, Gallup, New Mexico. April.
- Canonie Environmental Services Corp., 1989b, Ground Water Corrective Action, Annual Review - 1989, Church Rock Site, Gallup, New Mexico. December.
- Chester Engineers, 2011, A nnual Review Report 2010 Gr oundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. January 26, 2011.
- Chester Engineers, 2010, A nnual Review Report 2009 Gr oundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. January 28, 2010.
- Chester Engineers, 2009, A nnual Review Report 2008 Gr oundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. January 28, 2009. Ea rth Tech, 2002a, Ground Water Corrective Action, Annual Review - 2001, Church Rock Site, Gallup, New Mexico. January.
- Earth Tech, Inc., 2002a, Ground Water Corrective Action, Annual Review 2001, Church Rock Site, Gallup, New Mexico. January.
- Earth Tech, 2002b, Annual Review Report 2002 Groundwater Corrective Action, Church Rock Site, Gallup, New Mexico. December.



- Earth Tech, 2000, Ground W ater Corrective Action, Annual Review 2000, Church Rock Site, Gallup, New Mexico. December.
- Earth Tech, 1999, Ground W ater Corrective Action, Annual Review 1999, Church Rock Site, Gallup, New Mexico. December.
- Earth Tech, 1998, Ground W ater Corrective Action, Annual Review 1998, Church Rock Site, Gallup, New Mexico. December.
- Foster, S.A. and Chrostowski, P.C.,1987, Inhalation Exposures to Volatile Organic Contaminants in the Shower: Presentation at the 80th A nnual Meeting of APCA. New York, NY. June 21-26, 1987.
- Idaho Department of Environmental Quality, 2001, State Ground Water Program, Ground Water Quality Information Series No. 1, Nitrat es in Ground Water -- A Continuing Issue for Idaho Citizens: Idaho Departm ent of Environmental Quality State Ground Water Program.
- Institute of Medicine, 2001, Dietary Reference Intakes for Vitam in A, Vitam in K, Arsenic, Boron, Chromium, Copper, Iodine, Iron, Manganese, Molybdenum, Nickel, S ilicon, Vanadium, and Zinc: Panel on Micronutrients, Subcommittees on Upper Reference Levels of Nutrients and of In terpretation and Use of Dietary Reference Intakes, and the Standing Committee on the Scientific Evaluation of Dietary Reference Intakes; National Academies Press.
- King, L. J., 2007, Church Rock Uranium Monito ring Project, Presentation Materials for USEPA Community Involvement Conference. June.
- Lindsey, B. D. and S. W. Ator, 1996, Radon in ground water of the Lower Susquehanna and Potomac River basins, U.S. Geological Surv ey, Water-Resources Investigations Report 96-4156.
- N.A. Water Systems, 2008a, Annual Review Report 2007 Gr oundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. January 29, 2008.
- N.A. Water Systems, 2008b, Revised Subm ittal Calculation of Background Statistics with Comparison Values, U NC Church Rock Mill & Ta ilings Site, Church Rock, New Mexico. October 17, 2008.
- N.A. Water Systems, 2008c, Revised Subm ittal Estimated UCL95 Statistics and EPCs in Impacted Groundwater, UNC Church Rock M ill & Tailings Site, Church Rock, New Mexico. December 5, 2008.
- N.A. Water Systems, 2008d, Alternate Concentration Limits Application, Zone 1 of the Lower Gallup Sandstone, UNC Church Rock Site, C hurch Rock, New Mexico. December 29, 2008.



- N.A. Water Systems, 2007a, Annual Review Report 2006 Gr oundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. January 9, 2007.
- N.A. Water Systems, 2007b, Site-Wide Supplemental Feasibility Study, Part I, C hurch Rock Remediation Standards Update, Church Rock Site, Church Rock, New Mexico. February 19, 2007.
- N.A. Water Systems, 2005, Annual Review Re port -- 2005 Groundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. December 28, 2004.
- N.A. Water Systems, 2004, Annual Review Re port -- 2004 Groundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. December 30, 2004.
- National Research Cou ncil, 1995, Nitrate and Nitrite in Drinking W ater: Subcommittee on Nitrate and Nitrite in Drinking Water, Committee on Toxicology Board on Environmental Studies and Toxicology, Commission on Life Sciences, National Academy Press, Washington, D.C.
- Nuclear Regulatory Commission, 2003, Standard Review Plan for the Review of a Reclam ation Plan for Mill Tailings Sites Under Title II of the Uranium Mill Tailings Radiation Control Act of 1978 (NUREG 1620). June.
- Nuclear Regulatory Commission, 1999b, Letter to Roy Blickwedel from John Surmeier, Subject: "Consideration of Temporary Saturation of a portion of Zone 3 at the Church Rock Site." September 16, 1999.
- Nuclear Regulatory Commission, 1996, Evaluation of the Statisti cal Basis for E stablishing Background Levels and Re mediation Standards at the United Nuclear Corporation Church Rock Uranium Mill Tailings Disposal Facility, Gallup, New Mexico. June.
- Raymondi, R.R. and R. C. Conrad, 1983, Hydroge ology of Pipeline Canyon, Near Gallup, New Mexico: Ground Water, Vol. 21, No. 2, pp. 188-198.
- Rhodes, M.C., K.G. Keil, W.T. Frederick, J.S. Leithner, J.M. Peterson, M.M. MacDonell, 2006, Utilizing Isotopic Uranium Ratios in Ground-water Evaluations at NFSS: Proceedings of 2006 Waste Management Symposium, Global Accomplishments in Environmental and Radioactive Waste Management: Education and Opportunity for the Next Generation of Waste Management Professionals; February 26 - March 2, 2006, Tucson, Arizona.
- Rust Environment and Infrastructure, 1997, Gr ound Water Corrective Action, Annual Review 1997, Church Rock Site, Gallup, New Mexico. December.
- Smith Technology Corporation, 1996, Ground Water Corrective Action, Annual Review 1996, Church Rock Site, Gallup, New Mexico. December.



- Smith Technology Corporation (Smith Environmental Technologies Corporation), 1995, Ground
 Water Corrective Action, Annual Review 1995, Church Rock Site, Gallup, Ne w
 Mexico. December.
- State of Ohio Environmental Pro tection Agency, Nitrate in Public Drin king Water, Fact Sheet, Division of Drinking and Ground Waters, January 2005.
- Stone, W.J., 1981, Hydrogeology of the Gallup Sa ndstone, San Juan Basin, Northwest New Mexico: Ground Water, Vol. 19, No. 1, pp. 4-11.
- United Nuclear Corporation, 1989a, Corrective Action Plan, Church Rock Uranium Mill Tailings Facility. April.
- United Nuclear Corporation, 1989b, Remedial Action Plan, Church Rock Uranium Mill Tailings Facility. April.
- U.S. Environmental Protection Agency (EPA), 2010a, Letter from Katrina Higgins-Coltrain to Larry Bush [UNC], Subject: "Comments on the Site-Wide Supplemental Feasibility Study Part 2, Docum ent dated July 2009, for the UNC Chur ch Rock Mill, Gallup, New Mexico Superfund Site and related Source Materials L icense SUA-1475, Docket No: 040-08907 and General Site-wide considerations for Part 3." September 2, 2010.
- U.S. Environmental Protection Agency (EPA), 2010b, November 2010 EPA Risk Screening Level Table, <u>http://www.epa.gov/region9/superfund/prg/</u>. November.
- U.S. Environmental Protection Agency (EPA), 2009, Risk Assessment Guidance for Superfund Volume I: Hum an Health Evaluation Manua 1 (Part F, Supplem ental Guidance for Inhalation Risk Assessment), Office of Superfund Rem ediation and Technology Innovation, EPA-540-R-070-002, OSWER 9285.7-82. January.
- U.S. Environmental Protection Agency (EPA), 2008, Third Five-Year Review Report for the United Nuclear Corporation Ground W ater Operable Unit, Church Rock, McKinley County, New Mexico. September.
- U.S. Environmental Protection Agency (EPA), 2007, ProUCL Version 4.00.02 User Guide: Prepared by A. Singh, R. Maichle, A. K. Singh, S. Lee, N. Ar mbya, EPA/600/R-07/038. April.
- U.S. Environmental Protection Agency (EPA), 2004, Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) Final; Office of Emergency and Re medial Response, EPA/540/R/99/005, OSWER 9285.7-02EP PB99-963312. July.
- U.S. Environmental Protection Agency (EPA), 2003, Memorandum from Michael B. Cook, Director, "Human Health Toxicity Values in Superfund Risk Assessments"; Office of Solid Waste and Emergency Response, OSWER Directive 9285.7-53. Decem ber 5, 2003.



- U.S. Environmental Protection Agency (EPA), 2002, Role of Background in the CERCLA Cleanup Program; Office of Solid W aste and Emergency Response, OSWER 9285.6-07P. April 26, 2002.
- U.S. Environmental Protection Agency (EP A), 2001a, Health Effects Summ ary Tables (HEAST); National Center for r Environmental Assessment, <u>http://www.epa.gov/radiation/heast/</u>, updated as of April 2001.
- U.S. Environmental Protection Agency (EPA), 2001b, Risk Assessment Guidance for Superfund: Volume I Human Health Evaluation Manual (Part D, Standardized Planning, Reporting, and Review of Superfund Risk Assessments), Office of Emergency and Remedial Response, Publication 9285.7-47. December.
- U.S. Environmental Protection Agency (EPA), 1999, Cancer Risk Coefficients for Environmental Exposure to Radionuclides, Federal Guidance Report No. 13, Air and Radiation, EPA 402-R-99-001. September.
- U.S. Environmental Protection Agency (EPA), 1991a, Risk Assessment Guidance for Superfund: Volume I Hu man Health Ev aluation Manual (Part B, De velopment of Risk-based Preliminary Remediation Goals), Interim, Office of Emergency and Remedial Response EPA/540/R-92/003. December.
- U.S. Environmental Protection Agency (EPA), 1991b, Risk Assessment Guidance for Superfund. Volume 1: Human Health Evaluation Manual - Supplemental Guidance, Standard Default Exposure Factors, Interim Final; OSWER 9285.6-03.
- UU.S. Environmental Protection Agency (EPA), 1989, Risk Ass essment Guidance for Superfund: Volume I - Human Health Evaluation Manual (Part A), Interim Final, Office of Emergency and Remedial Response EPA/540/1-89/002. December.
- U.S. Environmental Protection Agency (EPA), 1988a, Remedial Investigation, United Nuclear Corporation Churchrock Site. August.
- U.S. Environmental Protection Agency (EPA), 1988b, United Nuclear Corporation Church Rock Site, Operable Unit Feasibility Study, Gallup, New Mexico. August.
- U.S. Environmental Protection Agency (EPA), 1988c, Record of Decision, United Nuclear Corporation, Ground Water Operable Unit, McKinley County, New Mexico. U.S. Environmental Protection Agency, Region VI, Dallas, Texas. September.
- USFilter, 2004a, Annual Review Re port 2003 Gr oundwater Corrective Action, Church Rock Site, Church Rock, New Mexico. January 14, 2004.
- USFilter, 2004b, Rationale and Field Inves tigation Work Plan to Evaluate Recharge and Potential Cell Sourcing to the Zone 3 Plum e, Church Rock Site, Gallup, New Mexico. January 19, 2004.





TABLE 1

Monitoring COPCs Impacted Water Quality, July 2006 - April 2008 UNC Church Rock Mill and Tailings Site

Sampling and Analys	is Plan Monitoring COPCs					
Aluminum	Chloride*					
Arsenic	Sulfate*					
Beryllium	Nitrate-Nitrogen*					
Cadmium	Uranium					
Cobalt	Chloroform					
Lead	Total Dissolved Solids*					
Manganese	Thorium-230					
Molybdenum	Lead-210					
Nickel	Gross Alpha*					
Selenium	Radium (including Ra-226,					
Vanadium	Ra-228, and total radium*)					
Historical Monitoring Para	neters - Trace Metals Plus Iron**					
Antimony	Mercury					
Barium	Silver					
Chromium	Thallium					
Copper	Zinc					
Iron						

* Common ion, general chemistry, and grouped parameters eliminated from consideration in the quantitative risk assessment calculations

** Previously dropped from the performance monitoring program. These parameters were eliminated from consideration in the quantitative risk assessment calculations

TABLE 2

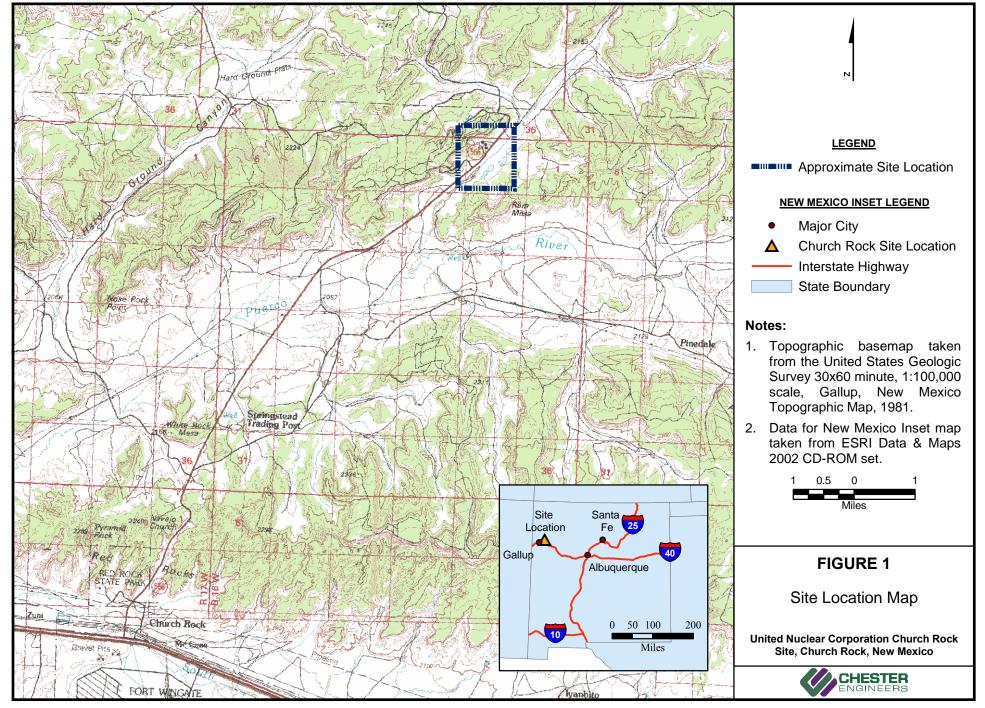
Wells Having Samples Representative of Impacted Water Quality, July 2006 - April 2008 UNC Church Rock Mill and Tailings Site

Southwest Alluvium	Zone 1	Zone 3
509 D (POC)	515 A*	504 B
624	604 (POC)*	517 (POC)
632 (POC)	614 (POC)*	613 (POC)*
801	EPA 5 (POC)	708 (POC)
802	EPA 7 (POC)	711 (POC)
803		717
808		719
EPA 23 (POC)		EPA 13
EPA 25		EPA 14
GW 1 (POC)		NBL 1
GW 2 (POC)		
GW 3 (POC)		

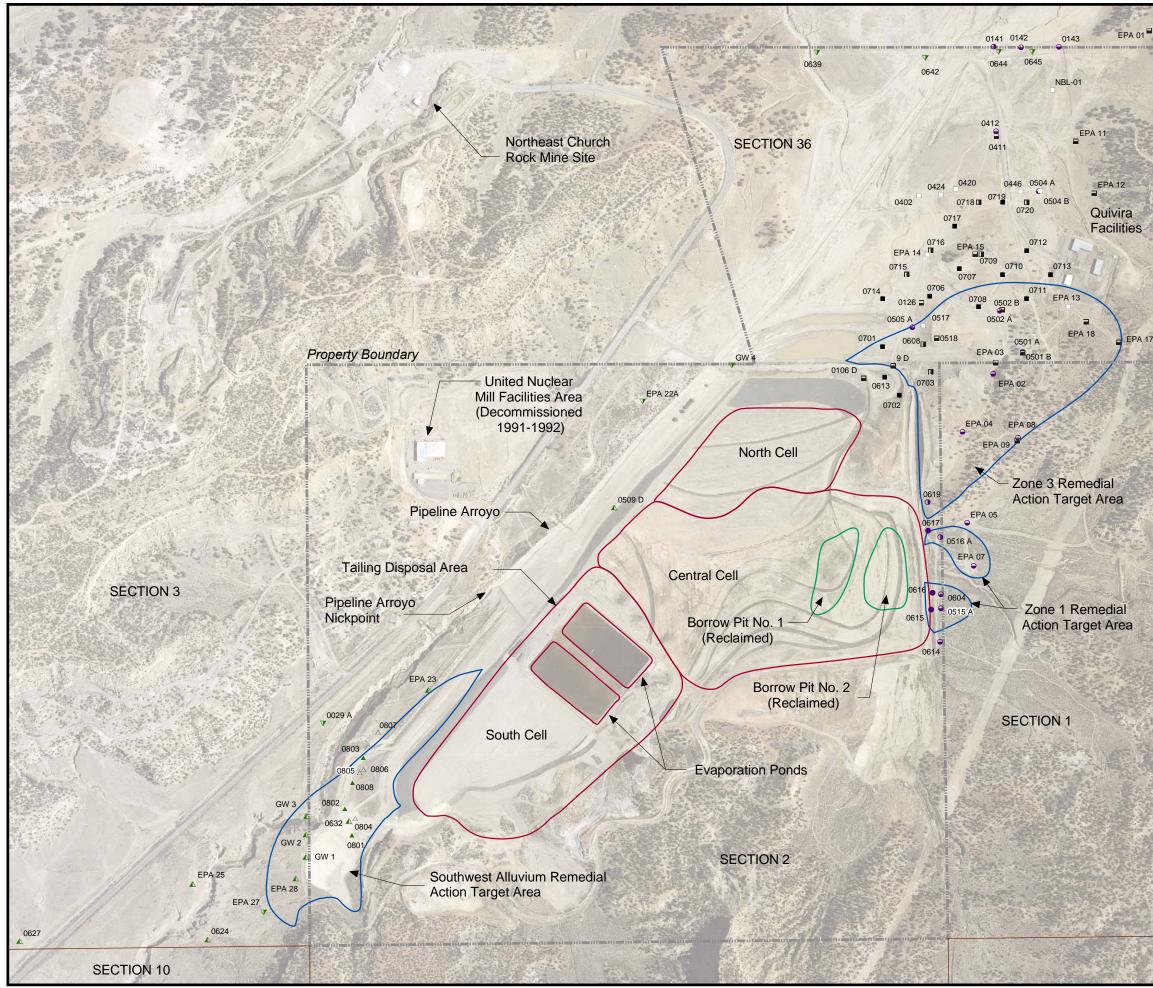
Notes:

* indicates well not included in the HHRA due to its location within Section 2 POC = NRC License Point-of-Compliance Well





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Legend

Southwest Alluvium

- Idled Extraction Well
- Monitoring Well
- Water Level Monitoring Well
- Dry Monitoring Well

Zone 3

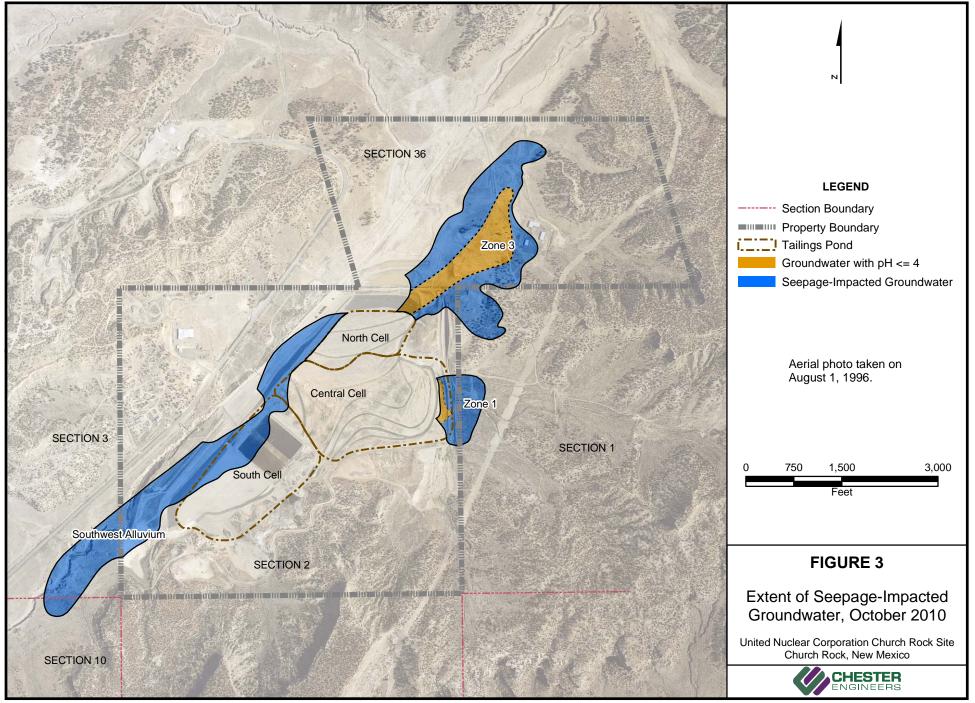
- Idled Extraction Well Used for Monitoring
- Decommissioned or Idle Extraction Well
- Monitoring Well
- Dry or Decommissioned Monitoring Well

Zone 1

- Decommissioned Extraction Well
- Decommissioned Monitoring Well
- Monitoring Well

0	400	800 Feet	1,6	600]						
FIGURE 2										
Site Layout and Performance Monitoring Well Locations 2010 Operating Year										
United Nu			Church Ro w Mexico	ock Site,	,					





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APPENDIX A

EPA Risk Assessment Tables – Seepage Impacted Water



TABLE 0 SITE RISK ASSESSMENT IDENTIFICATION INFORMATION UNC Church Rock Mill and Tailings Site

Site Name/OU:	UNC CHURCH ROCK MILL AND TAILINGS SITE
Region:	6
EPA ID Number:	NMD030443303
State:	New Mexico
Status:	RP Sitewide Supplemental Feasibility Study (ongoing)
Federal Facility (Y/N):	Ν
EPA Project Manager	Katrina Coltrain
EPA Risk Assessor:	Anna Milburn
Prepared by (Organization):	Chester Engineers
Prepared for (Organization):	United Nuclear Corporation
Document Title:	Human Health Risk Assessment for the UNC Church Rock Site
Document Date:	March 2011
Probabilistic Risk Assessment (Y/N):	Ν
Comments:	Groundwater at this site is impacted by seepage from uranium mill tailings disposal

TABLE 1 SELECTION OF EXPOSURE PATHWAYS UNC Church Rock Mill and Tailings Site

Scenario Timeframe	Medium	Exposure Medium	Exposure Point	Receptor Population	Receptor Age	Exposure Route	Type of Analysis	Rationale for Selection or Exclusion of Exposure Pathway ⁽¹⁾
Future	Groundwater -	Groundwater	SW Alluvium	Resident	Adult	Dermal	Quant	Future resident using domestic well in SW Alluvium
	Southwest Alluvium		Tapwater			Ingestion	Quant	Future resident using domestic well in SW Alluvium
					Child	Dermal	Quant	Future resident using domestic well in SW Alluvium
						Ingestion	Quant	Future resident using domestic well in SW Alluvium
					Adult/Child	Dermal	Quant	Future resident using domestic well in SW Alluvium
						Ingestion	Quant	Future resident using domestic well in SW Alluvium
		Air	Water Vapors from	Resident	Adult	Inhalation	Quant	Future resident using domestic well in SW Alluvium
			Showerhead /		Child	Inhalation	Quant	Future resident using domestic well in SW Alluvium
			Domestic Uses		Adult/Child	Inhalation	Quant	Future resident using domestic well in SW Alluvium
	Groundwater -	Groundwater	Zone 1	Resident	Adult	Dermal	Quant	Future resident using domestic well in Zone 1
	Zone 1		Tapwater			Ingestion	Quant	Future resident using domestic well in Zone 1
	Upper Gallup Fm				Child	Dermal	Quant	Future resident using domestic well in Zone 1
						Ingestion	Quant	Future resident using domestic well in Zone 1
					Adult/Child	Dermal	Quant	Future resident using domestic well in Zone 1
						Ingestion	Quant	Future resident using domestic well in Zone 1
		Air	Water Vapors from	Resident	Adult	Inhalation	Quant	Future resident using domestic well in Zone 1
			Showerhead /		Child	Inhalation	Quant	Future resident using domestic well in Zone 1
			Domestic Uses		Adult/Child	Inhalation	Quant	Future resident using domestic well in Zone 1
	Groundwater -	Groundwater	Zone 3	Resident	Adult	Dermal	Quant	Future resident using domestic well in Zone 3
	Zone 3		Tapwater			Ingestion	Quant	Future resident using domestic well in Zone 3
	Upper Gallup Fm				Child	Dermal	Quant	Future resident using domestic well in Zone 3
						Ingestion	Quant	Future resident using domestic well in Zone 3
					Adult/Child	Dermal	Quant	Future resident using domestic well in Zone 3
						Ingestion	Quant	Future resident using domestic well in Zone 3
	l í	Air	Water Vapors from	Resident	Adult	Inhalation	Quant	Future resident using domestic well in Zone 3
			Showerhead /		Child	Inhalation	Quant	Future resident using domestic well in Zone 3
			Domestic Uses		Adult/Child	Inhalation	Quant	Future resident using domestic well in Zone 3

(1) All domestic well scenarios are based on the assumption that a potential future domestic well would be located outside UNC property (Sections 2 and 36) where tailings seepage has impacted existing background groundwater.

TABLE 2.1 OCCURRENCE, DISTRIBUTION, AND SELECTION OF CHEMICALS OF POTENTIAL CONCERN UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Medium: Groundwater Exposure Medium: Groundwater

Exposure	CAS	Chemical	Minimum	Maximum	Units	Location	Detection	Range of	Concentration	-	Screening	Potential	Potential ARAR/TBC	COPC	Rationale for
Point	Number		Concentration (Qualifier) (1)	Concentration (Qualifier) (1)		of Maximum Concentration	Frequency	Detection Limits	Used for Screening (4)	Value (5)	Toxicity Value (N/C) (6)	ARAR/TBC Value	Source	Flag (Y/N)	Selection or Deletion (7)
SW Alluvium	7429-90-5	Aluminum	0.1	0.3	mg/L	808	6 / 96	0.1 - 0.1	0.3	0.107	3.7 (N)	5	NMWQCC-I	N	BSL
Tapwater	7440-38-2		0.001	0.01	mg/L	632, 801, 803, 808, EPA23, EPA25, GW1, GW3	13 / 96	0.001 - 0.001	0.01	0.00116	0.000045 (C)	0.01	MCL	Y	ASL
	7440-48-4	Cobalt	0.01	0.01	mg/L	GW 3	1 / 96	0.01 - 0.01	0.01	0.0121	0.0011 (N)	0.05	NMWQCC-I	Y	ASL
	7439-96-5	Manganese	0.03	5.4	mg/L	EPA23	96 / 96	0.01 -0.01	5.4	0.414	0.088 (N)	0.2	NMWQCC-O	Y	ASL
	7782-49-2	Selenium	0.001	0.001	mg/L	EPA 23	1 / 96	0.001 - 0.001	0.001	0.00516	0.018 (N)	0.05	MCL	N	BSL
	7440-61-1	Uranium	0.0229	0.2460	mg/L	509D	96 / 96	0.0003 - 0.0004	0.246	0.0459	0.011 (N)	0.03	MCL	Y	ASL
	13966-29-5	Uranium-234	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	15117-96-1	Uranium-235	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	7440-61-1	Uranium-238	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	67-66-3	Chloroform	0.00061	0.0155	mg/L	802	49 / 96	0.0005 -0.001	0.0155	ND	0.00019 (C)	0.08	MCL (TTHM) MCL (combined	Y	ASL
	13982-63-3	Radium-226 (3)	0.1	1	pCi/L	632	37 / 96	0.001 - 0.2	1	0.798	NA (C)	5	radium)	Y	DET
	15262-20-1	Radium-228 (3)	0.3	4.3	pCi/L	632	22 / 96	0.04 - 1	4.3	1.611	NA (C)	5	MCL (combined radium)	Y	DET
													MCL (gross		
		Thorium-230	0.2	1.6	pCi/L	632	10 / 96	0.2 - 0.2	1.6	0.509	NA (C)	15	alpha)	Y	DET
Zone 1		Aluminum	0.2	1.3	mg/L	EPA07	7 / 16	0.1 - 0.1	1.3	0.117	3.7 (N)	5	NMWQCC-I	N	BSL
Tapwater	7440-38-2	Arsenic	0.001	0.003	mg/L	EPA07	4 / 16	0.001 - 0.001	0.003	0.00117	0.000045 (C)	0.01	MCL	Y	ASL
	7440-48-4	Cobalt	0.02	0.06	mg/L	EPA05	16 / 16	0.01 - 0.01	0.06	0.0112	0.0011 (N)	0.05	NMWQCC-I	Y	ASL
	7439-96-5	Manganese	0.95	2.96	mg/L	EPA07	16 / 16	0.01 - 0.01	2.96	2.519	0.088 (N)	0.2	NMWQCC-O	Y	ASL
	7440-02-0	Nickel	0.05	0.06	mg/L	EPA05	3 / 16	0.05 - 0.05	0.06	0.0602	0.073 (N)	0.2	NMWQCC-I	N	BSL
	7782-49-2	Selenium	0.001	0.001	mg/L	EPA05	1 / 16	0.001 - 0.001	0.001	0.00107	0.018 (N)	0.05	MCL	N	BSL
	7440-62-2	Vanadium	0.2	0.2	mg/L	EPA07	1 / 16	0.1 - 0.1	0.2	ND	0.018 (N)	NA	NA	Y	ASL
	7440-61-1	Uranium	0.0012	0.0022	mg/L	EPA07	16 / 16	0.0003 - 0.0004	0.0022	0.0255	0.011 (N)	0.03	MCL	N	BSL
	13966-29-5	Uranium-234	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	15117-96-1	Uranium-235	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	7440-61-1	Uranium-238	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	67-66-3	Chloroform	0.0006	0.00076	mg/L	EPA07	2 / 16	0.0005 -0.001	0.00076	NBC	0.00019 (C)	0.08	MCL (TTHM)	Y	ASL
	13982-63-3	Radium-226 (3)	0.4	1.8	pCi/L	EPA05	13 / 16	0.2 - 0.2	1.8	1.314	NA (C)	5	MCL (combined radium)	Y	DET
	15262-20-1	Radium-228 (3)	1	4	pCi/L	EPA05	8 / 16	1.0 - 1.0	4	2.946	NA (C)	5	MCL (combined radium)	Y	DET
	14269-63-7	Thorium-230	0.6	0.7	pCi/L	EPA05	2 / 16	0.2 - 0.2	0.7	0.403	NA (C)	15	MCL (gross alpha)	Y	DET
Zone 3	7429-90-5	Aluminum	0.1	163	mg/L	EPA14	58 / 70	0.1 - 0.1	163	0.231	3.7 (N)	5	NMWQCC-I	Y	ASL
Tapwater	7440-38-2		0.001	2.5 D	mg/L	NBL-01	48 / 70	0.001 - 0.03	2.5	0.175	0.000045 (C)	0.01	MCL	Y	ASL
	7440-41-7		0.01	0.09	mg/L	EPA14	9 / 70	0.01 - 0.01	0.09	ND	0.0073 (N)	0.004	MCL	Y	ASL
	7440-43-9		0.005	1	mg/L	EPA14	16 / 70	0.005 - 0.005	1	0.0113	0.0018 (N)	0.005	MCL	Y	ASL
	7440-48-4		0.05	0.95 D	mg/L	EPA14	70 / 70	0.01 - 0.02	0.95	0.0877	0.0011 (N)	0.05	NMWQCC-I	Y	ASL
	-	Manganese	3.33	23.7	mg/L	717	70 / 70	0.01 - 0.01	23.7	3.436	0.088 (N)	0.2	NMWQCC-O	Y	ASL
		Molybdenum	0.1	5	mg/L	NBL-01	32 / 70	0.1 - 0.1	5	17.43	0.018 (N)	1	NMWQCC-I	Y	ASL
	7440-02-0		0.11	0.89	mg/L	EPA14	70 / 70	0.05 - 0.05	0.89	0.14	0.073 (N)	0.2	NMWQCC-I	Y	ASL
	7782-49-2 7440-62-2	Selenium Vanadium	0.001	0.01	mg/L mg/L	NBL-01 517, 708, EPA13, EPA14, NBL-01	3 / 70 5 / 70	0.001 - 0.001	0.01	0.00159 ND	0.018 (N) 0.018 (N)	0.05 NA	MCL NA	N Y	BSL ASL
	7440-61-1	Uranium	0.0011	0.138	mg/L	NBL-01	70 / 70	0.0003 - 0.002	0.138	0.107	0.011 (N)	0.03	MCL	Y	ASL
	13966-29-5	Uranium-234	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	15117-96-1	Uranium-235	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	7440-61-1	Uranium-238	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Y	DET
	67-66-3	Chloroform	0.00093	0.00676	mg/L	517	13 / 70	0.0005 -0.001	0.00676	NBC	0.00019 (C)	0.08	MCL (TTHM) MCL (combined	Y	ASL
	13982-63-3	Radium-226	2	27.6	pCi/L	EPA14	70 / 70	0.1 - 0.2	27.6	4.996	NA (C)	5	radium) MCL	Y	DET
		Radium-228	3.8	56.1	pCi/L	EPA14	70 / 70	1 - 1.4	56.1	4.509	NA (C)	5	(combined radium) MCL (gross	Y	DET
		Thorium-230	0.2	1.3	pCi/L	517	6 / 70	0.2 - 0.2	1.3	1.426	NA (C)	15	alpha) MCL (gross	Y	DET
	14255-04-0	Lead-210	1.8	8.1	pCi/L	719	6 / 70	1.0 - 1.0	8.1	1.618	NA (C)	4 mrem/y	beta)	Y	DET

(1) Qualifier codes used for the "Minimum Concentration" and "Maximum Concentration":

Definitions: NA = Not Applicable

D = the sample was diluted to facilitate analysis.

- (2) Uranium isotopes not analyzed. Isotope concentrations esimated from total uranium mass concentration (see Table 3.A.RME).
- (3) The summary statistics for Radium-226 and Radium-228 count the raw below-detection-limit values from the 2nd Quarter 2008 as detections, because they were treated as such in the calculation of EPC statistics using ProUCL.
- (4) Maximum concentration used for screening chemicals. No screening was conducted for radionuclides. All radionuclides detected are selected as COPCs.
- (5) Background values are mean (UCL95) calculated in N.A. Water Systems (2008b).
- (6) All compounds were screened against the November 2010 USEPA Risk Screening Level Table (http://www.epa.gov/region9/superfund/prg/).
 - For non-carcinogens: screening value = 0.1 x RSL tapwater value
 - For carcinogens : screening value = RSL tapwater value
- (7) Rationale Codes:

Notes:

Selection Reason	Above Screening Level (ASL)
	Detected in seepage-impacted groundwater at Site (DET)
Deletion Reason	Below Screening Level (BSL)

ND = Not Detected

- MCL = Maximum Contaminant level
- SMCL = Secondary Maximum Contaminant level
- NMWQCC = New Mexico Water Quality Control Commission Groundwater Standard
- NMWQCC-I = New Mexico Water Quality Control Commission Irrigation Standard
- NMWQCC-O = New Mexico Water Quality Control Commission Other Standard
- C = Carcinogen
- N = Noncarcinogen
- TTHM = 0.080 mg/l is the MCL for total trihalomethanes, of which chloroform is a single compound.
- NBC = No background concentration chloroform was not detected frequently enough in Zone 1 and Zone 3 background samples (less than 1%) to calculate background concentration

TABLE 2.2 OCCURRENCE, DISTRIBUTION, AND SELECTION OF CHEMICALS OF POTENTIAL CONCERN UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Medium: Groundwater Exposure Medium: Air

F	040	Observised	Minimum	Massianum	1.1	Lasatian	Detection	Dense of	Ormentertier	Declaration	Sereening	Potential	Potential	0000	Dationals (or
Exposure	CAS	Chemical	Minimum	Maximum	Units	Location	Detection	Range of	Concentration	0	Screening	ARAR/TBC	ARAR/TBC	COPC	Rationale for
Point	Number		Concentration	Concentration		of Maximum	Frequency	Detection	Used for	Value	Toxicity Value			Flag	Selection or
			(Qualifier) (1)	(Qualifier) (1)		Concentration		Limits	Screening (4)	(5)	(N/C) (6)	Value	Source	(Y/N)	Deletion (7)
SW Alluvium	7429-90-5	A				000	C / OC	0.1 0.1				5		N	
Water Vapors from		Aluminum	0.1	0.3	mg/L	808 632, 801, 803, 808,	6 / 96	0.1 - 0.1	0.3	0.107	3.7 (N)	5	NMWQCC-I	N	BSL
Showerhead /						EPA23, EPA25, GW1,									
Domestic Use	7440-38-2		0.001	0.01	mg/L	GW3	13 / 96	0.001 - 0.001	0.01	0.00116	0.000045 (C)	0.01	MCL	N	NV
	7440-48-4		0.01	0.01	mg/L	GW 3	1 / 96	0.01 - 0.01	0.01	0.0121	0.0011 (N)	0.05	NMWQCC-I	N	NV
		Manganese	0.03	5.4	mg/L	EPA23	96 / 96	0.01 -0.01	5.4	0.414	0.088 (N)	0.2	NMWQCC-O	N	NV
	7782-49-2		0.001	0.001	mg/L	EPA 23	1 / 96	0.001 - 0.001	0.001	0.00516	0.018 (N)	0.05	MCL	N	BSL
	7440-61-1	Uranium	0.0229	0.2460	mg/L	509D	96 / 96	0.0003 - 0.0004	0.246	0.0459	0.011 (N)	0.03	MCL	N	NV
	13966-29-5		NA (2)	NA (2)	pCi/L	NA (2)	NA (C)	NA (2)	NA	NA	NA (C)	NA	NA	N	NV
	15117-96-1	Uranium-235	NA (2)	NA (2)	pCi/L	NA (2)	NA (C)	NA (2)	NA	NA	NA (C)	NA	NA	N	NV
	7440-61-1	Uranium-238	NA (2)	NA (2)	pCi/L	NA (2)	NA (C)	NA (2)	NA	NA	NA (C)	NA	NA	N Y	NV
	67-66-3	Chloroform	0.00061	0.0155	mg/L	802	49 / 96	0.0005 - 0.001	0.0155	ND	0.00019 (C)	0.08	MCL (TTHM) MCL	T	ASL
													(combined		
	13982-63-3	Radium-226 (3)	0.1	1	pCi/L	632	37 / 96	0.001 - 0.2	1	0.798	NA (C)	5	radium)	Y	DET
													MCL (combined		
	15262-20-1	Radium-228 (3)	0.3	4.3	pCi/L	632	22 / 96	0.04 - 1	4.3	1.611	NA (C)	5	radium)	N	NV
	14060 60 7	Thorium 220	0.0	1.6		630	10/06	0.2.0.2	16	0.500		45	MCL (gross alpha)	N	NIV/
Zone 1		Thorium-230	0.2	1.6	pCi/L	632 EBA07	10/96	0.2 - 0.2	1.6	0.509	NA (C)	15		N	NV
	7429-90-5		0.2	1.3	mg/L	EPA07	7/16	0.1 - 0.1	1.3	0.117	3.7 (N)	5	NMWQCC-I	N	BSL
Water Vapors	7440-38-2		0.001	0.003	mg/L	EPA07	4/16	0.001 - 0.001	0.003	0.00117	0.000045 (C)	0.01	MCL	N	NV NV
from	7440-48-4		0.02	0.06	mg/L	EPA05	16/16	0.01 - 0.01	0.06	0.0112	0.011 (N)	0.05	NMWQCC-I	N	NV
Showerhead /		Manganese	0.95	2.96	mg/L	EPA07	16/16	0.01 - 0.01	2.96	2.519	0.088 (N)	0.2	NMWQCC-O	N	NV
Domestic Use	7440-02-0		0.05	0.06	mg/L	EPA05	3/16	0.05 - 0.05	0.06	0.0602	0.073 (N)	0.2	NMWQCC-I	N	BSL
	7782-49-2		0.001	0.001	mg/L	EPA05	1/16	0.001 - 0.001	0.001	0.00107 ND	0.018 (N)	0.05	MCL	N	BSL
					mg/L	EPA07	1/16	0.1 - 0.1			0.018 (N)	NA	NA	N	NV
		Uranium	0.0012	0.0022	mg/L	EPA07	16 / 16	0.0003 - 0.0004	0.0022	0.0255	0.011 (N)	0.03	MCL	N	BSL
		Uranium-234	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (C)	NA	NA	NA (C)	NA	NA	N	NV
		Uranium-235	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (C)	NA	NA	NA (C)	NA	NA	N	NV
	7440-61-1	Uranium-238	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (C)	NA	NA	NA (C)	NA	NA	N	NV
	67-66-3	Chloroform	0.0006	0.00076	mg/L	EPA07	2/16	0.0005 -0.001	0.00076	NBC	0.00019 (C)	0.08	MCL (TTHM) MCL	Y	ASL
													(combined		
	13982-63-3	Radium-226 (3)	0.4	1.8	pCi/L	EPA05	13 / 16	0.2 - 0.2	1.8	1.314	NA (C)	5	radium)	Y	DET
													MCL (combined		
	15262-20-1	Radium-228 (3)	1	4	pCi/L	EPA05	8 / 16	1.0 - 1.0	4	2.946	NA (C)	5	radium)	N	NV
	14269-63-7	Thorium-230	0.6	0.7	pCi/L	EPA05	2/16	0.2 - 0.2	0.7	0.403	NA (C)	15	MCL (gross alpha)	N	NV
Zone 3	7429-90-5		0.0	163	•	EPA14	58/70	0.1 - 0.1	163	0.231	3.7 (N)	5	NMWQCC-I	N	NV
Water Vapors	7429-90-3		0.001	2.5 D	mg/L mg/L	NBL-01	48/70	0.001 - 0.03	2.5	0.231	0.000045 (C)	0.01	MCL	N	NV
from	7440-41-7		0.01	0.09	mg/L	EPA14	9/70	0.01 - 0.01	0.09	ND	0.0073 (N)	0.004	MCL	N	NV
Showerhead /	7440-43-9		0.005	1	mg/L	EPA14	16/70	0.005 - 0.005	1	0.0113	0.0018 (N)	0.005	MCL	N	NV
Domestic Use	7440-48-4		0.05	0.95 D	mg/L	EPA14	70 / 70	0.01 - 0.02	0.95	0.0877	0.0010 (N)	0.05	NMWQCC-I	N	NV
		Manganese	3.33	23.7	mg/L	717	70/70	0.01 - 0.01	23.7	3.436	0.088 (N)	0.2	NMWQCC-O	N	NV
		Molybdenum	0.1	5	mg/L	NBL-01	32 / 70	0.1 - 0.1	5	17.43	0.018 (N)	1	NMWQCC-I	N	NV
	7440-02-0		0.11	0.89	mg/L	EPA14	70 / 70	0.05 - 0.05	0.89	0.14	0.073 (N)	0.2	NMWQCC-I	Ν	NV
	7782-49-2	Selenium	0.001	0.01	mg/L	NBL-01	3 / 70	0.001 - 0.001	0.01	0.00159	0.018 (N)	0.05	MCL	Ν	NV
						517, 708, EPA13, EPA14,									
	7440-62-2	Vanadium	0.1	0.2	mg/L	NBL-01	5 / 70	0.1 - 0.1	0.2	ND	0.018 (N)	NA	NA	Ν	NV
	7440-61-1	Uranium	0.0011	0.138	mg/L	NBL-01	70 / 70	0.0003 - 0.002	0.138	0.107	0.011 (N)	0.03	MCL	Ν	NV
	13966-29-5	Uranium-234	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Ν	NV
	15117-96-1	Uranium-235	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Ν	NV
	7440-61-1	Uranium-238	NA (2)	NA (2)	pCi/L	NA (2)	NA (2)	NA (2)	NA	NA	NA (C)	NA	NA	Ν	NV
	67-66-3	Chloroform	0.00093	0.00676	mg/L	517	13/70	0.0005 -0.001	0.00676	NBC	0.00019 (C)	0.08	MCL (TTHM)	Y	ASL
													MCL (combined		
	13982-63-3	Radium-226	2	27.6	pCi/L	EPA14	70 / 70	0.1 - 0.2	27.6	4.996	NA (C)	5	radium)	Y	DET
													MCL (combined		
	15262-20-1	Radium-228	3.8	56.1	pCi/L	EPA14	70 / 70	1 - 1.4	56.1	4.509	NA (C)	5	(combined radium)	N	NV
													MCL (gross		
	14269-63-7	Thorium-230	0.2	1.3	pCi/L	517	6 / 70	0.2 - 0.2	1.3	1.426	NA (C)	15	alpha)	N	NV
		1	1			1	1	1	1				MCL (gross	1	
	14255-04-0	Lead-210	1.8	8.1	pCi/L	719	6 / 70	1.0 - 1.0	8.1	1.618	NA (C)	4 mrem/y	beta)	N	NV

(1) Qualifier codes used for the "Minimum Concentration" and "Maximum Concentration".

D = the sample was diluted to facilitate analysis.

- (2) Uranium isotopes not analyzed. Isotope concentrations esimated from total uranium mass concentration (see Table 3.A.RME).
- (3) The summary statistics for Radium-226 and Radium-228 count the raw below-detection-limit values from the 2nd Quarter 2008 as detections, because they were treated as such in the calculation of EPC statistics using ProUCL.
- (4) Maximum concentration used for screening chemicals. No screening was conducted for radionuclides. All detected radionuclides considered volatile, or with volatile decay products, are selected as COPCs (i.e., only Ra-226).
- (5) Background value calculated in N.A. Water Systems (2008b).
- (6) All compounds were screened against the November 2010 USEPA Risk

Screening Level Table (http://www.epa.gov/region9/superfund/prg/).

- For non-carcinogens: screening value = 0.1 x RSL tapwater value
- For carcinogens : screening value = RSL tapwater value
- (7) Rationale Codes:

Selection Reason Above Screening Level (ASL)
Detected in seepage-impacted groundwater at Site (DET)

Deletion Reason

Detected in seepage-impacted ground Below Screening Level (BSL) Not volatile (NV) NA = Not Applicable ND = Not Detected MCL = Maximum Contaminant level SMCL = Secondary Maximum Contaminant level NMWQCC = New Mexico Water Quality Control Commission Groundwater Standard NMWQCC-I = New Mexico Water Quality Control Commission Irrigation Standard NMWQCC-O = New Mexico Water Quality Control Commission Other Standard C = Carcinogen N = Noncarcinogen TTHM = 0.080 mg/l is the MCL for total trihalomethanes, of which chloroform is a single compound. NBC = No background concentration - chloroform was not detected frequently enough in Zone 1 and Zone 3

background samples (less than 1%) to calculate background concentration

TABLE 3.1.RME EXPOSURE POINT CONCENTRATION SUMMARY REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Medium: Groundwater

Exposure Medium: Groundwater

Exposure Point	Chemical of	Units	Arithmetic	95% UCL	Maximum Concentration		Exposu	re Point Concentrati	on
	Potential Concern		Mean	(Distribution)	(Qualifier)	Value	Units	Statistic	Rationale
			(of Detected)	(1)	(3)			(4)	(5)
SW Alluvium	Arsenic	mg/L	0.00885	0.00256 (NP)	0.01	0.00256	mg/l	95% UCL	KM (t)
Tapwater	Cobalt	mg/L	NA	NA	0.01	0.01	mg/l	Max	(a)
	Manganese	mg/L	1.865	2.8 (O)	5.4	2.8	mg/l	97.5% UCL	Chebyshev, (b)
	Uranium	mg/L	0.104	0.128 (NP)	0.246	0.128	mg/l	95% UCL	Chebyshev
	Uranium-234 (2)	pCi/L	NA	NA	NA	4.37E+01	pCi/L	NA	(e)
	Uranium-235 (2)	pCi/L	NA	NA	NA	1.99E+00	pCi/L	NA	(e)
	Uranium-238 (2)	pCi/L	NA	NA	NA	4.27E+01	pCi/L	NA	(e)
	Chloroform	mg/L	0.00479	0.00338 (NP)	0.0155	0.00338	mg/l	95% UCL	KM (%)
	Radium-226	pCi/L	0.435	0.267 (N)	1	0.267	pCi/L	95% UCL	KM (%)
	Radium-228	pCi/L	1.786	0.86 (N)	4.3	0.86	pCi/L	95% UCL	KM (%)
	Thorium-230	pCi/L	0.69	0.29 (O)	1.6	0.29	pCi/L	95% UCL	KM (t)
Zone 1	Arsenic	mg/L	0.00175	0.00145 (N)	0.003	0.00145	mg/l	95% UCL	KM (t)
Tapwater	Cobalt	mg/L	0.0363	0.0557 (NP)	0.06	0.0557	mg/l	95% UCL	Chebyshev
	Manganese	mg/L	1.656	1.95 (T)	2.96	1.95	mg/l	95% UCL	Modified t
	Vanadium	mg/L	NA	NA	0.2	0.2	mg/l	Max	(a)
	Uranium (f)	mg/L	0.00161	0.00174 (N)	0.0022	0.00174	mg/l	95% UCL	Modified t
	Uranium-234 (2)	pCi/L	NA	NA	NA	5.94E-01	pCi/L	NA	(e)
	Uranium-235 (2)	pCi/L	NA	NA	NA	2.71E-02	pCi/L	NA	(e)
	Uranium-238 (2)	pCi/L	NA	NA	NA	5.80E-01	pCi/L	NA	(e)
	Chloroform	mg/L	0.00068	0.000639 (NP)	0.00076	0.00068	mg/l	Mean	(c)
	Radium-226	pCi/L	1.138	1.213 (N)	1.8	1.213	pCi/L	95% UCL	KM (%)
	Radium-228	pCi/L	2.286	2.087 (N)	4	2.087	pCi/L	95% UCL	KM (t)
	Thorium-230	pCi/L	0.65	0.621 (NP)	0.7	0.65	pCi/L	Mean	(c)
Zone 3	Aluminum	mg/L	16.14	39.15 (NP)	163	39.15	mg/l	97.5% UCL	KM (Chebyshev), (b)
Tapwater	Arsenic	mg/L	0.206	0.412 (NP)	2.5 D	0.412	mg/l	95% UCL	KM (Chebyshev), (b)
	Beryllium	mg/L	0.0589	0.0202 (N)	0.09	0.0202	mg/l	95% UCL	KM (t)
	Cadmium	mg/L	0.0713	0.0628 (NP)	1	0.0628	mg/l	95% UCL	KM (BCA)
	Cobalt	mg/L	0.381	0.439 (O)	0.95 D	0.439	mg/l	95% UCL	Gamma
	Manganese	mg/L	9.836	10.89 (NP)	23.7	10.89	mg/l	95% UCL	Modified t
	Molybdenum	mg/L	1.084	0.739 (NP)	5	0.739	mg/l	95% UCL	KM (BCA)
	Nickel	mg/L	0.377	0.489 (NP)	0.89	0.489	mg/l	95% UCL	Chebyshev
	Vanadium	mg/L	0.18	0.111 (NP)	0.2	0.18	mg/l	Mean	(d)
	Uranium	mg/L	0.0287	0.0431 (NP)	0.138	0.0431	mg/l	95% UCL	Chebyshev
	Uranium-234 (2)	pCi/L	NA	NA	NA	1.47E+01	pCi/L	NA	(e)
	Uranium-235 (2)	pCi/L	NA	NA	NA	6.71E-01	pCi/L	NA	(e)
	Uranium-238 (2)	pCi/L	NA	NA	NA	1.44E+01	pCi/L	NA	(e)
	Chloroform	mg/L	0.00441	0.00326 (N)	0.00676	0.00326	mg/l	95% UCL	KM (%)
	Radium-226	pCi/L	9.823	11.14 (O)	27.6	11.14	pCi/L	95% UCL	Gamma
	Radium-228	pCi/L	15.73	17.84 (O)	56.1	17.84	pCi/L	95% UCL	Gamma
	Thorium-230	pCi/L	0.533	0.259 (N)	1.3	0.259	pCi/L	95% UCL	KM (t)
	Lead-210	pCi/L	4.883	2.287 (N)	8.1	2.287	pCi/L	95% UCL	KM (t)

Notes:

(1) "95% UCL" term calculated using ProUCL ver 4.00.02 as described in N.A.Water Systems (2008c)

Codes describing the type of distribution for the "95% UCL" term.

(5) Codes describing the rationale the statistic is used to represent the EPC

KM (t) = UCL based upon Kaplan-Meier estimates using the Student's t-distribution cutoff value

KM (%) = UCL based upon Kaplan-Meier estimates using the percentile bootstrap method

n. Chebyshev = Chebyshev inequality-based UCL

N = Normal

T = Transformed (lognormal)

NP = Nonparametric

O = Other

(2) Uranium isotopes not analyzed. Isotope concentrations esimated

from total uranium mass concentration (see Table 3.A.RME).

(3) Qualifier codes used for the "Maximum Concentration".

D = the sample was diluted to facilitate analysis.

(4) Codes describing the "EPC Statistic".

95% UCL = 95% UCL Statistic

97.5% UCL = 97.5% UCL Statistic

Max = Maximum Detected Concentration

Mean = Arithmetic Mean of Detected Concetnrations

Definitions:

NA = Not Applicable

Modified t = UCL based on the modified t statistic

KM (Chebyshev) = UCL based upon Kaplan-Meier estimates using the Chebyshev inequality

KM (BCA) = UCL based upon Kaplan-Meier estimates using bias-corrected accelerated bootstrap method Gamma = 95% Approximate Gamma UCL

(a) Only one detection, 95% UCL not calculated, used maximum detected concentration.

(b) UCL calculated at 97.5% confidence level.

(c) Only two detections. 95% UCL of questionable reliability; used arithmetic mean of detected values.

(d) 95% UCL of questionable reliability; used arithmetic mean of detected values.

- (e) Uranium isotopes not analyzed. Isotope concentrations esimated from total uranium mass concentration (see Table 3.A.RME)
- (f) The total uranium mass concentration was screened out of the Zone 1 risk evaluation but appears in this table because it is the basis of the uranium isotope actitivies shown.

TABLE 3.2.RME EXPOSURE POINT CONCENTRATION SUMMARY REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future

Medium: Groundwater

Exposure Medium: Air

Exposure Point	Chemical of	Units	Arithmetic Mean	95% UCL (Distribution) (1)	Maximum Concentration	Exposure Point Concentration				
	Potential Concern				(Qualifier)	Value	Units	Statistic (2)	Rationale (3)	
SW Alluvium Water Vapors	Chloroform	mg/L	0.00479	0.00338 (NP)	0.0155	0.00338	mg/l	95% UCL	KM (%)	
from Showerhead / Domestic Use	Radium-226	pCi/L	0.435	0.267 (N)	1	0.267	pCi/L	95% UCL	KM (%)	
Zone 1 Water Vapors	Chloroform	mg/L	0.00068	0.000639 (NP)	0.00076	0.00068	mg/l	Mean	(a)	
from Showerhead / Domestic Use	Radium-226	pCi/L	1.138	1.213 (N)	1.8	1.213	pCi/L	95% UCL	KM (%)	
Zone 3 Water Vapors	Chloroform	mg/L	0.00441	0.00326 (N)	0.00676	0.00326	mg/l	95% UCL	KM (%)	
from Showerhead / Domestic Use	Radium-226	pCi/L	9.823	11.14 (O)	27.6	11.14	pCi/L	95% UCL	Gamma	

Notes:

(1) "95% UCL" term calculated using ProUCL ver 4.00.02 as described

in N.A.Water Systems (2008c).

Codes describing the type of distribution for the "95% UCL" term:

N = Normal

NP = Nonparametric

O = Other

(2) Codes describing the "EPC Statistic":

95% UCL = 95% UCL Statistic

97.5% UCL = 97.5% UCL Statistic

Max = Maximum Detected Concentration

Mean = Arithmetic Mean of Detected Concetnrations

(3) Codes describing the rationale the statistic is used to represent the EPC:

KM (%) = UCL based upon Kaplan-Meier estimates using the percentile bootstrap method

Gamma = 95% Approximate Gamma UCL

(a) Only two detections. 95% UCL lower than arithmetic mean of detected and of questionable reliability; used arithmetic mean.

Table 3.A.RME (SUPPLEMENTAL) ESTIMATION OF URANIUM ISOTOPE CONCENTRATIONS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

					Isotope Mass Concentration at		Isotope Activity (pCi/L) at
	Uranium		Mass-Based Uranium	Uranium Isotope Natural	EPC Concentration	Isotope Specific Activity	, , ,
Hydrostratigraphic Unit	Isotopes		EPC Concentration (mg/l)	Abundance (%) (1)	(mg/l)	(pCi/µg)	Concentration (2)
SWA	U-234	mg/L	0.128	0.000055	0.000007	6209	4.37E+01
SWA	U-235	mg/L	0.128	0.00720	0.00092	2.161	1.99E+00
SWA	U-238	mg/L	0.128	0.99270	0.12707	0.336	4.27E+01
Zone 1	U-234	mg/L	0.00174	0.000055	0.0000001	6209	5.94E-01
Zone 1	U-235	mg/L	0.00174	0.00720	0.00001	2.161	2.71E-02
Zone 1	U-238	mg/L	0.00174	0.99270	0.00173	0.336	5.80E-01
Zone 3	U-234	mg/L	0.0431	0.000055	0.0000024	6209	1.47E+01
Zone 3	U-235	mg/L	0.0431	0.00720	0.00031	2.161	6.71E-01
Zone 3	U-238	mg/L	0.0431	0.99270	0.04279	0.336	1.44E+01
Zone 3 (Background)	U-234	mg/L	0.107	0.000055	0.0000059	6209	3.65E+01
Zone 3 (Background)	U-235	mg/L	0.107	0.00720	0.00077	2.161	1.67E+00
Zone 3 (Background)	U-238	mg/L	0.107	0.99270	0.10622	0.336	3.57E+01

Notes:

(1) Source of natural abundance percentages: http://www.epa.gov/radiation/radionuclides/uranium.html

(2) Isotope activity (pCi/L) = Isotope Mass Concentration at EPC concentration (mg/L) x Isotope Specific Activity (pCi/µg) x Conversion Factor (1000 µg/mg)

TABLE 4.1.RME VALUES USED FOR DAILY INTAKE CALCULATIONS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future
Medium: Groundwater
Exposure Medium: Groundwater

F

Exposure Route	Receptor Population	Receptor Age	Exposure Point	Parameter	Parameter Definition	Value	Units	Rationale/	Intake Equation/
				Code				Reference (1)	Model Name (2)
Ingestion	Resident	Adult	SW Alluvium Zone 1 and	CW IRw	Chemical Concentration in Water Ingestion Rate of Water	See Table 3.1 2	mg/l I/day	See Table 3.1 EPA, 1991	Chronic Daily Intake (CDI) (mg/kg/day) = CW x IRw x EF x ED x 1/BW x 1/(ATn)
			Zone 3	EF	Exposure Frequency	350	days/year	EPA, 1991	
			Tapwater	ED BW	Exposure Duration Body Weight	24 70	years kg	EPA, 1991 EPA, 1991	
				ATn	Averaging Time - Non-Cancer	8,760	days	EPA, 1989	
		Child	SW Alluvium Zone 1 and	CW	Chemical Concentration in Water	See Table 3.1	mg/l	See Table 3.1 CalEPA, 1994	CDI (mg/kg/day) =
			Zone 1 and Zone 3	IRw EF	Ingestion Rate of Water Exposure Frequency	1 350	l/day days/year	EPA, 1994 EPA, 1991	CW x IRw x EF x ED x 1/BW x 1/(ATn)
			Tapwater	ED	Exposure Duration	6	years	EPA, 1991	
				BW ATn	Body Weight Averaging Time - Non-Cancer	15 2,190	kg days	EPA, 1991 EPA, 1989	
		Adult/Child	SW Alluvium	CW	Chemical Concentration in Water	See Table 3.1	mg/l	See Table 3.1	CDI (mg/kg/day) =
			Zone 1 and	IRwc	Ingestion Rate of Water - Child	1	l/day	CalEPA, 1994	CW x IRwadj x EF x 1/(ATc)
			Zone 3 Tapwater	IRwa IRwadj	Ingestion Rate of Water - Adult Ingestion Rate of Water - Age-Adjusted	2 1.09	l/day [L*yr]/[kg*d]	EPA, 1991 Calculated	IRwadj ([L*yr]/[kg*d])=
				EF	Exposure Frequency	350	days/year	EPA, 1991	EDc x IRwc x (1/BWc) + EDa x IRwa x (1/BWa)
				EDc	Exposure Duration - Child	6	years	EPA, 1991	
				EDa BW	Exposure Duration - Adult Body Weight - Child	24 15	years kg	EPA, 1991 EPA, 1991	
				BW	Body Weight - Adult	70	kg	EPA, 1991	
				ATc CWR	Averaging Time - Cancer	25,550	days	EPA, 1989	
				IRwadjR	Radionuclide Concentration in Water Ingestion Rate of Water - Adjusted - Radionuclides	See Table 3.1 1.8	pCi/l I/day	See Table 3.1 Calculated	Intake (pCi) = CWR x IRWadjR x EF x ED
				IRwc	Ingestion Rate of Water - Child	1	l/day	CalEPA, 1994	IRwadjR (L/d)= [EDc x IRwc + EDa x IRwa]/ED
				IRwa EF	Ingestion Rate of Water - Adult Exposure Frequency	2	l/day	EPA, 1991	
				EF EDc	Exposure Frequency Exposure Duration - Child	350 6	days/year years	EPA, 1991 EPA, 1991	
				EDa	Exposure Duration - Adult	24	years	EPA, 1991	
Dermal	Resident	Adult	SW Alluvium	ED CW	Exposure Duration Chemical Concentration in Water	30 See Table 3.1	years mg/l	EPA, 1991 See Table 3.1	Dermally Absorbed Dose (DAD) (mg/kg-day) =
Definidi	Resident	Addit	Zone 1 and	DA-event	Absorbed Dose per Event	Chemical Specific	mg/cm2-event	Calculated	DA-event x EV x ED x EF x SA x 1/BW x 1/(ATn)
			Zone 3	FA	Fraction Absorbed Water	Chemical Specific		EPA, 2004	where, for organic compounds with t-event $\leq t^*$:
			Tapwater	Kp SA	Permeability Constant Skin Surface Area	Chemical Specific 18000	cm/hr cm2	EPA, 2004 EPA, 2004	Absorbed Dose per Event (DA-event) (mg/cm2-event) = 2 FA x Kp x CW x CF x SQRT{(6 x tau-event x t-event)/
				tau-event	Lag time per event	Chemical Specific	hr/event	EPA, 2004	or
				t-event	Event Duration	0.58	hr/event	EPA, 2004	where, for organic compounds with t-event $> t^*$:
				t* B	Time to reach steady-state (hr) Ratio of permeability coefficient of a	Chemical Specific Chemical Specific	hrs 	EPA, 2004	DA-event = FA x Kp x CW x {(t-event/(1 + B)) + 2 x tau-event x ((1 + (3 x B) + (3 x B x B))/(1 + B)2)}
					compound through the stratum			,	and
					corneum relative to its permeability				where, for inorganic compounds,
				EV	coefficient across the viable epidermis Event Frequency	1	events/day	EPA, 2004	DA-event = Kp x CW x CF x t-event (see Tables 7.A.RME and 7.B.RME)
				EF	Exposure Frequency	350	days/year	EPA, 2004	(See Tables F.A.Rivie and F.B.Rivie)
				ED	Exposure Duration	24	years	EPA, 1991	
				CF BW	Volumetric Conversion Factor for Water Body Weight	0.001 70	l/cm3 kg	 EPA, 2004	
				ATn	Averaging Time - Non-Cancer	8,760	days	EPA, 2004	
		Child	SW Alluvium	CW	Chemical Concentration in Water	See Table 3.1	mg/l	See Table 3.1	Dermally Absorbed Dose (DAD) (mg/kg-day) =
			Zone 1 and Zone 3	DA-event FA	Absorbed Dose per Event Fraction Absorbed Water	Chemical Specific Chemical Specific	mg/cm2-event	Calculated EPA, 2004	DA-event x EV x ED x EF x SA x 1/BW x 1/(ATn) where, for organic compounds with t-event <= t*:
			Tapwater	Кр	Permeability Constant	Chemical Specific	cm/hr	EPA, 2004	Absorbed Dose per Event (DA-event) (mg/cm2-event)
				SA	Skin Surface Area	6600	cm2	EPA, 2004	2 FA x Kp x CW x CF x SQRT{(6 x tau-event x t-event)/
				tau-event	Lag time per event Event Duration	Chemical Specific	hr/event hr/event	EPA, 2004 EPA, 2004	or where, for organic compounds with t-event > t*:
				t*	Time to reach steady-state (hr)	Chemical Specific	hr	EPA, 2004	DA-event = FA x Kp x CW x {(t-event/(1 + B)) +
				В	Ratio of permeability coefficient of a compound through the stratum	Chemical Specific		EPA, 2004	2 x tau-event x ((1 + (3 x B) + (3 x B x B))/(1 + B)2)} and
					corneum relative to its permeability				where, for inorganic compounds,
					coefficient across the viable epidermis				DA-event = Kp x CW x CF x t-event
				EV EF	Event Frequency Exposure Frequency	1 350	events/day days/year	EPA, 2004 EPA, 2004	(see Tables 7.A.RME and 7.B.RME)
				ED	Exposure Duration	6	years	EPA, 2004	
				CF	Volumetric Conversion Factor for Water	0.001	I/cm3		
				BW ATn	Body Weight Averaging Time - Non-Cancer	15 2,190	kg days	EPA, 2004 EPA, 2004	
		Adult/Child	SW Alluvium	CW	Chemical Concentration in Water	See Table 3.1	mg/l	See Table 3.1	Dermally Absorbed Dose (DAD) (mg/kg-day) =
			Zone 1 and	DA-event-a	Absorbed Dose per Event - Adult	Chemical Specific	mg/cm2-event	Calculated	[(DA-event-a x EV x EDa x EF x SAa x 1/Bwa) +
			Zone 3 Tapwater	DA-event-c FA	Absorbed Dose per Event - Child Fraction Absorbed Water	Chemical Specific Chemical Specific	mg/cm2-event	Calculated EPA, 2004	[(DA-event-c x EV x EDc x EF x Sac x 1/Bwc)] x 1/(AT where, for organic compounds with t-event <= t*:
				Кр	Permeability Constant	Chemical Specific	cm/hr	EPA, 2004	Absorbed Dose per Event (DA-event) (mg/cm2-event)
				SAc SAa	Skin Surface Area - Child Skin Surface Area - Adult	6600 18000	cm2 cm2	EPA, 2004 EPA, 2004	2 FA x Kp x CW x CF x SQRT{(6 x tau-event x t-event) or
				tau-event	Lag time per event	Chemical Specific	hrs/event	EPA, 2004 EPA, 2004	where, for organic compounds with t-event > t*:
				t-event-a	Event Duration - Adult	0.58	hrs/event	EPA, 2004	DA-event = FA x Kp x CW x {(t-event/(1 + B)) +
				t-event-c	Event Duration - Child Time to reach steady-state (hr)	1 Chemical Specific	hrs/event hr	EPA, 2004 EPA, 2004	2 x tau-event x ((1 + (3 x B) + (3 x B x B))/(1 + B)2)} and
				В	Ratio of permeability coefficient of a	Chemical Specific		EPA, 2004 EPA, 2004	where, for inorganic compounds,
		1			compound through the stratum				DA-event = Kp x CW x CF x t-event
					corneum relative to its permeability				(see Tables 7.A.RME and 7.B.RME)
				EV	coefficient across the viable epidermis Event Frequency	1	events/day	EPA, 2004	Note: Adult and Child DA-event values are calculated us
				EF	coefficient across the viable epidermis	1 350	events/day days/year	EPA, 2004 EPA, 2004	Note: Adult and Child DA-event values are calculated u corresponding Adult or Child t-event value
				EF EDa	coefficient across the viable epidermis Event Frequency Exposure Frequency Exposure Duration - Adult		days/year years	EPA, 2004 EPA, 2004	
				EF	coefficient across the viable epidermis Event Frequency Exposure Frequency	350	days/year	EPA, 2004	
				EF EDa EDc	coefficient across the viable epidermis Event Frequency Exposure Frequency Exposure Duration - Adult Exposure Duration - Child	350 24 6	days/year years years	EPA, 2004 EPA, 2004 EPA, 2004	

Notes:

Dermal uptake is generally not an important route of uptake for radionuclides, which have small permeability constants (EPA, 1989a). Dermal uptake of radionuclides is not evaluated in this risk assessment.

(1) References:

California EPA (CalEPA). 1994. Preliminary Endangerment Assessment Guidance Manual, (PEA) Department of Toxic Substances Control, Sacramento, California.

EPA 1989: Risk Assessment Guidance for Superfund. Volume 1: Human Health Evaluation Manual, Part A. OERR EPA/540/1-89/002.

EPA 1991: Risk Assessment Guidance for Superfund. Volume 1: Human Health Evaluation Manual - Supplemental Guidance, Standard Default Exposure Factors. Interim Final. OSWER 9285.6-03.

EPA 2004: Risk Assessment Guidance for Superfund. Volume 1: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) Final.

(2) Intake for adults and children combined/adjusted for cancer risk calculations based on exposure duration.

TABLE 4.2.RME VALUES USED FOR DAILY INTAKE CALCULATIONS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Medium: Groundwater Exposure Medium: Air

Exposure Route	Receptor Population	Receptor Age	Exposure Point	Parameter Code	Parameter Definition	Value	Units	Rationale/ Reference	Intake Equation/ Model Name
Inhalation (1)	Resident	Adult	Water Vapors From Showerhead	(1)	(1)	(1)	(1)	(1)	Foster and Chrostowski Model
Inhalation (1)	Resident	Child	Water Vapors From Showerhead	(1)	(1)	(1)	(1)	(1)	Foster and Chrostowski Model
Inhalation (1)	Resident	Child/Adult	Water Vapors From Showerhead	(1)	(1)	(1)	(1)	(1)	Foster and Chrostowski Model
Inhalation (2)	Resident	Child/Adult	Water Vapors From Showerhead and Other Domestic Uses	(2)	(2)	(2)	(2)	(2)	Andelman Volatilization Factor

Footnote Instructions:

(1) Refer to Tables 7.C.RME and 7.D.RME and the Risk Assessment text for details on the modeled intake methodology and parameters used to calculate modeled intake values for the Foster and Chrostowski Shower Model. (2) Refer to the Table 7.E.RME and the Risk Assessment text for details on the modeled intake methodology and parameters used to calculate modeled intake values for the Andelman Volatilization Factor.

TABLE 5.1 NON-CANCER TOXICITY DATA -- ORAL/DERMAL UNC Church Rock Mill and Tailings Site

Chemical of Potential	Chronic/ Subchronic	Oral	RfD	Oral Absorption Efficiency for Dermal	Absorbed RfD	for Dermal (2)	Primary Target	Combined Uncertainty/Modifying	RfD:Targe	et Organ(s) (3)
Concern		Value	Units		Value	Units	Organ(s)	Factors	Source(s)	Date(s)
				(1)			Or Effects			(MM/DD/YYYY)
Aluminum	Chronic	1.0E+00	mg/kg-day	1	1.0E+00	mg/kg-day	Central nervous system	100	PPRTV	12/13/2010
Arsenic	Chronic	3.0E-04	mg/kg-day	1	3.0E-04	mg/kg-day	Skin	3	IRIS	12/13/2010
Beryllium	Chronic	2.0E-03	mg/kg-day	0.007	1.4E-05	mg/kg-day	Gastrointestinal	300	IRIS	12/13/2010
Cadmium	Chronic	5.0E-04	mg/kg-day	0.05	2.5E-05	mg/kg-day	Kidney	10	IRIS	12/13/2010
Chloroform	Chronic	1.0E-02	mg/kg-day	1	1.0E-02	mg/kg-day	Liver	100	IRIS	12/13/2010
Cobalt	Chronic	3.0E-04	mg/kg-day	1	3.0E-04	mg/kg-day	Thyroid	1000	PPRTV	08/25/2008
Lead-210	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Manganese	Chronic	2.4E-02	mg/kg-day	0.04	9.6E-04	mg/kg-day	Central nervous system	3	RSL (4)	11/2010
Molybdenum	Chronic	5.0E-03	mg/kg-day	1	5.0E-03	mg/kg-day	Increased uric acid (kidney)	30	IRIS	12/13/2010
Nickel	Chronic	2.0E-02	mg/kg-day	0.04	8.0E-04	mg/kg-day	Reduced organ and body weights	300	IRIS	12/13/2010
Radium-226	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Radium-228	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Thorium-230	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Uranium	Chronic	3.0E-03	mg/kg-day	1	3.0E-03	mg/kg-day	Kidney	1000	IRIS	12/13/2010
Uranium-234	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Uranium-235	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Uranium-238	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Vanadium	Chronic	5.0E-03	mg/kg-day	1	5.0E-03	mg/kg-day	Decreased hair cystine	100	RSL (5) (6)	12/13/2010

Notes:

(1) EPA 2004: Risk Assessment Guidance for Superfund. Volume 1: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) Final, Section 4.2 and Exhibit 4-1, except for vanadium, which was based on the GIABS value listed in the EPA RSL Table for "Vanadium and Compounds".

(2) Absorbed RfD for Dermal calculated by multiplying Oral RfD by Oral Absorption Efficiency for Dermal.

(3) RfD Date: The date shown for IRIS values is the date IRIS was searched; for PPRTV values, the date of the PPRTV report; for RSL values, the date of the RSL Table.

(4) Manganese RfD obtained from November 2010 EPA RSL Table; based on IRIS assessment recommendation to subtract dietary contribution from IRIS RfD when evaluating non-food exposures (e.g., drinking water or soil).

(5) Vanadium oral RfD obtained from November 2010 EPA RSL Table for "Vanadium and Compounds"; derived from IRIS oral RfD for vanadium pentoxide by factoring out the molecular wt of the oxide ion.

(6) Vanadium oral RfD for decreased hair cystine interpreted to have a metabolic system endpoint.

Definitions:

NA = Not Applicable

IRIS = Integrated Risk Information System

RSL = Regional Screening Level Table

PPRTV = Provisional Peer Reviewed Toxicity Value, Superfund Health Technical Support Center

TABLE 5.2 NON-CANCER TOXICITY DATA -- INHALATION UNC Church Rock Mill and Tailings Site

Chemical of Potential	Chronic/ Subchronic	Inhalation RfC		Primary Target	Combined Uncertainty/Modifying	RfC : Target Organ(s) (1)		
Concern		Value	Units	Organ(s)	Factors	Source(s)	Date(s) (MM/DD/YYYY)	
Chloroform Radium-226	Chronic NA	9.80E-02 NA	mg/m3 NA	Liver NA	100 NA	RSL/ATSDR NA	12-2010 / 09-1997 NA	

Notes:

(1) RfC Date: The date shown for RSL values, the date of the RSL Table; for ATSDR values (MRLs), the date of the ATSDR report.

Definitions:

NA = Not Applicable

RSL = Regional Screening Level Table

ATSDR = Agency for Toxic Substances and Disease Registry

TABLE 5.3 NON-CANCER TOXICITY DATA -- SPECIAL CASE CHEMICALS UNC Church Rock Mill and Tailings Site

Chemical of Potential	Chronic/ Subchronic	Parameter			Primary Target Organ(s)	Combined Uncertainty/Modifying	Parameter:Target Organ(s)	
Concern		Name	Value	Units		Factors	Source(s)	Date(s) (MM/DD/YYYY)
		NOT	APF	PLIC	ABLE			

TABLE 6.1 CANCER TOXICITY DATA - ORAL/DERMAL

UNC Church	Rock	Mill	and	Tailings	Site
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Chemical of Potential	Oral Cancer Slope Factor		Oral Absorption Efficiency for Dermal	Absorbed Cancer Slope Factor for Dermal		Weight of Evidence/ Cancer Guideline	Oral CSF	
Concern	Value	Units	(1)	Value (2)	Units	Description	Source(s)	Date(s) (MM/DD/YYYY)
Aluminum	NA	NA	NA	NA	NA	NA	NA	NA
Arsenic	1.50E+00	(mg/kg-day) ⁻¹	1	1.5E+00	(mg/kg-day) ⁻¹	А	IRIS	12/12/2010
Beryllium	NA	NA	NA	NA	NA	NA	NA	NA
Cadmium	NA	NA	NA	NA	NA	NA	NA	NA
Chloroform	3.10E-02	(mg/kg-day) ⁻¹	1	3.1E-02	(mg/kg-day) ⁻¹	B2	CalEPA	09/01/1990
Cobalt	NA	NA	NA	NA	NA	NA	NA	NA
Lead-210+D (Water Ingestion)	1.27E-09	Risk/pCi	NA	NA	NA	А	HEAST	04/16/2001
Manganese	NA	NA	NA	NA	NA	NA	NA	NA
Molybdenum	NA	NA	NA	NA	NA	NA	NA	NA
Nickel	NA	NA	NA	NA	NA	NA	NA	NA
Radium-226+D (Water Ingestion)	3.86E-10	Risk/pCi	NA	NA	NA	А	HEAST	04/16/2001
Radium-228+D (Water Ingestion)	1.04E-09	Risk/pCi	NA	NA	NA	А	HEAST	04/16/2001
Thorium-230 (Water Ingestion)	9.10E-11	Risk/pCi	NA	NA	NA	А	HEAST	04/16/2001
Uranium	NA	NA	NA	NA	NA	NA	NA	NA
Uranium-234 (Water Ingestion)	7.07E-11	Risk/pCi	NA	NA	NA	А	HEAST	04/16/2001
Uranium-235+D (Water Ingestion)	7.18E-11	Risk/pCi	NA	NA	NA	А	HEAST	04/16/2001
Uranium-238+D (Water Ingestion)	8.71E-11	Risk/pCi	NA	NA	NA	А	HEAST	04/16/2001
Vanadium	NA	NA	NA	NA	NA	NA	NA	NA

Notes:

(1) EPA 2004: Risk Assessment Guidance for Superfund. Volume 1: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) Final, Section 4.2 and Exhib

(2) Absorbed Cancer Slope Factor for Dermal calculated by multiplying Oral Cancer Slope Factor by Oral Absorption Efficiency for Dermal.

(3) Oral CSF Date: The date shown for IRIS values is the date IRIS was searched; for CalEPA values, date of CalEPA document; for HEAST values, date of HEAST document.

Definitions:

NA = Not Applicable

CalEPA = California EPA (identified on EPA RSL Table)

HEAST = Health Effects Summary Tables

IRIS = Integrated Risk Information System

RSL = Regional Screening Level Table

A = Known human carcinogen

B2 = Probable human carcinogen - indicates sufficient evidence in animals and inadequate or no evidence in humans

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TABLE 6.2 CANCER TOXICITY DATA - INHALATION UNC Church Rock Mill and Tailings Site

Chemical of Potential	Unit Risk (1)			er Slope Factor 2)	Weight of Evidence/ Cancer Guideline	Unit Risk : Inhalation CSF		
Concern	Value	Units	Value	Units	Description	Source(s)	Date(s) (MM/DD/YYYY)	
Chloroform Radium-226+D	2.3E-05 NA	(µg/m ³) ⁻¹ NA	8.1E-02 1.16E-08	(mg/kg-day) ⁻¹ risk/pCl	B2 A	IRIS HEAST	12/13/2010 04/16/2001	

(1) Inhalation Unit Risk (IUR) values used in risk calculations for chloroform

(2) Inhalation cancer slope factor used in risk calculations for radium-226+D

Definitions:

NA = Not Applicable

HEAST = Health Effects Summary Tables

IRIS = Integrated Risk Information System

A = Known human carcinogen

B2 = Probable human carcinogen - indicates sufficient evidence in animals and inadequate or no evidence in humans

TABLE 6.3 CANCER TOXICITY DATA - SPECIAL CASE CHEMICALS UNC Church Rock Mill and Tailings Site

Chemical of Potential		Parameters	Source(s)	Date(s) (MM/DD/YYYY)	
Concern	Name	Value	Units		
	NOT AP	PLICAE	SLE		

TABLE 6.4 CANCER TOXICITY DATA - EXTERNAL (RADIATION) UNC Church Rock Mill and Tailings Site

Chemical of Potential Concern	Cancer SI Value	ope Factor Units	Source(s)	Date(s) (MM/DD/YYYY)
	NOT A	PPLIC	ABLE	

TABLE 7.1.RME CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Future
Resident
Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	E	PC		Cance	er Risk Calculati	ons			Non-Cano	cer Hazard Cal	culations		
				Potential Concern	Value	Units	Intake/Exposure	Concentration	CSF/Unit Risk		Cancer Risk	Intake/Exposure Concentration		RfD/RfC		Hazard Quotier	
					(1)		Value	Units	Value	Units		Value	Units	Value	Units		
Groundwater	Groundwater	Tapwater	Ingestion	Arsenic	0.00256	mg/l	NA	NA	NA	NA	NA	7.01E-05	mg/kg/d	3.0E-04	mg/kg/d	0.23	
SW Alluvium				Cobalt	0.01	mg/l	NA	NA	NA	NA	NA	2.74E-04	mg/kg/d	3.0E-04	mg/kg/d	0.91	
				Manganese	2.8	mg/l	NA	NA	NA	NA	NA	7.67E-02	mg/kg/d	2.4E-02	mg/kg/d	3.2	
				Uranium	0.128	mg/l	NA	NA	NA	NA	NA	3.51E-03	mg/kg/d	3.0E-03	mg/kg/d	1.2	
				Chloroform	0.00338	mg/l	NA	NA	NA	NA	NA	9.26E-05	mg/kg/d	1.0E-02	mg/kg/d	0.0091	
			Exp. Route Total								NA					5.5	
			Dermal	Arsenic	0.00256	mg/l	NA	NA	NA	NA	NA	3.7E-07	mg/kg/d	3.0E-04	mg/kg/d	0.0012	
				Cobalt	0.01	mg/l	NA	NA	NA	NA	NA	1.4E-06	mg/kg/d	3.0E-04	mg/kg/d	0.0048	
				Manganese	2.8	mg/l	NA	NA	NA	NA	NA	4.0E-04	mg/kg/d	9.6E-04	mg/kg/d	0.42	
				Uranium	0.128	mg/l	NA	NA	NA	NA	NA	1.8E-05	mg/kg/d	3.0E-03	mg/kg/d	0.0061	
				Chloroform	0.00338	mg/l	NA	NA	NA	NA	NA	8.4E-06	mg/kg/d	1.02E-02	mg/kg/d	0.0008	
			Exp. Route Total								NA					0.43	
		Exposure Point Total									NA					6.0	
	Exposure Medium Total	Exposure Medium Total									NA					6.0	
	Air	Water Vapors from Showerhead	Inhalation (1)	Chloroform	0.0070	mg/m3	NA	NA	NA	NA	NA	1.6E-04	mg/m3	9.8E-02	mg/m3	0.0017	
			Exp. Route Total								NA					0.0017	
		Exposure Point Total									NA					0.0017	
	Exposure Medium Total										NA					0.0017	
oundwater Total - S	W Alluvium										NA					6.0	
									Total of Receptor Risks Across All Media					Total of Receptor Hazards Across All Media			

Notes:

(1) Inhalation EPC represents chemical air concentration (mg/m3) calculated from shower exposure model in Table 7.C.RME (Supplemental)

Definitions: NA = Not Applicable

TABLE 7.2.RME CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	EF	Ň		Cano	er Risk Calculat	ions		Non-Cancer Hazard Calculations				
				Potential Concern	Value	Units	Intake/Exposure	e Concentration	Concentration CSF/Uni		Cancer Risk	Intake/Exposure Concentration		RfD/RfC		Hazard Quotier
					(1)	(1)	Value	Units	Value	Units		Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Arsenic	0.00256	mg/l	NA	NA	NA	NA	NA	1.64E-04	mg/kg/d	3.0E-04	mg/kg/d	0.55
SW Alluvium				Cobalt	0.01	mg/l	NA	NA	NA	NA	NA	6.39E-04	mg/kg/d	3.0E-04	mg/kg/d	2.1
				Manganese	2.8	mg/l	NA	NA	NA	NA	NA	1.79E-01	mg/kg/d	2.4E-02	mg/kg/d	7.5
				Uranium	0.128	mg/l	NA	NA	NA	NA	NA	8.18E-03	mg/kg/d	3.0E-03	mg/kg/d	2.7
				Chloroform	0.00338	mg/l	NA	NA	NA	NA	NA	2.16E-04	mg/kg/d	1.0E-02	mg/kg/d	0.021
			Exp. Route Total								NA					12.9
			Dermal	Arsenic	0.00256	mg/l	NA	NA	NA	NA	NA	1.1E-06	mg/kg/d	3.0E-04	mg/kg/d	0.0036
				Cobalt	0.01	mg/l	NA	NA	NA	NA	NA	4.2E-06	mg/kg/d	3.0E-04	mg/kg/d	0.014
				Manganese	2.8	mg/l	NA	NA	NA	NA	NA	1.2E-03	mg/kg/d	9.6E-04	mg/kg/d	1.2
				Uranium	0.128	mg/l	NA	NA	NA	NA	NA	5.4E-05	mg/kg/d	3.0E-03	mg/kg/d	0.018
				Chloroform	0.00338	mg/l	NA	NA	NA	NA	NA	1.9E-05	mg/kg/d	1.02E-02	mg/kg/d	0.0019
			Exp. Route Total								NA					1.3
		Exposure Point Total									NA					14.2
	Exposure Medium Total	Exposure Medium Total									NA					14.2
	Air	Water Vapors from Showerhead	Inhalation	Chloroform	0.010	mg/m3	NA	NA	NA	NA	NA	4.1E-04	mg/m3	9.8E-02	mg/m3	0.0041
			Exp. Route Total								NA					0.0041
		Exposure Point Total									NA					0.0041
	Exposure Medium Total										NA					0.0041
oundwater Total - S	W Alluvium										NA					14.2
								Total of Rece	eptor Risks Acro	ss All Media	NA		Total of Recept	or Hazards Ac	ross All Media	14.2

Notes:

(1) Inhalation EPC represents chemical air concentration (mg/m3) calculated from shower exposure model in Table 7.D.RME (Supplemental)

Definitions:

NA = Not Applicable

TABLE 7.3.RME CALCULATION OF CHEMICAL CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	EI	PC		Canc	er Risk Calcula	tions		Non-Cancer Hazard Calculations				
				Potential Concern	Value	Units	Intake/Exposure	Concentration	CSF/L	nit Risk	Cancer Risk	Intake/Exposure Concentration		n RfD/RfC		Hazard Quotie
					(1)		Value	Units	Value	Units		Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Arsenic	0.00256	mg/l	3.82E-05	mg/kg/d	1.5E+00	(mg/kg-day)-1	5.7E-05	NA	NA	NA	NA	NA
SW Alluvium				Cobalt	0.01	mg/l	1.49E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	2.8	mg/l	4.18E-02	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Uranium	0.128	mg/l	1.91E-03	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	0.00338	mg/l	5.05E-05	mg/kg/d	3.1E-02	(mg/kg-day) ⁻¹	1.6E-06	NA	NA	NA	NA	NA
			Exp. Route Total								5.9E-05					NA
			Dermal	Arsenic	0.00256	mg/l	2.2E-07	mg/kg/d	1.5E+00	(mg/kg-day)-1	3.3E-07	NA	NA	NA	NA	NA
				Cobalt	0.01	mg/l	8.5E-07	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	2.8	mg/l	2.4E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Uranium	0.128	mg/l	1.1E-05	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	0.00338	mg/l	4.5E-06	mg/kg/d	3.1E-02	(mg/kg-day)-1	1.4E-07	NA	NA	NA	NA	NA
			Exp. Route Total								4.7E-07					NA
		Exposure Point Total									5.9E-05	NA	NA	NA	NA	NA
	Exposure Medium Total										5.9E-05					NA
	Air	Water Vapors from Showerhead	Inhalation	Chloroform	0.0076	mg/m3	9.1E-05	mg/m3	2.3E-05	(µg/m3)-1	2.1E-06	NA	NA	NA	NA	NA
			Exp. Route Total								2.1E-06					NA
		Exposure Point Total									2.1E-06					NA
	Exposure Medium Total										2.1E-06					NA
oundwater Total - S	W Alluvium										6.1E-05					NA
							Total of R	eceptor Risks Ac	ross All Media		6.1E-05		Total of Recep	tor Hazards A	cross All Media	NA

Notes:

(1) Inhalation EPC represents time weighted chemical air concentration (mg/m3) calculated from shower exposure model in Tables 7.C.RME Supplement C and 7.D.RME Supplement D (note that "Exposure Concentration" is calculated independently).

Definitions: NA = Not Applicable

TABLE 7.4.RME CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	E	PC		Can	cer Risk Calcula	itions						
				Potential Concern	Value	Units	Intake/Exposur	e Concentration	CSF/L	Init Risk	Cancer Risk	Intake/Exposure Concentratio		Rf	D/RfC	Hazard Quotient
					(1)		Value	Units	Value	Units		Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Arsenic	0.00145	mg/l	NA	NA	NA	NA	NA	3.97E-05	mg/kg/d	3.0E-04	mg/kg/d	0.13
Zone 1				Cobalt	0.0557	mg/l	NA	NA	NA	NA	NA	1.53E-03	mg/kg/d	3.0E-04	mg/kg/d	5.1
				Manganese	1.95	mg/l	NA	NA	NA	NA	NA	5.34E-02	mg/kg/d	2.4E-02	mg/kg/d	2.2
				Vanadium	0.2	mg/l	NA	NA	NA	NA	NA	5.48E-03	mg/kg/d	5.0E-03	mg/kg/d	1.1
				Chloroform	0.00068	mg/l	NA	NA	NA	NA	NA	1.86E-05	mg/kg/d	1.0E-02	mg/kg/d	0.0018
			Exp. Route Total								NA					8.5
			Dermal	Arsenic	0.00145	mg/l	NA	NA	NA	NA	NA	2.1E-07	mg/kg/d	3.0E-04	mg/kg/d	0.0007
				Cobalt	0.0557	mg/l	NA	NA	NA	NA	NA	8.0E-06	mg/kg/d	3.0E-04	mg/kg/d	0.027
				Manganese	1.95	mg/l	NA	NA	NA	NA	NA	2.8E-04	mg/kg/d	9.6E-04	mg/kg/d	0.29
				Vanadium	0.2	mg/l	NA	NA	NA	NA	NA	2.9E-05	mg/kg/d	5.0E-03	mg/kg/d	0.0057
				Chloroform	0.00068	mg/l	NA	NA	NA	NA	NA	1.7E-06	mg/kg/d	1.02E-02	mg/kg/d	0.0002
			Exp. Route Total								NA					0.32
		Exposure Point Total									NA					8.9
	Exposure Medium Total										NA					8.9
	Air	Water Vapors from Showerhead	Inhalation	Chloroform	0.0014	mg/m3	NA	NA	NA	NA	NA	3.3E-05	mg/m3	9.8E-02	mg/m3	0.0003
			Exp. Route Total								NA					0.0003
		Exposure Point Total									NA					0.0003
	Exposure Medium Total										NA					0.0003
Groundwater Total - Zo	one 1										NA					8.9
								Total of Rece	eptor Risks Acro	ss All Media	NA		Total of Recept	tor Hazards Ac	cross All Media	8.9

Notes:

(1) Inhalation EPC represents chemical air concentration (mg/m3) calculated from shower exposure model in Table 7.D.RME (Supplemental)

Definitions:

NA = Not Applicable

TABLE 7.5.RME CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	E	PC		Can	cer Risk Calcula	tions			Non-Can	cer Hazard Ca	lculations	
				Potential Concern	Value	Units	Intake/Exposure	Concentration	CSF/U	nit Risk	Cancer Risk	Intake/Exposure	e Concentration	Rf	D/RfC	Hazard Quotien
					(1)		Value	Units	Value	Units		Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Arsenic	0.00145	mg/l	NA	NA	NA	NA	NA	9.27E-05	mg/kg/d	3.0E-04	mg/kg/d	0.31
Zone 1				Cobalt	0.0557	mg/l	NA	NA	NA	NA	NA	3.56E-03	mg/kg/d	3.0E-04	mg/kg/d	11.9
				Manganese	1.95	mg/l	NA	NA	NA	NA	NA	1.25E-01	mg/kg/d	2.4E-02	mg/kg/d	5.2
				Vanadium	0.2	mg/l	NA	NA	NA	NA	NA	1.28E-02	mg/kg/d	5.0E-03	mg/kg/d	2.6
				Chloroform	0.00068	mg/l	NA	NA	NA	NA	NA	4.35E-05	mg/kg/d	1.0E-02	mg/kg/d	0.0043
			Exp. Route Total								NA					19.9
			Dermal	Arsenic	0.00145	mg/l	NA	NA	NA	NA	NA	6.1E-07	mg/kg/d	3.0E-04	mg/kg/d	0.0020
				Cobalt	0.0557	mg/l	NA	NA	NA	NA	NA	2.4E-05	mg/kg/d	3.0E-04	mg/kg/d	0.078
				Manganese	1.95	mg/l	NA	NA	NA	NA	NA	8.2E-04	mg/kg/d	9.6E-04	mg/kg/d	0.86
				Vanadium	0.2	mg/l	NA	NA	NA	NA	NA	8.4E-05	mg/kg/d	5.0E-03	mg/kg/d	0.017
				Chloroform	0.00068	mg/l	NA	NA	NA	NA	NA	3.8E-06	mg/kg/d	1.02E-02	mg/kg/d	0.0004
			Exp. Route Total								NA					0.95
		Exposure Point Total									NA					20.9
	Exposure Medium Total										NA					20.9
	Air	Water Vapors from Showerhead	Inhalation	Chloroform	0.0020	mg/m3	NA	NA	NA	NA	NA	8.2E-05	mg/m3	9.8E-02	mg/m3	0.0008
			Exp. Route Total								NA					0.0008
	<u> </u>	Exposure Point Total									NA					0.0008
	Exposure Medium Total										NA					0.0008
Groundwater Total - Zo	one 1						Ν									20.9
							Total of Receptor Risks Across All Media N/					Total of Receptor Hazards Across All Media				20.9

Notes:

(1) Inhalation EPC represents chemical air concentration (mg/m3) calculated from shower exposure model in Table 7.D.RME (Supplemental)

TABLE 7.6.RME CALCULATION OF CHEMICAL CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	E	PC		Can	cer Risk Calcula	itions			Non-Can	cer Hazard Ca	lculations	
				Potential Concern	Value	Units	Intake/Exposure	Concentration	CSF/L	Init Risk	Cancer Risk			Hazard Quotient		
					(1)		Value	Units	Value	Units		Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Arsenic	0.00145	mg/l	2.17E-05	mg/kg/d	1.5E+00	(mg/kg-day)-1	3.2E-05	NA	NA	NA	NA	NA
Zone 1				Cobalt	0.0557	mg/l	8.32E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	1.95	mg/l	2.91E-02	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Vanadium	0.2	mg/l	2.99E-03	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	0.00068	mg/l	1.02E-05	mg/kg/d	3.1E-02	(mg/kg-day)-1	3.1E-07	NA	NA	NA	NA	NA
			Exp. Route Total								3.3E-05					NA
			Dermal	Arsenic	0.00145	mg/l	1.24E-07	mg/kg/d	1.5E+00	(mg/kg-day)-1	1.9E-07	NA	NA	NA	NA	NA
				Cobalt	0.0557	mg/l	4.7E-06	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	1.95	mg/l	1.7E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Vanadium	0.2	mg/l	1.7E-05	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	0.00068	mg/l	9.1E-07	mg/kg/d	3.1E-02	(mg/kg-day)-1	2.8E-08	NA	NA	NA	NA	NA
			Exp. Route Total								2.1E-07					NA
		Exposure Point Total									3.3E-05					NA
	Exposure Medium Total		-								3.3E-05					NA
	Air	Water Vapors from Showerhead	Inhalation	Chloroform	0.0015	mg/m3	1.8E-05	mg/m3	2.3E-05	(µg/m3)-1	4.2E-07	NA	NA	NA	NA	NA
			Exp. Route Total								4.2E-07					NA
		Exposure Point Total									4.2E-07					NA
	Exposure Medium Total										4.2E-07					NA
Groundwater Total - Zo	one 1										3.3E-05					NA
								Total of Rece	eptor Risks Acro	ss All Media	3.3E-05		Total of Recept	or Hazards Ac	ross All Media	NA

Notes:

(1) Inhalation EPC represents time weighted chemical air concentration (mg/m3) calculated from shower exposure model in Tables 7.C.RME Supplement C and 7.D.RME Supplement D (note that "Exposure Concentration" is calculated independently.

TABLE 7.7.RME CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	E	PC		Cano	er Risk Calcula	tions			Non-Can	cer Hazard Ca	lculations	
				Potential Concern	Value	Units	Intake/Exposure	e Concentration	CSF/U	nit Risk	Cancer Risk	Intake/Exposure	Concentration	Rf	D/RfC	Hazard Quotient
					(1)		Value	Units	Value	Units		Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Aluminum	39.15	mg/l	NA	NA	NA	NA	NA	1.07E+00	mg/kg/d	1.0E+00	mg/kg/d	1.1
Zone 3				Arsenic	0.412	mg/l	NA	NA	NA	NA	NA	1.13E-02	mg/kg/d	3.0E-04	mg/kg/d	37.6
				Beryllium	0.0202	mg/l	NA	NA	NA	NA	NA	5.53E-04	mg/kg/d	2.0E-03	mg/kg/d	0.28
				Cadmium	0.0628	mg/l	NA	NA	NA	NA	NA	1.72E-03	mg/kg/d	5.0E-04	mg/kg/d	3.4
				Cobalt	0.439	mg/l	NA	NA	NA	NA	NA	1.20E-02	mg/kg/d	3.0E-04	mg/kg/d	40.1
				Manganese	10.89	mg/l	NA	NA	NA	NA	NA	2.98E-01	mg/kg/d	2.4E-02	mg/kg/d	12.4
				Molybdenum	0.739	mg/l	NA	NA	NA	NA	NA	2.02E-02	mg/kg/d	5.0E-03	mg/kg/d	4.0
				Nickel	0.489	mg/l	NA	NA	NA	NA	NA	1.34E-02	mg/kg/d	2.0E-02	mg/kg/d	0.67
				Vanadium	0.18	mg/l	NA	NA	NA	NA	NA	4.93E-03	mg/kg/d	5.0E-03	mg/kg/d	1.0
				Uranium	0.0431	mg/l	NA	NA	NA	NA	NA	1.18E-03	mg/kg/d	3.0E-03	mg/kg/d	0.39
				Chloroform	0.00326	mg/l	NA	NA	NA	NA	NA	8.93E-05	mg/kg/d	1.0E-02	mg/kg/d	0.0088
			Exp. Route Total								NA					101
			Dermal	Aluminum	39.15	mg/l	NA	NA	NA	NA	NA	5.6E-03	mg/kg/d	1.0E+00	mg/kg/d	0.0056
				Arsenic	0.412	mg/l	NA	NA	NA	NA	NA	5.9E-05	mg/kg/d	3.0E-04	mg/kg/d	0.20
				Beryllium	0.0202	mg/l	NA	NA	NA	NA	NA	2.9E-06	mg/kg/d	1.4E-05	mg/kg/d	0.21
				Cadmium	0.0628	mg/l	NA	NA	NA	NA	NA	9.0E-06	mg/kg/d	2.5E-05	mg/kg/d	0.36
				Cobalt	0.439	mg/l	NA	NA	NA	NA	NA	6.3E-05	mg/kg/d	3.0E-04	mg/kg/d	0.21
				Manganese	10.89	mg/l	NA	NA	NA	NA	NA	1.6E-03	mg/kg/d	9.6E-04	mg/kg/d	1.6
				Molybdenum	0.739	mg/l	NA	NA	NA	NA	NA	1.1E-04	mg/kg/d	5.0E-03	mg/kg/d	0.021
				Nickel	0.489	mg/l	NA	NA	NA	NA	NA	1.4E-05	mg/kg/d	8.0E-04	mg/kg/d	0.017
				Vanadium	0.18	mg/l	NA	NA	NA	NA	NA	2.6E-05	mg/kg/d	5.0E-03	mg/kg/d	0.0051
				Uranium	0.0431	mg/l	NA	NA	NA	NA	NA	6.2E-06	mg/kg/d	3.0E-03	mg/kg/d	0.0021
				Chloroform	0.00326	mg/l	NA	NA	NA	NA	NA	8.1E-06	mg/kg/d	1.0E-02	mg/kg/d	0.0008
			Exp. Route Total							-	NA				-	2.6
		Exposure Point Total									NA					104
	Exposure Medium Total										NA					104
]	Air	Water Vapors from Showerhead	Inhalation	Chloroform	0.0068	mg/m3	NA	NA	NA	NA	NA	1.6E-04	mg/m3	9.8E-02	mg/m3	0.0016
		<u> </u>	Exp. Route Total								NA					0.0016
		Exposure Point Total									NA					0.0016
	Exposure Medium Total										NA					0.0016
Groundwater Total - Zo	one 3						N				NA					104
								Total of Rece	ptor Risks Acro	ss All Media	NA		Total of Recept	or Hazards Ac	ross All Media	104

Notes:

(1) Inhalation EPC represents chemical air concentration (mg/m3) calculated from shower exposure model in Table 7.D.RME (Supplemental)

TABLE 7.8.RME CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child

Medium	Exposure Medium					tions		Non-Cancer Hazard Calculations								
				Potential Concern	Value	Units	Intake/Exposure	e Concentration	CSF/L	nit Risk	Cancer Risk	Intake/Exposure	e Concentration	RfE	D/RfC	Hazard Quotient
					(1)		Value	Units	Value	Units		Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Aluminum	39.15	mg/l	NA	NA	NA	NA	NA	2.50E+00	mg/kg/d	1.0E+00	mg/kg/d	2.5
Zone 3				Arsenic	0.412	mg/l	NA	NA	NA	NA	NA	2.63E-02	mg/kg/d	3.0E-04	mg/kg/d	87.8
				Beryllium	0.0202	mg/l	NA	NA	NA	NA	NA	1.29E-03	mg/kg/d	2.0E-03	mg/kg/d	0.65
				Cadmium	0.0628	mg/l	NA	NA	NA	NA	NA	4.01E-03	mg/kg/d	5.0E-04	mg/kg/d	8.0
				Cobalt	0.439	mg/l	NA	NA	NA	NA	NA	2.81E-02	mg/kg/d	3.0E-04	mg/kg/d	93.5
				Manganese	10.89	mg/l	NA	NA	NA	NA	NA	6.96E-01	mg/kg/d	2.4E-02	mg/kg/d	29.0
				Molybdenum	0.739	mg/l	NA	NA	NA	NA	NA	4.72E-02	mg/kg/d	5.0E-03	mg/kg/d	9.4
				Nickel	0.489	mg/l	NA	NA	NA	NA	NA	3.13E-02	mg/kg/d	2.0E-02	mg/kg/d	1.6
				Vanadium	0.18	mg/l	NA	NA	NA	NA	NA	1.15E-02	mg/kg/d	5.0E-03	mg/kg/d	2.3
				Uranium	0.0431	mg/l	NA	NA	NA	NA	NA	2.76E-03	mg/kg/d	3.0E-03	mg/kg/d	0.92
				Chloroform	0.00326	mg/l	NA	NA	NA	NA	NA	2.08E-04	mg/kg/d	1.0E-02	mg/kg/d	0.020
			Exp. Route Total								NA					236
			Dermal	Aluminum	39.15	mg/l	NA	NA	NA	NA	NA	1.7E-02	mg/kg/d	1.0E+00	mg/kg/d	0.017
				Arsenic	0.412	mg/l	NA	NA	NA	NA	NA	1.7E-04	mg/kg/d	3.0E-04	mg/kg/d	0.58
				Beryllium	0.0202	mg/l	NA	NA	NA	NA	NA	8.5E-06	mg/kg/d	1.4E-05	mg/kg/d	0.61
				Cadmium	0.0628	mg/l	NA	NA	NA	NA	NA	2.6E-05	mg/kg/d	2.5E-05	mg/kg/d	1.1
				Cobalt	0.439	mg/l	NA	NA	NA	NA	NA	1.9E-04	mg/kg/d	3.0E-04	mg/kg/d	0.62
				Manganese	10.89	mg/l	NA	NA	NA	NA	NA	4.6E-03	mg/kg/d	9.6E-04	mg/kg/d	4.8
				Molybdenum	0.739	mg/l	NA	NA	NA	NA	NA	3.1E-04	mg/kg/d	5.0E-03	mg/kg/d	0.062
				Nickel	0.489	mg/l	NA	NA	NA	NA	NA	4.1E-05	mg/kg/d	8.0E-04	mg/kg/d	0.052
				Vanadium	0.18	mg/l	NA	NA	NA	NA	NA	7.6E-05	mg/kg/d	5.0E-03	mg/kg/d	0.015
				Uranium	0.0431	mg/l	NA	NA	NA	NA	NA	1.8E-05	mg/kg/d	3.0E-03	mg/kg/d	0.0061
				Chloroform	0.00326	mg/l	NA	NA	NA	NA	NA	1.8E-05	mg/kg/d	1.0E-02	mg/kg/d	0.0018
			Exp. Route Total							-	NA				-	7.8
		Exposure Point Total									NA					244
l.	Exposure Medium Total										NA					244
Ţ	Air	Water Vapors from Showerhead	Inhalation	Chloroform	0.0098	mg/m3					NA	3.9E-04	mg/m3	9.8E-02	mg/m3	0.0040
			Exp. Route Total								NA					0.0040
		Exposure Point Total									NA					0.0040
	Exposure Medium Total						NA									0.0040
Groundwater Total - Zor	ne 3						NA				NA					244
								Total of Rece	ptor Risks Acro	ss All Media	NA		Total of Recept	or Hazards Ac	ross All Media	244

Notes:

(1) Inhalation EPC represents chemical air concentration (mg/m3) calculated from shower exposure model in Table 7.D.RME (Supplemental)

TABLE 7.9.RME CALCULATION OF CHEMICAL CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

uture
Resident
hild/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	El	PC		Can	cer Risk Calcula	ations			Non-Cano	er Hazard Ca	lculations	
			1	Potential Concern	Value	Units	Intake/Exposure	e Concentration	CSF/L	Jnit Risk	Cancer Risk	Intake/Exposur	e Concentration	RfD	D/RfC	Hazard Quotier
					(1)		Value	Units	Value	Units		Value	Units	Value	Units	1
iroundwater	Groundwater	Tapwater	Ingestion	Aluminum	39.15	mg/l	5.8E-01	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
Zone 3				Arsenic	0.412	mg/l	6.2E-03	mg/kg/d	1.5E+00	(mg/kg-day)-1	9.2E-03	NA	NA	NA	NA	NA
				Beryllium	0.0202	mg/l	3.0E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cadmium	0.0628	mg/l	9.4E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cobalt	0.439	mg/l	6.6E-03	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	10.89	mg/l	1.6E-01	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Molybdenum	0.739	mg/l	1.1E-02	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Nickel	0.489	mg/l	7.3E-03	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Vanadium	0.18	mg/l	2.7E-03	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Uranium	0.0431	mg/l	6.4E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	0.00326	mg/l	4.9E-05	mg/kg/d	3.1E-02	(mg/kg-day)-1	1.5E-06	NA	NA	NA	NA	NA
			Exp. Route Total								9.2E-03					NA
			Dermal	Aluminum	39.15	mg/l	3.34E-03	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Arsenic	0.412	mg/l	3.51E-05	mg/kg/d	1.5E+00	(mg/kg-day)-1	5.3E-05	NA	NA	NA	NA	NA
				Beryllium	0.0202	mg/l	1.72E-06	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cadmium	0.0628	mg/l	5.35E-06	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cobalt	0.439	mg/l	3.74E-05	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	10.89	mg/l	9.28E-04	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Molybdenum	0.739	mg/l	6.30E-05	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Nickel	0.489	mg/l	8.33E-06	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Vanadium	0.18	mg/l	1.53E-05	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Uranium	0.0431	mg/l	3.67E-06	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	0.00326	mg/l	4.35E-06	mg/kg/d	3.1E-02	(mg/kg-day)-1	1.3E-07	NA	NA	NA	NA	NA
			Exp. Route Total		-	-		-		-	5.3E-05					NA
		Exposure Point Total									9.3E-03					NA
	Exposure Medium Total										9.3E-03					NA
	Air	Water Vapors from Showerhead	Inahalation	Chloroform	0.0074	mg/m3	8.7E-05	mg/m3	2.3E-05	(µg/m3)-1	2.0E-06					NA
			Exp. Route Total								2.0E-06					NA
		Exposure Point Total									2.0E-06					NA
	Exposure Medium Total										2.0E-06					NA
dwater Total - Zo	one 3										9.3E-03					NA
								Total of Rece	eptor Risks Acro	oss All Media	9.3E-03		Total of Recept	or Hazards Ac	ross All Media	a NA

Notes:

(1) Inhalation EPC represents time weighted chemical air concentration (mg/m3) calculated from shower exposure model in Tables 7.C.RME Supplement C and 7.D.RME Supplement D (note that "Exposure Concentration" is calculated independently.

TABLE 7.A.RME (SUPPLEMENTAL) CALCULATION OF DAevent FUTURE RESIDENT ADULT UNC Church Rock Mill and Tailings Site

	Chemical	Groundwater	Permeability		Lag		Fraction	Duration	
	of Potential	Concentration	Coefficient ⁽²⁾		Time ⁽²⁾		Absorbed Water ⁽²⁾	of Event ⁽²⁾	
Hydrostratigraphic	Concern	(CW)	(Kp)	B ⁽²⁾	(τ_{event})	t* ⁽²⁾	(FA)	(tevent)	DAevent
Unit	(1)	mg/L	(cm/hr)	(dimensionless)	(hr)	(hr)	(dimensionless)	(hr)	(mg/cm ² -event)
SW Alluvium	Arsenic (arsenite)	2.56E-03	1.0E-03	NA	NA	NA	NA	0.58	1.5E-09
SW Alluvium	Cobalt	1.00E-02	1.0E-03	NA	NA	NA	NA	0.58	5.8E-09
SW Alluvium	Manganese	2.80E+00	1.0E-03	NA	NA	NA	NA	0.58	1.6E-06
SW Alluvium	Uranium	1.28E-01	1.0E-03	NA	NA	NA	NA	0.58	7.4E-08
SW Alluvium	Chloroform	3.38E-03	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	0.58	3.4E-08
Zone 1	Arsenic (arsenite)	1.45E-03	1.0E-03	NA	NA	NA	NA	0.58	8.4E-10
Zone 1	Cobalt	5.57E-02	1.0E-03	NA	NA	NA	NA	0.58	3.2E-08
Zone 1	Manganese	1.95E+00	1.0E-03	NA	NA	NA	NA	0.58	1.1E-06
Zone 1	Vanadium	2.00E-01	1.0E-03	NA	NA	NA	NA	0.58	1.2E-07
Zone 1	Uranium	1.74E-03	1.0E-03	NA	NA	NA	NA	0.58	1.0E-09
Zone 1	Chloroform	6.80E-04	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	0.58	6.9E-09
Zone 3	Aluminum	3.92E+01	1.0E-03	NA	NA	NA	NA	0.58	2.3E-05
Zone 3	Arsenic (arsenite)	4.12E-01	1.0E-03	NA	NA	NA	NA	0.58	2.4E-07
Zone 3	Beryllium	2.02E-02	1.0E-03	NA	NA	NA	NA	0.58	1.2E-08
Zone 3	Cadmium (water)	6.28E-02	1.0E-03	NA	NA	NA	NA	0.58	3.6E-08
Zone 3	Cobalt	4.39E-01	1.0E-03	NA	NA	NA	NA	0.58	2.5E-07
Zone 3	Manganese	1.09E+01	1.0E-03	NA	NA	NA	NA	0.58	6.3E-06
Zone 3	Molybdenum	7.39E-01	1.0E-03	NA	NA	NA	NA	0.58	4.3E-07
Zone 3	Nickel	4.89E-01	2.0E-04	NA	NA	NA	NA	0.58	5.7E-08
Zone 3	Vanadium	1.80E-01	1.0E-03	NA	NA	NA	NA	0.58	1.0E-07
Zone 3	Uranium	4.31E-02	1.0E-03	NA	NA	NA	NA	0.58	2.5E-08
Zone 3	Chloroform	3.26E-03	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	0.58	3.3E-08

Notes:

(1) Radionuclide COPCs not included in dermal evaluation because the ingested dose significantly outweighs the dermally absorbed dose.

(2) Values from EPA 2004, Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment - Final). EPA/540/R/99/005. Per the guidance, the default Kp value of 1E-06 was assigned to inorganics without designated Kp values.

Definitions:

NA = Not Applicable

TABLE 7.B.RME (SUPPLEMENTAL) CALCULATION OF DAevent FUTURE RESIDENT CHILD UNC Church Rock Mill and Tailings Site

	Chemical of Potential	Groundwater Concentration	Permeability Coefficient ⁽²⁾		Lag Time ⁽²⁾		Fraction Absorbed Water ⁽²⁾	Duration of Event ⁽²⁾	
Hydrostratigraphic	Concern	(CW)	(Kp)	B ⁽²⁾	(τ_{event})	t* ⁽²⁾	(FA)	(tevent)	DAevent
Unit	(1)	mg/L	(cm/hr)	(dimensionless)	(hr)	(hr)	(dimensionless)	(hr)	(mg/cm ² -event)
SW Alluvium	Arsenic (arsenite)	2.56E-03	1.0E-03	NA	NA	NA	NA	1	2.6E-09
SW Alluvium	Cobalt	1.00E-02	1.0E-03	NA	NA	NA	NA	1	1.0E-08
SW Alluvium	Manganese	2.80E+00	1.0E-03	NA	NA	NA	NA	1	2.8E-06
SW Alluvium	Uranium	1.28E-01	1.0E-03	NA	NA	NA	NA	1	1.3E-07
SW Alluvium	Chloroform	3.38E-03	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	1	4.5E-08
Zone 1	Arsenic (arsenite)	1.45E-03	1.0E-03	NA	NA	NA	NA	1	1.5E-09
Zone 1	Cobalt	5.57E-02	1.0E-03	NA	NA	NA	NA	1	5.6E-08
Zone 1	Manganese	1.95E+00	1.0E-03	NA	NA	NA	NA	1	2.0E-06
Zone 1	Vanadium	2.00E-01	1.0E-03	NA	NA	NA	NA	1	2.0E-07
Zone 1	Uranium	1.74E-03	1.0E-03	NA	NA	NA	NA	1	1.7E-09
Zone 1	Chloroform	6.80E-04	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	1	9.0E-09
Zone 3	Aluminum	3.92E+01	1.0E-03	NA	NA	NA	NA	1	3.9E-05
Zone 3	Arsenic (arsenite)	4.12E-01	1.0E-03	NA	NA	NA	NA	1	4.1E-07
Zone 3	Beryllium	2.02E-02	1.0E-03	NA	NA	NA	NA	1	2.0E-08
Zone 3	Cadmium (water)	6.28E-02	1.0E-03	NA	NA	NA	NA	1	6.3E-08
Zone 3	Cobalt	4.39E-01	1.0E-03	NA	NA	NA	NA	1	4.4E-07
Zone 3	Manganese	1.09E+01	1.0E-03	NA	NA	NA	NA	1	1.1E-05
Zone 3	Molybdenum	7.39E-01	1.0E-03	NA	NA	NA	NA	1	7.4E-07
Zone 3	Nickel	4.89E-01	2.0E-04	NA	NA	NA	NA	1	9.8E-08
Zone 3	Vanadium	1.80E-01	1.0E-03	NA	NA	NA	NA	1	1.8E-07
Zone 3	Uranium	4.31E-02	1.0E-03	NA	NA	NA	NA	1	4.3E-08
Zone 3	Chloroform	3.26E-03	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	1	4.3E-08

Notes:

(1) Radionuclide COPCs not included in dermal evaluation because the ingested dose significantly outweighs the dermally absorbed dose.

(2) Values from EPA 2004, Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment - Final). EPA/540/R/99/005. Per the guidance, the default Kp value of 1E-06 was assigned to inorganics without designated Kp values.

Definitions:

NA = Not Applicable

TABLE 7.C.RME (SUPPLEMENTAL)

INHALATION EXPOSURE CONCENTRATIONS FROM FOSTER AND CHROSTOWSKI SHOWER MODEL

FUTURE RESIDENT ADULT

UNC Church Rock Mill and Tailings Site

Hydrostratigraphic Unit	Chemical of Potential Concern	Exposure Point Concentration Cwo (µg/L)	Molecular weight (MW) (g/mole)	Henry's Law Constant (H) (atm-m ³ /mole)	Kg (VOC) (cm/hr)	KI (VOC) (cm/hr)	KL (cm/hr)	Kal (cm/hr)	Cwd (µg/L)	S (µg/m ³ -min)	Ca (mg/m ³)
SW Alluvium	Chloroform	3.4E+00	119.38	3.67E-03	1.2E+03	1.2E+01	1.1E+01	1.5E+01	4.1E-01	3.4E-01	7.0E-03
Zone 1	Chloroform	6.8E-01	119.38	3.67E-03	1.2E+03	1.2E+01	1.1E+01	1.5E+01	8.2E-02	6.8E-02	1.4E-03
Zone 3	Chloroform	3.3E+00	119.38	3.67E-03	1.2E+03	1.2E+01	1.1E+01	1.5E+01	3.9E-01	3.3E-01	6.8E-03

Kg(VOC) = gas-film mass transfer coefficient									
rig(ree) gae intrinace trailerer eeenielent	cm/hr	Solved by Eq 1							
KI(VOC) = liquid-film mass transfer coefficient	cm/hr	Solved by Eq 2							
KL = overall mass transfer coefficient	cm/hr	Solved by Eq 3							
Kal = adjusted overall mass transfer coeff.	cm/hr	Solved by Eq 4							
TI = Calibration temp. of water	K (20C +273)	293							
Ts = Shower water temperature	k (45C)	318							
Us = water viscosity at Ts	centipoise	0.596							
UI = water viscosity at TI	ср	1.002							
Cwd = conc. leaving droplets after time sdt	μg/l	Solved by Eq 5							
sdt = shower droplet drop time	sec	0.5							
d = shower droplet diameter	mm	1							
FR = shower water flow rate	l/min	10							
SV = shower room air volume	m³								
S = indoor VOC generation rate	μg/m³-min Solve								
VR = ventilation rate	I/min								
BW = body weight	kg								
Ds = duration of shower	min	35							
Dt = total duration in shower room	min	60							
R = Universal gas constant	atm-m3/mol-°K	8.20E-05							
Rae = air exchange rate	min ⁻¹	0.0083							
Ca = indoor air concentration of VOCs	µg/m³	Solved by Eq 7							
Equation 1: Kg(VOC) =	3000 * (18 / MW) ^{0.5}								
Equation 2: KI(VOC) =	20 * (44 / MW) ^{0.5}								
Equation 3: KL =	((1 / KI(VOC)) + (0.024 / (Kg (VOC) * H)))								
Equation 4: Kal =	$((T + U) + ((T + U)) + (T + U))^{-U.5})$								
Equation 5: Cwd =	(Cwo * (1-EXP((-1 * Kal * sdt)/(60 * d))))								
Equation 6: S =	(Cwd * FR / SV)								
Equation 7: Ca =	lf t>Ds [(S / Rae) * (Ds + (EXF	P(-Rae * Dt) / Rae)							
	-(EXP(Rae *(Ds - Dt)) / Rae)] / Dt * 1/1000								

Notes:

Inhalation Exposure Concentrations calculated based on Foster, Sarah A., and Paul C. Chrostowski. 1987. Inhalation Exposures to Volatile Organic Contaminants in the Shower.

In The Proceedings of the 80th Annual Meeting of the Air Pollution Control Association (APCA), June 21-26, New York. Air Pollution Control Association.

TABLE 7.D.RME (SUPPLEMENTAL) INHALATION EXPOSURE CONCENTRATIONS FROM FOSTER AND CHROSTOWSKI SHOWER MODEL FUTURE RESIDENT CHILD

UNC Church Rock Mill and Tailings Site

Hydrostratigraphic Unit	Chemical of Potential Concern	Exposure Point Concentration Cwo (µg/L)	Molecular weight (MW) (g/mole)	Henry's Law Constant (H) (atm-m ³ /mole)	Kg (VOC) (cm/hr)	KI (VOC) (cm/hr)	KL (cm/hr)	Kal (cm/hr)	Cwd (µg/L)	S (µg/m ³ -min)	Ca (mg/m³)
SW Alluvium	Chloroform	3.4E+00	119.38	3.7E-03	1.2E+03	1.2E+01	1.1E+01	1.5E+01	4.1E-01	3.4E-01	1.0E-02
Zone 1	Chloroform	6.8E-01	119.38	3.7E-03	1.2E+03	1.2E+01	1.1E+01	1.5E+01	8.2E-02	6.8E-02	2.0E-03
Zone 3	Chloroform	3.3E+00	119.38	3.7E-03	1.2E+03	1.2E+01	1.1E+01	1.5E+01	3.9E-01	3.3E-01	9.8E-03

	Variables		Units	Exposure Assumptions					
Kg(VOC) = gas-f	ilm mass transfer coefficient		cm/hr	Calculated using Eq 1					
KI(VOC) = liquid-	film mass transfer coefficient		cm/hr	Calculated using Eq 2					
KL = overall mas	s transfer coefficient		cm/hr	Calculated using Eq 3					
Kal = adjusted ov	verall mass transfer coeff.		cm/hr	Calculated using Eq 4					
TI = Calibration te	emp. of water		K (20C +273)	293					
Ts = Shower wat	er temperature		k (45C)	318					
Us = water viscos	sity at Ts		ср	0.596					
UI = water viscos	ity at TI		ср	1.002					
Cwd = conc. leav	ring droplets after time sdt		μg/l	Calculated using Eq 5					
sdt = shower dro	plet drop time		sec	0.5					
d = shower drop	let diameter		mm	1					
FR = shower wat	er flow rate		l/min						
SV = shower roo	m air volume		m³						
S = indoor VOC	generation rate		µg/m³-min	Calculated using Eq 6					
Ds = duration of	shower		min	60					
Dt = total duration	n in shower room		min	80					
R = Universal ga	s constant		atm-m3/mol-°K	8.20E-05					
Rae = air exchan	ge rate		min ⁻¹	0.0083					
Ca = indoor air c	oncentration of VOCs		µg/m³	Calculated using Eq 7					
Equation 1:	Kg(VOC) =	3000 * (18 / MW) ^{0.}	5						
Equation 2:	KI(VOC) =	20 * (44 / MW) ^{u.o}							
Equation 3:	KL =	((1 / KI(VOC)) + (R	*TI / (Kg (VOC) * H)))-						
Equation 4:	Kal =	(KL * (((TI * Us) / ($((L) + (((TI * Us) / (TS * UI))^{-U.5}))$						
Equation 5:	Cwd =	(Cwo * (1-EXP((-1							
Equation 6:	S =	(Cwd * FR / SV)							
Equation 7:	Ca =	If t>Ds [(S / Rae) * (Ds + (EXP(-Rae * Dt) / Rae)							
			-(EXP(Rae *(Ds - I	Dt)) / Rae)] / Dt * 1/1000					
Notes:									

Notes:

Inhalation Exposure Concentrations calculated based on Foster, Sarah A., and Paul C. Chrostowski. 1987. Inhalation Exposures to Volatile Organic Contaminants in the Shower.

In The Proceedings of the 80th Annual Meeting of the Air Pollution Control Association (APCA), June 21-26, New York. Air Pollution Control Association.

TABLE 7.E.RME (SUPPLEMENTAL) CALCULATION OF INHALATION INTAKE USING ANDELMAN VOLATILIZATION FACTOR FUTURE RESIDENT CHILD/ADULT UNC Church Rock Mill and Tailings Site

			Intake Through
	Hydrostratigraphic	EPC Activity	Vapor Inhalation
Compound	Unit	pCi/L	pCi
Radium 226+D	SWA	0.267	2.52E+04
Radium 226+D	Zone 1	1.213	1.15E+05
Radium 226+D	Zone 3	11.14	1.05E+06

$$Intake(pCi) = CW_{R}(\frac{pCi}{L}) \times EF(\frac{350d}{y}) \times ED(30yr) \times IRAadj(\frac{18m3}{d}) \times ETr(\frac{24hr}{d}) \times (\frac{1d}{24h}) \times K(\frac{0.5L}{m3}) \times ETr(\frac{1}{24h}) \times K(\frac{1}{24h}) \times K(\frac{$$

Name	Abbr	Value	Units
Radionuclide concentration (activity) in water	CW _R	Chem. specific	pCi/L
Exposure Frequency	EF	350	day/year
Exposure Duration	ED	30	year
Age-adjusted inhalation rate	IRAadj	18	m3/day
Andelman volatilization factor	К	0.5	L/day
Exposure Time-residential	ETr	24	hrs/day
Conversion	Constant (1 day/24 hours)	0.041666667	day/hours

Notes:

The inhalation exposure route is only calculated for Ra-226+D. Volatilization in the equation comes from household uses of water (e.g., showering,

laundering, dishwashing)

TABLE 7.F.RME (SUPPLEMENTAL) CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	EP	C (1)		Cano	cer Risk Calcula	tions			Non-Can	cer Hazard Ca	lculations	
				Potential Concern	Value	Units	Intake/Exposure	e Concentration	CSF/U	nit Risk	Cancer Risk	Intake/Exposure	e Concentration	Rf	D/RfC	Hazard Quotien
							Value	Units	Value	Units		Value	Units	Value	Units	
Background	Groundwater	Tapwater	Ingestion	Aluminum	0.231	mg/l	NA	NA	NA	NA	NA	6.33E-03	mg/kg/d	1.0E+00	mg/kg/d	0.0063
Groundwater				Arsenic	0.175	mg/l	NA	NA	NA	NA	NA	4.79E-03	mg/kg/d	3.0E-04	mg/kg/d	16.0
Zone 3				Beryllium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	2.0E-03	mg/kg/d	NC
				Cadmium	0.0113	mg/l	NA	NA	NA	NA	NA	3.10E-04	mg/kg/d	5.0E-04	mg/kg/d	0.62
				Cobalt	0.0877	mg/l	NA	NA	NA	NA	NA	2.40E-03	mg/kg/d	3.0E-04	mg/kg/d	8.0
				Manganese	3.436	mg/l	NA	NA	NA	NA	NA	9.41E-02	mg/kg/d	2.4E-02	mg/kg/d	3.9
				Molybdenum	17.43	mg/l	NA	NA	NA	NA	NA	4.78E-01	mg/kg/d	5.0E-03	mg/kg/d	95.5
				Nickel	0.14	mg/l	NA	NA	NA	NA	NA	3.84E-03	mg/kg/d	2.0E-02	mg/kg/d	0.19
				Vanadium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	5.0E-03	mg/kg/d	NC
				Uranium	0.107	mg/l	NA	NA	NA	NA	NA	2.93E-03	mg/kg/d	3.0E-03	mg/kg/d	0.98
				Chloroform	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	1.0E-02	mg/kg/d	NC
			Exp. Route Total								NA					125
			Dermal	Aluminum	0.231	mg/l	NA	NA	NA	NA	NA	3.3E-05	mg/kg/d	1.0E+00	mg/kg/d	0.00003
				Arsenic	0.175	mg/l	NA	NA	NA	NA	NA	2.5E-05	mg/kg/d	3.0E-04	mg/kg/d	0.083
				Beryllium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	1.4E-05	mg/kg/d	NC
				Cadmium	0.0113	mg/l	NA	NA	NA	NA	NA	1.6E-06	mg/kg/d	2.5E-05	mg/kg/d	0.065
				Cobalt	0.0877	mg/l	NA	NA	NA	NA	NA	1.3E-05	mg/kg/d	3.0E-04	mg/kg/d	0.042
				Manganese	3.436	mg/l	NA	NA	NA	NA	NA	4.9E-04	mg/kg/d	9.6E-04	mg/kg/d	0.51
				Molybdenum	17.43	mg/l	NA	NA	NA	NA	NA	2.5E-03	mg/kg/d	5.0E-03	mg/kg/d	0.50
				Nickel	0.14	mg/l	NA	NA	NA	NA	NA	4.0E-06	mg/kg/d	8.0E-04	mg/kg/d	0.0050
				Vanadium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	5.0E-03	mg/kg/d	NC
				Uranium	0.107	mg/l	NA	NA	NA	NA	NA	1.5E-05	mg/kg/d	3.0E-03	mg/kg/d	0.0051
				Chloroform	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	1.0E-02	mg/kg/d	NC
			Exp. Route Total								NA					1.2
ļ		Exposure Point Total									NA					126
	Exposure Medium Total										NA					126
	Air	Water Vapors from Showerhead	Inhalation	Chloroform	ND	mg/m3	NA	NA	NA	NA	NA	NC	mg/m3	9.8E-02	mg/m3	NC
			Exp. Route Total								NA					NC
		Exposure Point Total									NA					NC
	Exposure Medium Total										NA					NC
Groundwater Total - Zor	indwater Total - Zone 3										NA					126
Notes:								Total of Rece	eptor Risks Acro	ss All Media	NA		Total of Recept	or Hazards Ac	cross All Media	126

Notes:

(1) EPC concentration shown is "95% UCL" term for background groundwater calculated using ProUCL ver 4.00.02 as described in N.A.Water Systems (2008b). Chloroform is shown as not detected because it was detected in only one of 186 Zone 3 background samples.

Calculations do not include chemicals detected in background that are not COPCs in impacted water

Definitions: NA = Not Applicable

NC = Not calculated, COPC was not detected in background water

TABLE 7.G.RME (SUPPLEMENTAL) CALCULATION OF CHEMICAL NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	EPO	C (1)		Cano	er Risk Calcula	tions			Non-Can	cer Hazard Ca	lculations	
				Potential Concern	Value	Units	Intake/Exposure	e Concentration	CSF/U	nit Risk	Cancer Risk	Intake/Exposure	e Concentration	Rf	D/RfC	Hazard Quotien
							Value	Units	Value	Units		Value	Units	Value	Units	
Background	Groundwater	Tapwater	Ingestion	Aluminum	0.231	mg/l	NA	NA	NA	NA	NA	1.48E-02	mg/kg/d	1.0E+00	mg/kg/d	0.015
Groundwater				Arsenic	0.175	mg/l	NA	NA	NA	NA	NA	1.12E-02	mg/kg/d	3.0E-04	mg/kg/d	37.3
Zone 3				Beryllium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	2.0E-03	mg/kg/d	NC
				Cadmium	0.0113	mg/l	NA	NA	NA	NA	NA	7.22E-04	mg/kg/d	5.0E-04	mg/kg/d	1.4
				Cobalt	0.0877	mg/l	NA	NA	NA	NA	NA	5.61E-03	mg/kg/d	3.0E-04	mg/kg/d	18.7
				Manganese	3.436	mg/l	NA	NA	NA	NA	NA	2.20E-01	mg/kg/d	2.4E-02	mg/kg/d	9.2
				Molybdenum	17.43	mg/l	NA	NA	NA	NA	NA	1.11E+00	mg/kg/d	5.0E-03	mg/kg/d	223
				Nickel	0.14	mg/l	NA	NA	NA	NA	NA	8.95E-03	mg/kg/d	2.0E-02	mg/kg/d	0.45
				Vanadium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	5.0E-03	mg/kg/d	NC
				Uranium	0.107	mg/l	NA	NA	NA	NA	NA	6.84E-03	mg/kg/d	3.0E-03	mg/kg/d	2.3
				Chloroform	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	1.0E-02	mg/kg/d	NC
			Exp. Route Total								NA					292
			Dermal	Aluminum	0.231	mg/l	NA	NA	NA	NA	NA	9.7E-05	mg/kg/d	1.0E+00	mg/kg/d	0.0001
				Arsenic	0.175	mg/l	NA	NA	NA	NA	NA	7.4E-05	mg/kg/d	3.0E-04	mg/kg/d	0.25
				Beryllium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	1.4E-05	mg/kg/d	NC
				Cadmium	0.0113	mg/l	NA	NA	NA	NA	NA	4.8E-06	mg/kg/d	2.5E-05	mg/kg/d	0.19
				Cobalt	0.0877	mg/l	NA	NA	NA	NA	NA	3.7E-05	mg/kg/d	3.0E-04	mg/kg/d	0.12
				Manganese	3.436	mg/l	NA	NA	NA	NA	NA	1.4E-03	mg/kg/d	9.6E-04	mg/kg/d	1.5
				Molybdenum	17.43	mg/l	NA	NA	NA	NA	NA	7.4E-03	mg/kg/d	5.0E-03	mg/kg/d	1.5
				Nickel	0.14	mg/l	NA	NA	NA	NA	NA	1.2E-05	mg/kg/d	8.0E-04	mg/kg/d	0.015
				Vanadium	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	5.0E-03	mg/kg/d	NC
				Uranium	0.107	mg/l	NA	NA	NA	NA	NA	4.5E-05	mg/kg/d	3.0E-03	mg/kg/d	0.015
				Chloroform	ND	mg/l	NA	NA	NA	NA	NA	NC	mg/kg/d	1.0E-02	mg/kg/d	NC
			Exp. Route Total								NA					3.6
		Exposure Point Total									NA					296
	Exposure Medium Total										NA					296
	Air	Water Vapors from Showerhead	Inhalation	Chloroform	ND	mg/m3					NA	NC	mg/m3	9.8E-02	mg/m3	NC
			Exp. Route Total								NA					NC
		Exposure Point Total									NA					NC
	Exposure Medium Total										NA					NC
Groundwater Total - Zo	undwater Total - Zone 3										NA					296
Notes:								Total of Rece	ptor Risks Acro	ss All Media	NA		Total of Recept	or Hazards Ac	ross All Media	296

Notes:

(1) EPC concentration shown is "95% UCL" term for background groundwater calculated using ProUCL ver 4.00.02 as described in N.A.Water Systems (2008b). Chloroform is shown as not detected because it was detected in only one of 186 Zone 3 background samples.

Calculations do not include chemicals detected in background that are not COPCs in impacted water Definitions:

NA = Not Applicable

NC = Not calculated, COPC was not detected in background water

TABLE 7.H.RME (SUPPLEMENTAL) CALCULATION OF CHEMICAL CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Chemical of	EP	C (1)		Can	cer Risk Calcula	itions			Non-Can	er Hazard Ca	lculations	
				Potential Concern	Value	Units	Intake/Exposure	e Concentration	CSF/L	Init Risk	Cancer Risk	Intake/Exposure	Concentration	Rf	D/RfC	Hazard Quotien
							Value	Units	Value	Units		Value	Units	Value	Units	
Background	Groundwater	Tapwater	Ingestion	Aluminum	0.231	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
Groundwater				Arsenic	0.175	mg/l	2.6E-03	mg/kg/d	1.5E+00	(mg/kg-day)-1	3.9E-03	NA	NA	NA	NA	NA
Zone 3				Beryllium	ND	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cadmium	0.0113	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cobalt	0.0877	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	3.436	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Molybdenum	17.43	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Nickel	0.14	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Vanadium	ND	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Uranium	0.107	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	ND	mg/l	NC	mg/kg/d	3.1E-02	(mg/kg-day)-1	NC	NA	NA	NA	NA	NA
			Exp. Route Total	<u> </u>							3.9E-03					NA
			Dermal	Aluminum	0.231	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Arsenic	0.175	mg/l	1.49E-05	mg/kg/d	1.5E+00	(mg/kg-day)-1	2.2E-05	NA	NA	NA	NA	NA
				Beryllium	ND	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cadmium	0.0113	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Cobalt	0.0877	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Manganese	3.436	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Molybdenum	17.43	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Nickel	0.14	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Vanadium	ND	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Uranium	0.107	mg/l	NA	mg/kg/d	NA	NA	NA	NA	NA	NA	NA	NA
				Chloroform	ND	mg/l	NC	mg/kg/d	3.1E-02	(mg/kg-day)-1	NC	NA	NA	NA	NA	NA
			Exp. Route Total								2.2E-05					NA
l		Exposure Point Total									3.9E-03					NA
	Exposure Medium Total		1.1.1.2		ND			1.0	0.05.05	((0) (3.9E-03		1		1	NA
	Air	Water Vapors from Showerhead	Inahalation	Chloroform	ND	mg/m3	NC	mg/m3	2.3E-05	(µg/m3)-1	NC					NA
			Exp. Route Total								NC					NA
	European Madium 7 11	Exposure Point Total									NC					NA
O	Exposure Medium Total										NC					NA
Groundwater Total - Zo	water lotal - Zone 3							T / F		AU 14 F	3.9E-03		T (D			NA
lotos:								I otal of Rece	eptor Risks Acro	iss all Media	3.9E-03		Total of Recept	or Hazards Ac	ross All Media	NA

Notes:

(1) EPC concentration shown is "95% UCL" term for background groundwater calculated using ProUCL ver 4.00.02 as described in N.A.Water Systems (2008b). Chloroform is shown as not detected because it was detected in only one of 186 Zone 3 background samples. Calculations do not include chemicals detected in background that are not COPCs in impacted water

Calculations do not include chemicals detected in background that are not COP Definitions:

NA = Not Applicable

NC = Not calculated, COPC was not detected in background water

TABLE 7.I.RME (SUPPLEMENTAL) CALCULATION OF DAevent Future Resident Adult UNC Church Rock Mill and Tailings Site

	Chemical	Groundwater	Permeability		Lag		Fraction	Duration	
	of Potential	Concentration	Coefficient ⁽²⁾		Time ⁽²⁾		Absorbed Water ⁽²⁾	of Event ⁽²⁾	
Hydrostratigraphic	Concern	(CW)	(Kp)	B ⁽²⁾	(τ_{event})	t* ⁽²⁾	(FA)	(tevent)	DAevent
Unit	(1)	mg/L	(cm/hr)	(dimensionless)	(hr)	(hr)	(dimensionless)	(hr)	(mg/cm ² -event)
Zone 3 - Background	Aluminum	2.31E-01	1.0E-03	NA	NA	NA	NA	0.58	1.3E-07
Zone 3 - Background	Arsenic (arsenite)	1.75E-01	1.0E-03	NA	NA	NA	NA	0.58	1.0E-07
Zone 3 - Background	Beryllium	ND	1.0E-03	NA	NA	NA	NA	0.58	NC
Zone 3 - Background	Cadmium (water)	1.13E-02	1.0E-03	NA	NA	NA	NA	0.58	6.6E-09
Zone 3 - Background	Cobalt	8.77E-02	1.0E-03	NA	NA	NA	NA	0.58	5.1E-08
Zone 3 - Background	Manganese	3.44E+00	1.0E-03	NA	NA	NA	NA	0.58	2.0E-06
Zone 3 - Background	Molybdenum	1.74E+01	1.0E-03	NA	NA	NA	NA	0.58	1.0E-05
Zone 3 - Background	Nickel	1.40E-01	2.0E-04	NA	NA	NA	NA	0.58	1.6E-08
Zone 3 - Background	Vanadium	ND	1.0E-03	NA	NA	NA	NA	0.58	NC
Zone 3 - Background	Uranium	1.07E-01	1.0E-03	NA	NA	NA	NA	0.58	6.2E-08
Zone 3 - Background	Chloroform	ND	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	0.58	NC

Notes:

(1) Radionuclide COPCs not included in dermal evaluation because the ingested dose significantly outweighs the dermally absorbed dose.

(2) Groundwater concentration shown is "95% UCL" term for background groundwater calculated using ProUCL ver 4.00.02 as described in N.A.Water Systems (2008b). Chloroform is shown as not detected because it was detected in only one of 186 Zone 3 background samples.

(3) Values from EPA 2004, Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment - Final). EPA/540/R/99/005. Per the guidance, the default Kp value of 1E-06 was assigned to inorganics without designated Kp values.

Definitions:

NA = Not Applicable

NC = Not calculated, COPC was not detected in background water

TABLE 7.J.RME (SUPPLEMENTAL) CALCULATION OF DAevent Future Resident Child UNC Church Rock Mill and Tailings Site

Hydrostratigraphic Unit	Chemical of Potential Concern (1)	Groundwater Concentration (CW) mg/L	Permeability Coefficient ⁽³⁾ (Kp) (cm/hr)	B ⁽³⁾ (dimensionless)	Lag Time ⁽³⁾ (τ _{event}) (hr)	t* ⁽³⁾ (hr)	Fraction Absorbed Water ⁽³⁾ (FA) (dimensionless)	Duration of Event ⁽³⁾ (tevent) (hr)	DAevent (mg/cm ² -event)
Zone 3 - Background	Aluminum	2.31E-01	1.0E-03	NA	NA	NA	NA	1	2.3E-07
Zone 3 - Background	Arsenic (arsenite)	1.75E-01	1.0E-03	NA	NA	NA	NA	1	1.8E-07
Zone 3 - Background	Beryllium	ND	1.0E-03	NA	NA	NA	NA	1	NC
Zone 3 - Background	Cadmium (water)	1.13E-02	1.0E-03	NA	NA	NA	NA	1	1.1E-08
Zone 3 - Background	Cobalt	8.77E-02	1.0E-03	NA	NA	NA	NA	1	8.8E-08
Zone 3 - Background	Manganese	3.44E+00	1.0E-03	NA	NA	NA	NA	1	3.4E-06
Zone 3 - Background	Molybdenum	1.74E+01	1.0E-03	NA	NA	NA	NA	1	1.7E-05
Zone 3 - Background	Nickel	1.40E-01	2.0E-04	NA	NA	NA	NA	1	2.8E-08
Zone 3 - Background	Vanadium	ND	1.0E-03	NA	NA	NA	NA	1	NC
Zone 3 - Background	Uranium	1.07E-01	1.0E-03	NA	NA	NA	NA	1	1.1E-07
Zone 3 - Background	Chloroform	ND	6.8E-03	2.9E-02	5.0E-01	1.2E+00	1.0E+00	1	NC

Notes:

(1) Radionuclide COPCs not included in dermal evaluation because the ingested dose significantly outweighs the dermally absorbed dose.

(2) Groundwater concentration shown is "95% UCL" term for background groundwater calculated using ProUCL ver 4.00.02 as described in N.A.Water Systems (2008b). Chloroform is shown as not detected because it was detected in only one of 186 Zone 3 background samples.

(3) Values from EPA 2004, Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment - Final). EPA/540/R/99/005. Per the guidance, the default Kp value of 1E-06 was assigned to inorganics without designated Kp values.

Definitions:

NA = Not Applicable

NC = Not calculated, COPC was not detected in background water

TABLE 8.1.RME CALCULATION OF RADIATION CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Radionuclide of Potential Concern	EI	PC 04	Risk Calculation		Ca	ancer Risk Calcu	lations	
					Value	Units	Approach	Intake/	/Activity	C	SF	Cancer Risk
					(1)			Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Uranium-234	4.4E+01	pCi/l	USEPA RAGS	8.3E+05	pCi	7.1E-11	Risk/pCi	5.8E-05
SW Alluvium				Uranium-235+D	2.0E+00	pCi/l	USEPA RAGS	3.8E+04	pCi	7.2E-11	Risk/pCi	2.7E-06
				Uranium-238+D	4.3E+01	pCi/l	USEPA RAGS	8.1E+05	pCi	8.7E-11	Risk/pCi	7.0E-05
				Radium-226+D	2.7E-01	pCi/l	USEPA RAGS	5.0E+03	pCi	3.9E-10	Risk/pCi	1.9E-06
				Radium-228+D	8.6E-01	pCi/l	USEPA RAGS	1.6E+04	pCi	1.0E-09	Risk/pCi	1.7E-05
				Thorium-230	2.9E-01	pCi/l	USEPA RAGS	5.5E+03	pCi	9.1E-11	Risk/pCi	5.0E-07
			Exp. Route Total							-		1.5E-04
		Exposure Point Total										1.5E-04
	Exposure Medium Total											1.5E-04
	Air	Water Vapors from Domestic Use	Inhalation	Radium-226+D	1.3E-01	pCi/m3	USEPA RAGS	2.5E+04	pCi	1.2E-08	Risk/pCi	2.9E-04
			Exp. Route Total						•		•	2.9E-04
		Exposure Point Total										2.9E-04
	Exposure Medium Total								2.9E-04			
Medium Total												4.4E-04

Notes:

(1) Inhalation EPC represents the air concentration (pCi/m3) calculated using the Andelman volatilization factor K (0.5 L/m3)

Total of Receptor Risks Across All Media

4.4E-04

TABLE 8.2.RME CALCULATION OF RADIATION CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Radionuclide of Potential Concern	El	С	Risk Calculation		C	ancer Risk Calcu	llations	
					Value	Units	Approach	Intake/	Activity	C	SF	Cancer Risk
					(1)			Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Uranium-234	5.9E-01	pCi/l	USEPA RAGS	1.1E+04	pCi	7.1E-11	Risk/pCi	7.9E-07
Zone 1				Uranium-235+D	2.7E-02	pCi/l	USEPA RAGS	5.1E+02	pCi	7.2E-11	Risk/pCi	3.7E-08
				Uranium-238+D	5.8E-01	pCi/l	USEPA RAGS	1.1E+04	pCi	8.7E-11	Risk/pCi	9.6E-07
				Radium-226+D	1.2E+00	pCi/l	USEPA RAGS	2.3E+04	pCi	3.9E-10	Risk/pCi	8.8E-06
				Radium-228+D	2.1E+00	pCi/l	USEPA RAGS	3.9E+04	pCi	1.0E-09	Risk/pCi	4.1E-05
				Thorium-230	6.5E-01	pCi/l	USEPA RAGS	1.2E+04	pCi	9.1E-11	Risk/pCi	1.1E-06
			Exp. Route Total									5.3E-05
		Exposure Point Total										5.3E-05
	Exposure Medium Total											5.3E-05
	Air	Water Vapors from Domestic Use	Inhalation	Radium-226+D	6.1E-01	pCi/m3	USEPA RAGS	1.1E+05	pCi	1.2E-08	Risk/pCi	1.3E-03
			Exp. Route Total									1.3E-03
		Exposure Point Total										1.3E-03
	Exposure Medium Total											1.3E-03
Medium Total												1.4E-03

Notes:

(1) Inhalation EPC represents the air concentration (pCi/m3) calculated using the Andelman volatilization factor K (0.5 L/m3)

Total of Receptor Risks Across All Media

1.4E-03

TABLE 8.3.RME CALCULATION OF RADIATION CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Radionuclide of Potential Concern	E	PC	Risk Calculation		С	ancer Risk Calcu	ulations	
					Value	Units	Approach	Intake	/Activity	C	SF	Cancer Risk
					(1)			Value	Units	Value	Units	
Groundwater	Groundwater	Tapwater	Ingestion	Uranium-234	1.5E+01	pCi/l	USEPA RAGS	2.8E+05	pCi	7.1E-11	Risk/pCi	2.0E-05
Zone 3				Uranium-235+D	6.7E-01	pCi/l	USEPA RAGS	1.3E+04	pCi	7.2E-11	Risk/pCi	9.1E-07
				Uranium-238+D	1.4E+01	pCi/l	USEPA RAGS	2.7E+05	pCi	8.7E-11	Risk/pCi	2.4E-05
				Radium-226+D	1.1E+01	pCi/l	USEPA RAGS	2.1E+05	pCi	3.9E-10	Risk/pCi	8.1E-05
				Radium-228+D	1.8E+01	pCi/l	USEPA RAGS	3.4E+05	pCi	1.0E-09	Risk/pCi	3.5E-04
				Thorium-230	2.6E-01	pCi/l	USEPA RAGS	4.9E+03	pCi	9.1E-11	Risk/pCi	4.5E-07
				Lead-210+D	2.3E+00	pCi/l	USEPA RAGS	4.3E+04	pCi	1.3E-09	Risk/pCi	5.5E-05
			Exp. Route Total									5.3E-04
		Exposure Point Total										5.3E-04
	Exposure Medium Tota	1										5.3E-04
	Air	Water Vapors from Domestic Use	Inhalation	Radium-226+D	6E+00	pCi/m3	USEPA RAGS	1.1E+06	pCi	1.2E-08	Risk/pCi	1.2E-02
			Exp. Route Total									1.2E-02
		Exposure Point Total										1.2E-02
	Exposure Medium Tota											1.2E-02
Medium Total												1.3E-02

Notes:

(1) Inhalation EPC represents the air concentration (pCi/m3) calculated using the Andelman volatilization factor K (0.5 L/m3)

Total of Receptor Risks Across All Media

1.3E-02

TABLE 8.A.RME (SUPPLEMENTAL) CALCULATION OF RADIATION CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Exposure Route	Radionuclide of Potential Concern	EP	C (1)	Risk Calculation		C	ancer Risk Calcu	ulations	
					Value	Units	Approach	Intake	Activity	С	SF	Cancer Risk
					(2)			Value	Units	Value	Units	
Background	Groundwater	Tapwater	Ingestion	Uranium-234	3.7E+01	pCi/l	USEPA RAGS	6.9E+05	pCi	7.1E-11	Risk/pCi	4.9E-05
Groundwater				Uranium-235+D	1.7E+00	pCi/l	USEPA RAGS	3.1E+04	pCi	7.2E-11	Risk/pCi	2.3E-06
Zone 3				Uranium-238+D	3.6E+01	pCi/l	USEPA RAGS	6.7E+05	pCi	8.7E-11	Risk/pCi	5.9E-05
				Radium-226+D	5.0E+00	pCi/l	USEPA RAGS	9.4E+04	pCi	3.9E-10	Risk/pCi	3.6E-05
				Radium-228+D	4.5E+00	pCi/l	USEPA RAGS	8.5E+04	pCi	1.0E-09	Risk/pCi	8.9E-05
				Thorium-230	1.4E+00	pCi/l	USEPA RAGS	2.7E+04	pCi	9.1E-11	Risk/pCi	2.5E-06
				Lead-210+D	1.6E+00	pCi/l	USEPA RAGS	3.1E+04	pCi	1.3E-09	Risk/pCi	3.9E-05
			Exp. Route Total									2.8E-04
		Exposure Point Total										2.8E-04
	Exposure Medium Total											2.8E-04
	Air	Water Vapors from Showerhead	Inhalation	Radium-226+D	2.5E+00	pCi/m3	USEPA RAGS	4.7E+05	pCi	1.2E-08	Risk/pCi	5.5E-03
			Exp. Route Total						•			5.5E-03
		Exposure Point Total										5.5E-03
	Exposure Medium Total											5.5E-03
Medium Total												5.8E-03

Notes:

(1) Groundwater EPC represents UCL95 background water concentration per N.A. Water System (2008b)

(2) Inhalation EPC represents the air concentration (pCi/m3) calculated using the Andelman volatilization factor K (0.5 L/m3)

Total of Receptor Risks Across All Media

5.8E-03

TABLE 9.1.RME SUMMARY OF RECEPTOR HAZARDS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident

Receptor Age: Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	c Risk		Nc	on-Carcinogeni	c Hazard Quot	ient	
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total
Groundwater	Groundwater	Tapwater	Arsenic						Skin	0.23		0.0012	0.24
SW Alluvium			Cobalt						Thyroid	0.91		0.0048	0.92
			Manganese						Central nervous system	3.2		0.42	3.6
			Uranium						Kidney	1.2		0.0061	1.2
			Chloroform						Liver	0.0091		0.0008	0.010
			Chemical Total							5.5		0.43	6.0
			Uranium-234										
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D										
			Radium-228+D										
			Thorium-230										
			Radionuclide Total										Į
		Exposure Point Total											6.0
		ledium Total	1										6.0
	Air	Water Vapors	Arsenic										
		from Showerhead	Cobalt										
			Manganese										
			Uranium										
			Chloroform						Liver		0.0017		0.0017
			Chemical Total								0.0017		0.0017
			Uranium-234										
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D										
			Radium-228+D										
			Thorium-230										
		Exposure Point Total	Radionuclide Total										0.0017
	Exposure										0.0017		
Madium Tat /	Exposure Medium Total								<u></u>				0.0017
Medium Total						B	. D. I. T. I. I.		1		-		6.0
Receptor Total	ptor Total					Recepto	r Risk Total				Rece	eptor HI Total	6.0

Total Skin HI Across All Media = 0.24

Total Thyroid HI Across All Media = 0.92

3.6

1.2

0.012

Total Central Nervous System HI Across All Media =

Total Kidney HI Across All Media =

Total Liver HI Across All Media =

TABLE 9.2.RME SUMMARY OF RECEPTOR HAZARDS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	ic Risk		Nc	on-Carcinogeni	c Hazard Quot	ient		
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total	
Groundwater	Groundwater	Tapwater	Arsenic						Skin	0.55		0.0036	0.55	
SW Alluvium			Cobalt						Thyroid	2.1		0.014	2.1	
			Manganese						Central nervous system	7.5		1.2	8.7	
			Uranium						Kidney	2.7		0.018	2.7	
			Chloroform						Liver	0.021		0.0019	0.023	
			Chemical Total			-				12.9		1.3	14.2	
			Uranium-234				Ī							
			Uranium-235+D											
			Uranium-238+D											
			Radium-226+D											
			Radium-228+D											
			Thorium-230											
			Radionuclide Total											
		Exposure Point Total											14.2	
ļ	Exposure N	Nedium Total											14.2	
	Air	Water Vapors	Arsenic											
		from Showerhead	Cobalt											
			Manganese											
			Uranium											
			Chloroform						Liver		0.0041		0.0041	
			Chemical Total								0.0041		0.0041	
			Uranium-234											
			Uranium-235+D											
			Uranium-238+D											
			Radium-226+D											
			Radium-228+D											
			Thorium-230											
			Radionuclide Total											
		Exposure Point Total										0.0041		
	Exposure N													
Medium Total													14.2 14.2	
eceptor Total						Recepto	r Risk Total			Receptor HI To				

Total Skin HI Across All Media = 0.55 Total Thyroid HI Across All Media = 2.1 Total Central Nervous System HI Across All Media = 8.7 Total Kidney HI Across All Media = 2.7

Total Liver HI Across All Media = 0.027

TABLE 9.3.RME SUMMARY OF RECEPTOR RISKS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child/Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	ic Risk		No	on-Carcinogeni	c Hazard Quot	ient	
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total
Groundwater	Groundwater	Tapwater	Arsenic	5.7E-05		3.3E-07		5.8E-05					
SW Alluvium			Cobalt										
			Manganese										
			Uranium										
			Chloroform	1.6E-06		1.4E-07		1.7E-06					
			Chemical Total	5.9E-05		4.7E-07		5.9E-05					
			Uranium-234	5.8E-05			1	5.8E-05					
			Uranium-235+D	2.7E-06				2.7E-06					
			Uranium-238+D	7.0E-05				7.0E-05					
			Radium-226+D	1.9E-06				1.9E-06					
			Radium-228+D	1.7E-05				1.7E-05					
			Thorium-230	5.0E-07				5.0E-07					
			Radionuclide Total	1.5E-04				1.5E-04					
		Exposure Point Total						2.1E-04					
	Exposure Medium Total							2.1E-04		-		-	
	Air	Water Vapors	Arsenic										
		from Showerhead	Cobalt										
			Manganese										
			Uranium										
			Chloroform		2.1E-06			2.1E-06					
			Chemical Total		2.1E-06			2.1E-06					
			Uranium-234										
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D		2.9E-04			2.9E-04					
			Radium-228+D										
			Thorium-230										
			Radionuclide Total		2.9E-04			2.9E-04					
		Exposure Point Total						2.9E-04					
	Exposure N	Medium Total						2.9E-04					
Medium Total								5.0E-04					
Receptor Total						Recepto	r Risk Total	5.0E-04			Rece	eptor HI Total	

TABLE 9.4.RME SUMMARY OF RECEPTOR HAZARDS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident

Receptor Age: Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	c Risk	Non-Carcinogenic Hazard Quotient						
			Concern	Ingestion	Inhalation	Dermal	External	Exposure	Primary	Ingestion	Inhalation	Dermal	Exposure	
							(Radiation)	Routes Total	Target Organ(s)				Routes Total	
Groundwater	Groundwater	Tapwater	Arsenic						Skin	0.13		0.0007	0.13	
Zone 1			Cobalt						Thyroid	5.1		0.027	5.1	
			Manganese						Central nervous system	2.2		0.29	2.5	
			Vanadium						Decreased hair cystine	1.1		0.0057	1.1	
			Chloroform						Liver	0.0018		0.0002	0.0020	
			Chemical Total							8.5		0.32	8.9	
			Uranium-234											
			Uranium-235+D											
			Uranium-238+D											
			Radium-226+D											
			Radium-228+D											
			Thorium-230											
		1	Radionuclide Total											
		Exposure Point Total											8.9	
	Exposure N	ledium Total	-										8.9	
	Air	Water Vapors	Arsenic											
		from Showerhead	Cobalt											
			Manganese											
			Vanadium											
			Chloroform						Liver		0.0003		0.0003	
			Chemical Total								0.0003		0.0003	
			Uranium-234											
			Uranium-235+D											
			Uranium-238+D											
			Radium-226+D											
			Radium-228+D											
			Thorium-230											
			Radionuclide Total											
		Exposure Point Total											0.0003	
<u> </u>	Exposure N	ledium Total											0.0003	
Medium Total													8.9	
Receptor Total						Recepto	r Risk Total				Rece	ptor HI Total	8.9	

Total Skin HI Across All Media = 0.13

Total Thyroid HI Across All Media =

Total Central Nervous System HI Across All Media =

Total Hair Cystine (Metabolic System) HI Across All Media =

Total Liver HI Across All Media = 0.0023

5.1

2.5

1.1

TABLE 9.5.RME SUMMARY OF RECEPTOR HAZARDS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child

Medium Exposure Exposure Chemical Carcinogenic Risk Medium Point of Potential

	Medium	Point	of Potential								1		1
			Concern	Ingestion	Inhalation	Dermal	External	Exposure	Primary	Ingestion	Inhalation	Dermal	Exposure
							(Radiation)	Routes Total	Target Organ(s)				Routes Total
Groundwater	Groundwater	Tapwater	Arsenic						Skin	0.31		0.0020	0.31
Zone 1			Cobalt						Thyroid	11.9		0.078	11.9
			Manganese						Central nervous system	5.2		0.86	6.1
			Vanadium						Decreased hair cystine	2.6		0.017	2.6
			Chloroform						Liver	0.0043		0.0004	0.0046
			Chemical Total							19.9		0.95	20.9
			Uranium-234		-								
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D										
			Radium-228+D										
			Thorium-230										
			Radionuclide Total										
		Exposure Point Total											20.9
	Exposure Medium Total												20.9
Ī	Air	Air Water Vapors	Arsenic										
		from Showerhead	Cobalt										
			Manganese										
			Vanadium										
			Chloroform						Liver		0.0008		0.0008
			Chemical Total								0.0008		0.0008
			Uranium-234										
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D										
			Radium-228+D										
			Thorium-230										
			Radionuclide Total			-							
		Exposure Point Total											0.0008
	Exposure N	ledium Total											0.0008
Medium Total													20.9
Receptor Total						Recepto	r Risk Total				Rece	eptor HI Total	20.9

Total Skin HI Across All Media = 0.31 Total Thyroid HI Across All Media = 11.9 Total Central Nervous System HI Across All Media 6.1

2.6

Total Hair Cystine (Metabolic System) HI Across All Media

Non-Carcinogenic Hazard Quotient

Total Liver HI Across All Media = 0.0055

TABLE 9.6.RME SUMMARY OF RECEPTOR RISKS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child/Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogen	ic Risk		Nc	n-Carcinogeni	c Hazard Quot	ient	
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total
Groundwater	Groundwater	Tapwater	Arsenic	3.2E-05		1.9E-07		3.3E-05					
Zone 1			Cobalt										
			Manganese										
			Vanadium										
			Chloroform	3.1E-07		2.8E-08		3.4E-07					
			Chemical Total	3.3E-05		2.1E-07		3.3E-05					
			Uranium-234	7.9E-07				7.9E-07					
			Uranium-235+D	3.7E-08				3.7E-08					
			Uranium-238+D	9.6E-07				9.6E-07					
			Radium-226+D	8.8E-06				8.8E-06					
			Radium-228+D	4.1E-05				4.1E-05					
			Thorium-230	1.1E-06				1.1E-06					
			Radionuclide Total	5.3E-05				5.3E-05					
		Exposure Point Total						8.6E-05					
	Exposure N	Exposure Medium Total						8.6E-05					
	Air	Water Vapors	Arsenic										
		from Showerhead	Cobalt										
			Manganese										
			Vanadium										
			Chloroform		4.2E-07			4.2E-07					
			Chemical Total		4.2E-07			4.2E-07					
			Uranium-234										
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D		1.3E-03			1.3E-03					
			Radium-228+D										
			Thorium-230										
			Radionuclide Total		1.3E-03			1.3E-03					
		Exposure Point Total						1.3E-03					
	Exposure N	Nedium Total						1.3E-03					
Medium Total								1.4E-03					
Receptor Total						Recepto	r Risk Total	1.4E-03			Rece	eptor HI Total	

TABLE 9.7.RME SUMMARY OF RECEPTOR HAZARDS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Receptor Age: Adult	

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogen	ic Risk		Non-	Carcinogenic	Hazard Quotie	nt	
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Tot
Groundwater	Groundwater	Tapwater	Aluminum					-	Central nervous system	1.1		0.0056	1.1
Zone 3			Arsenic						Skin	37.6		0.20	37.8
			Beryllium					-	Gastrointestinal	0.28		0.21	0.48
			Cadmium					-	Kidney	3.4		0.36	3.8
			Cobalt						Thyroid	40.1		0.21	40.3
			Manganese						Central nervous system	12.4		1.6	14.1
			Molybdenum						Increased uric acid (kidney)	4.0		0.021	4.1
			Nickel						Reduced organ and body weights	0.67		0.017	0.69
			Vanadium						Decreased hair cystine	1.0		0.0051	1.0
			Uranium						Kidney	0.39		0.0021	0.40
			Chloroform						Liver	0.0088		0.0008	0.01
			Chemical Total							101		2.6	104
			Uranium-234										
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D										
			Radium-228+D										
			Thorium-230										
			Lead-210+D										
			Radionuclide Total										
		Exposure Point Total			1								104
i	Exposure N	Aedium Total		1									104
ļ.	Air	Water Vapors	Aluminum										
		from Showerhead	Arsenic										
		nom onowernead	Beryllium					_					
			Cadmium					-	-				
			Cobalt					-					
			Manganese						-				
			Molybdenum										
			Nickel										
			Vanadium										
			Uranium										
			Chloroform	-	-		-		 Liver	-		0.0016	0.0016
			Chemical Total	-	-				Liver			0.0010	0.0010
			Uranium-234										0.002
			Uranium-235+D										
			Uranium-235+D Uranium-238+D										
										-			
			Radium-226+D										
			Radium-228+D										
		1	Thorium-230	-								-	
			Lead-210+D										
			Lead-210+D Radionuclide Total	-									
		Exposure Point Total	<u>h</u>										0.0016
dium Total	Exposure N	Exposure Point Total Aedium Total	<u>h</u>				-					-	0.0016 0.0016 104

Total Skin HI Across All Media =	37.8
	37.0
Total Thyroid HI Across All Media =	40.3
Total Central Nervous System HI Across All Media =	15.1
Total Kidney HI Across All Media =	8.3
Total Liver HI Across All Media =	0.011
Total Gastrointestinal HI Across All Media =	0.48
Total Reduced Body and Organ Weights HI Across All Media =	0.69
Total Hair Cystine (Metabolic System) HI Across All Media =	1.0

TABLE 9.8.RME SUMMARY OF RECEPTOR HAZARDS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe:	Future
Receptor Population:	Resident
Recorder Age: Child	

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	ic Risk		Non-	Carcinogenic	Hazard Quotie	nt	
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Tota
Groundwater	Groundwater	Tapwater	Aluminum			-		-	Central nervous system	2.5		0.017	2.5
Zone 3			Arsenic					-	Skin	87.8		0.58	88.4
			Beryllium					-	Gastrointestinal	0.65		0.61	1.3
			Cadmium						Kidney	8.0		1.1	9.1
			Cobalt						Thyroid	93.5	-	0.62	94.2
			Manganese					-	Central nervous system	29.0		4.8	33.8
			Molybdenum						Increased uric acid (kidney)	9.4		0.062	9.5
			Nickel						Reduced organ and body weights	1.6	-	0.052	1.6
			Vanadium						Decreased hair cystine	2.3		0.015	2.3
			Uranium						Kidney	0.92		0.0061	0.92
			Chloroform						Liver	0.020		0.0018	0.022
			Chemical Total							236		7.8	244
			Uranium-234										
			Uranium-235+D										
			Uranium-238+D										
			Radium-226+D										
			Radium-228+D										
			Thorium-230						_				
			Lead-210+D										
			Radionuclide Total										
		Exposure Point Total											244
Ē	Exposure N	Vedium Total										244	
, P	Air	Water Vapors	Aluminum								_		
		from Showerhead	Arsenic								_		
		nom Snowerneau	Beryllium										
			-										
			Cadmium										
			Cadmium					-	-				
			Cobalt					-	-	-	-	-	
			Cobalt Manganese				-	-			-		-
			Cobalt Manganese Molybdenum		 		-	-				 	
			Cobalt Manganese Molybdenum Nickel				-			-	-		
			Cobalt Manganese Molybdenum Nickel Vanadium					-			-		
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium								-		
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform						 Liver		-		0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chernical Total					-	Liver			 0.0040	0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chemical Total Uranium-234									 0.0040 	0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chloroform Chemical Total Uranium-234 Uranium-235+D						Liver			 0.0040	0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chemical Total Uranium-234 Uranium-234+D						Liver			 0.0040 	0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chemical Total Uranium-234 Uranium-235+D Uranium-238+D Radium-226+D						Liver			 0.0040 	0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Ehemical Total Uranium-234 Uranium-238+D Radium-228+D Radium-228+D						Liver			 0.0040 	0.0040 0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chloroform Chloroform Uranium-234 Uranium-238+D Radium-228+D Radium-228+D Thorium-230					-	Liver			 0.0040 	0.0040 0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chemical Total Uranium-234 Uranium-238+D Radium-228+D Radium-228+D Thorium-230 Lead-210+D									 0.0040 	0.0040 0.0040
			Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chloroform Chloroform Uranium-234 Uranium-238+D Radium-228+D Radium-228+D Thorium-230					-	Liver			 0.0040 	0.0040 0.0040
		Exposure Point Total	Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chemical Total Uranium-234 Uranium-238+D Radium-228+D Radium-228+D Thorium-230 Lead-210+D									 0.0040 	0.0040 0.0040 -
	Exposure N	Exposure Point Total Medium Total	Cobalt Manganese Molybdenum Nickel Vanadium Uranium Chloroform Chemical Total Uranium-234 Uranium-238+D Radium-228+D Radium-228+D Thorium-230 Lead-210+D									 0.0040 	0.0040 0.0040

Total Skin HI Across All Media =	88.4
Total Thyroid HI Across All Media =	94.2
Total Central Nervous System HI Across All Media =	36.3
Total Kidney HI Across All Media =	19.5
Total Liver HI Across All Media =	0.026
Total Gastrointestinal HI Across All Media =	1.3
Total Reduced Body and Organ Weights HI Across All Media =	1.6
Total Hair Cystine (Metabolic System) HI Across All Media =	2.3

TABLE 9.9.RME SUMMARY OF RECEPTOR RISKS FOR COPCs REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timefr	ame: Future
Receptor Popula	ation: Resident
Receptor Age:	Child/Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	c Risk		Non-	-Carcinogenic F	Hazard Quotier	t	
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total
Groundwater	Groundwater	Tapwater	Aluminum										
Zone 3			Arsenic	9.2E-03		5.3E-05		9.3E-03					
			Beryllium										
			Cadmium										
			Cobalt										
			Manganese										
			Molybdenum										
			Nickel										
			Vanadium										
			Uranium										
			Chloroform	1.5E-06		1.3E-07		1.6E-06					
			Chemical Total	9.2E-03		5.3E-05		9.3E-03					
			Uranium-234	2.0E-05				2.0E-05					
			Uranium-235+D	9.1E-07				9.1E-07					
			Uranium-238+D	2.4E-05				2.4E-05					
			Radium-226+D	8.1E-05				8.1E-05					
			Radium-228+D	3.5E-04				3.5E-04					
			Thorium-230	4.5E-07				4.5E-07					
		Lead-210+D	5.5E-05				5.5E-05						
	i		Radionuclide Total	5.3E-04				5.3E-04					
		Exposure Point Total						9.8E-03					
		ledium Total					-	9.8E-03		1	1		
	Air	Water Vapors	Aluminum										
		from Showerhead	Arsenic				-						
			Beryllium										
			Cadmium										
			Cobalt										
			Manganese										
			Molybdenum Nickel										
			Vanadium										
			Uranium										
			Chloroform		 2.0E-06			 2.0E-06					
			Chemical Total		2.0E-06			2.0E-06					-
			Uranium-234 Uranium-235+D					-					
1 1					-								
			Uranium-238+D										
			Uranium-238+D Radium-226+D		1.2E-02			1.2E-02	-				-
			Uranium-238+D Radium-226+D Radium-228+D		1.2E-02 								
			Uranium-238+D Radium-226+D Radium-228+D Thorium-230		1.2E-02 			1.2E-02 					
			Uranium-238+D Radium-226+D Radium-228+D Thorium-230 Lead-210+D		1.2E-02 			1.2E-02 					
		Exposure Point Total	Uranium-238+D Radium-226+D Radium-228+D Thorium-230		1.2E-02 			1.2E-02 1.2E-02					
	Exposure M	Exposure Point Total edium Total	Uranium-238+D Radium-226+D Radium-228+D Thorium-230 Lead-210+D		1.2E-02 			1.2E-02 1.2E-02 1.2E-02					
Medium Total	Exposure M		Uranium-238+D Radium-226+D Radium-228+D Thorium-230 Lead-210+D		1.2E-02 			1.2E-02 1.2E-02					

TABLE 10.1.RME RISK ASSESSMENT SUMMARY - NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential		Carcinogenic Risk Non-Carcinogenic Hazard Quotient								
			Concern	Ingestion	Ingestion Inhalation Dermal External				Primary	Ingestion	Inhalation	Dermal	Exposure
				(Radiation) Routes Total Target Organ(s)							Routes Total		
Groundwater	Groundwater	Tapwater	Cobalt						Thyroid	0.91		0.005	0.92
SW Alluvium			Manganese						Central nervous system	3.20		0.42	3.6
			Uranium						Kidney	1.17		0.006	1.2
			Chemical Total							5.3		0.43	5.7
		Exposure Point Total											5.7
	Exposure Medium Total												5.7
Medium Total	ledium Total												5.7
Receptor Total	eceptor Total					Recepto	r Risk Total				Rece	ptor HI Total	5.7

 Total Thyroid HI Across All Media =
 0.92

 Total Central Nervous System HI Across All Media =
 3.6

1.2

Total Kidney HI Across All Media =

TABLE 10.2.RME

RISK ASSESSMENT SUMMARY - NON-CANCER HAZARDS

REASONABLE MAXIMUM EXPOSURE

UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child

Medium	Exposure Medium	Exposure Point	Chemical of Potential		Carcinogenic Risk Non-Carcinogenic Hazard Quotient								
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total
Groundwater	Groundwater	Tapwater	Cobalt						Thyroid	2.1		0.014	2.1
SW Alluvium			Manganese						Central nervous system	7.5		1.2	8.7
			Uranium						Kidney	2.7		0.018	2.7
			Chemical Total							12.3		1.3	13.6
		Exposure Point Total											13.6
	Exposure N	Medium Total											13.6
Medium Total	Medium Total												13.6
Receptor Total	eceptor Total					Recepto	r Risk Total				Rece	eptor HI Total	13.6

Total Thyroid HI Across All Media = 2.1 8.7

2.7

Total Central Nervous System HI Across All Media =

Total Kidney HI Across All Media =

TABLE 10.3.RME RISK ASSESSMENT SUMMARY - CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child/Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential		Carcinogenic Risk Non-Carcinogenic Hazard Quotient								
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total
Groundwater	Groundwater	Tapwater	Arsenic	5.7E-05		3.3E-07		5.8E-05					
SW Alluvium			Chloroform	1.6E-06		1.4E-07		1.7E-06					
			Chemical Total	5.9E-05		4.7E-07		5.9E-05					
			Uranium-234	5.8E-05				5.8E-05					
			Uranium-235+D	2.7E-06				2.7E-06					
			Uranium-238+D	7.0E-05				7.0E-05					
			Radium-226+D	1.9E-06				1.9E-06					
			Radium-228+D	1.7E-05				1.7E-05					
			Radionuclide Total	1.5E-04				1.5E-04					
		Exposure Point Total						2.1E-04					
	Exposure N	Medium Total						2.1E-04					
Ī		Water Vapors	Chloroform		2.1E-06			2.1E-06					
		from Showerhead	Chemical Total		2.1E-06			2.1E-06					
			Radium-226+D		2.9E-04			2.9E-04					
			Radionuclide Total		2.9E-04			2.9E-04					
		Exposure Point Total						2.9E-04					
	Exposure N						2.9E-04						
Medium Total						5.0E-04							
Receptor Total					Recepto	r Risk Total	5.0E-04			Rece	eptor HI Total		

TABLE 10.4.RME RISK ASSESSMENT SUMMARY - NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential		Carcinogenic Risk Non-Carcinogenic Hazard Quotient								
			Concern	Ingestion	Inhalation	Dermal	External	Exposure	Primary	Ingestion	Inhalation	Dermal	Exposure
							(Radiation)	Routes Total	Target Organ(s)				Routes Total
Groundwater	Groundwater	Tapwater	Cobalt						Thyroid	5.1		0.03	5.1
Zone 1			Manganese						Central nervous system	2.2		0.3	2.5
			Vanadium						Decreased hair cystine	1.1		0.006	1.1
			Chemical Total							8.4		0.32	8.7
		Exposure Point Total											8.7
	Exposure N											8.7	
Medium Total	ledium Total												8.7
Receptor Total	eceptor Total					Recepto	r Risk Total				Rece	ptor HI Total	8.7

Total Thyroid HI Across All Media = 5.1 2.5 1.1

Total Central Nervous System HI Across All Media =

Decreased Hair Cystine (Metabolic System) HI Across All Media =

TABLE 10.5.RME RISK ASSESSMENT SUMMARY - NON-CANCER HAZARDS

REASONABLE MAXIMUM EXPOSURE

UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child

Medium	Exposure Medium	Exposure Point	Chemical of Potential		Carcinogenic Risk Non-Carcinogenic Hazard Quotient							ient	
			Concern	Ingestion Inhalation Dermal External				Exposure	Primary	Ingestion	Inhalation	Dermal	Exposure
					(Radiation) Routes Total Target Organ(s)								Routes Total
Groundwater	Groundwater	Tapwater	Cobalt						Thyroid	11.9		0.078	11.9
Zone 1			Manganese						Central nervous system	5.2		0.86	6.1
			Vanadium						Decreased hair cystine	2.6		0.017	2.6
			Chemical Total							19.6		0.95	20.6
		Exposure Point Total											20.6
	Exposure Medium Total												20.6
Medium Total	ledium Total												20.6
Receptor Total	eceptor Total					Recepto	r Risk Total				Rece	ptor HI Total	20.6

id HI Across All Media = 11.9

Total Thyroid HI Across All Media =

Total Central Nervous System HI Across All Media = 6.1

Decreased Hair Cystine (Metabolic System) HI Across All Media = 2.6

TABLE 10.6.RME RISK ASSESSMENT SUMMARY - CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child/Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	ic Risk		N	on-Carcinogeni	ic Hazard Quot	ient	
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary Target Organ(s)	Ingestion	Inhalation	Dermal	Exposure Routes Total
Groundwater	Groundwater	Tapwater	Arsenic	3.2E-05		1.9E-07		3.3E-05					
Zone 1			Chemical Total	3.2E-05		1.9E-07		3.3E-05					
			Radium-226+D	8.8E-06				8.8E-06					
			Radium-228+D	4.1E-05				4.1E-05					
			Thorium-230	1.1E-06				1.1E-06					
			Radionuclide Total	5.1E-05				5.1E-05					
		Exposure Point Total	-					8.4E-05					
	Exposure N	ledium Total						8.4E-05					
	Air	Water Vapors	Radium-226+D		1.3E-03			1.3E-03					
		from Domestic Use	Radionuclide Total		1.3E-03			1.3E-03					
		Exposure Point Total						1.3E-03					
	Exposure N	/ledium Total						1.3E-03					
Medium Total	dium Total							1.4E-03					
Receptor Total	ceptor Total					Recepto	r Risk Total	1.4E-03			Rece	eptor HI Total	

TABLE 10.7.RME RISK ASSESSMENT SUMMARY - NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	c Risk		Non-(Carcinogenic I	Hazard Quotier	nt	
			Concern	Ingestion Inhalation Dermal External				Exposure	Primary	Ingestion	Inhalation	Dermal	Exposure
							(Radiation)	Routes Total	Target Organ(s)				Routes Total
Groundwater	Groundwater	Tapwater	Aluminum						Central nervous system	1.1		0.0056	1.1
Zone 3			Arsenic						Skin	37.6		0.20	37.8
			Cadmium						Kidney	3.4		0.36	3.8
			Cobalt						Thyroid	40.1		0.21	40.3
			Manganese						Central nervous system	12.4		1.6	14.1
			Molybdenum						Increased uric acid (kidney)	4.0		0.021	4.1
			Nickel						Reduced organ and body weights	0.67		0.017	0.69
			Vanadium						Decreased hair cystine	1.0		0.0051	1.0
			Uranium						Kidney	0.39		0.0021	0.40
			Chemical Total							101		2.44	103
		Exposure Point Total											103
	Exposure Medium Total												103
Medium Total													
Receptor Total					Recepto	r Risk Total				Rece	eptor HI Total	103	

 Total Skin HI Across All Media =
 37.8

 Total Thyroid HI Across All Media =
 40.3

 Total Central Nervous System HI Across All Media =
 15.1

 Total Kidney HI Across All Media =
 8.3

 Total Reduced Body and Organ Weights HI Across All Media =
 0.69

 Decreased Hair Cystine (Metabolic System) HI Across All Media =
 1.0

TABLE 10.8.RME RISK ASSESSMENT SUMMARY - NON-CANCER HAZARDS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	c Risk		Non-(Carcinogenic H	Hazard Quotier	nt	
			Concern	Ingestion Inhalation Dermal External				Exposure	Primary	Ingestion	Inhalation	Dermal	Exposure
							(Radiation)	Routes Total	Target Organ(s)				Routes Total
Groundwater	Groundwater	Tapwater	Aluminum						Central nervous system	2.5		0.017	2.5
Zone 3			Arsenic						Skin	87.8		0.58	88.4
			Cadmium						Kidney	8.0		1.1	9.1
			Cobalt						Thyroid	93.5		0.6	94.2
			Manganese						Central nervous system	29.0		4.8	33.8
			Molybdenum						Increased uric acid (kidney)	9.4		0.062	9.5
			Nickel						Reduced organ and body weights	1.6		0.052	1.6
			Vanadium						Decreased hair cystine	2.3		0.015	2.3
			Uranium						Kidney	0.9		0.0061	0.92
			Chemical Total							235.1		7.2	242
		Exposure Point Total											242
	Exposure M										242		
Medium Total													242
Receptor Total					Recepto	r Risk Total				Rece	ptor HI Total	242	

 Total Skin HI Across All Media =
 88.4

 Total Thyroid HI Across All Media =
 94.2

 Total Central Nervous System HI Across All Media =
 36.3

 Total Kidney HI Across All Media =
 19.5

 Total Reduced Body and Organ Weights HI Across All Media =
 1.6

 Decreased Hair Cystine (Metabolic System) HI Across All Media =
 2.3

TABLE 10.9.RME RISK ASSESSMENT SUMMARY - CANCER RISKS REASONABLE MAXIMUM EXPOSURE UNC Church Rock Mill and Tailings Site

Scenario Timeframe: Future Receptor Population: Resident Receptor Age: Child/Adult

Medium	Exposure Medium	Exposure Point	Chemical of Potential			Carcinogeni	c Risk		Non-Carcinogenic Hazard Quotient						
			Concern	Ingestion	Inhalation	Dermal	External (Radiation)	Exposure Routes Total	Primary	Ingestion	Inhalation	Dermal	Exposure Routes Total <t< td=""></t<>		
One we down to a	Groundwater	Townstow	Arsenic	0.05.00		5 0F 05	,		Target Organ(s)						
Groundwater	Groundwater	Tapwater	Chloroform	9.2E-03		5.3E-05		9.3E-03							
Zone 3				1.5E-06		1.3E-07		1.6E-06							
			Chemical Total	9.2E-03		5.3E-05		9.3E-03	I						
			Uranium-234	2.0E-05				2.0E-05							
			Uranium-238+D	2.4E-05				2.4E-05							
			Radium-226+D	8.1E-05				8.1E-05							
			Radium-228+D	3.5E-04				3.5E-04							
			Lead-210+D	5.5E-05				5.5E-05							
			Radionuclide Total	5.3E-04				5.3E-04							
		Exposure Point Total						9.8E-03							
	Exposure N	ledium Total						9.8E-03							
	Air	Water Vapors	Chloroform		2.0E-06			2.0E-06							
		from Showerhead	Chemical Total		2.0E-06			2.0E-06							
			Radium-226+D		1.2E-02			1.2E-02							
			Radionuclide Total		1.2E-02			1.2E-02							
		Exposure Point Total	1		•		·	1.2E-02		•					
	Exposure N	ledium Total						1.2E-02							
Medium Total	<u> </u>							2.2E-02							
Receptor Total				u		Receptor	r Risk Total	2.2E-02			Rece	ptor HI Total			

APPENDIX B

Revised Submittal – Estimated UCL95 Statistics and EPCs in Impacted Groundwater, UNC Church Rock Mill and Tailings Site, Church Rock, New Mexico. December 5, 2008



N.A. WATER SYSTEMS

December 5, 2008

This Submittal Delivered by Email Only

Ref. No. 56007746 GE Church Rock Project

Mr. Mark Purcell Remedial Project Manager U.S. Environmental Protection Agency 1445 Ross Ave., Suite 1200 (6SF-LP) Dallas, TX 75202-2733

Re: Revised Submittal Estimated UCL95 Statistics and EPCs in Impacted Groundwater UNC Church Rock Mill & Tailings Site, Church Rock, New Mexico

Dear Mr. Purcell:

N.A. Water Systems (N.A.WS) is pleased to provide this revised report on the calculation of statistics for the estimation of exposure point concentrations (EPCs) in impacted groundwater at UNC's Church Rock Mill & Tailings Site in Church Rock, New Mexico. This report includes descriptions of the methods used to classify sample data, the statistical methods, and the estimation results.

The post-mining/pre-tailings water is referred to as background water, and the post-mining/post-tailings water is referred to as impacted water.

Introduction

Statistical analyses for the task of estimating exposure point concentrations (EPCs) have been completed for impacted groundwater in each of the three hydrostratigraphic zones at the Church Rock Site. Statistics were calculated for the 21 contaminants of potential concern (COPCs) included in the Church Rock Sampling and Analysis Plan. The estimates were made using the current version of the EPA's ProUCL software (ver. 4.00.02) as prescribed by the EPA and reiterated in the teleconference on June 27, 2008. The teleconference participants included representatives of U.S. Environmental Protection Agency (EPA), New Mexico Environment Department (NMED), and N.A.WS. The methodology is summarized by the following steps:

1. Classify sample data for the purpose of forming logical groupings for EPC estimation. The criteria used for these classifications include the sampled



hydrostratigraphic unit (i.e. the Southwest Alluvium (SWA), Zone 1, and Zone 3), determination of unequivocal impact from seepage fluids, representation of recent conditions (i.e. the most recent eight quarters of sampling), and location relative to administrative boundaries.

- 2. Use ProUCL software to estimate the upper confidence limits (UCL95) for the population means of COPC concentrations from sample groups determined to be representative of impacted groundwater quality.
- 3. Estimate EPCs in impacted groundwater for those COPCs for which valid UCL95 statistics have been estimated.

Classification of Samples

Identification of Samples Representative of Impacted Groundwater Quality

With respect to water quality, three exclusive classes of groundwater samples are germane to the estimation of EPCs. Those classes are post-mining/pre-tailings (background), post-mining/post-tailings (impacted), and other. For present purposes, the "other" class is meant to represent any samples that are not clearly representative of either background or impacted quality. These may include water whose quality is interpreted to be transitional or that is representative of pre-mining conditions.

Samples representative of background groundwater quality were identified for the SWA and Zone 1 in the license amendment request for changing the Groundwater Protection Standard for radium (N.A. Water Systems, February 2006, *Technical Analysis Report in Support of License Amendment Request for Changing the Method of Deter mining Exceedances of the Com bined Radium Groundwater Protection Standard in Source Materials License SUA-1475 (T AC LU0092), Groundwater Corrective Action Program , Church Rock Site, Church Rock, New Mexico, pp. 3-6). The same methods were used to identify samples from Zone 3 that are representative of background water quality (N.A. Water Systems, October 17, 2008, letter to Mark Purcell (EPA), <i>Calculation of Background Statistics with Comparison Values, UNC Church Rock Mill & Tailings Site, Church Rock, New Mexico*).

The methods used to identify wells having background water quality for the February 2006 and October 2008 submittals had as their essential criterion the absence of evidence of seepage impact. By extension, the same methods may be used to identify evidence of seepage impact. Samples where evidence of seepage impact was equivocal or clearly absent were excluded from the calculations presented in this report.

Mark Purcell U.S. EPA December 5, 2008

The data sets used in calculations made for this report are from the period July 2006 through April 2008 inclusive, which represents the most recent eight quarters of sampling available at the time of the calculations. This time frame was selected to be representative of recent conditions, while providing at least the minimum recommended number of samples to satisfy the requirements of the statistical methods. For this reason, the estimation of UCL95 statistics and EPCs extend only to the 21 current COPCs and do not include trace metals (plus iron) that had previously been dismissed as COPCs (EPA, *August 1988, Draft Final Remedial Investigation, United Nuclear Church Rock Site*). Table 1 lists wells interpreted as having samples representative of impacted groundwater during the most recent 8 quarters.

Grouping of Samples by Hydrostratigraphic Zone and by Administrative Area

The data sets used to calculate statistics were subdivided by hydrostratigraphic zone and by geographic location. The three hydrostratigraphic zones by which sample data were grouped are the SWA, Zone 1, and Zone 3. The geographic grouping resulted in the elimination from Zone 1 and Zone 3 datasets of sample data from wells within Section 2 of Township 16 North, Range 16 West. This discrimination of Section 2 data was based on two considerations. One consideration is that Section 2 encompasses the tailings disposal area, which will eventually be administered by the U.S Department of Energy (DOE). As such, groundwater exposure within Section 2 will be prohibited by DOE controls. The second consideration is that the more extreme effects of seepage impact evident in Zone 1 and Zone 3 wells proximal to the tailings disposal cells are not expected to migrate and occupy areas outside of Section 2. This judgment is based on the following conclusions:

- 1. The tailings cells are no longer a source of measurable quantities of seepage fluid (US Filter, January 19, 2004, *Rationale and Field Investigation Work Plan to Evaluate Recharge and Potential Cell So urcing to the Zone 3 Plum e, Church Rock Site, Gallup, New Mexico*).
- 2. Reductions of saturated thickness and diminishment of porosity and hydraulic conductivity (by geochemical reactions) will continue to reduce groundwater flux across the boundary of Sections 2 and 36 to less than the 0.5 gallons per minute estimated to have occurred in January 2005 (N.A. Water Systems, April 25, 2008, Recommendations and Summary of Hydrogeologic Analysis, Evaluation of Groundwater Flow in Zone 3 for the Desi gn of a Pumping System to Intercept Impacted Groundwater, United Nuclear Corporation's Church Rock Tailings Site, Gallup, New Mexico).

3. Evidence from groundwater sampling indicates that water quality in Zone 1 Point of Compliance Wells has been improving since the third quarter of 1989 (N.A. Water Systems, February 2006, Technical Analysis Report in Support of License Amendment Request for Changing the Me thod of Determining Ex ceedances of the Combined Radium Groundwater Prot ection Standard in Source Materials License SUA-1475 (TAC LU0092), Groun dwater Corrective Action Program, Church Rock Site, Church Rock, New Mexico; and N.A. Water Systems, January 2008, Annual Review Report 2007 – Groundwater er Corrective Action, Church Rock Site, Church Rock, New Mexico).

Results

Basic Statistics and Upper Confidence Limits for Means

Tables 2 through 4 list summary and UCL95 statistics for all COPCs calculated from the impacted data sets from wells in the SWA, Zone 1, and Zone 3. The data sets include only primary samples (i.e. no QA/QC samples). The samples were collected over the most recent eight quarters (July 2006 through April 2008) of data presently available. Eight quarters were selected as a compromise between the objective of representing current (or recent) conditions and the objective of having sample populations of sufficient size to estimate meaningful statistics. Probability (normal) plots of each dataset are provided in Appendix B.

All of the statistics were calculated using ProUCL software (Singh et al., April 2007, ProUCL Version 4.00.02 User Guide, EPA/600/R-07/038). The UCL95 estimates were selected from values recommended by the ProUCL software. One exception was made for a recommended UCL statistic (for nitrate as nitrogen, NO3 as N, in Zone 1, see Table 3) that exceeded the maximum detected value. Summary tables of the output of UCL95 estimates are provided in Appendix A. In cases where two alternative estimates of UCL95 statistics are provided by ProUCL, the higher value was selected and is listed in Tables 2 through 4, except in those cases where the software issued a warning that the higher value may be unreliable (typically because of the limitations of bootstrap methods with small sample sizes). In one case (manganese, Mn, in Zone 1, see Tables 3, A.2) ProUCL recommended three alternative UCL statistics. In this case the highest value was not selected, because it was based on an assumption that the population followed a log-normal distribution. Prior testing of a much larger background sample data set indicated that this distributional assumption is probably incorrect. The higher values were selected as conservative estimates, consistent with the use of these same statistics as estimators of exposure point concentrations (EPCs).

The numbers of distinct detected values were too few to calculate UCL95 estimates for a significant fraction of COPCs: nine in the southwest alluvium (SWA, Table 2), seven in Zone 1 (Table 3), and one in Zone 3 (Table 4). UCL95 statistics also could not be estimated for any of the trace metals not included among the analytes in the past eight quarters of sample analyses. Current COPCs lacking sufficient data to estimate UCL95 statistics are summarized by hydrostratigraphic zone in Table 5. Table 6 summarizes current COPCs having UCL95 estimates whose reliability may be suspect, according to warnings issued by ProUCL. Such warnings typically apply to datasets having four or fewer distinct detected values.

Discussion

The constituents listed in Table 5 lack sufficient data (numbers of detections) to statistically quantify EPCs. The constituents listed in Table 6 have estimated UCL95 statistics that may not be sufficiently reliable to estimate EPCs. However, the constituents listed in Tables 5 and 6 either have not been detected in the past two years or have been detected infrequently and for the most part at concentrations below MCLs (or other applicable standards, if lacking MCLs). The exceptions are vanadium in Zone 1 (detected once at 0.2 mg/L), and vanadium in Zone 3 (detected four times at 0.2 mg/L). Therefore, the most of these constituents would be unlikely to present an unacceptable risk, even if there was a basis to quantify their EPC concentrations. Furthermore, with the few noted exceptions, these constituents would not be characterized as COPCs in their respective hydrostratigraphic zones if sampling data of the most recent two years were used to make such a determination.

Conclusion

This submittal provides estimates of UCL95 statistics for COPCs in impacted groundwater applicable to all areas in the SWA and to areas outside Section 2 in Zones 1 and 3. The estimated statistics provide a basis for estimating EPCs for those COPCs more likely to make a substantive contribution to quantifiable risk. Those COPCs for which UCL95 statistics cannot be reliably estimated were found, with few exceptions, to be unlikely to contribute substantively to quantifiable risk.

The UCL95 statistics provided in this submittal complement those submitted previously for background groundwater (N.A. Water Systems, October 17, 2008), which are bases for estimating EPCs in areas of the three hydrostratigraphic zones characterized as having background groundwater quality.

Mark Purcell U.S. EPA December 5, 2008

Very Truly Yours,

an a G5

James Ewart, Ph.D., P.G. Technical Consultant

- JE: abc-191
- cc: Roy Blickwedel, GE Larry Bush, UNC Earle C. Dixon, NMED

Attachments

N.A. WATER SYSTEMS

Tables

VEOLIA WATER Solutions & Technologies

N.A. Water Systems, LLC Airside Business Park, 250 Airside Drive Moon Township, Pennsylvania 15108-2793 USA Tel: 412-809-6000 • Fax: 412-809-6075 Web site: www.nawatersystems.com

Wells Having Samples Representative of Impacted Water Quality, July 2006 - April 2008

Southwest Alluvium	Zone 1	Zone 3
509 D (POC)	515 A	504 B
624	604 (POC)	517 (POC)
632 (POC)	614 (POC)	613 (POC)
801	EPA 5 (POC)	708 (POC)
802	EPA 7 (POC)	711 (POC)
803		717
808		719
EPA 23 (POC)		EPA 13
EPA 25		EPA 14
GW 1 (POC)		NBL 1
GW 2 (POC)		
GW 3 (POC)		

Notes:

POC = Point-of-Compliance Well.

			-			-		
		Total	Percent	Minimum	Maximum			
Parameter	Units	Data	Nondetect	Detected	Detected		Detected	of Mean
AI	mg/L	96		0.1			0.15	0.109
As	mg/L	96		0.001			0.01	0.00256
Be	mg/L	96		N/A		N/A	N/A	N/A
Cd	mg/L	96	100.0%	N/A	N/A	N/A	N/A	N/A
Со	mg/L	96	99.0%	0.01	0.01	N/A	N/A	N/A
Pb	mg/L	96	100.0%	N/A	N/A	N/A	N/A	N/A
Mn	mg/L	96	0.0%	0.03	5.4	1.865	1.83	2.8
Мо	mg/L	96	100.0%	N/A	N/A	N/A	N/A	N/A
Ni	mg/L	96	100.0%	N/A	N/A	N/A	N/A	N/A
Se	mg/L	96	99.0%	0.001	0.001	N/A	N/A	N/A
V	mg/L	96	100.0%	N/A	N/A	N/A	N/A	N/A
CI	mg/L	96	0.0%	79	374	187.8	181	199.6
SO4	mg/L	96	0.0%	1510	4330	2745	2820	2867
NO3_as_N	mg/L	96	0.0%	0.3	160	65.08	76	94.42
U	mg/L	96	0.0%	0.0229	0.246	0.104	0.111	0.128
Chloroform	mg/L	96	49.0%	0.00061	0.0155	0.00479	0.00309	0.00338
Lab_TDS	mg/L	96	0.0%	3880	8250	6044	6245	6250
Rad-226	pCi/L	96	61.5%	0.1	1	0.435	0.4	0.267
Rad-228	pCi/L	96	77.1%	0.3		1.786	1.75	0.86
Rad_totl	pCi/L	96	57.3%	0.1	5.2	1.351	0.7	0.828
Th-230	pCi/L	96	89.6%	0.2	1.6	0.69	0.5	0.29
Pb-210	pCi/L	96	100.0%	N/A	N/A	N/A	N/A	N/A
Gross_Alpha	pCi/L	96	69.79%	1	2.4	1.317	1.2	1.141
Sb	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ba	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Cr	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Cu	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Fe	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Hg	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ag	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
TĪ	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Zn	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A

Summary Statistics for COPCs and Trace Metals in Southwest Alluvium Impacted Groundwater

Notes:

N/A - insufficient data to make an estimate.

UCL95 statistics highlighted in yellow may be of questionable reliability.

Listed UCL statistics for Mn and NO3_as_N are at 97.5% confidence level

		Total	Percent	Minimum	Maximum		Median of	UCL95
Parameter	Units	Data	Nondetect		Detected	Detected	Detected	of Mean
AI	mg/L	16		0.2	1.3		0.3	0.44
As	mg/L	16	75.0%	0.001	0.003			0.00145
Be	mg/L	16	100.0%	N/A	N/A	N/A	N/A	N/A
Cd	mg/L	16	100.0%	N/A	N/A	N/A	N/A	N/A
Co	mg/L	16	0.0%	0.02	0.06	0.0363	0.03	0.0557
Pb	mg/L	16	100.0%	N/A	N/A	N/A	N/A	N/A
Mn	mg/L	16	0.0%	0.95	2.96	1.656	1.47	1.95
Мо	mg/L	16	100.0%	N/A	N/A	N/A	N/A	N/A
Ni	mg/L	16	81.3%	0.05	0.06	0.0533	0.05	0.0519
Se	mg/L	16	93.8%	0.001	0.001	N/A	N/A	N/A
V	mg/L	16	93.8%	0.2	0.2	N/A	N/A	N/A
CI	mg/L	16	0.0%	48	221	131.5	128.5	214.3
SO4	mg/L	16	0.0%	2960	4760	3778	3955	4049
NO3_as_N	mg/L	16	0.0%	16.2	200	80.5	72.75	152*
U	mg/L	16	0.0%	0.0012	0.0022	0.00161	0.0015	0.00174
Chloroform	mg/L	16	87.5%	0.0006	0.00076	0.00068	0.00068	0.00063873
Lab_TDS	mg/L	16	0.0%	4620	7860	6208	6120	6843
Rad-226	pCi/L	16	18.8%	0.4	1.8	1.138	1.2	1.213
Rad-228	pCi/L	16	56.3%	1	4	2.286	1.9	2.087
Rad_totl	pCi/L	16	12.5%	0.6	5.2	2.2	1.6	2.8
Th-230	pCi/L	16	87.5%	0.6	0.7	0.65	0.65	0.621
Pb-210	pCi/L	16	100.0%	N/A	N/A	N/A	N/A	N/A
Gross_Alpha	pCi/L	16	18.8%	1.2	4.1	2.146	2	2.319
Sb	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ba	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Cr	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Cu	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Fe	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Hg	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ag	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
TĪ	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Zn	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A

Summary Statistics for COPCs and Trace Metals in Zone 1 Impacted Groundwater, Recent 8 Quarters 3rd Qtr. 2006 - 2nd Qtr. 2008, Excluding Samples from Section 2 Wells 0515 A, 0604, 0614

Notes:

*95% Chebyshev (Mean, Sd) UCL chosen for NO3_as_N in lieu of ProUCL recommended UCL 99 statistic,

which exceeded the maximum observed detection.

N/A - insufficient data to make an estimate.

UCL95 statistics highlighted in yellow may be of questionable reliability.

					-			
		Total	Percent		Maximum	Mean of	Median of	UCL95
Parameter	Units	Data	Nondetect	Detected	Detected	Detected	Detected	of Mean
AI	mg/L	70	17.1%	0.1	163	16.14	2.45	39.15
As	mg/L	70	31.4%	0.001	2.5	0.206	0.025	0.412
Be	mg/L	70	87.1%	0.01	0.09	0.0589	0.06	0.0202
Cd	mg/L	70	77.1%	0.005	1	0.0713	0.0095	0.0628
Co	mg/L	70	0.0%	0.05	0.95	0.381	0.35	0.439
Pb	mg/L	70	100.0%	N/A	N/A	N/A	N/A	N/A
Mn	mg/L	70	0.0%	3.33	23.7	9.836	7.485	10.89
Мо	mg/L	70	54.3%	0.1	5	1.084	0.3	0.739
Ni	mg/L	70	0.0%	0.11	0.89	0.377	0.31	0.489
Se	mg/L	70	95.7%	0.001	0.01	0.00433	0.002	0.0014
V	mg/L	70	92.9%	0.1	0.2	0.18	0.2	0.111
CI	mg/L	70	0.0%	14	98	43.66	37.5	48.01
SO4	mg/L	70	0.0%	2630	5260	3599	3545	3717
NO3_as_N	mg/L	70	61.4%	0.1	44.8	17.15	24	16.09
U	mg/L	70	0.0%	0.0011	0.138	0.0287	0.0219	0.0431
Chloroform	mg/L	70	81.4%	0.00093	0.00676	0.00441	0.00444	0.00326
Lab_TDS	mg/L	70	0.0%	3980	6680	5289	5290	5441
Rad-226	pCi/L	70	0.0%	2	27.6	9.823	7.9	11.14
Rad-228	pCi/L	70	0.0%	3.8	56.1	15.73	13.55	17.84
Rad_totl	pCi/L	70	0.0%	6.8	73.3	25.55	20.8	29.14
Th-230	pCi/L	70	91.4%	0.2	1.3	0.533	0.4	0.259
Pb-210	pCi/L	70	91.4%	1.8	8.1	4.883	4.9	2.287
Gross_Alpha	pCi/L	70	0.0%	2.4	35.2	12.62	10.55	14.25
Sb	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ва	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Cr	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Cu	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Fe	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Hg	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ag	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
TI	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Zn	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A

Summary Statistics for COPCs and Trace Metals in Zone 3 Impacted Groundwater, Recent 8 Quarters 3rd Qtr. 2006 - 2nd Qtr. 2008, Exluding Samples from Section 2 Well 0613

Notes:

N/A - insufficient data to make an estimate.

UCL95 statistics highlighted in yellow may be of questionable reliability.

Listed UCL statistics for AI, As, and NO3_as_N are at 97.5% confidence level

COPCs Lacking Sufficient Data to Estimate UCL95 Statistics for Impacted Water Quality, July 2006 - April 2008

Southwest Alluvium	Zone 1	Zone 3
Be*	Be*	Pb*
Cd*	Cd*	
Co***	Pb*	
Pb*	Mo*	
Mo*	Se**	
Ni*	V	
Se**	Pb-210*	
V*		
Pb-210*		

Notes:

* no detected results in 8 quarters of sampling.

** one detected result at or below MCL in 8 quarters of sampling.

*** one detected result at or below New Mexico WQCC standard in 8 quarters of sampling.

COPCs Having Estimated UCL95 Statistics of Questionable Reliability for Impacted Water Quality, July 2006 - April 2008

Southwest Alluvium	Zone 1	Zone 3
AI***	Ni** Chloroform* Th-230*	Se** V

Notes:

* 2 detected results at or below MCL or NRC compliance license standard in 8 quarters of sampling.

** 3 detected results at or below MCL or New Mexico WQCC standard in 8 quarters of sampling.

*** 6 detected results at or below New Mexico WQCC standard in 8 quarters of sampling.

N.A. WATER SYSTEMS

Appendix A

Summary Tables of Output from ProUCL for UCL95 Statistics



N.A. Water Systems, LLC Airside Business Park, 250 Airside Drive Moon Township, Pennsylvania 15108-2793 USA Tel: 412-809-6000 • Fax: 412-809-6075 Web site: www.nawatersystems.com

UCL95 Statistics for Southwest Alluvium Impacted Data Sets with Non-Detects

TABLE A1

	Al	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	v	CI	SO4	NO3 as N	U	Chloroform	Lab TDS	Rad-226
Total Number of Data	96	-	-	96	96	96	96	96		96	96	96	96	96	96		96	96
Number of Non-Detect Data	90			96	95	96	0	96		95	96	0	0	0	0	47	0	59
Number of Detected Data (or Distinct Obs. If zero							-										-	
nondetect)	6	13	0	0	1	0	79	0	0	1	0	72	80	75	76	49	82	37
Minimum Detected	0.1	0.001			0.01		0.03			0.001		79	1510	0.3	0.0229	0.00061	3880	0.1
Maximum Detected	0.3	0.01			0.01		5.4			0.001		374	4330	160	0.246	0.0155	8250	1
Percent Non-Detects	93.75%	86.46%	100.0%	100.0%	99.0%	100.0%	0.00%	100.0%	100.0%	99.00%	100.0%	0.0%	0.0%	0.00%	0.00%	48.96%	0.0%	61.46%
Minimum Non-detect	0.1	0.001	0.01	0.005	0.01	0.05		0.01	0.05	0.001	0.1					0.0005		0.001
Maximum Non-detect	0.1	0.001	0.01	0.005	0.01	0.05		0.01	0.05	0.001	0.1					0.001		0.2
																		Í
Mean of Detected Data	0.167	0.00885					1.865					187.8	2745	65.08	0.104	0.00479	6044	0.435
Median of Detected Data	0.15	0.01					1.83					181	2820	76	0.111	0.00309	6245	0.4
Variance of Detected Data	0.00667	0.000006141					2.151					4459	521381	2118	0.00299	1.5975E-05	1483184	0.0596
SD of Detected Data	0.0816	0.00248					1.467					66.78	722.1	46.03	0.0546	0.004	1218	0.244
CV of Detected Data	0.49	0.28					0.787					0.356	0.263	0.707	0.525	0.834	0.202	0.561
Skewness of Detected Data	0.86	-3.05					0.696					0.78	0.05	-0.0574	0.632	1.28	-0.00205	0.536
Mean of Log-Transformed Detected Data	-1.888	-4.833					0.0407					5.172	7.881	3.523	-2.427	-5.686	8.686	-1.019
SD of Log-Transformed Detected Data	0.477	0.629					1.414					0.364	0.275	1.569	0.619	0.881	0.208	0.67
Discernable Distribution (0.05) of Detected Data	normal	none					none					gamma	none	none	none	none	none	normal
																		Í
Kaplan-Meier (KM) Method																		
Mean	0.104	0.00206					ł		1							0.00275		0.229
SD	0.0247	2.82E-03														0.00351		0.221
Standard Error of Mean	0.00276	3.00E-04														0.00036246		0.0229
95% KM (t) UCL	0.109	0.00256														0.00335		0.267
95% KM (z) UCL	0.109	0.00256														0.00334		0.267
95% KM (BCA) UCL	N/A	0.00908														0.00338		0.272
95% KM (Percentile Bootstrap) UCL	N/A	0.00905														0.00338		0.267
95% KM (Chebyshev) UCL	0.116	0.00337														0.00433		0.329
97.5% KM (Chebyshev) UCL	0.121	0.00394														0.00501		0.372
99% KM (Chebyshev) ÚCL	0.132	0.00505														0.00635		0.457
																		í Í
Datasets without Nondetects																		í Í
Student's-t UCL							2.113					199.1	2867	72.88	0.113		6250	í l
95% UCLs (Adjusted for Skewness)																		í Í
95% Adjusted-CLT UCL							2.122					199.6	2866	72.78	0.114		6248	í l
95% Modified-t UCL							2.115					199.2	2867	72.88	0.113		6250	
Non-Parametric UCLs																		í l
95% CLT UCL							2.111					199	2866	72.81	0.113		6248	í l
95% Jackknife UCL							2.113					199.1	2867	72.88	0.113		6250	
95% Standard Bootstrap UCL							2.106					198.9	2867	72.55	0.113		6246	
95% Bootstrap-t UCL							2.121					199.9	2864	73.19	0.113		6252	í l
95% Hall's Bootstrap UCL							2.12					199.9	2870	72.41	0.114		6257	
95% Percentile Bootstrap UCL							2.104					198.9	2866	72.69	0.114		6250	í
95% BCA Bootstrap UCL					İ		2.129					199.5	2862	72.77	0.114		6244	
95% Chebyshev(Mean, Sd) UCL							2.517					217.5	3066	85.56	0.128		6586	
97.5% Chebyshev(Mean, Sd) UCL							2.8					230.3	3205	94.42	0.139		6820	
99% Chebyshev(Mean, Sd) UCL					İ		3.354					255.6	3478	111.8	0.16		7281	
Potential UCL to Use																		
95% KM (t) UCL	0.109	0.00256														0.00335		0.267
95% KM (z) UCL																		
95% KM (BCA) UCL																		
95% KM (Percentile Bootstrap) UCL		0.00905														0.00338		0.267
95% KM (Chebyshev) UCL																		
97.5% KM (Chebyshev) UCL																		
99% KM (Chebyshev) UCL																		
95% Student's-t UCL							t t				l l		2867				6250	i i
95% Modified-t UCL							1						2867				6250	r i
95% Chebyshev(Mean, Sd) UCL									<u> </u>						0.128			
97.5% Chebyshev(Mean, Sd) UCL							2.8		<u> </u>					94.42				
95% Approximate Gamma UCL												199.6						
											l l							
Notes	2	5	4	4	3	4	6	4	4	3	4			6		1		1
		. .				-	-	-										

Notes:

1. Data have multiple DLs - Use of KM Method is recommended.

2. Warning: There are only 3 Distinct Detected Values in this data set.

The number of detected data may not be adequate enough to perform GOF tests, bootstrap, and ROS methods.

3. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values

determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).4. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs!

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV). 5. Warning: There are only 4 Distinct Detected Values in this data. It should be noted that bootstrap calculations may not be reliable enough to draw conclusions.

6. Potential UCL to use is at 97.6% confidence level

6 96 96 96 96 96 96 96 97 9 74 55 86 96 6 6 7 22 41 10 0 2 1 0.3 0.1 0.2 1 0 2 1 0.3 0.1 0.2 1 1 4.3 5.2 1.6 2.1 0.2 1 1.3 1 1.3 1 1.3 1 1.2 1.09 1.5 1.1 1.09 1.5 1.2 <td< th=""><th></th><th></th><th></th><th></th><th></th><th></th></td<>						
9 74 55 86 96 6 7 22 41 10 0 2 1 0.3 0.1 0.2 1 0.3 0.1 1 4.3 5.2 1.6 2 1 6 77.08% 57.29% 89.58% 100.0% 69.79% 1 0.04 0.2 0.2 1 0.2 2 1 0.2 0.2 1 0.2 5 1.786 1.351 0.69 1.31 4 1.75 0.7 0.5 1.1 6 0.449 1.12 1.09 1.5 9 0.316 -0.208 -0.607 0.24 1 normal none gamma non 9 0.693 0.634 0.251 1.09 1 0.799 1.039 0.216 0.24 9 0.6837 0.813 0.29 1.13 2 0.826 0.439 1.14 1 0.826 0.501 <td>5</td> <td>Rad-228</td> <td>Rad_totl</td> <td>Th-230</td> <td>Pb-210</td> <td>Gross_Alpha</td>	5	Rad-228	Rad_totl	Th-230	Pb-210	Gross_Alpha
7 22 41 10 0 2 1 0.3 0.1 0.2 1 1 4.3 5.2 1.6 2. 6 77.08% 57.29% 89.58% 100.0% 69.799 1 0.04 0.2 0.2 1 0.2 2 1 0.2 0.2 1 0.2 5 1.786 1.351 0.69 1.31 4 1.75 0.7 0.5 1. 6 1.269 1.673 0.257 0.13 4 1.126 1.293 0.507 0.36 1 0.631 0.957 0.734 0.24 7 0.829 1.085 0.729 0.24 7 0.835 0.634 0.251 1.09 1 0.799 1.039 0.216 0.24 9 0.693 0.634 0.252 0.025 7 0.836 0.813 0.29 1.13 1 0.828 0.439 1.14 <	6	96	96	96	96	96
1 0.3 0.1 0.2 1 4.3 5.2 1.6 2. 6 77.08% 57.29% 89.58% 100.0% 69.799 1 0.4 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 4 1.75 0.7 0.5 1.1 6 0.449 1.12 1.09 1.5 9 0.316 -0.208 -0.607 0.24 7 0.829 1.039 0.216 0.24 9 0.693 0.634 0.251 1.09 1 0.799 1.039 0.216 0.24 9 0.835 0.811 0.289 1.13 2 0.835 0.813 0.29 1.	9	74	55	86	96	67
1 0.3 0.1 0.2 1 4.3 5.2 1.6 2. 6 77.08% 57.29% 89.58% 100.0% 69.799 1 0.4 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 4 1.75 0.7 0.5 1.1 6 0.449 1.12 1.09 1.5 9 0.316 -0.208 -0.607 0.24 7 0.829 1.039 0.216 0.24 9 0.693 0.634 0.251 1.09 1 0.799 1.039 0.216 0.24 9 0.835 0.811 0.289 1.13 2 0.835 0.813 0.29 1.						
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1 0.04 0.2 0.2 1 0. 2 1 0.2 0.2 1 0. 5 1.786 1.351 0.69 1.31 4 1.75 0.7 0.5 1. 6 1.269 1.673 0.257 0.13 4 1.126 1.293 0.507 0.36 1 0.631 0.957 0.734 0.27 6 0.449 1.12 1.09 1.5 9 0.316 -0.208 -0.607 0.24 7 0.829 1.085 0.729 0.24 1 normal norma non 9 0.693 0.634 0.251 1.09 1 0.799 1.039 0.216 0.24 9 0.866 0.107 0.0232 0.025 7 0.835 0.828 0.439 1.14 7 0.866 0.828 0.439 1.25 7 1.102 0.352 1.20 1 1		77.08%			100.0%	69.79%
2 1 0.2 0.2 1 5 1.786 1.351 0.69 1.31 4 1.75 0.7 0.5 1. 6 1.269 1.673 0.257 0.13 4 1.126 1.293 0.507 0.36 1 0.631 0.957 0.734 0.27 6 0.449 1.12 1.09 1.5 9 0.316 -0.208 -0.607 0.24 7 0.829 1.085 0.729 0.24 1 normal none gamma non 9 0.693 0.634 0.251 1.09 1 0.799 1.039 0.216 0.24 9 0.0866 0.107 0.0232 0.025 7 0.837 0.813 0.29 1.13 2 0.835 0.828 0.501 1.14 9 1.07 1.102 0.352 1.20 2 1.234 1.305 0.396 1.25 7						0.9
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UCL95 Statistics for Zone 1 Impacted Data Sets with Non-Detects

TABLE A2

Total Number of Data	AI 16	As 16	Be	Cd	Co 16	Pb	Mn 16	Mo	Ni	Se	V 16	CI		NO3_as_N		Chloroform				-
Total Number of Data	16		16	16		16	16	16		16	16		16	16	16			16	16	
Number of Non-Detect Data Number of Detected Data (or Distinct Obs. If zero	9	12	16	16	0	16	0	16	13	15	15	0	0	4	0	14	0	3	8	
nondetect)	7	4	0	0	4	0	15	0	3	1	1	15	16	12	8	2	15	13	8	
Minimum Detected	0.2	0.001			0.02		0.95		0.05	0.001	0.2	48	2960	16.2	0.0012	0.0006	4620	0.4	1	(
Maximum Detected	1.3	0.003			0.06		2.96		0.06	0.001	0.2	221	4760	200	0.0022	0.00076	7860	1.8	4	Ę
Percent Non-Detects	56.25%	75.00%	100.0%	100.0%	0.00%	100.0%	0.00%	100.0%	81.25%	93.75%	93.8%	0.0%	0.0%	0.00%	0.00%	87.50%	0.0%	18.75%	50.00%	12.50
Minimum Non-detect	0.1	0.001	0.01	0.005		0.05		0.1	0.05	0.001	0.1					0.0005		0.2	1	(
Maximum Non-detect	0.1	0.001	0.01	0.005		0.05		0.1	0.05	0.001	0.1					0.001		0.2	1	(
Mean of Detected Data	0.457	0.00175			0.0363		1.656		0.0533			131.5	3778	80.5	0.00161	0.00068	6208	1.138	2.275	2.3
Median of Detected Data	0.3	0.0015			0.03		1.47		0.05			128.5	3955	72.75	0.0015	0.00068	6120	1.2	2.05	1.
Variance of Detected Data	0.15				0.000318		0.436		3.33E-05			5780	382536	4300	9E-08	1.28E-08	2099270	0.136	1.171	2.1
SD of Detected Data	0.387				0.0178		0.66		0.00577			76.02	618.5	65.57	0.0003		1449	0.369	1.082	1.4
CV of Detected Data	0.846				0.492		0.399		0.108			0.578	0.164	0.815	0.187	0.166		0.324	0.476	0.
Skewness of Detected Data	2.25				0.33		0.649		1.73			0.02	0.02	0.259	0.948	N/A	0.0162	-0.491	0.718	0.
Mean of Log-Transformed Detected Data SD of Log-Transformed Detected Data	-0.996				-3.436 0.506		0.433		-2.935 0.105			4.69 0.656	8.224 0.166	3.946 1.037	-6.449 0.178	-7.3 0.167	8.707 0.237	0.0661	0.723	0.6
Discernable Distribution (0.05) of Detected Data	gamma	normal			none	10	ognormal		none			none	normal	none	normal	none	none	normal	normal	gamr
Discernable Distribution (0.05) of Detected Data	yamma	normai			TIONE	K	Jynomai		none			none	normai	none	normai	TIONE	TIONE	normai	normai	yann
Kaplan-Meier (KM) Method																				
Mean	0.313						-		0.0506							0.00061143		1	1.638	2.1
SD	0.269	5.27E-04							0.00242							4.1206E-05		0.43	0.958	1.4
Standard Error of Mean	0.0726	1.52E-04							7.41E-04							1.5575E-05		0.112	0.256	0.3
95% KM (t) UCL	0.44	0.00145							0.0519							0.00063873		1.196	2.087	2.7
95% KM (z) UCL 95% KM (BCA) UCL	0.432	0.00144							0.0518 N/A							0.00063705		1.184	2.059 2.388	2.7 2.8
95% KM (Percentile Bootstrap) UCL	0.469								N/A							0.00076 N/A		1.230	2.366	2.0
95% KM (Chebyshev) UCL	0.43								0.0539							0.00067932		1.488	2.754	3.7
97.5% KM (Chebyshev) UCL	0.766								0.0553							0.00070869		1.699	3.237	4.4
99% KM (Chebyshev) UCL	1.035								0.058							0.00076639		2.114	4.186	5.8
Datasets without Nondetects																				
Student's-t UCL					0.0441		1.946					164.8	4049	109.2	0.00174		6843			
95% UCLs (Adjusted for Skewness)																				
95% Adjusted-CLT UCL					0.044		1.956					162.9	4033	108.6	0.00175		6805			
95% Modified-t UCL Non-Parametric UCLs					0.0441		1.95					164.8	4049	109.4	0.00174		6843			
95% CLT UCL					0.0436		1.928					162.8	4032	107.5	0.00173		6804			
95% Jackknife UCL					0.0430		1.926					164.8	4032	107.3	0.00173		6843			
95% Standard Bootstrap UCL		1			0.0432		1.92					161.1	4024	106.7	0.00172		6780			
95% Bootstrap-t UCL					0.0445		1.991					162.8	4043	110.7	0.00177		6828			
95% Hall's Bootstrap UCL					0.0431		1.923					160	4008	106.9	0.00175		6735			
95% Percentile Bootstrap UCL					0.0438		1.917					161.1	4018	107	0.00173		6760			
95% BCA Bootstrap UCL					0.0431		1.954					161.3	4018	109.1	0.00174		6764			
95% Chebyshev(Mean, Sd) UCL					0.0557		2.376		↓ ↓			214.3	4452	152	0.00193		7787			
97.5% Chebyshev(Mean, Sd) UCL					0.0641		2.687		┟───┤			250.2	4744	182.9	0.00207		8470			
99% Chebyshev(Mean, Sd) UCL					0.0806		3.299		+			320.6	5317	243.6	0.00235		9812			
Potential UCL to Use																				
95% KM (t) UCL	0.44	0.00145							0.0519							0.00063873		1.196	2.087	2.7
95% KM (z) UCL																				
95% KM (BCA) UCL 95% KM (Percentile Bootstrap) UCL		│			├				├								├	4 040	0.050	
95% KM (Percentile Bootstrap) UCL 95% KM (Chebyshev) UCL					├				╂────┤									1.213	2.256	
95% KM (Chebyshev) UCL 97.5% KM (Chebyshev) UCL					├				╂───┼											
99% KM (Chebyshev) UCL					├				<u>├</u>											
95% Student's-t UCL					├		1.946		 				4049				6843			
95% Modified-t UCL							1.95		<u>† </u>						0.00174	1	6843			
95% Chebyshev(Mean, Sd) UCL					0.0557				† †			214.3		152		İ				
97.5% Chebyshev(Mean, Sd) UCL																				
99% Chebyshev(Mean, Sd) UCL														243.6						
95% Approximate Gamma UCL																				
95% H-UCL							2.02													
No.4		_														1.0.5			-	
Notes	5	5	4	4	5	4	7	4	2, 5	3	3			6, 8		1, 2, 5			5	

Notes:

1. Data have multiple DLs - Use of KM Method is recommended.

2. There may not be adequate detected values to compute meaningful and reliable test statistics and estimates.

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

3. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values

determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).

4. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs!

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

5. Warning: There are less than 10 Distinct Detected Values in this data. It should be noted that bootstrap calculations may not be reliable enough to draw conclusions.

6. Recommended 99% Chebyshev(Mean, Sd) UCL exceeds the maximum observation (log transformed standard deviation of 1.037 barely exceeds the threshold of 1.0 for using the 99% Chebyshev UCL)

95% Chebyshev(Mean, Sd) UCL chosen alternative

7. 95% Modified-t UCL selected instead of 95% H-UCL because of relatively small sample size and evidence from larger background Mn dataset that the population distribution may not be log-normal

Rad_totl	Th-230	Pb-210	Gross_Alpha
16	16	16	16
2	14	16	3
_			
14	2	0	13
		0	
0.6	0.6		1.2
5.2	0.7		4.1
12.50%	87.50%	100.0%	18.75%
0.2	0.2	1	1
0.2	0.2	1	1
2.357	0.65		2.146
	0.65		
1.85	0.65		2
2.133	0.005		0.549
1.461	0.0707		0.741
0.62	0.109		0.345
0.75	N/A		1.60
0.668	-0.434		0.716
0.658	0.109		0.311
gamma	none		normal
2.138	0.606		1.969
1.439	0.0242		0.741
0.373	0.00856		0.193
2.792	0.621		2.307
2.752	0.62		2.286
2.813	N/A		2.394
2.010	N/A		2.319
	0.644		
3.765			2.809
4.469	0.66		3.172
5.852	0.691		3.886
2.792	0.621		2.307
2.8			2.319
	2, 5	4	
	_, •	-	

UCL95 Statistics for Zone 3 Impacted Data Sets with Non-Detects

TABLE A3

	AI	As	Be	Cd	Co	-	Mn	Мо	Ni	Se	V	CI		NO3_as_N		Chloroform	-		Rad-228	Rad_te
Total Number of Data	70	-		70	_	70	70	70	70	70	70	70	70	70	70	70		70	70	
Number of Non-Detect Data	12	22	61	54	0	70	0	38	0	67	65	0	0	43	0	57	0	0	0	
Number of Detected Data (or Distinct Obs. If																				
zero nondetect)	58		-	16	46	0	64	32	31	3	5	39	58	27	63	13	60	58	59	
Minimum Detected	0.1	0.001		0.005	0.05		3.33	0.1	0.11	0.001	0.1	14	2630	0.1	0.0011	0.00093	3980	2	3.8	(
Maximum Detected	163			1	0.95		23.7	5	0.89	0.01	0.2	98	5260	44.8	0.138	0.00676	6680	27.6	56.1	73
Percent Non-Detects	17.14%		87.14%	77.14%	0.00%	100.0%	0.00%	54.29%	0.00%	95.71%	92.86%	0.0%	0.0%	61.43%	0.00%	81.43%	0.0%	0.00%	0.00%	0.00
Minimum Non-detect	0.1	0.001	0.01	0.005		0.05		0.1		0.001	0.1			0.1		0.0005				
Maximum Non-detect	0.1	0.001	0.01	0.005		0.05		0.1		0.001	0.1			0.1		0.001				
Mean of Detected Data	16.14	0.206		0.0713	0.381		9.836	1.084	0.377	0.00433	0.18	43.66	3599	17.15	0.0287	0.00441	5289	9.823	15.73	25.
Median of Detected Data	2.45	0.025		0.0095	0.35		7.485	0.3	0.31	0.002	0.2	37.5	3545	24	0.0219	0.00444	5290	7.9	13.55	20
Variance of Detected Data	1393	0.177	0.000586	0.0613	0.0572		27.12	1.533	0.0459	2.43E-05	0.002	468.3	349814	227.1	0.000764	3.3795E-06	585903	36.79	109.4	23
SD of Detected Data	37.32	0.421	0.0242	0.248	0.239		5.208	1.238	0.214	0.00493	0.0447	21.64	591.5	15.07	0.0276	0.00184	765.4	6.065	10.46	15.
CV of Detected Data	2.312	2.04	0.411	3.473	0.627		0.529	1.142	0.568	1.138	0.248	0.496	0.164	0.879	0.964	0.417	0.145	0.617	0.665	0.6
Skewness of Detected Data	2.66	3.83	-0.65	4.00	0.68		1.079	1.40	1.13	1.652	-2.24	0.75	0.18	-0.0511	2.408	-0.39	-0.191	0.921	1.671	1.
Mean of Log-Transformed Detected Data	0.9	-3.518	-2.966	-4.416	-1.206		2.163	-0.642	-1.119	-5.909	-1.748	3.656	8.175	1.277	-3.922	-5.539	8.563	2.091	2.561	3.0
SD of Log-Transformed Detected Data	1.832	2.251	0.657	1.216	0.772		0.489	1.271	0.536	1.181	0.31	0.498	0.166	2.572	0.928	0.559	0.149	0.647	0.632	0.5
Discernable Distribution (0.05) of Detected Data	none	none	normal	none	gamma		none	none	none	normal	none	none	normal	none	none	normal	normal	gamma	gamma	lognorr
Kaplan-Meier (KM) Method		İ																		
Mean	13.39	0.142	0.0163	0.0202				0.55		0.00114	0.106			6.677		0.00158				
SD	34.22			0.118				0.959		0.00107	0.0232			12.38		0.00155				
Standard Error of Mean	4.125	4.32E-02		1.46E-02				0.116		1.57E-04	0.0031			1.508		0.00019315				
95% KM (t) UCL	20.27	0.214		0.0444		<u> </u>		0.744		0.0014	0.111			9.191		0.0019				
95% KM (z) UCL	20.18	0.213		0.0441				0.742		0.0014	0.111			9.157		0.00189				
95% KM (BCA) UCL	20.79	0.224		0.0628				0.739		N/A	N/A			9.191		0.00339				
95% KM (Percentile Bootstrap) UCL	20.16	0.22		0.0486				0.747		N/A	N/A			9.246		0.00326				
95% KM (Chebyshev) UCL	31.37	0.33		0.0400				1.058		0.00183	0.119			13.25		0.00242				
97.5% KM (Chebyshev) UCL	39.15	0.33		0.0000				1.277		0.00212	0.115			16.09		0.00242				
99% KM (Chebyshev) UCL	54.44	0.572		0.165				1.709		0.00212	0.123			21.68		0.00270				
33 / RM (Chebyshev) OCL	J4.44	0.572	0.0394	0.105				1.709		0.00271	0.157			21.00		0.0033				
Datasets without Nondetects																				
Student's-t UCL					0.429		10.87		0.42			47.97	3717		0.0342		5441	11.03	17.81	28.
95% UCLs (Adjusted for Skewness)					0.429		10.07		0.42			41.31	5/1/		0.0342		3441	11.05	17.01	20.
95% Adjusted-CLT UCL					0.431		10.95		0.423			48.16	3717		0.0351		5437	11.1	18.05	28.
95% Modified-t UCL		ł			0.431		10.95		0.423			48.10	3717		0.0343		5437	11.04	17.85	28.
Non-Parametric UCLs					0.429		10.09		0.421			40.01	3/1/		0.0343		0441	11.04	17.05	20.
95% CLT UCL		ł			0.428		10.86		0.419			47.91	3715		0.0341		5439	11.02	17.78	28.
95% Jackknife UCL		ł			0.428		10.80		0.419			47.91	3713		0.0341		5439	11.02	17.78	28.
95% Standard Bootstrap UCL					0.429		10.87		0.42			47.86	3713		0.0342		5438	10.97	17.01	
					0.428		10.85		0.42			48.63	3713		0.0361		5436	11.13	17.74	28. 28.
95% Bootstrap-t UCL 95% Hall's Bootstrap UCL					0.432		10.95		0.424			46.63	3710		0.0361		5430	11.13	18.31	20.
					0.432		10.9		0.423			40 47.96	3719		0.0357		5447	11.14	17.81	29.
95% Percentile Bootstrap UCL																	5432			
95% BCA Bootstrap UCL					0.43		10.96		0.424			48.06	3710 3907		0.0356		-	11.05	17.99	28.
95% Chebyshev(Mean, Sd) UCL					0.506 0.56		12.55		0.489			54.93			0.0431 0.0493		5688	12.98	21.18	33.
97.5% Chebyshev(Mean, Sd) UCL							13.72		0.537			59.81	4040 4302				5860	14.35	23.53	4;
99% Chebyshev(Mean, Sd) UCL					0.666		16.03		0.632			69.39	4302		0.0615		6199	17.04	28.16	4,
Detertial UCL to Upp			├																	
Potential UCL to Use			0.0000							0.0044	0.444		1			0.0040				
95% KM (t) UCL			0.0202							0.0014	0.111					0.0019	├			
95% KM (z) UCL		ļ			L												┝────┤			
95% KM (BCA) UCL			0.000	0.0628		 		0.739												
95% KM (Percentile Bootstrap) UCL			0.0529										1			0.00326				
95% KM (Chebyshev) UCL																				
97.5% KM (Chebyshev) UCL	39.15	0.412												16.09						
99% KM (Chebyshev) UCL		ļ																		
95% Student's-t UCL		ļ					10.87					47.97	3717				5441			
95% Modified-t UCL							10.89					48.01								
95% Chebyshev(Mean, Sd) UCL									0.489						0.0431					
97.5% Chebyshev(Mean, Sd) UCL																				
99% Chebyshev(Mean, Sd) UCL																				
95% Approximate Gamma UCL					0.439					-								11.14	17.84	
95% H-UCL										-										29.
											-									
Notes	6	6	5			4				3, 5	2, 5			6		1				

Notes:

1. Data have multiple DLs - Use of KM Method is recommended.

2. There may not be adequate distinct detected values to compute meaningful and reliable test statistics and estimates

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

3. Warning: There are only 3 Distinct Detected Values in this data set.

The number of detected data may not be adequate enough to perform GOF tests, bootstrap, and ROS methods.

4. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs!

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

5. Warning: There are less than 10 Distinct Detected Values in this data. It should be noted that bootstrap calculations may not be reliable enough to draw conclusions.

6. Potential UCL to use is at 97.6% confidence level

			-
Rad_totl	Th-230	Pb-210	Gross_Alpha
70	70	70	70
0	64	64	0
61	6	6	62
6.8	0.2	1.8	2.4
73.3	1.3	8.1	35.2
0.00%	91.43%	91.43%	0.00%
	0.2	1	
	0.2	1	
25.55	0.533	4.883	12.62
20.8	0.4	4.9	10.55
235.5	0.159	4.078	60.29
15.35	0.398	2.019	7.765
0.601	0.747	0.414	0.615
1.22	1.88	0.14	1.13
3.078	-0.816	1.497	2.355
0.572	0.641	0.496	0.617
lognormal	normal	normal	gamma
	0.229	2.064	
	0.142	1.018	
	0.0185	0.133	
	0.259	2.287	
	0.259	2.284	
	0.439	5.093	
	0.421	4.941	
	0.309	2.645	
	0.344	2.897	
	0.344	3.39	
	0.413	5.59	
20.61			14 17
28.61			14.17
20.05			44.00
28.85			14.28
28.65			14.19
00.57			
28.57			14.15
28.61			14.17
28.53			14.12
28.91			14.18
29.03			14.29
28.53			14.16
28.56			14.24
33.54			16.67
37			18.42
43.8			21.85
	0.259	2.287	
	0.421	4.941	
			14.25
29.14			
20.14			
	5	5	
	5	5	

N.A. WATER SYSTEMS

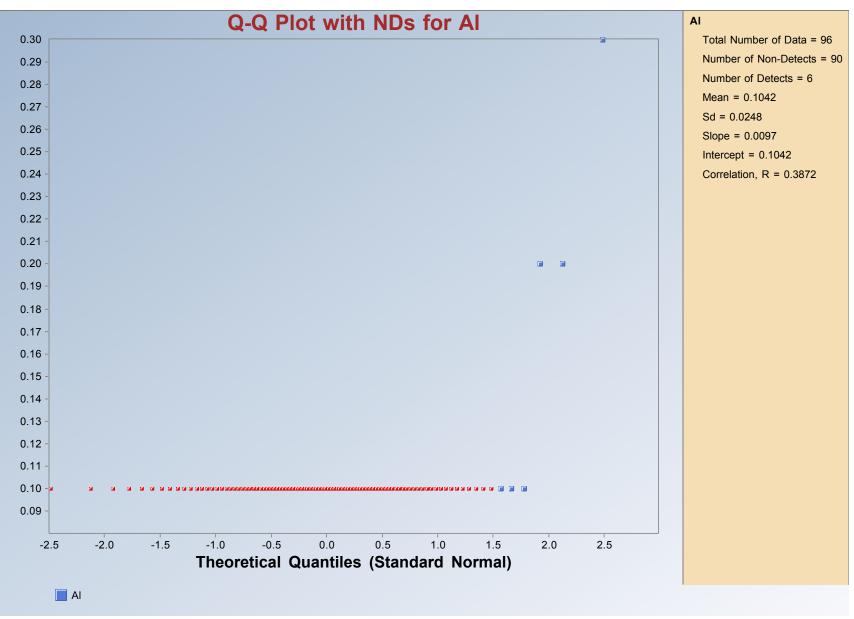
Appendix B

Probability Plots

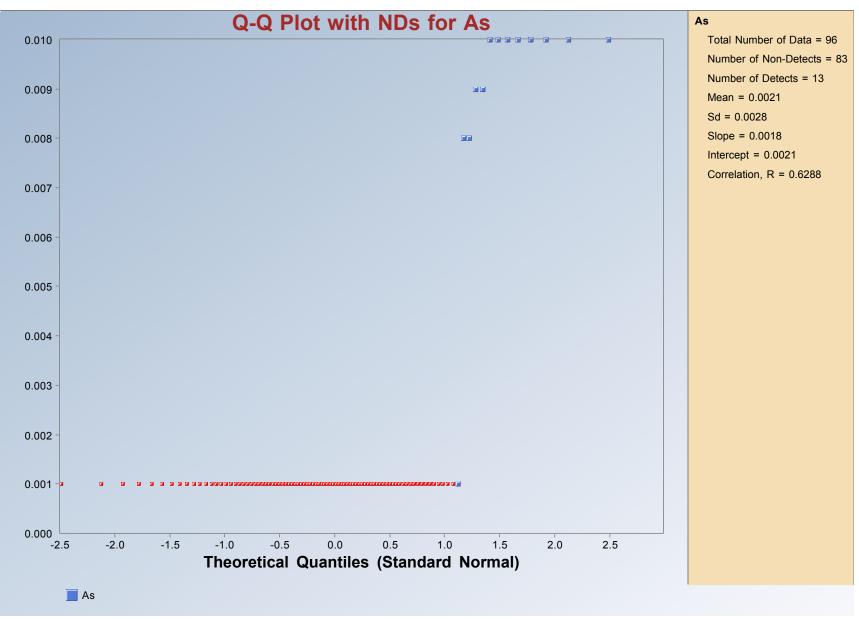


N.A. Water Systems, LLC Airside Business Park, 250 Airside Drive Moon Township, Pennsylvania 15108-2793 USA Tel: 412-809-6000 • Fax: 412-809-6075 Web site: www.nawatersystems.com

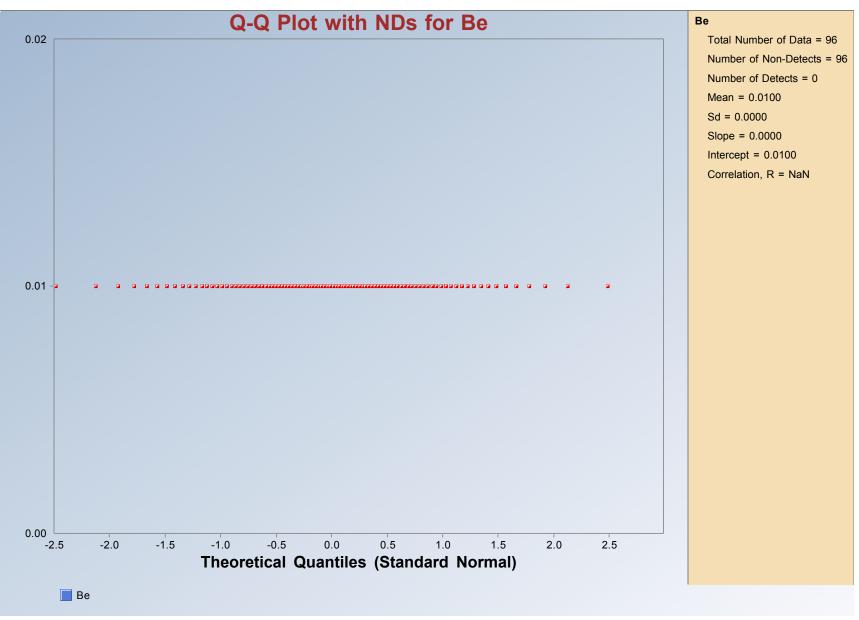
Probability Plot of Aluminum in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



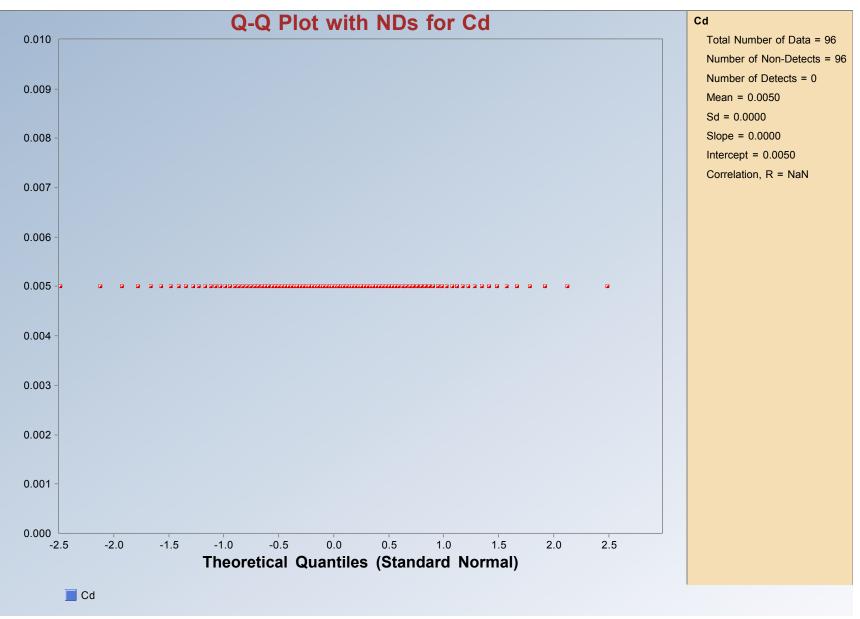
Probability Plot of Arsenic in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



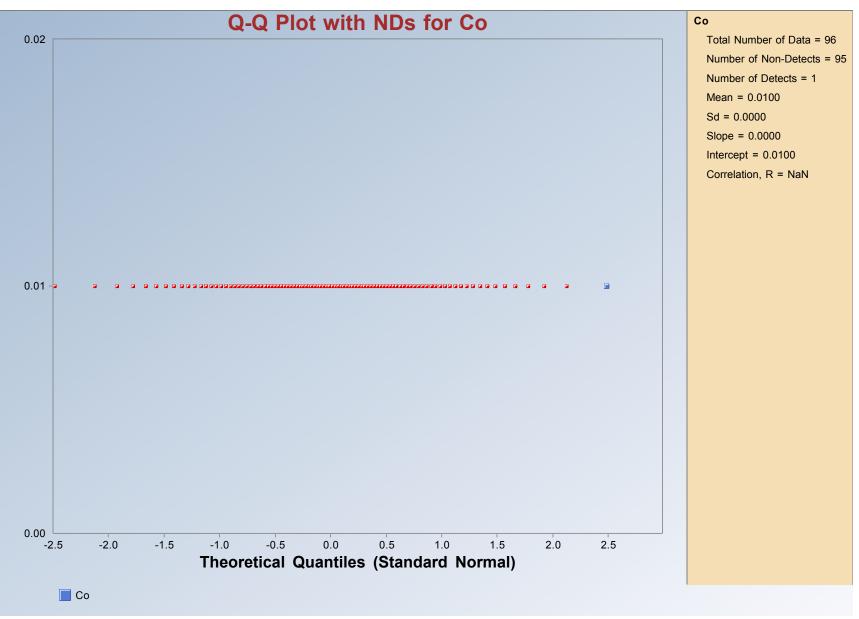
Probability Plot of Beryllium in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



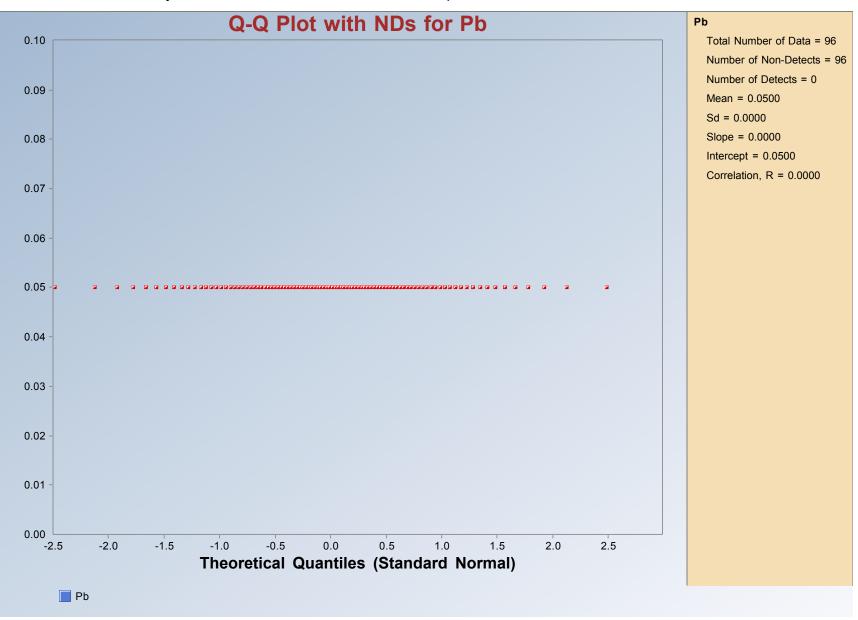
Probability Plot of Cadmium in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



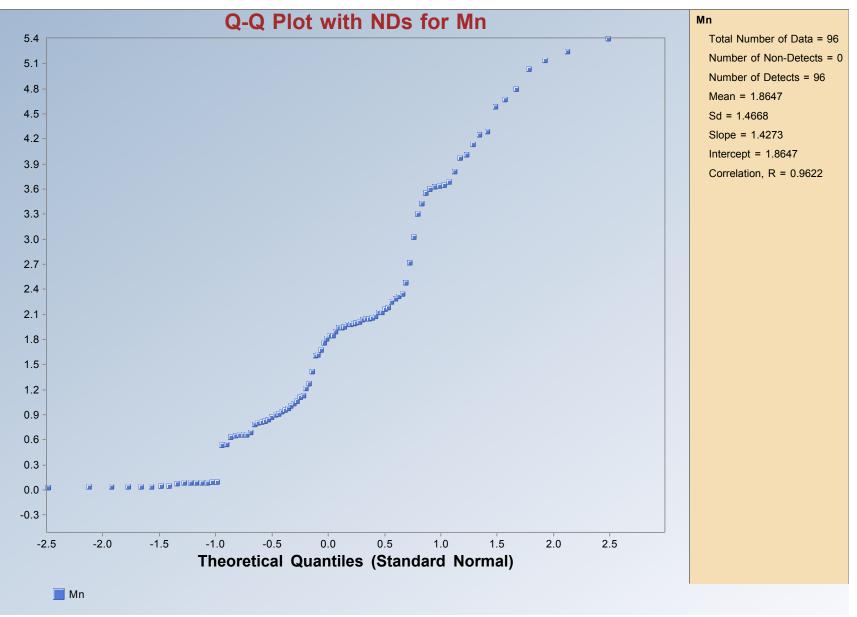
Probability Plot of Cobalt in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



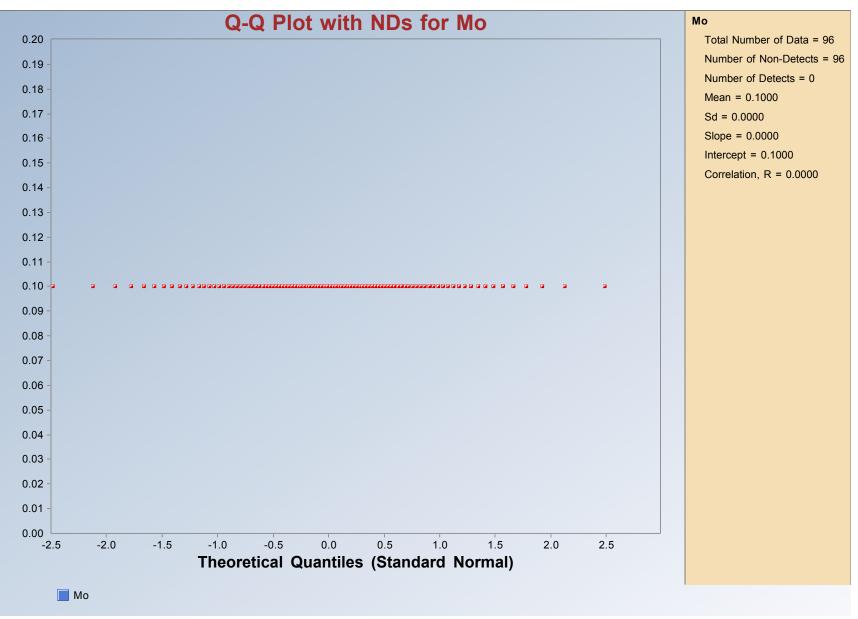
Probability Plot of Lead in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



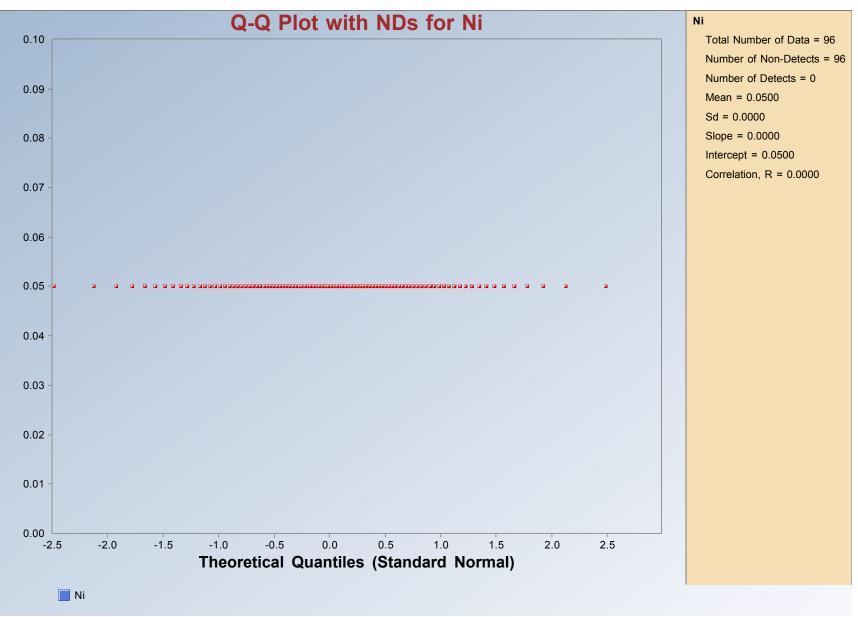
GRAPH B 1.7 Probability Plot of Manganese in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



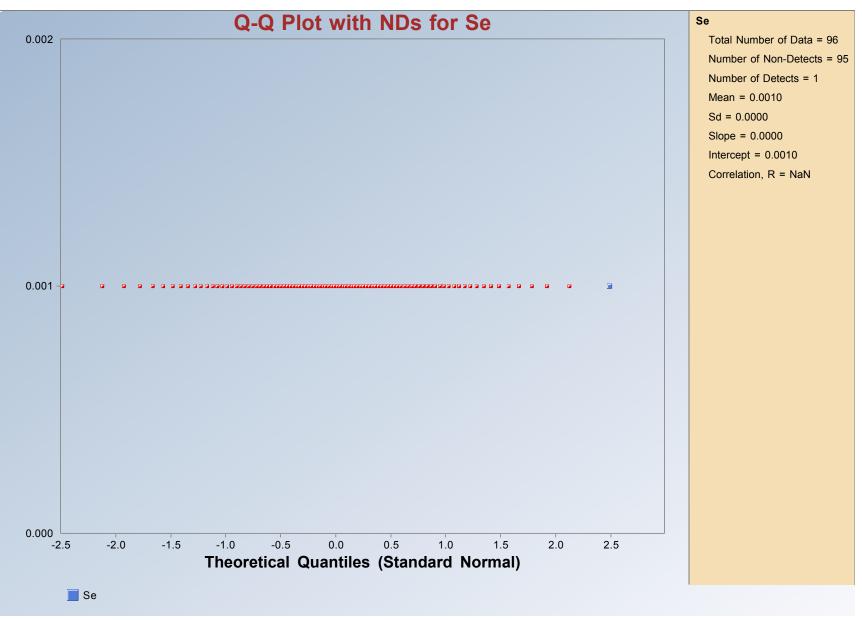
Probability Plot of Molybdenum in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



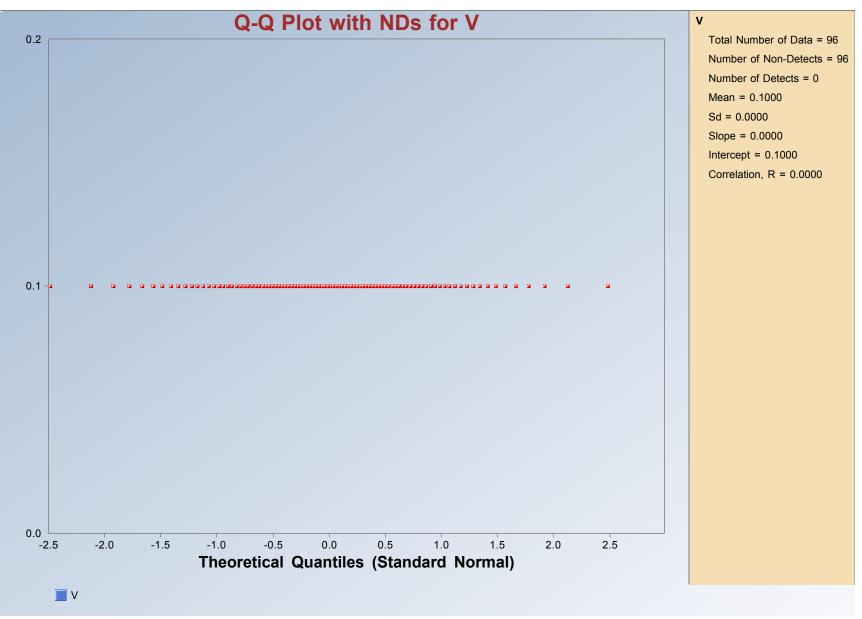
Probability Plot of Nickel in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



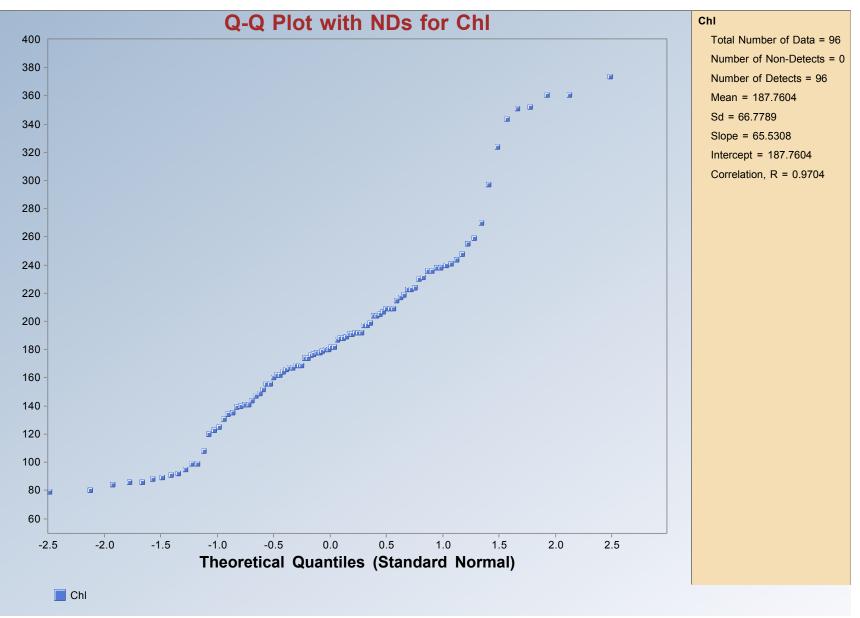
Probability Plot of Selenium in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



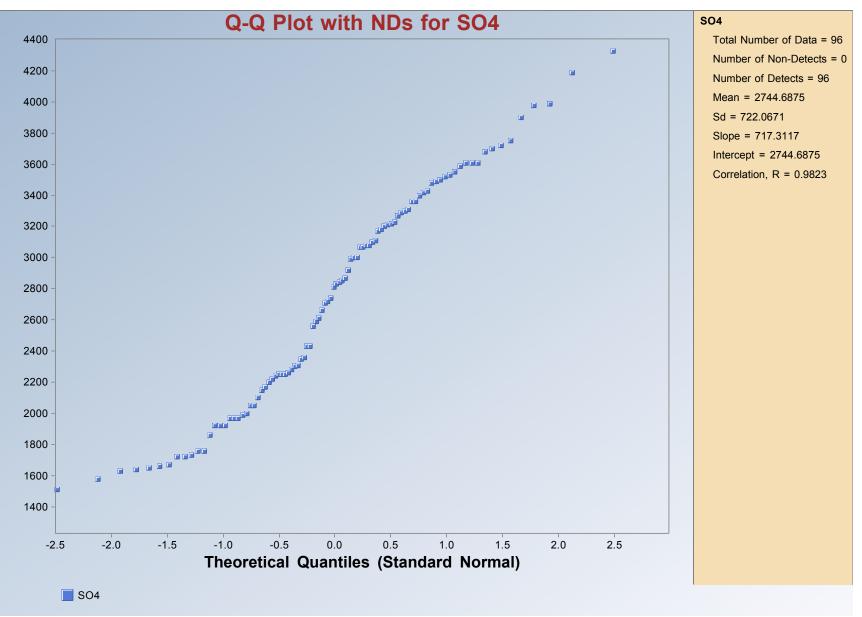
Probability Plot of Vandadium in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



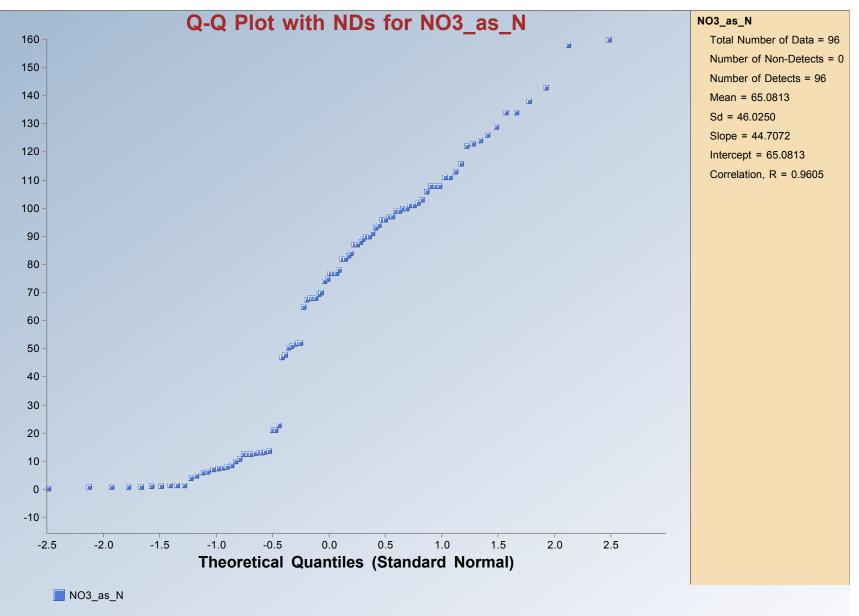
GRAPH B 1.12 Probability Plot of Chloride in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



GRAPH B 1.13 Probability Plot of Sulfate in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



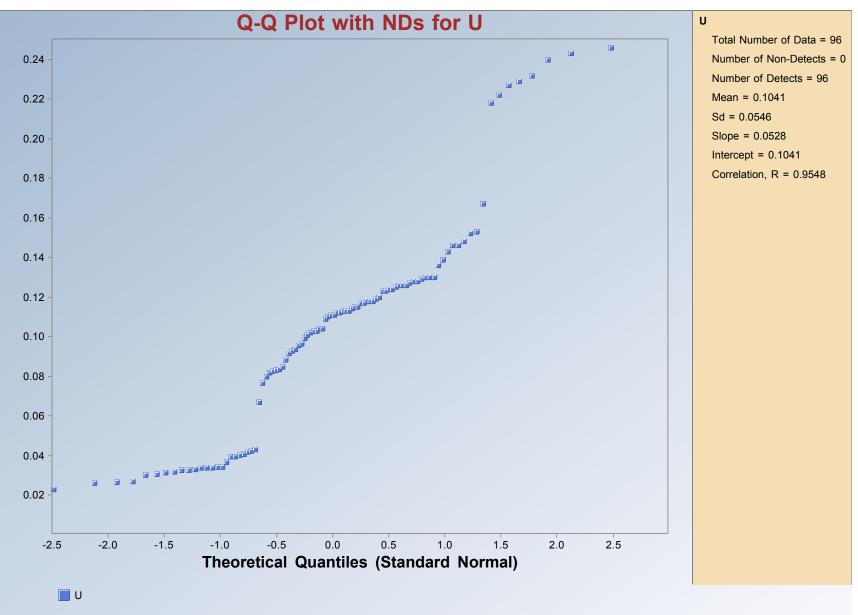
GRAPH B 1.14 Probability Plot of Nitrate in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



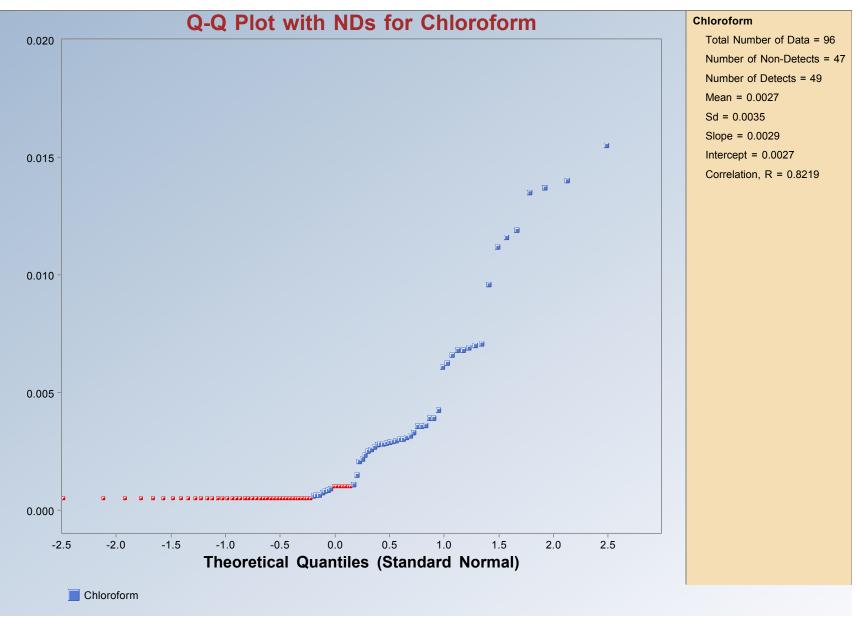
(concentrations in milligrams per liter)

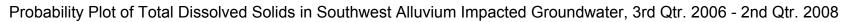
GRAPH B 1.15

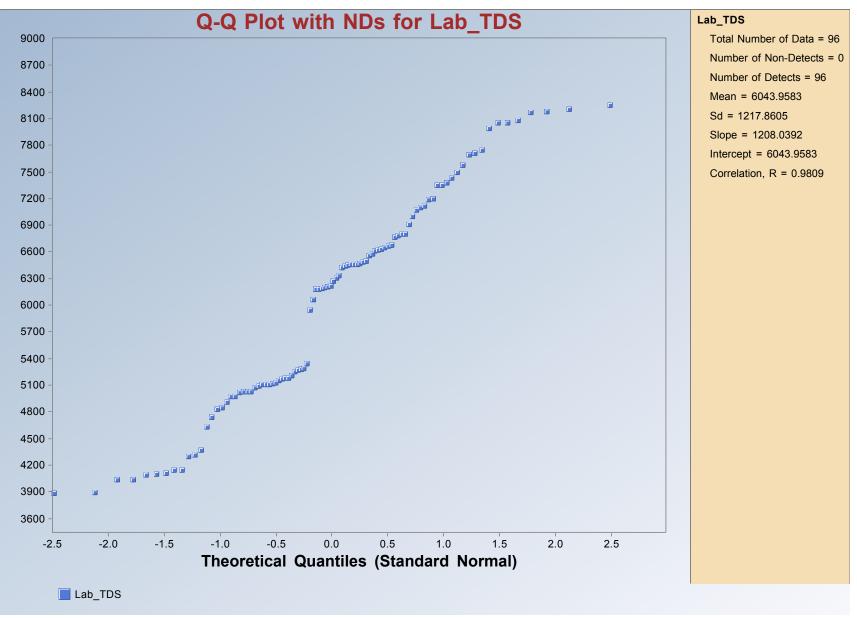
Probability Plot of Uranium in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



Probability Plot of Chloroform in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008

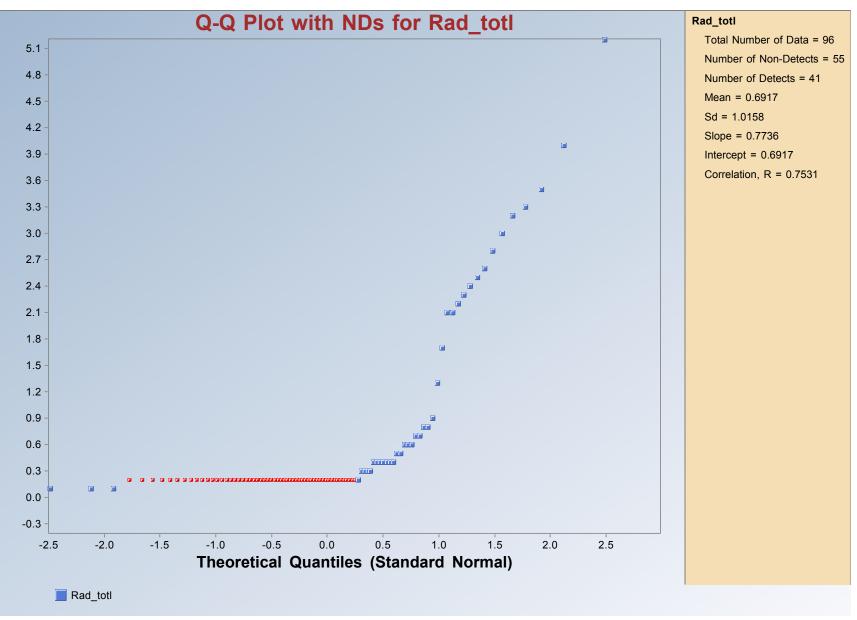




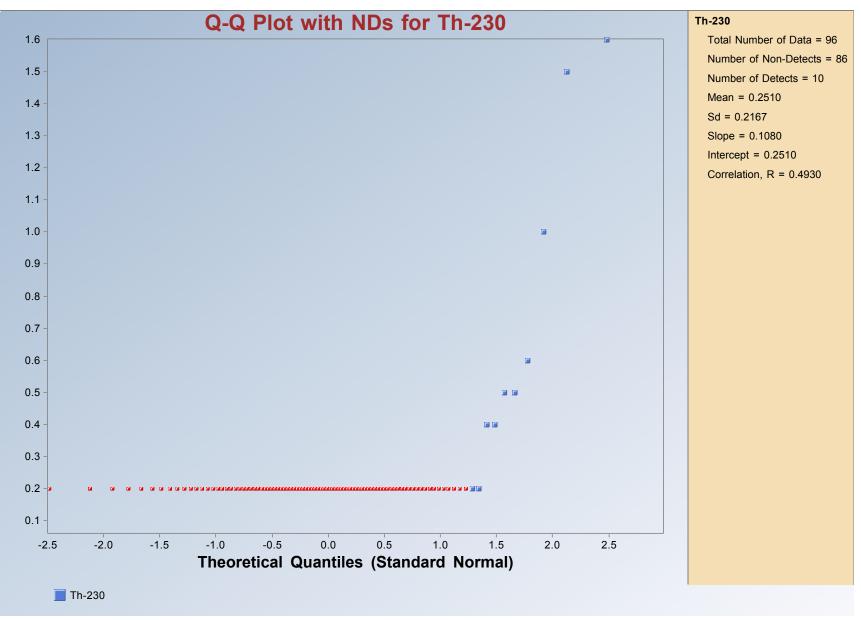


(concentrations in milligrams per liter)

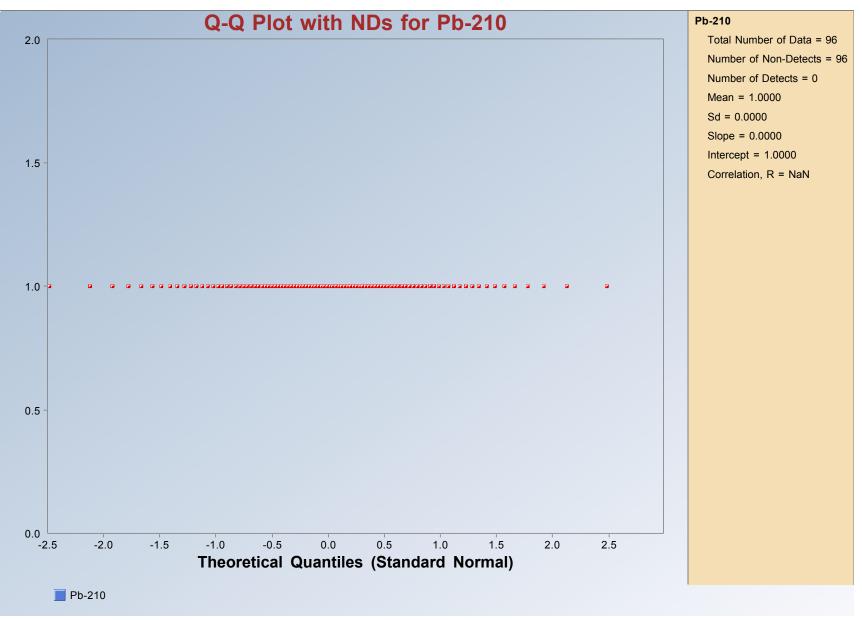
Probability Plot of Total Radium in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008



Probability Plot of Thorium-230 in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008

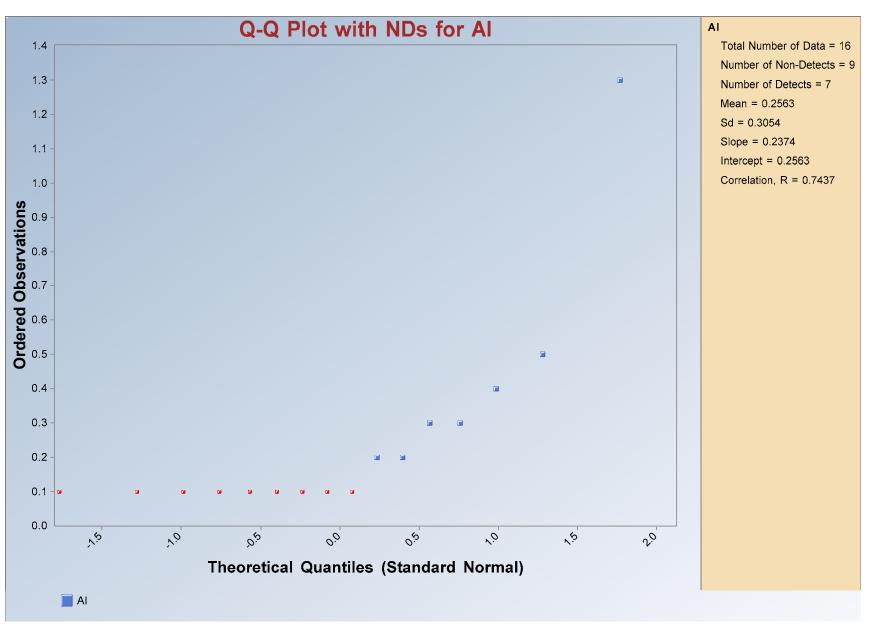


Probability Plot of Lead-210 in Southwest Alluvium Impacted Groundwater, 3rd Qtr. 2006 - 2nd Qtr. 2008

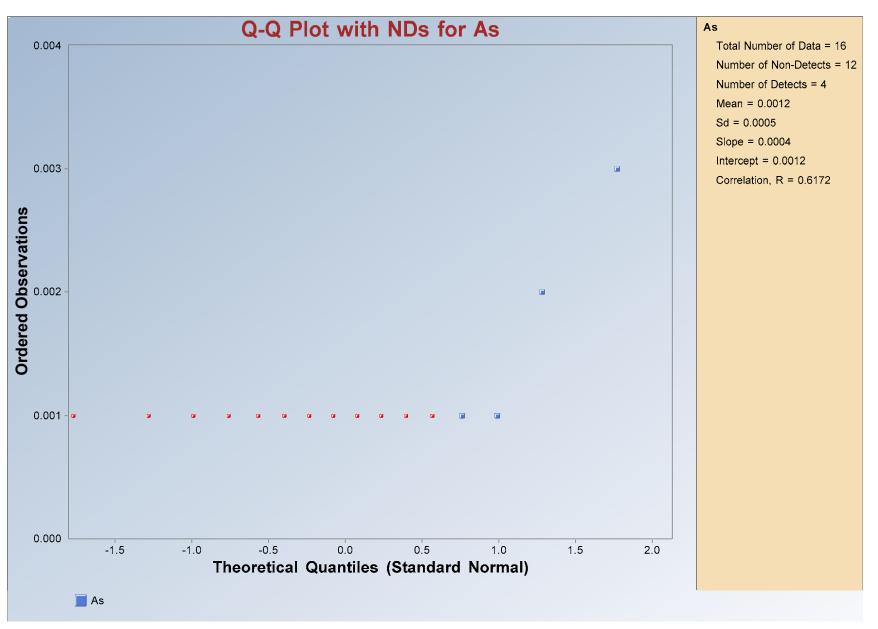


(concentrations in pico curies per liter)

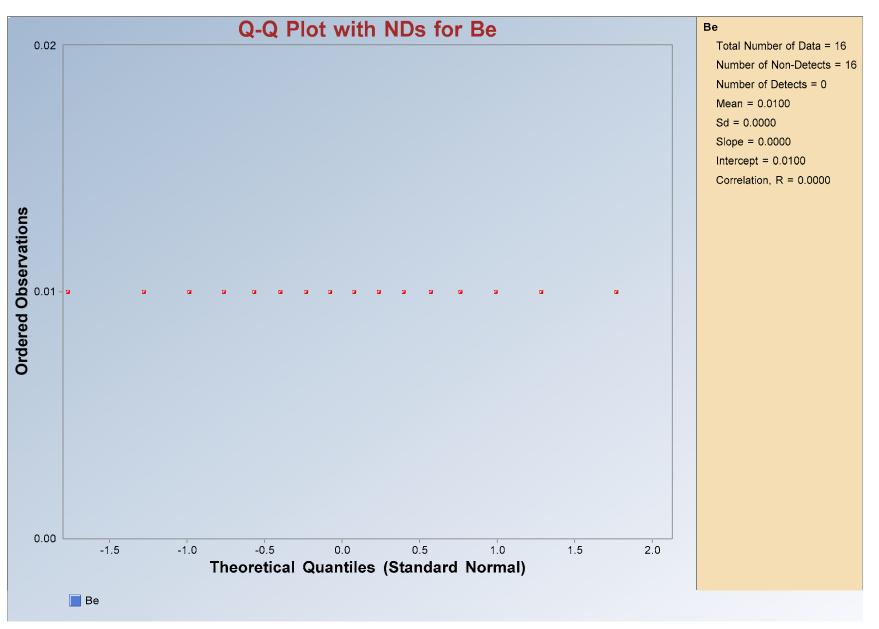
GRAPH B 2.1 Probability Plot of Aluminum in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



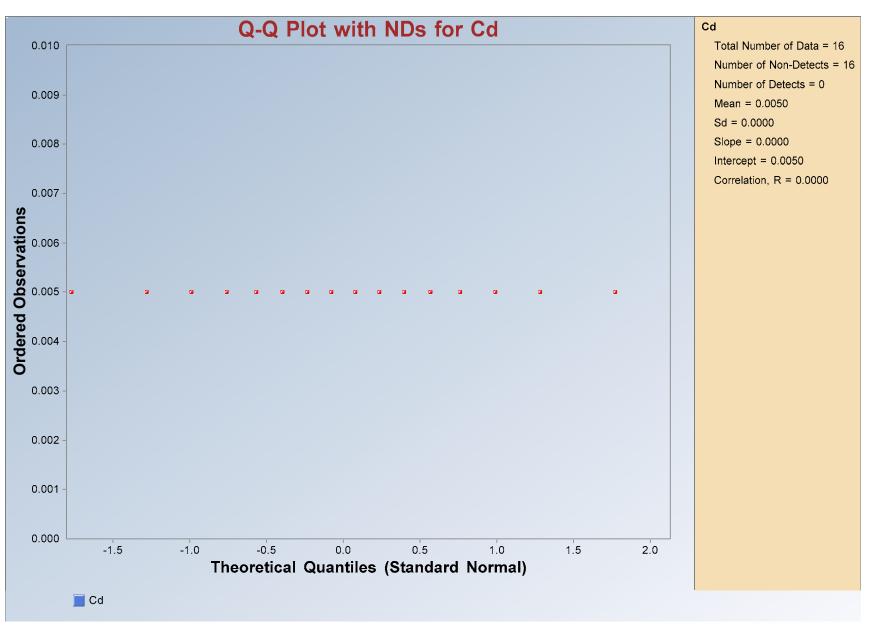
GRAPH B 2.2 Probability Plot of Arsenic in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



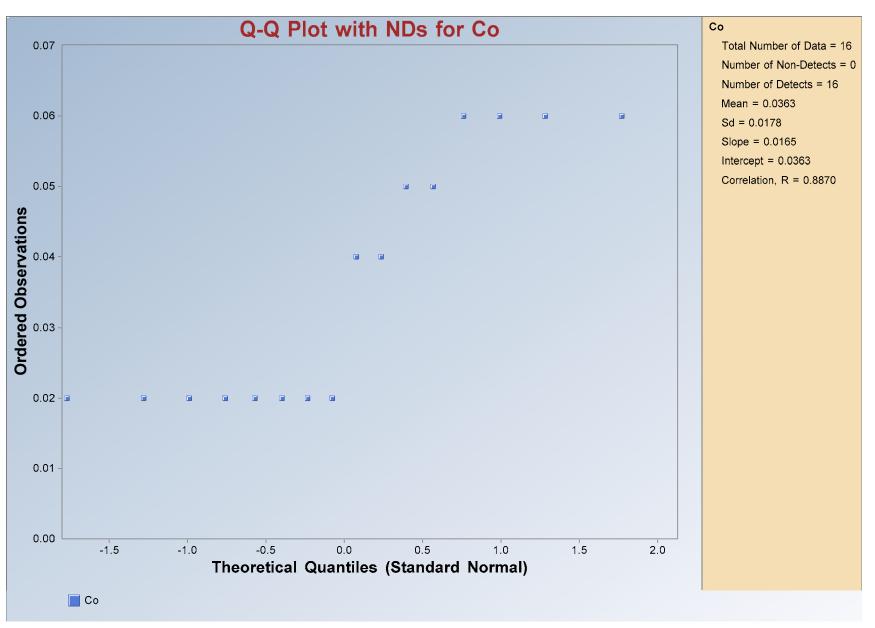
GRAPH B 2.3 Probability Plot of Beryllium in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



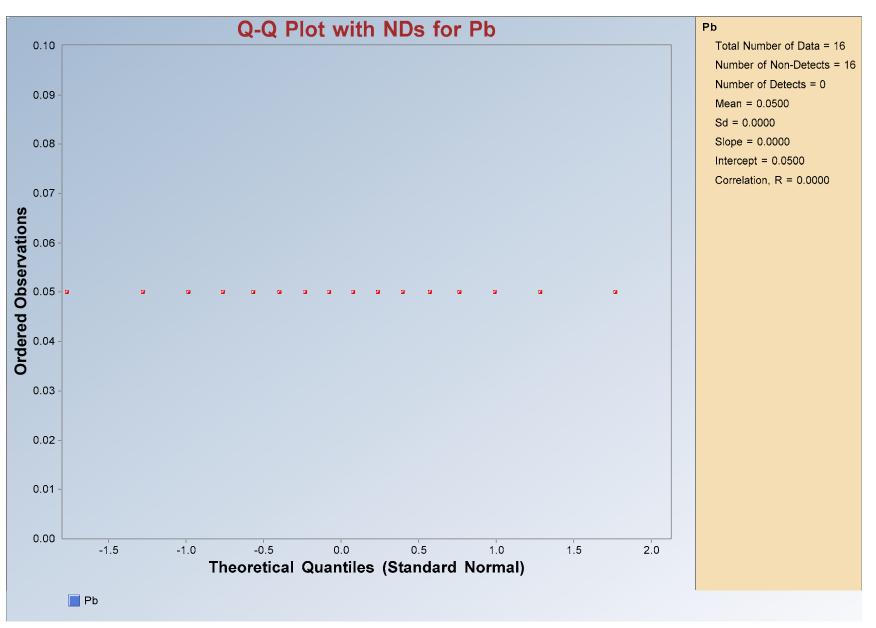
GRAPH B 2.4 Probability Plot of Cadmium in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



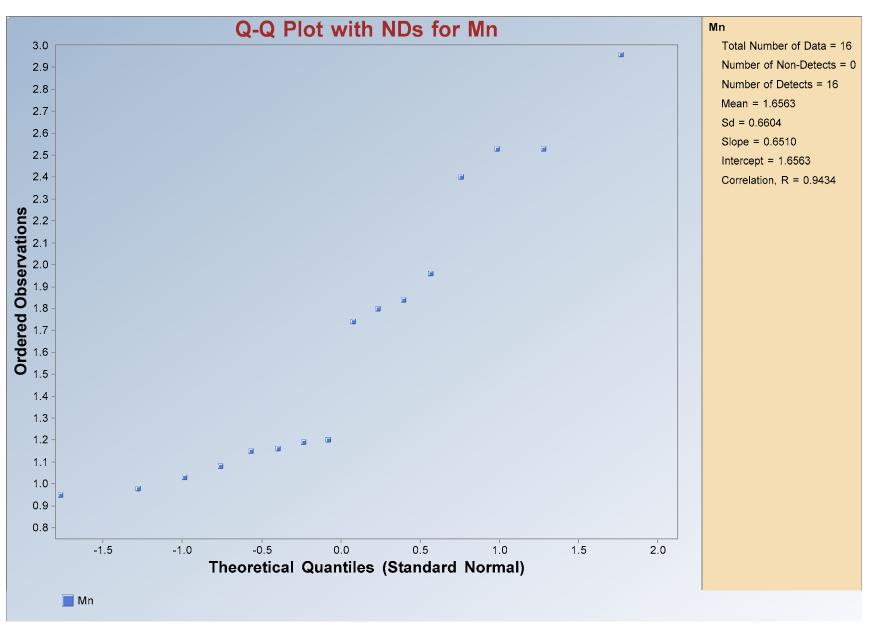
GRAPH B 2.5 Probability Plot of Cobalt in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



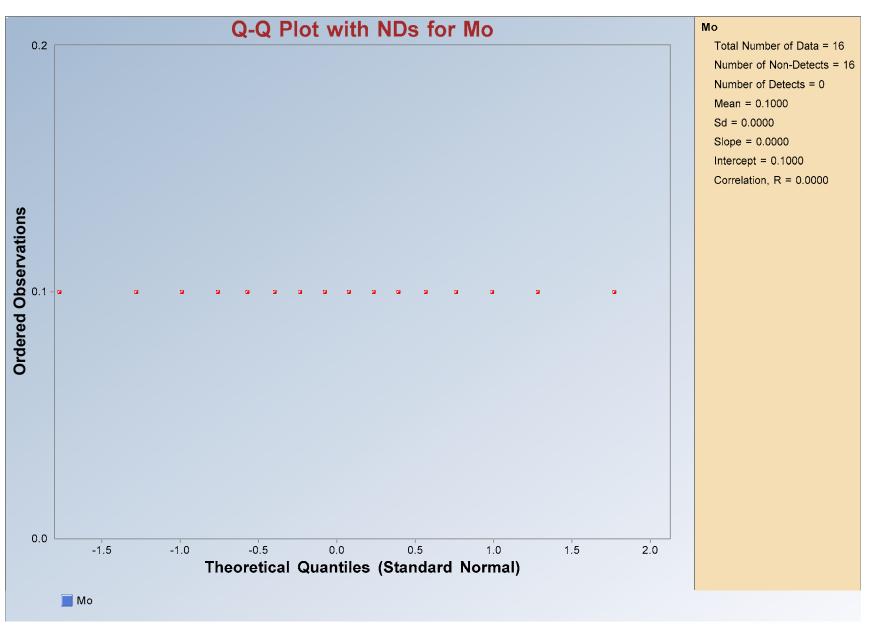
GRAPH B 2.6 Probability Plot of Lead in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



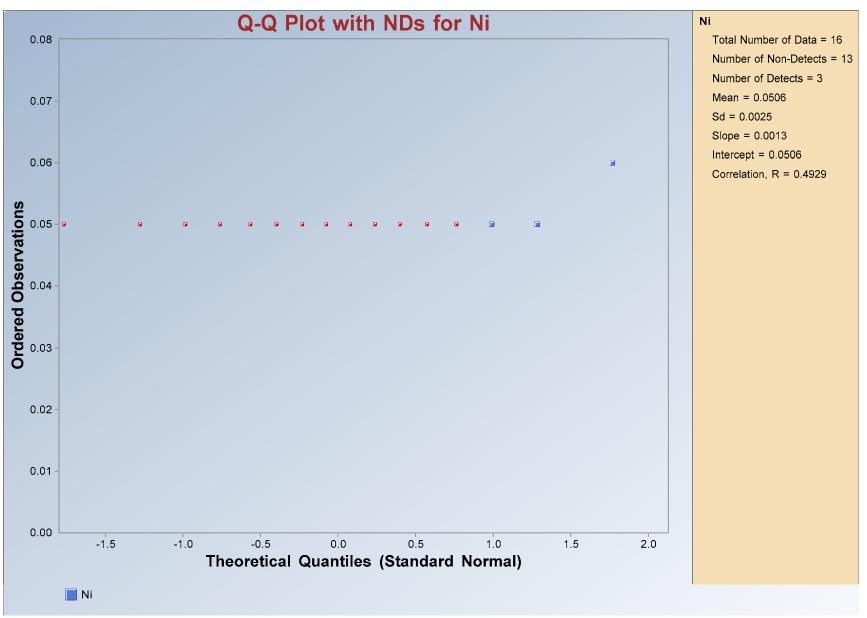
GRAPH B 2.7 Probability Plot of Manganese in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



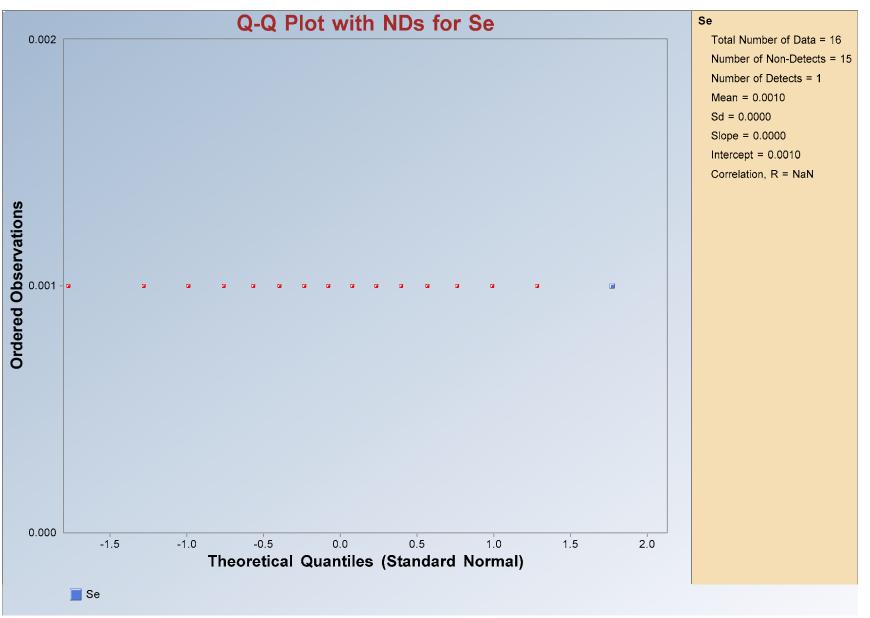
GRAPH B 2.8 Probability Plot of Molybdenum in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



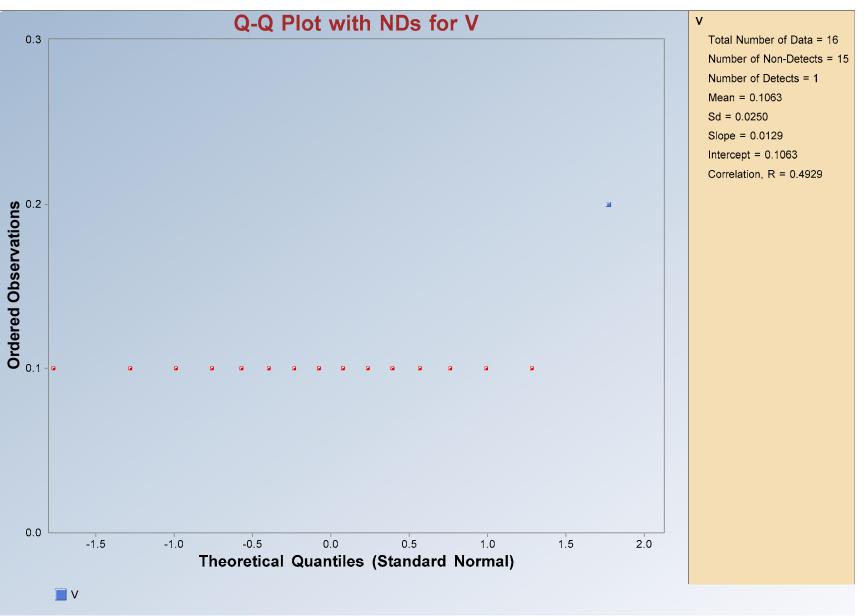
Probability Plot of Nickel in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



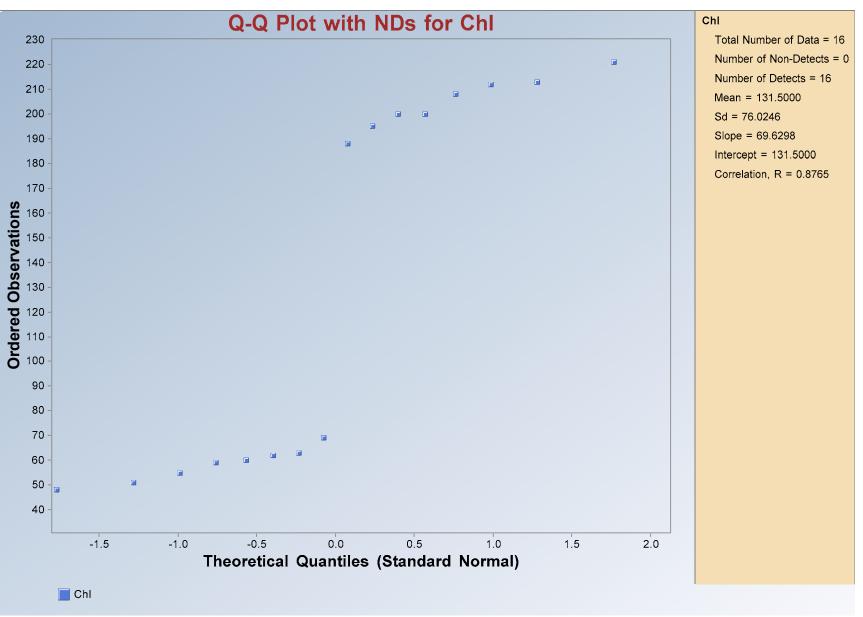
Probability Plot of Selenium in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



Probability Plot of Vanadium in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

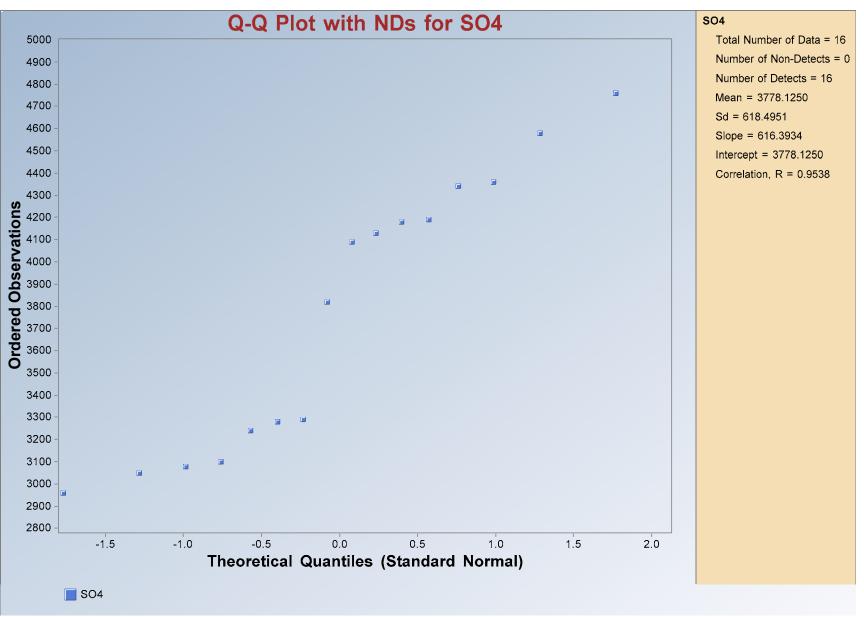


Probability Plot of Chloride in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

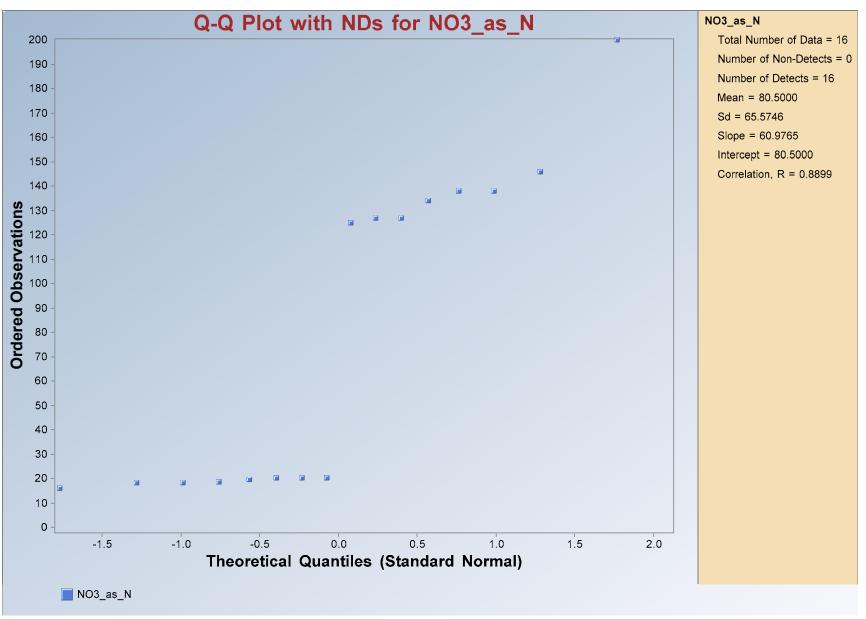


(concentrations in milligrams per liter)

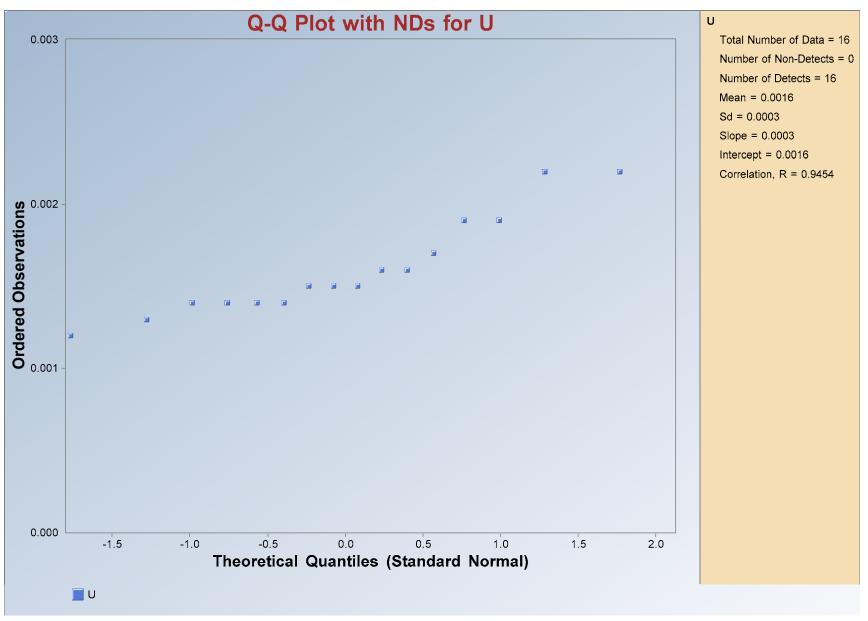
Probability Plot of Sulfate in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



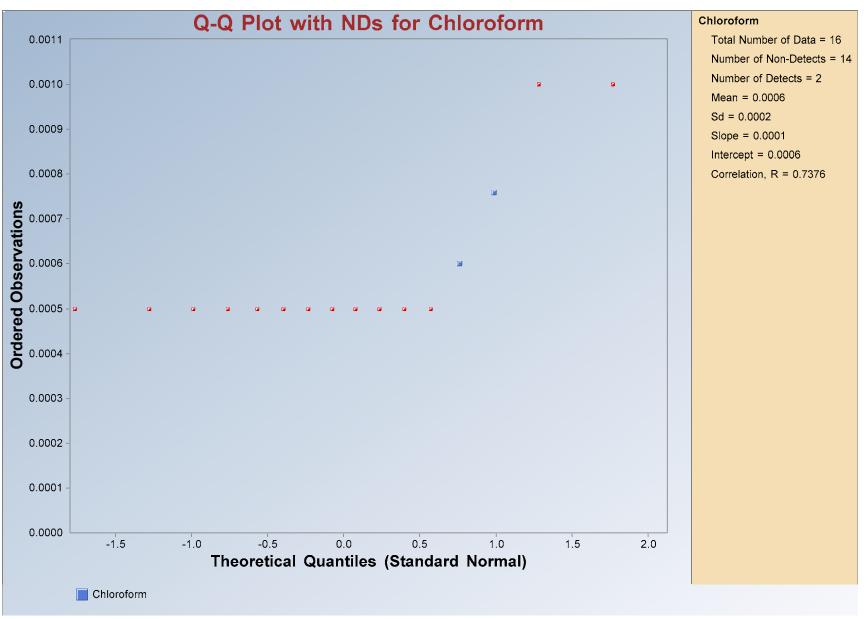
Probability Plot of Nitrate in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



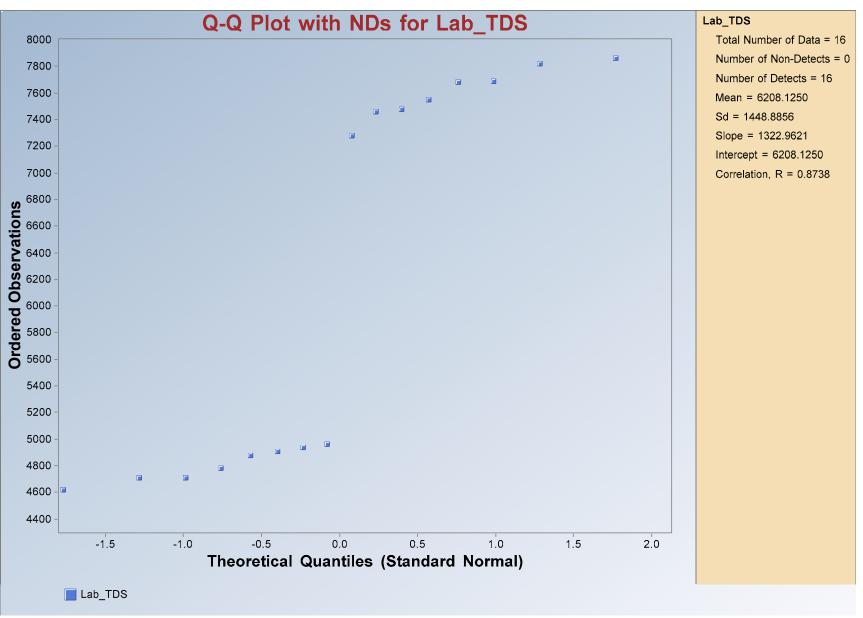
Probability Plot of Uranium in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



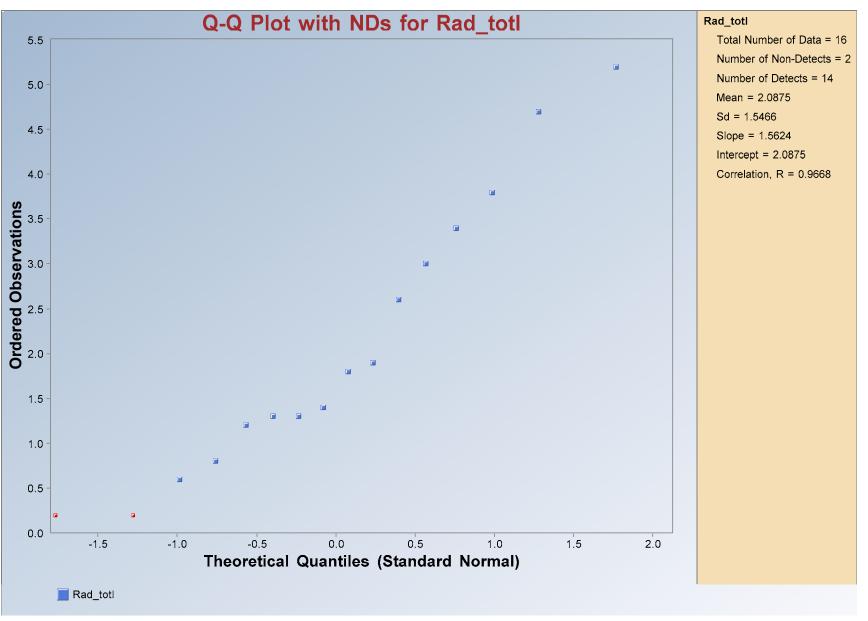
Probability Plot of Chloroform in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



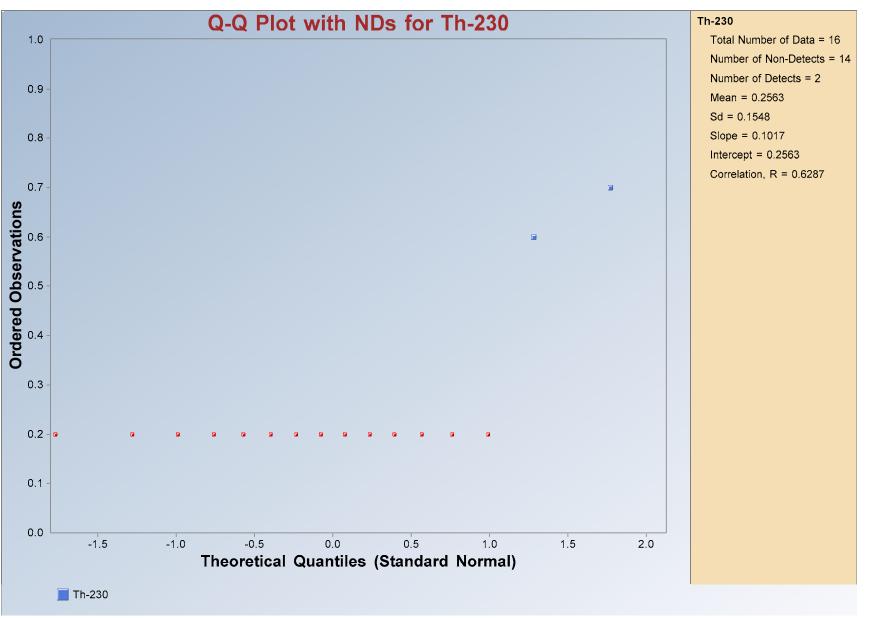
Probability Plot of Total Dissolved Solids in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



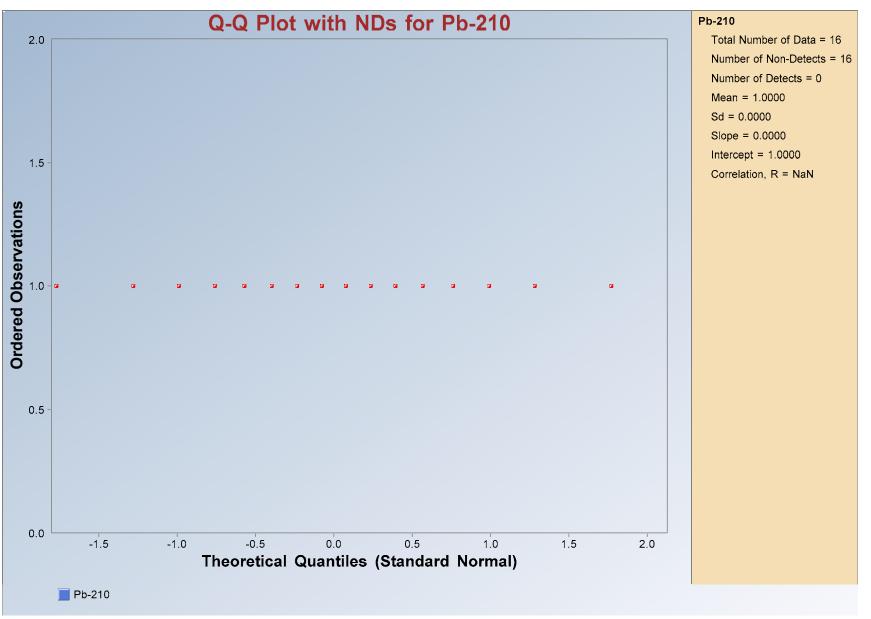
Probability Plot of Total Radium in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



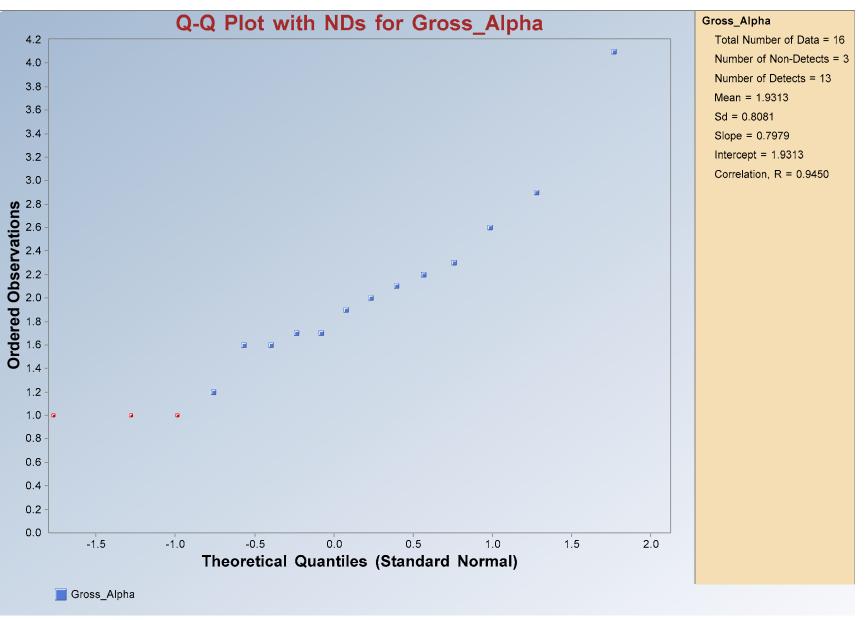
Probability Plot of Thorium-230 in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

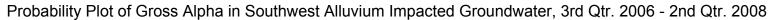


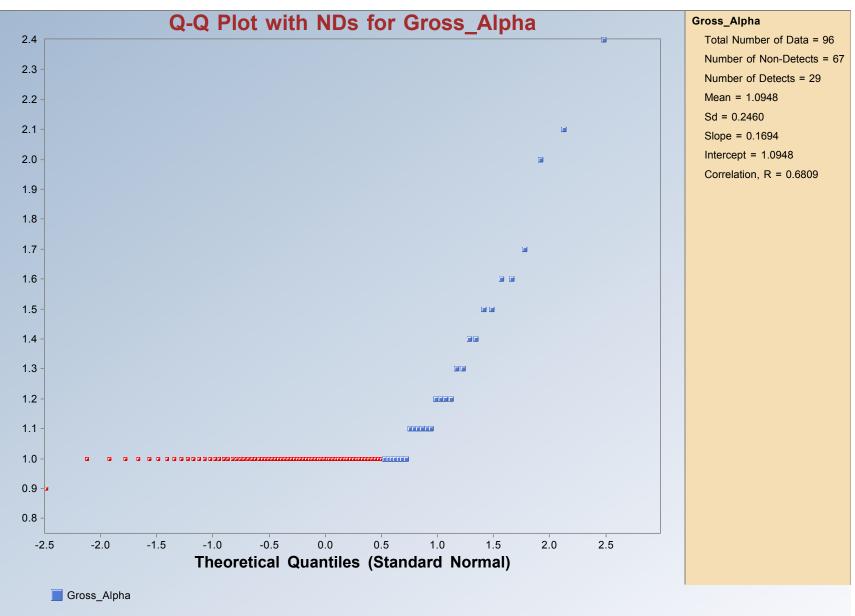
Probability Plot of Lead-210 in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



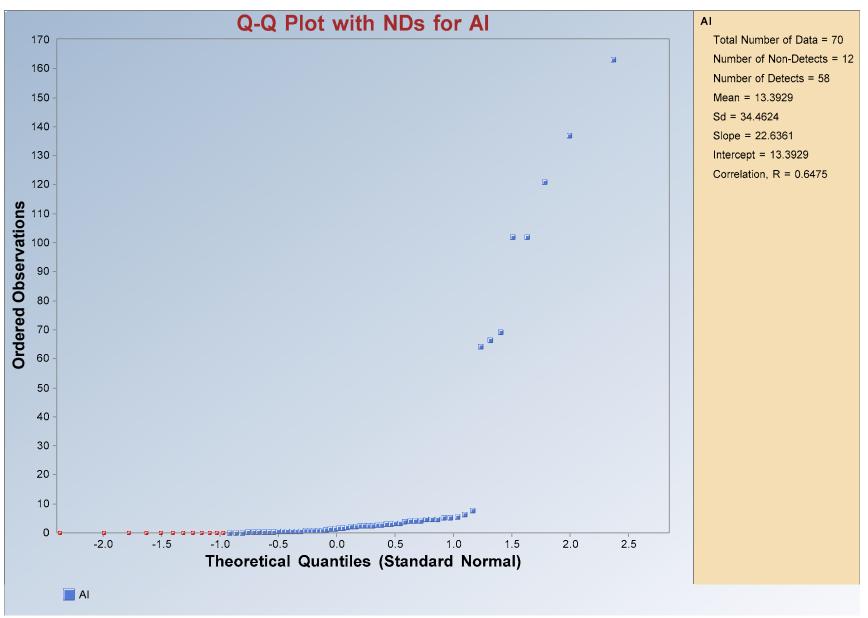
Probability Plot of Gross Alpha in Zone 1 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



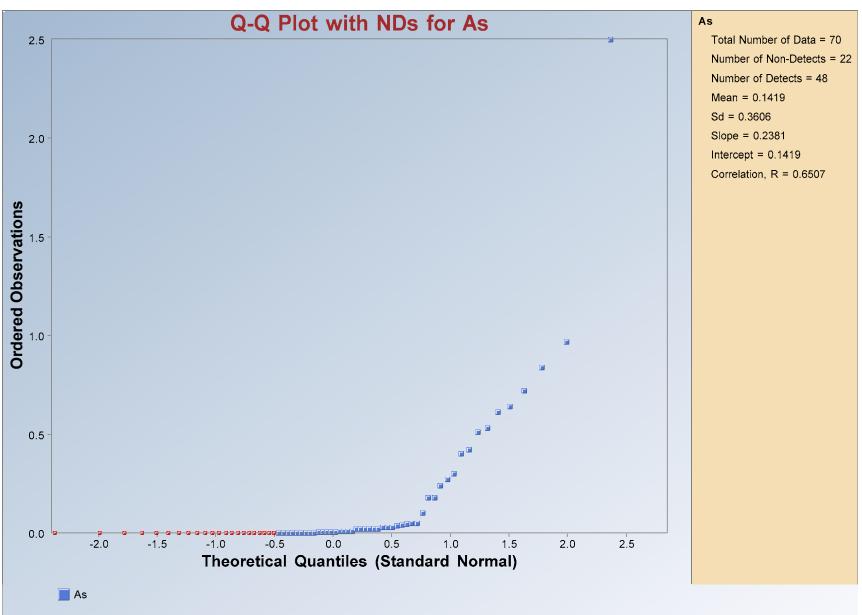




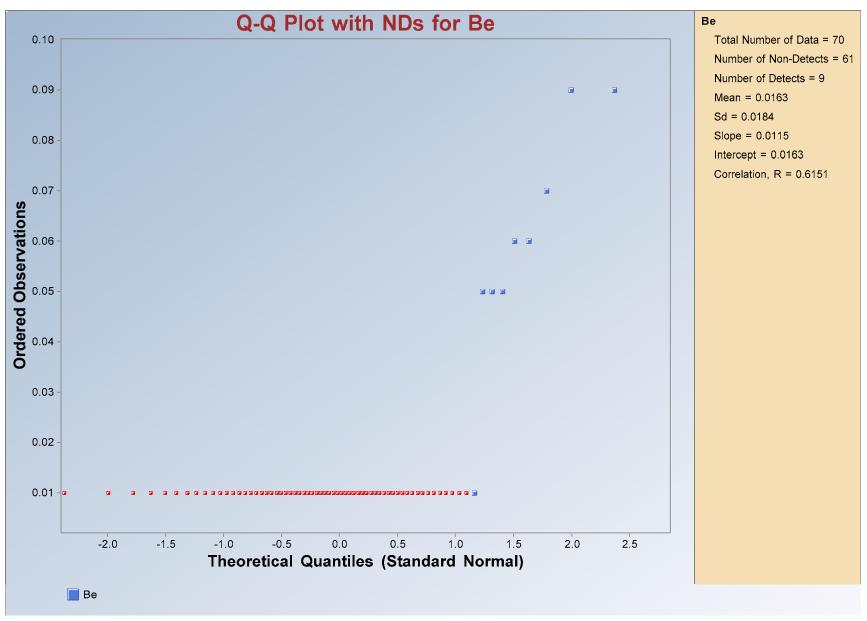
Probability Plot of Aluminum in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



Probability Plot of Arsenic in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

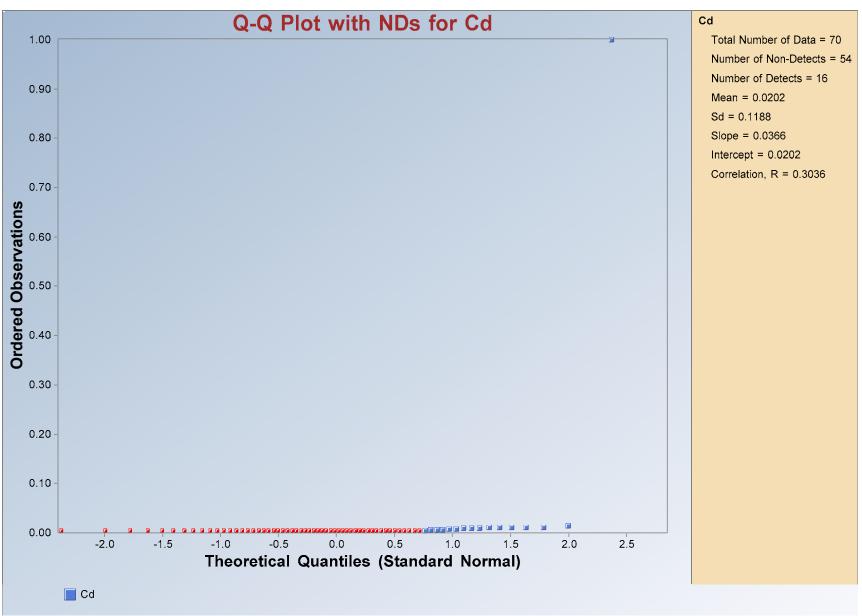


Probability Plot of Beryllium in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

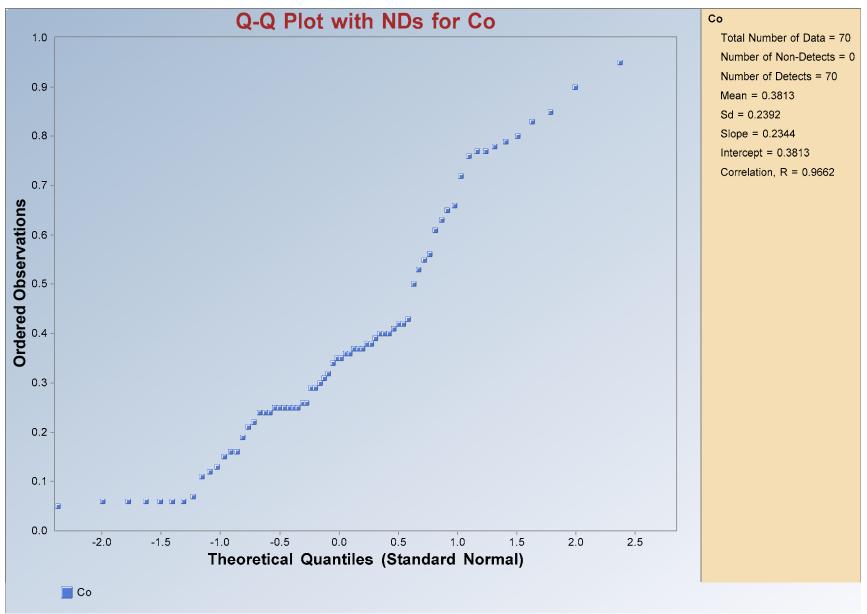


(concentrations in milligrams per liter)

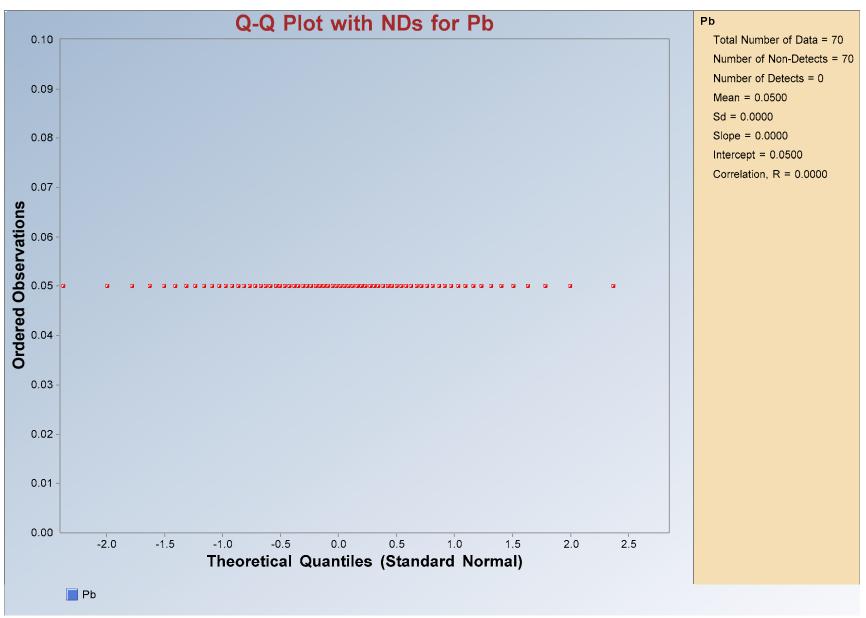
Probability Plot of Cadmium in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



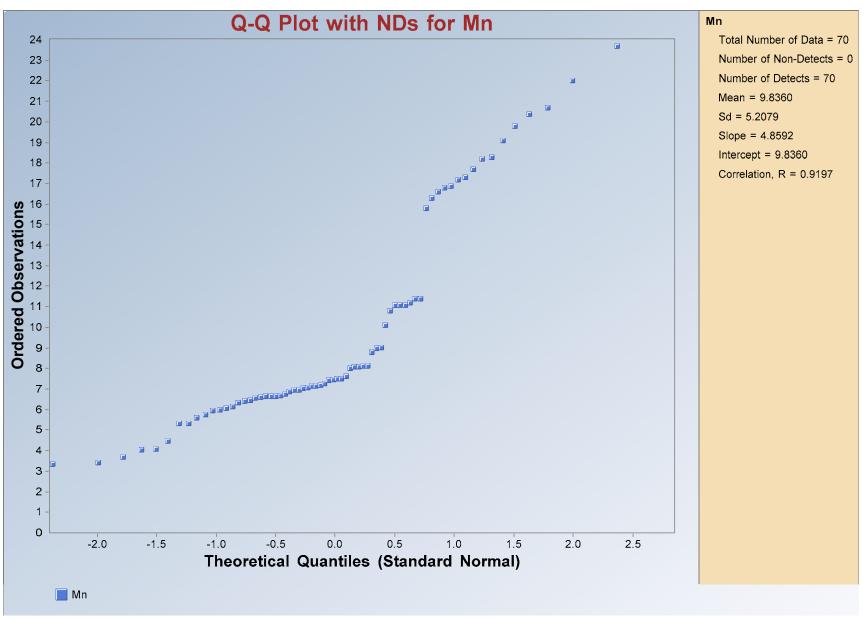
GRAPH B 3.5 Probability Plot of Cobalt in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



Probability Plot of Lead in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

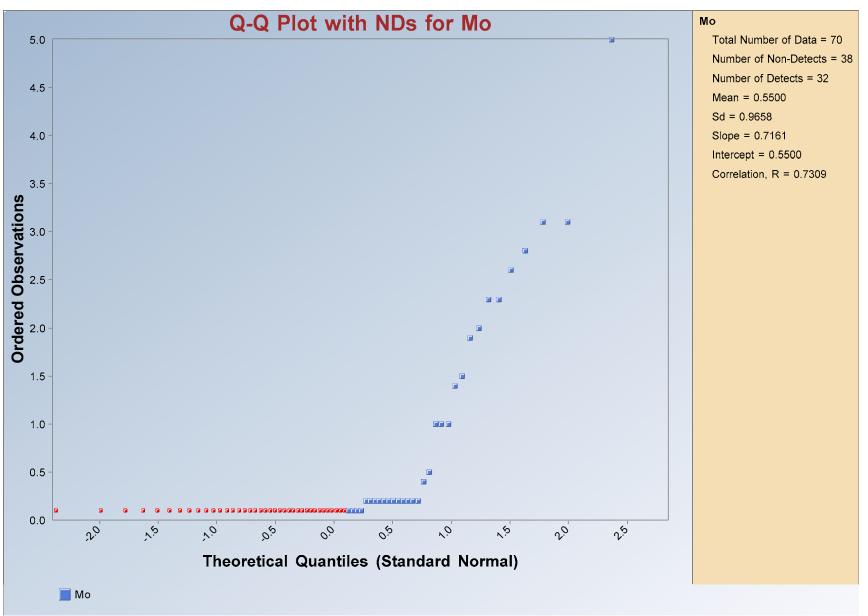


Probability Plot of Manganese in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

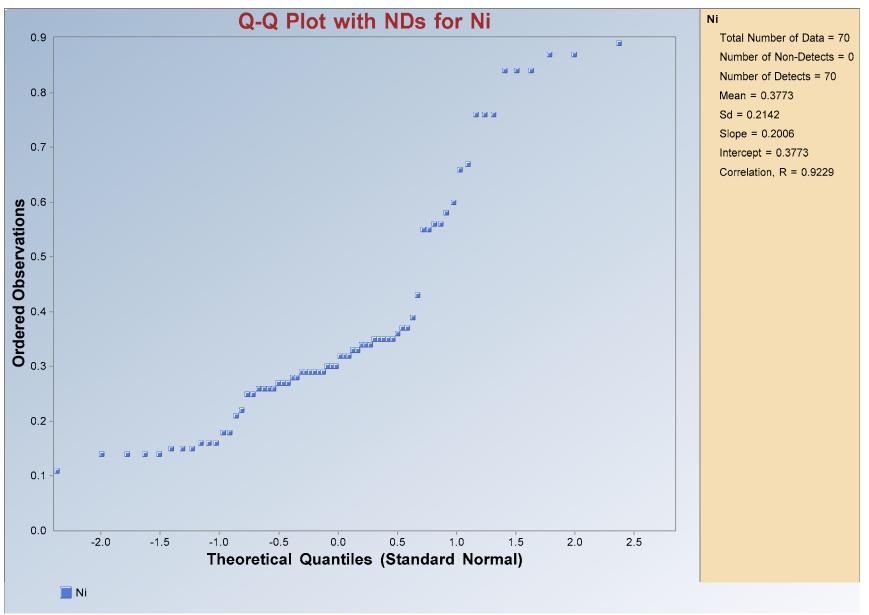


(concentrations in milligrams per liter)

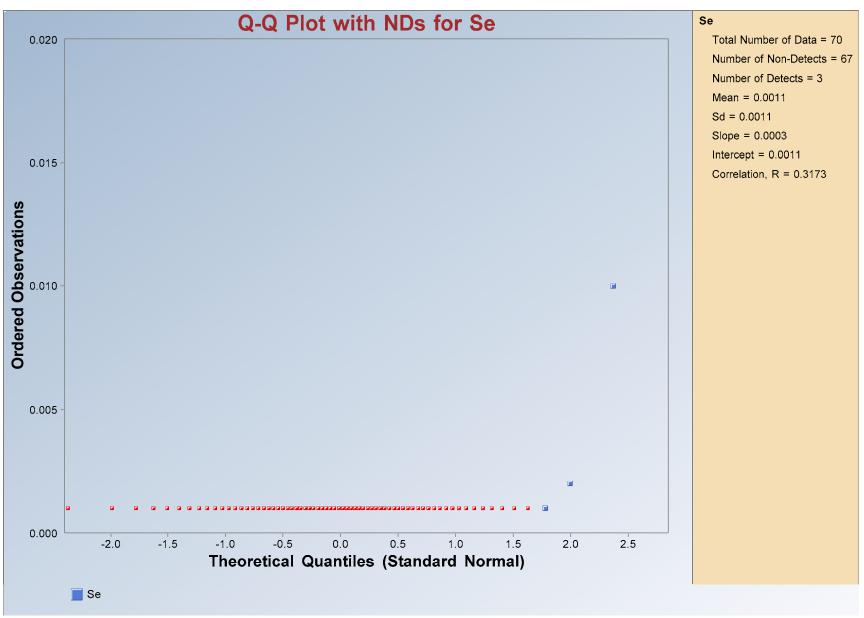
Probability Plot of Molybdenium in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



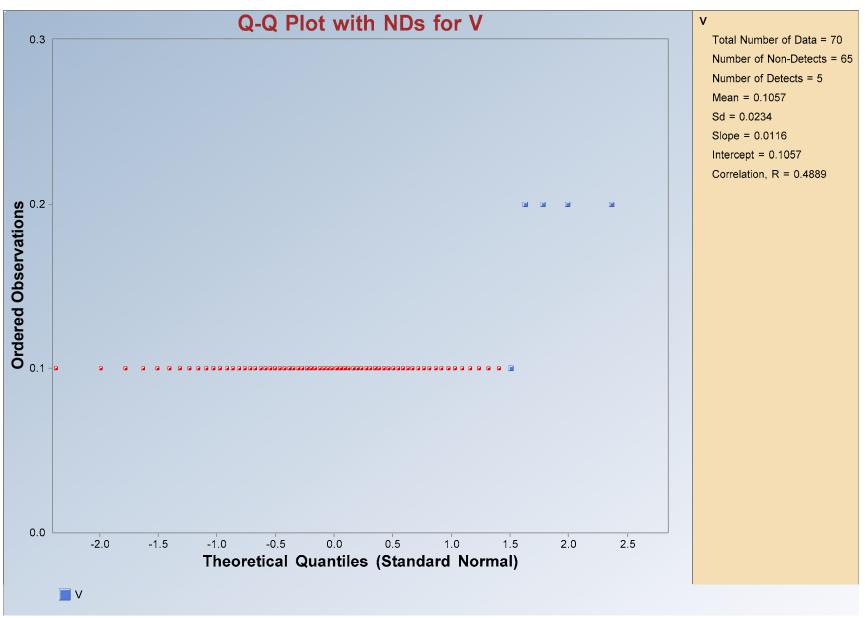
GRAPH B 3.9 Probability Plot of Nickel in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

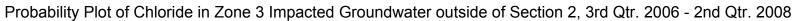


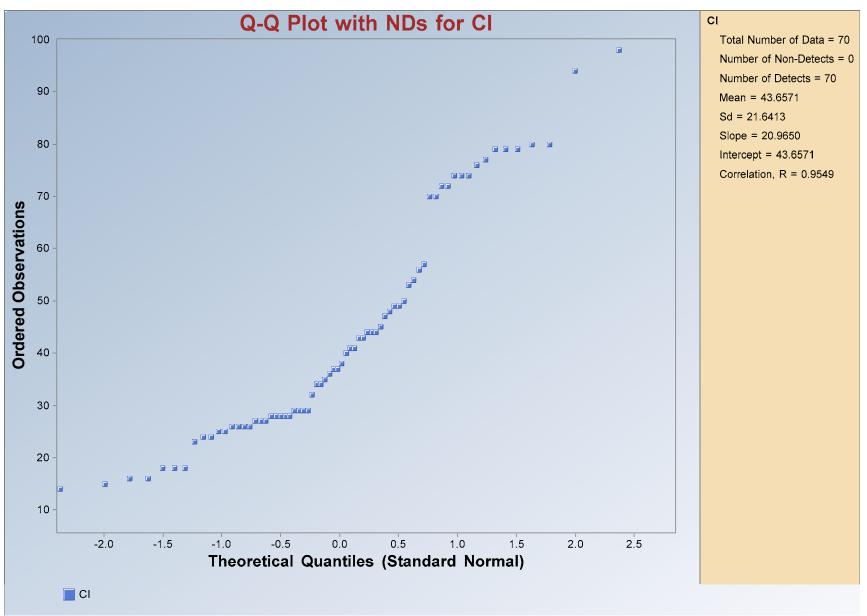
Probability Plot of Selenium in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



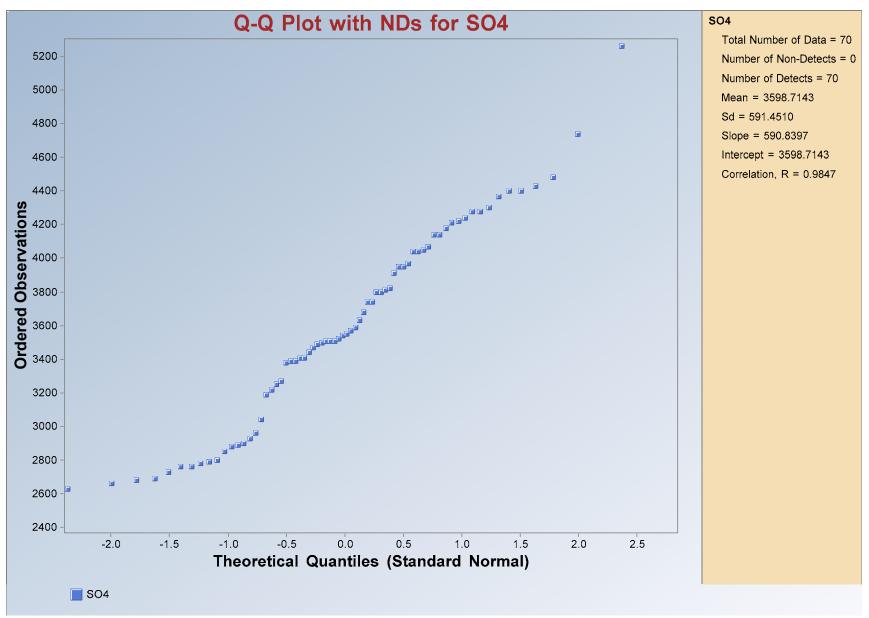
Probability Plot of Vanadium in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



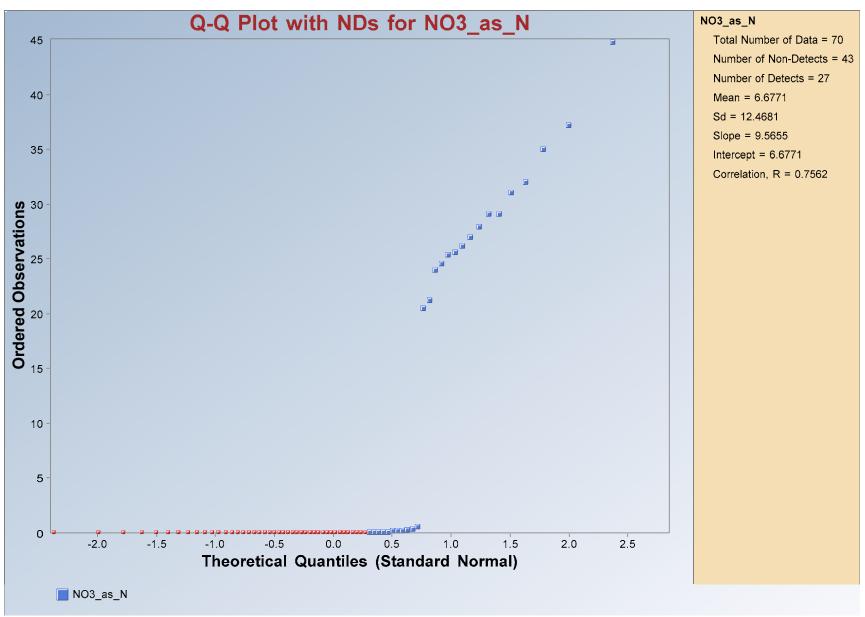




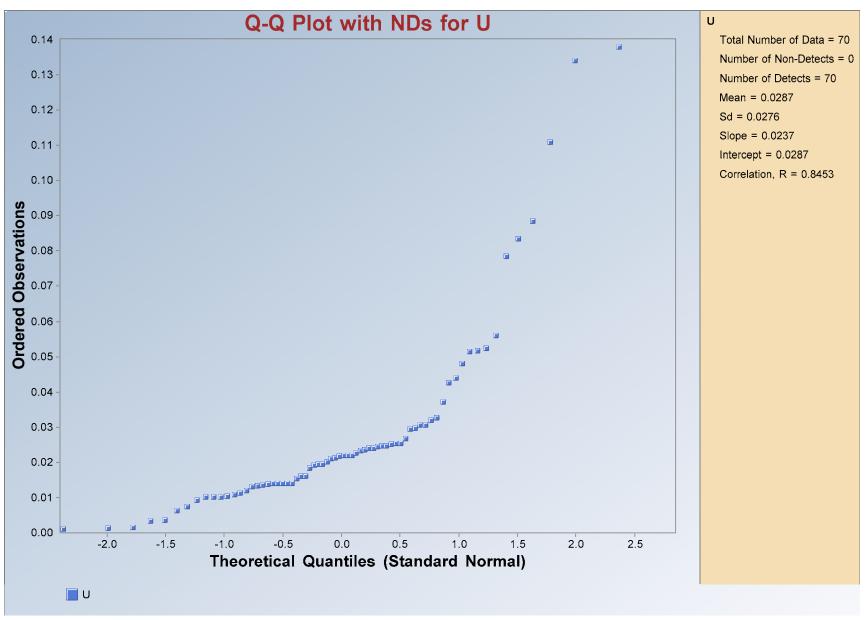
Probability Plot of Sulfate in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



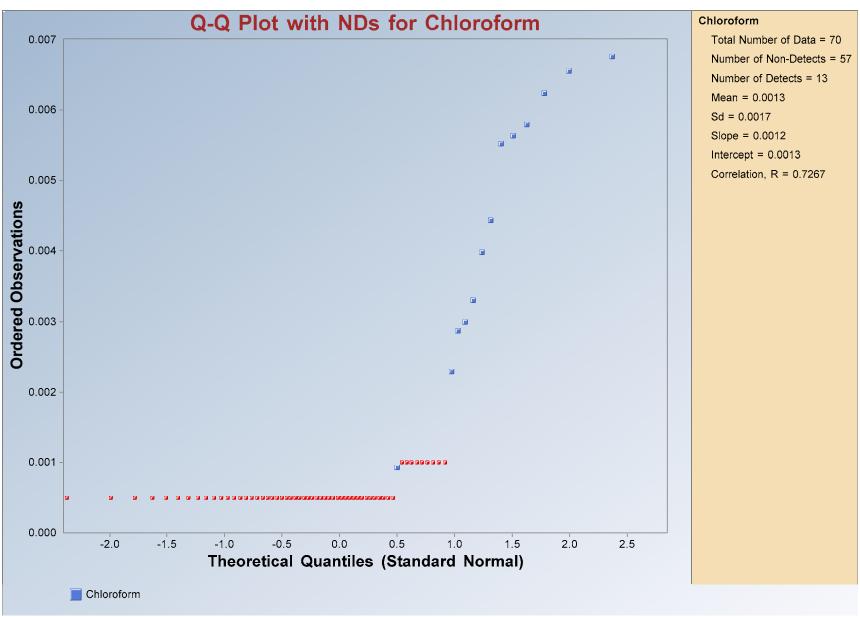
Probability Plot of Nitrate in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



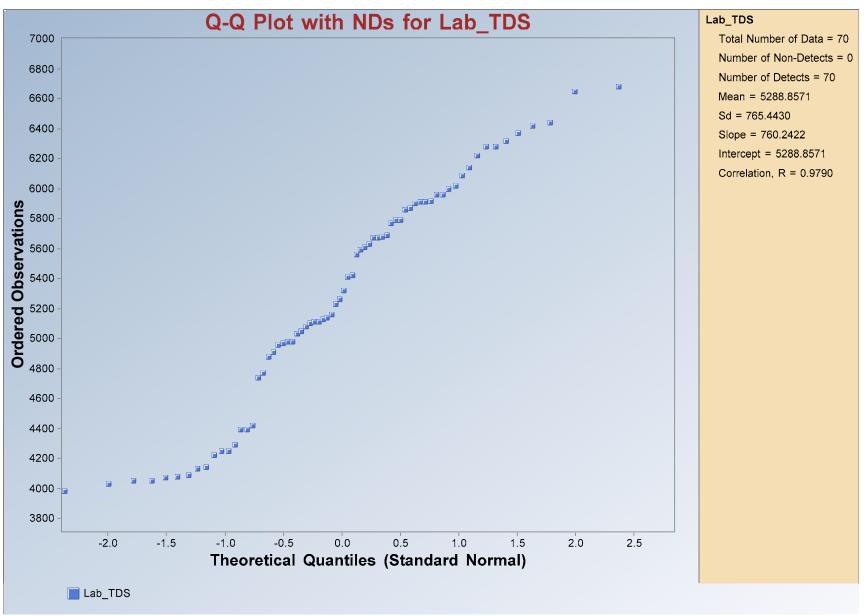
Probability Plot of Uranium in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



Probability Plot of Chloroform in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

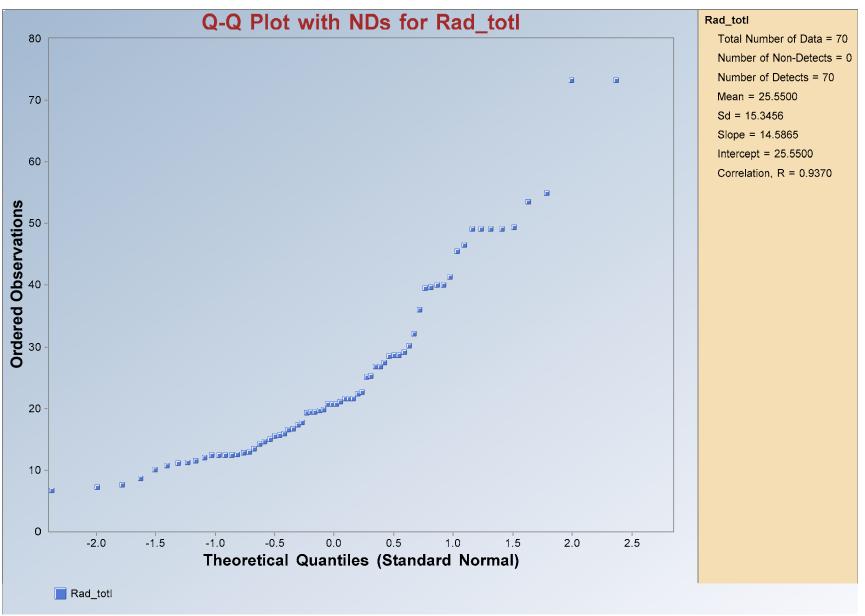




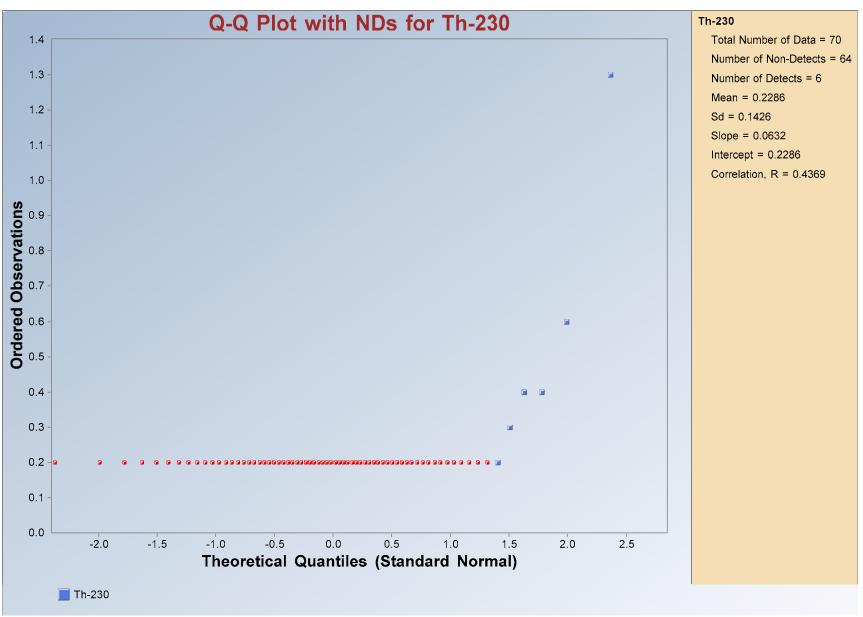


(concentrations in milligrams per liter)

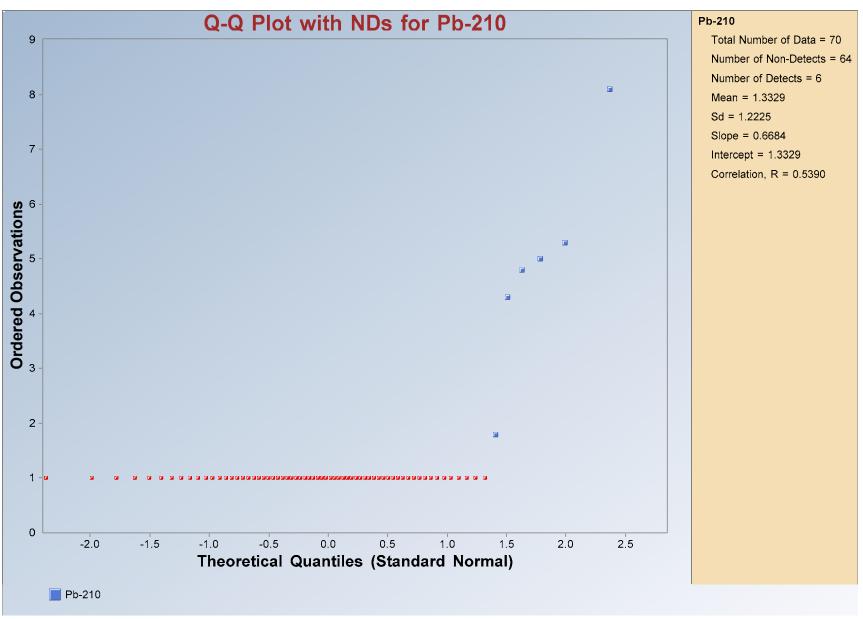




Probability Plot of Thorium-230 in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008

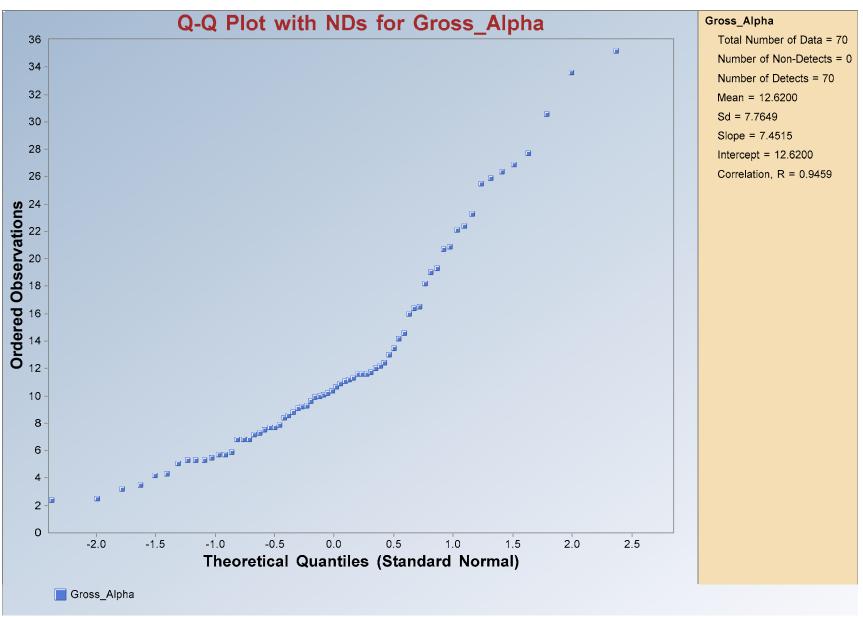


Probability Plot of Lead-210 in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



(concentrations in pico curies per liter)

Probability Plot of Gross Alpha in Zone 3 Impacted Groundwater outside of Section 2, 3rd Qtr. 2006 - 2nd Qtr. 2008



APPENDIX C

Revised Submittal – Calculation of Background Statistics with Comparison Values, UNC Church Rock Mill & Tailings Site, Church Rock, New Mexico. October 17, 2008



N.A. WATER SYSTEMS

This Submittal Delivered by Email Only

October 17, 2008

Ref. No: 56007746 GE Church Rock Project

Mr. Mark Purcell Remedial Project Manager U.S. Environmental Protection Agency 1445 Ross Ave., Suite 1200 (6SF-LP) Dallas, TX 75202-2733

Re: Revised Submittal Calculation of Background Statistics with Comparison Values UNC Church Rock Mill & Tailings Site, Church Rock, New Mexico

Dear Mr. Mark Purcell:

N.A. Water Systems (N.A.WS) is pleased to provide this revised report regarding the calculation of background water statistics with comparison values for the UNC Church Rock Mill & Tailings Site in Church Rock, New Mexico. This report includes revisions to the August 26, 2008 submittal based on comments received from Dennis Beal of Science Applications International Corp. (Beal, SAIC, email communication, Sept. 19, 2008), and other reviewers (Mark Purcell, EPA, and Earle Dixon, NMED) during a teleconference of September 30, 2008.

Introduction

Calculations of background statistics have been completed for the Church Rock project. These calculations were made using results for COPCs (Constituents of Potential Concern) in samples collected from July 1989 through October 2007, inclusive. Similar calculations were made for trace and major metal results obtained from samples collected from May 1988 through April 1989, inclusive. Methods used to calculate the statistics were consistent with those discussed by and agreed to in the teleconference on June 27, 2008. The teleconference participants included representatives of U.S. Environmental Protection Agency (EPA), New Mexico Environment Department (NMED), and N.A.WS. The methodology agreed to in that meeting is summarized by the following steps:



- Use ProUCL software to estimate the upper confidence limits (UCL95) for the means of background populations of COPC concentrations from samples determined to be representative of background groundwater quality. (Background sample sets for the Southwest Alluvium and Zone 1 were established in the February 2006 license amendment request for changing the Groundwater Protection Standard (GWPS) for combined radium. Determination of background sample sets for Zone 3 and for the older trace metal data are presented below).
- 2. Determine which COPCs have higher median concentrations in background groundwater than the comparison values (these are presented below). The method of testing recommended during the June 27 teleconference by the expert consultant to the EPA (Dennis Beal of SAIC) was the single sample hypothesis test. Of the three nonparametric methods available in the ProUCL software, he recommended that the Wilcoxon Signed Rank test be used, rather than the Sign Test or the Test of Proportions.
- 3. Select for consideration as potential modifications to cleanup levels those background UCL95 statistics associated with COPCs that are determined (from single sample hypothesis tests) to have median concentrations equal or exceeding the comparison values. The selected UCL95 statistics (if adopted) would be single-valued standards that will be representative of background UCL95 (i.e., upper confidence limit on the mean at the 95% confidence level). Note that the UCL95 statistics presented in this document as candidates for consideration as modifications to cleanup levels are based solely on statistical calculations.

One of the conclusions of the June 27 teleconference was that the preferred method of comparing site samples to revised background-based cleanup levels is two-sample hypothesis testing (e.g., of a compliance data set against the background data set from which the revised cleanup level was derived). Therefore, one of the objectives of the current work is to define appropriate background data sets for those future comparisons.

Identification of Samples Representative of Background Groundwater Quality

The process used to identify samples representative of background groundwater quality was identical to that described in the license amendment request for changing the GWPS for radium (N.A. Water Systems, February 2006, *Technical Analysis Report in Support of License Am endment Request for Changing the Method of Deter mining Exceedances of the Com bined Radium Groundwater Protection Standard in Source*

Materials License SUA-1475 (T AC LU0092), Groundwater Corrective Action Program, Church Rock Site, Church Rock, New Mexico, pp. 3-6). As such, the wells selected for the purposes of this report as having samples representative of background quality in Zone 1 and in the Southwest Alluvium are the same as those identified in the February 2006 report. One difference is that the data sets used in calculations made for this report are from the period July 1989 through October 2007 inclusive, whereas the February 2006 submittal only included samples collected through October 2005. The methods used to identify wells having background water quality for the February 2006 submittal were used to verify that the designation remained valid for samples collected through October 2007. Table 1 lists wells and sample dates representative of background.

A second difference from the February 2006 report is that the current calculations have been applied to all COPCs, as well as a group of trace metals (plus iron) that had previously been dismissed as COPCs (EPA, August 1988, Draft Final Remedial Investigation, United Nuclear Church Rock Site). The inclusion of former COPCs, which had not been a part of the site's approved Sampling and Analysis Plan, required that the much smaller pool of pre-plan (pre-July 1989) sample results be included in the statistical calculations for these metals. Well samples collected from May 1985 through 1989 (including those from the Remedial Investigation sampling) were evaluated for evidence of background water quality. This resulted in the addition of sample results for metals from wells GW-4 and 623 for the Southwest Alluvium and from well EPA-5 for Zone 1. Time series graphs of indicator parameters for wells GW-4, 623, and EPA-5 are included in Appendix A. Metals results from other background wells (identified in the February 2006 report) were also verified to have come from samples representative of background water quality. In other words, those wells identified as having background water quality subsequent to July 1989 (see Table 1) were found, as expected, to have had background water quality prior to July 1989.

Zone 3 groundwater was not a subject of the February 2006 report; therefore, the identification of samples representative of background water quality in Zone 3 is new to this report. Table 2 lists wells and sample dates representative of background in Zone 3. The following criteria have been used to distinguish background versus impacted groundwater quality in Zone 3:

• pH < 5 and bicarbonate < 100 and > 500 mg/L are useful (but not always definitive) indicators of seepage impact (e.g., see discussion of these empirically derived

criteria in the 2007 annual review report). See Figure 1 for box-and-whiskers plots of bicarbonate and pH for the background wells.

- Time-series of these two indicator parameters are very helpful, sometimes essential. See Appendix A for time-series of pH and bicarbonate for the background wells.
- Well locations within the overall area impacted by seepage (e.g., see Figure 35 in the 2007 annual report).
- Time trends in the concentrations of major ions; in particular, decreasing ratios of Ca:Mg are associated with degrading groundwater quality (see Appendix B in the 2007 annual report; e.g. well EPA-14).
- Time trends in the concentrations of many metals and radionuclides will usually increase as the water quality degrades in Zone 3 (see Appendix B in the 2007 annual report; e.g. well EPA-15).

Invariably, some wells (or certain time spans at some wells) are difficult to classify because their groundwater chemistry tends to be gradational. For example, during the period of time of relevance for present purposes (1989 to October 2007), the geochemistry associated with well 420 is "borderline" – therefore, we have excluded it from the dataset associated with background water quality.

The time-series included in Appendix A show the inferred dates of the onset of seepage impact for those wells whose sampling regime spanned such a transition. Also shown on time series spanning the date is the May 3, 2000 transition to low-flow, unfiltered sampling from multiple-well-volume, purge-and-filter sampling. This transition date does not coincide with any of the inferred onset dates of seepage impact. However, May 2000 appears to coincide with changes of indicator parameter trends at two wells, EPA-5 and EPA-14. It is unlikely that the change of sampling method initiated the multi-year concentration trends noted at these two wells. Other groundwater parameter changes, post-filtration, can be gleaned from a review of the tabulated historic water quality data in the appendices of the 2007 annual review report (N.A. Water Systems, 2008), and these changes cannot be ascribed to the absence of field filtration.

The background sample sets used to make the current calculations have been revised by the removal of small numbers of sample results having unusually high reporting limits. This culling of data affected the sample data sets for each hydrostratigraphic zone. However, it involved less than approximately one percent of the sample data (typically no more than three data points for a particular COPC) and a relatively small number of COPCs. These data were removed because they were discovered to have undesirable consequences on the results of the single sample hypothesis tests (particularly with the recommended Wicoxon Signed Rank Tests). The causes of these problems and the rationale for removing these data were discussed with Dennis Beal by James Ewart. These problems and our solution for them are described below in the discussion of the results.

Results

Basic Statistics and Upper Confidence Limits for Means

Tables 3 through 5 list basic statistics for all COPCs and additional metals calculated from the background data sets from wells in the Southwest Alluvium, Zone 1, and Zone 3. The data sets include only primary samples (i.e. no QA/QC samples). Also listed are upper confidence limits at the 95% confidence level for the means (UCL95). All of the statistics were calculated using ProUCL software (Singh et al., April 2007, *ProUCL Version 4.00.02 User Guide*, EPA/600/R-07/038). The UCL95 estimates were selected from values recommended by the ProUCL software. Summary tables of the output of UCL95 estimates are provided in Appendix B. In cases where two alternative estimates of UCL95 statistics are provided by ProUCL, the higher value was selected and is listed in Tables 3 through 5. The higher values were selected as conservative estimates, consistent with the use of these same statistics as estimators of exposure point concentrations (EPCs).

Comparisons of Background COPC Concentrations with Comparison Values

Table 6 is a compilation of site cleanup goals (EPA, *September 1988, Record of Decision for the Church Rock Site*) at Church Rock and other more recently developed information sources and standards for COPCs and metals. Comparison values were selected from Table 6 (see the green highlighted values) in consultation with Mark Purcell (EPA). Tables 7 through 9 summarize the background concentrations versus comparison values for COPCs and metals in each of the three hydrostratigraphic zones. The results, as presented, deviate in one significant way from the methodology described in the three steps outlined in the Introduction. The selection of candidates for consideration as new background-based cleanup levels (shown in the last column of each table) was based solely on the estimated UCL95 statistics and comparison values (see column 6, titled UCL95>=CV?, meaning "is the UCL95 value greater than or equal to the comparison value?"). The results of single sample hypothesis testing, which are shown in the tables for information purposes, were not used in this determination. This methodological difference was based on an evaluation of the algorithms employed by

the single sample hypothesis test methods (as implemented in ProUCL), and particularly how the accuracy of these methods are affected by the characteristics of the Church Rock datasets. Note that in the following discussion the term nondetect is used as a catchall for censored data, which in the case of the Church Rock data represents a result less than its reporting limit.

Datasets having the following characteristics tended to result in adverse consequences for the single sample hypothesis tests:

- 1. High percentages of results below reporting limits (nondetects).
- 2. Multiple values of reporting limits within datasets (i.e. for a particular COPC).
- 3. Nondetect results having values greater than other results reported as detected within a dataset.
- 4. Highly skewed distributions.

The Church Rock datasets for several COPCs commonly have one or more of these characteristics. The adverse consequences from characteristics 1 through 3 arise largely because of the handling of nondetect values by the algorithms employed by the single sample hypothesis tests in ProUCL. These consequences typically affected the results of the Wilcoxon Signed Rank Tests more than those of the Sign Tests. The reason for this is that the Wilcoxon Signed Rank Test replaces all nondetect results with a value equal to half the detection limit (or reporting limit). Furthermore, the method treats any result (detected or not) less than the highest nondetect result in a dataset as a nondetect result, and accordingly reduces its value by one half. Therefore, having even a single highly valued nondetect result can have a profound (and undesirable) affect on the outcome of the test. This is the reason that large nondetect results were removed from the datasets for these calculations.

In contrast to the Wilcoxon Signed Rank Test, the Sign Test retains nondetect values at their reported (reporting limit) value. The Sign Test also discards any nondetect result that exceeds the value of the comparison value, thereby typically avoiding the problems created by the handling of large nondetect values by the Wilcoxon Signed Rank Test. However, the Sign Test also resulted in questionable "Do Not Reject" outcomes in some cases where 100% of the data were nondetect results. (Note that the null hypothesis used in all the tests is that the median of the background dataset equals or exceeds the comparison value.)

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For example, in cases where more than 50% of the results are nondetects and the reporting limit equals the comparison value, the Sign Test will fail to reject the null hypothesis even though the majority of the data are clearly less than the reporting limit (and the comparison value). This occurs because the Sign Test records a nondetect equaling the comparison value as a tie. The very different handling of nondetects by these two methods is illustrated by the significant differences of outcomes for the two tests shown in Tables 7 through 9. Finally, highly skewed distributions, a characteristic common to many of the Church Rock datasets, is described by the ProUCL documentation as a factor reducing the accuracy of the Wilcoxon Signed Rank Test.

The methodologies used by ProUCL for the Sign Test and Wilcoxon Signed Rank Test were tested by hand calculations. This was done using algorithms published in EPA statistical guidance (EPA, February 2006, *Data Quality Assessment: Statistical Methods for Practitioners*, EPA QA/G-9S, pp. 60-61). The same document is referenced by the ProUCL documentation as a source of its algorithms. The handling of nondetect results by ProUCL was determined to be faithful to the published algorithm for the Wilcoxon Signed Rank Test and numerically accurate. This procedure includes the substitution of values equal to one half of the detection limit (DL/2) for nondetects. The description of the Sign Test in EPA (February, 2006) does not explicitly mention any substitutions for nondetect results. However, the example calculation provided for the Sign Test (EPA, February 2006, Box 3-17, p. 63) indicates the use of the same DL/2 substitution used for the Wilcoxon Signed Rank Test. In this respect the authors of the ProUCL software may have misinterpreted the intentions of their source for the Sign Test algorithm, resulting in an inappropriate treatment of nondetects having the same value as the standard of comparison.

The solution of ignoring the single sample hypothesis tests in favor of direct comparisons of the estimated UCL95 statistics with comparison values avoids the problems described above, and has other additional advantages. Direct comparison of two values has the intuitive advantage of simplicity, and it also avoids the logical inconsistency of concluding (as would have been the case in some instances) that a UCL95 statistic should be considered lower than a comparison value when it obviously is not.

Equally important is the observation that the methodologies employed by the ProUCL software for estimating UCL95 statistics are highly advanced relative to the algorithms used by the single sample hypothesis tests. For example, a battery of more than 20 independent algorithms is employed by ProUCL to estimate UCL95 statistics. The

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software automatically sifts through these methods to recommend the better one or two estimates according to a variety of dataset characteristics, including number of samples, numbers and values of nondetect results, and shape of the distribution (including skewness). In particular, the handling of nondetect values by the Kaplan-Meier methods (for estimating UCL95 statistics) is more sophisticated than the methods used by the single sample hypothesis tests.

In the teleconference of September 30, 2008, it was agreed that this use of UCL95 statistics was an acceptable alternative to the single sample hypothesis test, for evaluating the background data sets versus the comparison values.

Note that direct evaluation of background UCL95 statistics versus comparison values are being made solely for the purpose of determining whether those statistics are numerically greater than the respective COPC comparison values. It would be inappropriate to use the same methodology for comparisons of compliance samples to cleanup levels, because of the much smaller size of compliance well sample sets (relative to background sample sets). For such comparisons single- or two-sample hypothesis testing is preferable.

Conclusion

We have concluded that direct evaluation of UCL95 statistics versus comparison values is the preferable method of determining which UCL95 statistics should be selected as candidates for consideration for modifying cleanup levels to reflect background concentrations. The UCL95 statistics presented in this document as candidates for consideration as modified cleanup levels are based solely on statistical calculations. One of the conclusions of the June 27, 2008, teleconference was that the preferred method of comparing compliance samples to background-based cleanup levels is twosample hypothesis testing (e.g. of a compliance data set against the defined background data set). Therefore, one of the objectives of the current work has been to define appropriate background data sets for those future comparisons. Mark Purcell U.S. EPA October 17, 2008

Very Truly Yours,

an a G5

James Ewart, Ph.D., P.G. Technical Consultant

- JE: abc-220-mj
- cc: Roy Blickwedel, GE Larry Bush, UNC Earle C. Dixon, NMED

Attachments

N.A. WATER SYSTEMS

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Figures and Tables



N.A. Water Systems, LLC Airside Business Park, 250 Airside Drive Moon Township, Pennsylvania 15108-2793 USA Tel: 412-809-6000 • Fax: 412-809-6075 Web site: www.nawatersystems.com



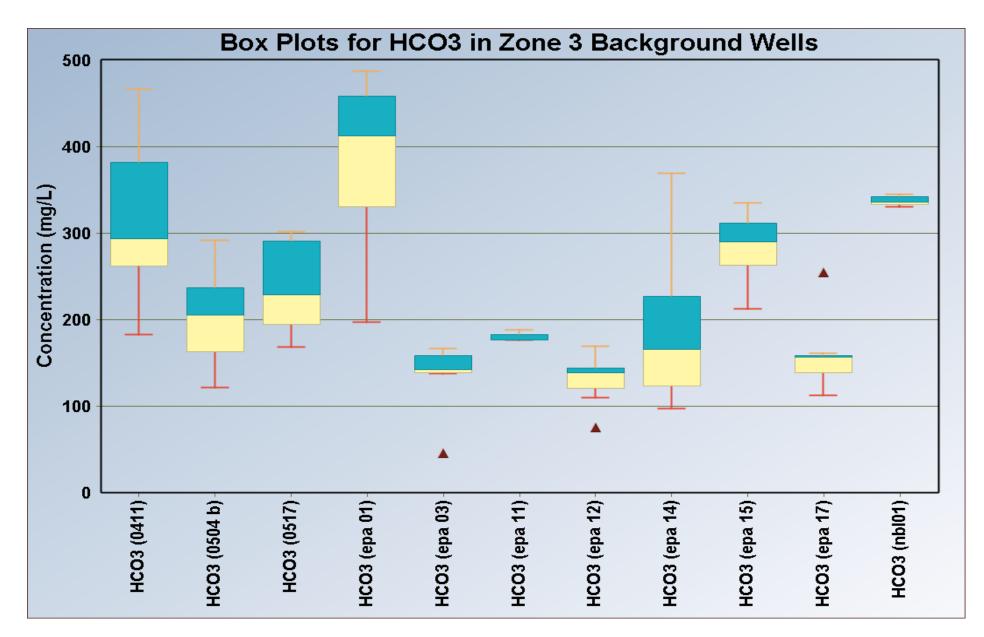


FIGURE 1B

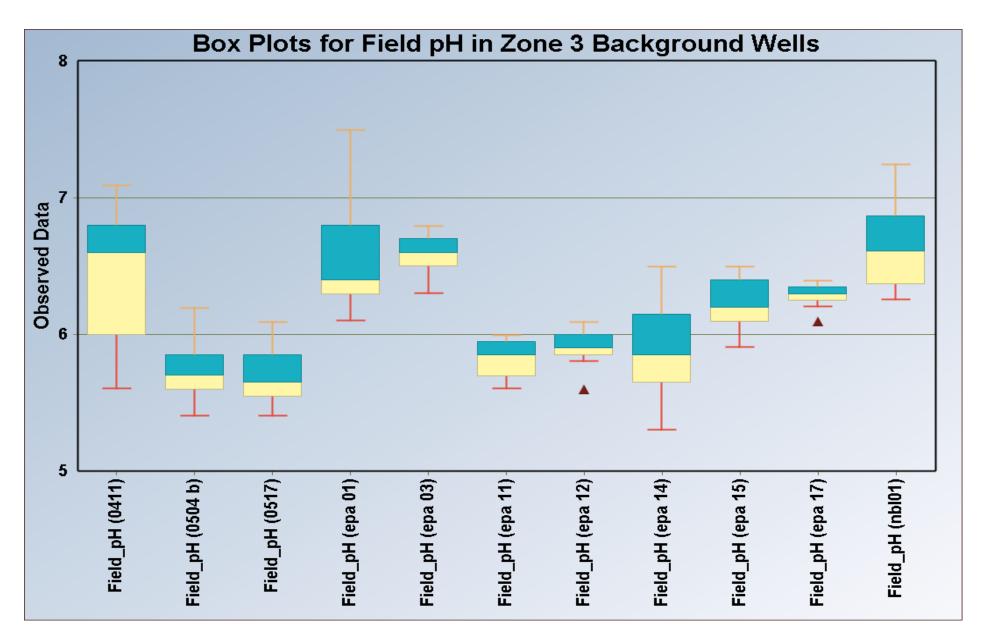


TABLE 1

Southwest Alluvium and Zone 1 Wells Having Samples Representative of Background Water Quality

Southwest Alluvium	Zone 1
29 A	619
624 (Jul 89 - Oct 95)	EPA 2
627	EPA 4 (POC)
639	EPA 8
642	
645	
EPA 22 A	
EPA 25 (Jul 89 - Oct 95)	
EPA 27	
EPA 28 (POC)	
SBL 1	

Notes:

POC = Point-of-Compliance Well. The following wells were included only for the pre-July 1989 metals results: GW 4 and 623 (SWA) EPA 5 (Zone 1)

TABLE 2

Zone 3 Wells Having Samples Representative of Background Water Quality

Well	Sampled Time Period
411	Jul 89 - Jan 98
504 B	Jul 89 - Apr 92
517 (POC)	Jul 89 - Apr 91
EPA 01	Jul 89 - Oct 97
EPA 03	Jul 89 - Oct 91
EPA 11	Jul 89 - Apr 90
EPA 12	Jul 89 - Apr 92
EPA 14	Jul 89 - Apr 95
EPA 15	Jul 89 - Apr 95
EPA 17	Jul 89 - Apr 92
NBL-01	Aug 01 - Jan 04

Note: POC = Point-of-Compliance well.

 TABLE 3

 Summary Statistics for COPCs and Trace Metals in Southwest Alluvium Background Groundwater

					-	-		
		Total	Percent		Maximum			
Parameter	Units	Data	Nondetect	Detected	Detected		Detected	of Mean
AI	mg/L	391	94.6%	0.1	0.6		0.14	0.107
As	mg/L	391	93.1%	0.001	0.01		0.001	0.00116
Be	mg/L	389	100.0%	N/A	N/A	N/A	N/A	N/A
Cd	mg/L	391	96.9%	0.006	0.07	0.0255	0.01	0.0108
Co	mg/L	391	81.6%	0.01	0.06	0.0186	0.02	0.0121
Pb	mg/L	388	99.5%	0.05	0.07	0.06	0.06	0.0502
Mn	mg/L	389	11.8%	0.01	3.35	0.339	0.13	0.414
Мо	mg/L	391	99.5%	0.03	0.03	N/A	N/A	N/A
Ni	mg/L	391	96.4%	0.05	0.17	0.08	0.08	0.0613
Se	mg/L	390	50.5%	0.001	0.195	0.00708	0.003	
V	mg/L	391	100.0%	N/A	N/A	N/A	N/A	N/A
CI	mg/L	391	0.0%	9.8	169	74.82	67.8	83.72
SO4	mg/L	391	0.0%	605	5830	2401	2420	2468
NO3_as_N	mg/L	391	1.3%	0.09	1225	99.54	74.1	137.4
U	mg/L	390	0.3%	0.001	0.367	0.0419	0.031	0.0459
Chloroform	ug/L	391	100.0%	N/A	N/A	N/A	N/A	N/A
Lab_TDS	mg/L	390	0.0%	1310	10530	4630	4795	4745
Rad-226	pCi/L	391	34.3%	0.2	9.4	0.979	0.6	0.798
Rad-228	pCi/L	391	67.8%	1	7	2.55	2.2	1.611
Rad_totl	pCi/L	391	25.3%	0.2	12	1.9	1.3	1.621
Th-230	pCi/L	391	91.8%	0.2		2.841	1.6	0.509
Pb-210	pCi/L	391	78.3%	1	14.2	2.845	2.2	1.513
Gross_Alpha	pCi/L	391	70.6%	0.4	17.8	3.35	2.1	1.693
Sb	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ва	mg/L	26	100.0%	N/A	N/A	N/A	N/A	N/A
Cr	mg/L	37	97.3%	0.29	0.29	N/A	N/A	N/A
Cu	mg/L	13	84.6%	0.01	0.01	N/A	N/A	N/A
Fe	mg/L	19	79.0%	0.06	1.4	0.418	0.105	0.275
Hg	mg/L	8	100.0%	N/A	N/A	N/A	N/A	N/A
Ag	mg/L	21	100.0%	N/A	N/A	N/A	N/A	N/A
TĪ	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Zn	mg/L	25	40.0%	0.02	0.429	0.0891	0.05	0.0949

 TABLE 4

 Summary Statistics for COPCs and Trace Metals in Zone 1 Background Groundwater

		Total	Percent		Maximum			
Parameter	Units	Data	Nondetect		Detected		Detected	
AI	mg/L	234		0.1	0.6	0.185		0.117
As	mg/L	234	83.8%	0.001	0.004			0.00117
Be	mg/L	234	100.0%	N/A	N/A	N/A	N/A	N/A
Cd	mg/L	234	98.7%	0.005	0.01	0.00733	0.007	0.0051
Co	mg/L	234	89.7%	0.01	0.06	0.0171	0.01	0.0112
Pb	mg/L	234	99.6%	0.05	0.05	N/A	N/A	N/A
Mn	mg/L	234	0.4%	0.66	4.15	2.434	2.65	2.519
Мо	mg/L	234	97.9%	0.03	0.27	0.12	0.13	0.132
Ni	mg/L	230	98.7%	0.06	0.07	0.0667	0.07	0.0602
Se	mg/L	234	95.7%	0.001	0.004	0.0019	0.0015	0.00107
V	mg/L	234	100.0%	N/A	N/A	N/A	N/A	N/A
CI	mg/L	234	0.0%	19.4	252	37.13		39.03
SO4	mg/L	234	0.0%	1410	3882	2703	2952	2773
NO3_as_N	mg/L	233	71.7%	0.01	51.8	1.767	0.16	1.754
U	mg/L	233	16.7%	0.0004	0.975	0.00862	0.0013	0.0255
Chloroform	ug/L	234	99.6%	0.91	0.91	N/A	N/A	N/A
Lab_TDS	mg/L	234	0.0%	2490	5610	4225	4569	4319
Rad-226	pCi/L	233	1.7%	0.2	5.4	1.269	1.2	1.314
Rad-228	pCi/L	234	29.9%	1	13.8	3.457	3.1	2.946
Rad_totl	pCi/L	234	0.9%	0.2	14.8	3.618	3.35	3.841
Th-230	pCi/L	234	91.9%	0.2	4.9	0.974	0.7	0.403
Pb-210	pCi/L	234	80.8%	1.1	9.1	2.58		1.579
Gross_Alpha	pCi/L	234	35.0%	0.9	14	2.757	2	2.361
Sb	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ва	mg/L	14	78.6%	0.079	0.091	0.0847	0.084	0.091
Cr	mg/L	11	100.0%	N/A	N/A	N/A	N/A	N/A
Cu	mg/L	4	75.0%	0.026		N/A	N/A	N/A
Fe	mg/L	12	8.3%	0.25	14	6.386	6.2	8.701
Hg	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Ag	mg/L	11	100.0%	N/A	N/A	N/A	N/A	N/A
TĪ	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Zn	mg/L	16	56.3%	0.01	5	0.784	0.046	3.583

 TABLE 5

 Summary Statistics for COPCs and Trace Metals in Zone 3 Background Groundwater

		Total	Percent		Maximum			
Parameter	Units	Data	Nondetect				Detected	
AI	mg/L	186	68.28%	0.1	1.68		0.31	0.231
As	mg/L	186	26.88%	0.001			0.0235	0.175
Be	mg/L	186	100.00%	N/A	N/A	N/A	N/A	N/A
Cd	mg/L	186	95.16%	0.01	0.09	0.02	0.01	0.0113
Co	mg/L	186	9.14%	0.01	0.53	0.0835	0.06	0.0877
Pb	mg/L	185	97.84%	0.05		0.065	0.065	0.0701
Mn	mg/L	186	0.54%	0.42	7.5	3.25	3.3	3.436
Мо	mg/L	184	14.13%	0.02	75	11.88	3.76	17.43
Ni	mg/L	186	39.25%	0.05	0.67	0.173	0.12	0.14
Se	mg/L	186	77.42%	0.001	0.015	0.0026	0.001	0.00159
V	mg/L	186	100.00%	N/A	N/A	N/A	N/A	N/A
CI	mg/L	186	0%	15	66	31.62	30.85	32.65
SO4	mg/L	186	0%	1319	4674	2588	2651	2674
NO3_as_N	mg/L	186	17.20%	0.01	61	11.34	4.785	15.61
U	mg/L	186	1.08%	0.0007	0.38	0.0791	0.039	0.107
Chloroform	ug/L	186	99.46%	1.1	1.1	N/A	N/A	N/A
Lab_TDS	mg/L	186	0%	2244	6930	4115	4237	4239
Rad-226	pCi/L	186	11.83%	0.2	23.7	5.01	4.5	4.996
Rad-228	pCi/L	185	29.19%	1	22.3	5.34	4.3	4.509
Rad_totl	pCi/L	185	9.73%	0.2	40.9	9.099	7.9	10.66
Th-230	pCi/L	186	89.78%	0.2	57	6.705	2.3	1.426
Pb-210	pCi/L	186	69.35%	1	11	2.549	2	1.618
Gross_Alpha	pCi/L	186	15.59%	1	69	8.191	5.4	8.217
Sb	mg/L	1	100.0%	N/A	N/A	N/A	N/A	N/A
Ba	mg/L	36	94.4%	0.54		N/A	N/A	N/A
Cr	mg/L	37	100.0%	N/A	N/A	N/A	N/A	N/A
Cu	mg/L	13	76.9%	0.028			0.038	0.06
Fe	mg/L	23	39.1%	0.03	67	9.682	1.45	12.16
Hg	mg/L	4	100.0%	N/A	N/A	N/A	N/A	N/A
Ag	mg/L	29	100.0%	N/A	N/A	N/A	N/A	N/A
TĪ	mg/L	0	N/A	N/A	N/A	N/A	N/A	N/A
Zn	mg/L	31	19.4%	0.02	6.859	0.766	0.193	3.539

TABLE 6 Contaminant-Specific Groundwater Cleanup Levels and Other Comparison Values United Nuclear Corporation, Church Rock Site Church Rock, New Mexico

	ROD (September 198	,	-Year Review	NRC Source							pared to in 2007				
Source		(September 20	003, Table 3-1)		Materials	Materials Post-ROD Promulgated ARARs			Annual Review		Current Health-Based Criteria (+)		_		
Contaminant	New Mexico WQCC Standards	Health-based	Maximum Concentration Limit (MCL)	Background Level	License Compliance Standards	NRC Appendix List*	New Mexico WQCC Standards		EPA I MCL	Orinking Water Other**	ЕРА	NRC	Health-Based Criterion	Source	
Sulfate				2160			2125***				2125***				SO ₂
Total Dissolved Solids				3170			4800***				4800***				TDS
NO3 as N				30			190***		10		190***		10	R6HHSL, MCL	NO
Manganese				2.6			0.2	0			2.6		1.7	R6HHSL	Mn
Chloride	250						250	0			250				Cl2
Aluminum	5						5	Ι			5		37	R6HHSL	Al
Antimony		0.014							0.006				0.006	MCL	Sb
Arsenic			0.05		0.05	0.05	0.1	HH	0.01		0.05	0.05	0.01	MCL	As
Barium	1		1			1	1	HH	2				2	MCL	Ba
Beryllium		0.017			0.05				0.004		0.017	0.05	0.004	MCL	Be
Cadmium	0.01		0.01		0.01	0.01	0.01	HH	0.005		0.01	0.01	0.005	MCL	Cd
Chromium	0.05		0.05			0.05	0.05	HH	0.1				0.1	MCL	Cr
Cobalt	0.05						0.05	Ι			0.05		0.73	R6HHSL	Со
Copper	1						1	0	1.3	MCLG & TT			1.3	MCL	Cu
Iron				5.5			1	0					26	R6HHSL	Fe
Lead	0.05		0.05		0.05	0.05	0.05	HH	0.015	MCLG & TT	0.05	0.05	0.015	MCL	Pb
Mercury	0.002		0.002			0.002	0.002	HH	0.002				0.002	MCL	Hg
Molybdenum	1						1	Ι			1		0.18	R6HHSL	Mo
Nickel	0.2				0.05		0.2	Ι			0.2	0.05	0.73	R6HHSL	Ni
Selenium			0.01		0.01	0.01	0.05	HH	0.05		0.01	0.01	0.05	MCL	Se
Silver	0.05		0.05			0.05	0.05	HH					0.18	R6HHSL	Ag
Thallium		0.014							0.002	MCLG = 0.0005			0.002	MCL	T1
Vanadium		0.7			0.1						0.7	0.1	0.18	R6HHSL	Zn
Zinc	10						10	0					11	R6HHSL	V
TTHMs****					0.08		0.1	HH	0.08	MCLG = 0.07****		0.08	0.08	MCL	TTH
Uranium	5				0.3		0.03	HH	0.03		5	0.3	0.03	MCL	U
Radium 226 and 228			5 pCi/l		****	5 pCi/l	30 pCi/l	HH			5 pCi/l	****	5 pCi/l	MCL	comb
Lead-210			<u> </u>		1 pCi/l		Ŧ				1 · · · · ·	1 pCi/l	0.0541 pCi/l	PRG	Pb-2
Thorium-230			15 pCi/l		5 pCi/L							5 pCi/l	0.523 pCi/l	PRG	Th-2
Gross Alpha			15 pCi/l		15 pCi/l	15 pCi/l			15 pCi/l		15 pCi/l	15 pCi/l	15 pCi/l	MCL	GA

Notes:

Units = mg/L unless otherwise noted

Yellow or shaded cells = constituents not analyzed since site active remediation started in 1989, per EPA FS (August 1988) and ROD (September 1988)

* 10 CFR Appendix A to Part 40

** "Other" includes non-zero Maximum Contaminant Level Goals (MCLG) or Treatment Technology Action Levels (TT)

*** New Mexico Environment Department recommended background values (letter to EPA of January 6, 1998); EPA has not formally adopted these revisions

**** TTHMs (total trihalomethanes) include chloroform; TTHMs MCL = 0.08 mg/L; in addition, chloroform has an MCLG = 0.07 mg/L

***** Combined radium NRC Site Groundwater Protection Standards are 5.0 pCi/L for Zone 3; 5.2 pCi/L for Southwest Alluvium (background); and 9.4 pCi/L for Zone 1 (background)

(+) Sources of health-based criteria include EPA Region 6 Human Health Medium-Specific Screening Levels (R6HHSL) and EPA Preliminary Remediation Goals for Radionuclides (PRGs). For those contaminants with federal MCLs, the MCL is shown as the health-based screening level, per January 25, 2008 letter from EPA to UNC (General Comment 5).

HH = Human Health Standard

I = Irrigation Standard

O = Other Standards for domestic water supply

"Comparison Values" column in N.A. Water Systems report: Calculation of Background Statistics with Comparison Values (Tables 7 to 9)

		Comparison				-			Hypothesis Test ³ edian >= CV	Potential Background
Parameter	Units	Value ¹	Max RL ²	UCL95	UCL95>CV?	Max RL>=CV?	Percent < RL	Sign Test	Wilcoxon Signed	Level
AI	mg/L	5	0.1	0.107	NO	NO	95%	Reject	Reject	
As	mg/L	0.01	0.001	0.00116	NO	NO	93%	Reject	Reject	
Be	mg/L	0.004	0.1	NA	N/A	YES	100%	N/A	N/A	
Cd	mg/L	0.005	0.01	0.0108	YES	YES	97%	Do not Reject	Reject	0.0108
Со	mg/L	0.05	0.01	0.0121	NO	NO	82%	Reject	Reject	
Pb	mg/L	0.05	0.05	0.0502	YES	YES	99%	Do not Reject	Reject	0.0502
Mn	mg/L	0.2	0.01	0.414	YES	NO	12%	Reject	Reject	0.414
Мо	mg/L	1	0.1	NA	N/A	NO	99%	Reject	Reject	
Ni	mg/L	0.2	0.05	0.0613	NO	NO	96%	Reject	Reject	
Se	mg/L	0.05	0.001	0.00516	NO	NO	51%	Reject	Reject	
V	mg/L	0.1	0.1	NA	N/A	YES	100%	N/A	N/A	
CI	mg/L	250	N/A	83.72	NO	N/A	0%	Reject	Reject	
SO4	mg/L	2125	N/A	2468	YES	N/A	0%	Do not Reject	Do not Reject	2468
NO3_as_N	mg/L	30	0.1	137.4	YES	NO	1%	Do not Reject	Do not Reject	137.4
Ū	mg/L	0.03	0.0003	0.0459	YES	NO	0%	Do not Reject	Do not Reject	0.0459
Chloroform	ug/L	80	1	NA	N/A	NO	100%	N/A	N/A	
Lab_TDS	mg/L	3170	N/A	4745	YES	N/A	0%	Do not Reject	Do not Reject	4745
Rad_totl	pCi/L	5	0.2	1.621	NO	NO	25%	Reject	Reject	
Th-230	pCi/L	5	0.2	0.509	NO	NO	92%	Reject	Reject	
Pb-210	pCi/L	1	1	1.513	YES	YES	78%	Do not Reject	Reject	1.513
Gross_Alpha	pCi/L	15	1	1.693	NO	NO	71%	Reject	Reject	
Sb	mg/L	0.006	N/A	NA	N/A	N/A	N/A	no data	no data	
Ва	mg/L	2	0.1	NA	N/A	NO	100%	Reject	Reject	
Cr	mg/L	0.05	0.05	NA	N/A	YES	97%	Reject	Reject	
Cu	mg/L	1	0.02	NA	N/A	NO	85%	Reject	Reject	
Fe	mg/L	1	0.1	0.275	NO	NO	79%	Reject	Reject	
Hg	mg/L	0.002	0.001	NA	N/A	NO	100%	N/A	N/A	
Ag	mg/L	0.05	0.05	NA	N/A	YES	100%	N/A	N/A	
TĬ	mg/L	0.002	N/A	NA	N/A	N/A	N/A	no data	no data	
Zn	mg/L	10	0.1	0.0949	NO	NO	40%	Reject	Reject	

 TABLE 7

 Summary comparisons of Parameter Concentrations in Southwest Alluvium Background Groundwater to Comparison Values

Note:

1. See Table 6 for sources of Comparison Values (CV)

2. RL is an abbreviation of reporting limit

3. Single sample hypotheses tests are not applicable to datasets having 100% censored data

		Comparison						• •	Hypothesis Test ³ edian >= CV	Potential Background
Parameter	Units	Value ¹	Max RL ²	UCL95	UCL95>CV?	Max RL>=CV?	Percent < RL	Sign Test	Wilcoxon Signed	Level
Al	mg/L	5	0.1	0.117	NO	NO	87%	Reject	Reject	
As	mg/L	0.01	0.001	0.00117	NO	NO	84%	Reject	Reject	
Be	mg/L	0.004	0.05	N/A	N/A	YES	100%	N/A	N/A	
Cd	mg/L	0.005	0.01	0.0051	YES	YES	99%	Do not Reject	Reject	0.0051
Co	mg/L	0.05	0.01	0.0112	NO	NO	90%	Reject	Reject	
Pb	mg/L	0.05	0.05	N/A	N/A	YES	100%	N/A	N/A	
Mn	mg/L	0.2	0.01	2.519	YES	NO	0%	Do not Reject	Do not Reject	2.519
Мо	mg/L	1	0.1	0.132	NO	NO	98%	Reject	Reject	
Ni	mg/L	0.2	0.05	0.0602	NO	NO	99%	Reject	Reject	
Se	mg/L	0.05	0.001	0.00107	NO	NO	96%	Reject	Reject	
V	mg/L	0.1	0.1	N/A	N/A	YES	100%	N/A	N/A	
CI	mg/L	250	N/A	39.03	NO	N/A	0%	Reject	Reject	
SO4	mg/L	2125	N/A	2773	YES	N/A	0%	Do not Reject	Do not Reject	2773
NO3_as_N	mg/L	30	0.1	1.754	NO	NO	72%	Reject	Reject	
U	mg/L	0.03	0.0004	0.0255	NO	NO	17%	Reject	Reject	
Chloroform	ug/L	80	1	N/A	N/A	N/A	100%	N/A	N/A	
Lab_TDS	mg/L	3170	N/A	4319	YES	N/A	0%	Do not Reject	Do not Reject	4319
Rad_totl	pCi/L	5	0.2	3.841	NO	NO	1%	Reject	Reject	
Th-230	pCi/L	5	0.2	0.403	NO	NO	92%	Reject	Reject	
Pb-210	pCi/L	1	1	1.579	YES	YES	81%	Do not Reject	Reject	1.579
Gross_Alpha	pCi/L	15	1	2.361	NO	NO	35%	Reject	Reject	
Sb	mg/L	0.006	N/A	N/A	N/A	N/A	N/A	no data	no data	
Ва	mg/L	2	0.1	0.091	NO	NO	79%	Reject	Reject	
Cr	mg/L	0.05	0.05	N/A	N/A	YES	100%	N/A	N/A	
Cu	mg/L	1	0.02	N/A	N/A	NO	75%	Do not Reject	Do not Reject	
Fe	mg/L	1	0.1	8.701	YES	NO	8%	Do not Reject	Do not Reject	8.701
Hg	mg/L	0.002	N/A	N/A	N/A	N/A	N/A	no data	no data	
Ag	mg/L	0.05	0.05	N/A	N/A	N/A	100%	N/A	N/A	
TĬ	mg/L	0.002	N/A	N/A	N/A	N/A	N/A	no data	no data	
Zn	mg/L	10	0.1	3.583	NO	NO	56%	Reject	Reject	

 TABLE 8

 Summary Comparisons of Parameter Concentrations in Zone 1 Background Groundwater to Comparison Values

Note:

1. See Table 6 for sources of Comparison Values (CV)

2. RL is an abbreviation of reporting limit

3. Single sample hypotheses tests are not applicable to datasets having 100% censored data

		Comparison				1			Hypothesis Test ³ edian >= CV	Potential Background
Parameter	Units	Value ¹	Max RL ²	UCL95	UCL95>CV?	Max RL>=CV?	Percent < RL	Sign Test	Wilcoxon Signed	Level
Al	mg/L	5	0.1	0.231	NO	NO	68%	Reject	Reject	
As	mg/L	0.01	0.001	0.175	YES	NO	27%	Do Not Reject	Do Not Reject	0.175
Be	mg/L	0.004	0.05	N/A	N/A	YES	100%	N/A	N/A	
Cd	mg/L	0.005	0.01	0.0113	YES	YES	95%	Do Not Reject	Reject	0.0113
Co	mg/L	0.05	0.01	0.0877	YES	NO	9%	Do Not Reject	Do Not Reject	0.0877
Pb	mg/L	0.05	0.05	0.0701	YES	YES	98%	Do Not Reject	Reject	0.0701
Mn	mg/L	0.2	0.01	3.436	YES	NO	1%	Do Not Reject	Do Not Reject	3.436
Мо	mg/L	1	0.1	17.43	YES	NO	14%	Do Not Reject	Do Not Reject	17.43
Ni	mg/L	0.2	0.05	0.14	NO	NO	39%	Reject	Reject	
Se	mg/L	0.05	0.001	0.00159	NO	NO	77%	Reject	Reject	
V	mg/L	0.1	0.1	N/A	N/A	YES	100%	N/A	N/A	
CI	mg/L	250	N/A	32.65	NO	N/A	0%	Reject	Reject	
SO4	mg/L	2125	N/A	2674	YES	N/A	0%	Do Not Reject	Do Not Reject	2674
NO3_as_N	mg/L	30	0.1	15.61	NO	NO	17%	Reject	Reject	
U	mg/L	0.03	0.0003	0.107	YES	NO	1%	Do Not Reject	Do Not Reject	0.107
Chloroform	ug/L	80	1	N/A	N/A	NO	99%	Reject	Reject	
Lab_TDS	mg/L	3170	N/A	4239	YES	N/A	0%	Do Not Reject	Do Not Reject	4239
Rad_totl	pCi/L	5	0.2	10.66	YES	NO	10%	Do Not Reject	Do Not Reject	10.66
Th-230	pCi/L	5	0.2	1.426	NO	NO	90%	Reject	Reject	
Pb-210	pCi/L	1	1	1.618	YES	YES	69%	Do Not Reject	Reject	1.618
Gross_Alpha	pCi/L	15	1	8.217	NO	NO	16%	Reject	Reject	
Sb	mg/L	0.006	0.05	N/A	N/A	YES	100%	N/A	N/A	
Ва	mg/L	2	0.1	N/A	N/A	NO	94%	Reject	Reject	
Cr	mg/L	0.05	0.1	N/A	N/A	YES	100%	Reject	Reject	
Cu	mg/L	1	0.02	0.06	NO	NO	77%	Reject	Reject	
Fe	mg/L	1	0.1	12.16	YES	NO	39%	Do not Reject	Do not Reject	12.16
Hg	mg/L	0.002	0.0002	N/A	N/A	NO	100%	N/A	N/A	
Ag	mg/L	0.05	0.05	N/A	N/A	YES	100%	N/A	N/A	
TĬ	mg/L	0.002	N/A	N/A	N/A	N/A	N/A	no data	no data	
Zn	mg/L	10	0.1	3.539	NO	NO	19%	Reject	Reject	

 TABLE 9

 Summary Comparisons of Parameter Concentrations in Zone 3 Background Groundwater to Comparison Values

Note:

1. See Table 6 for sources of Comparison Values (CV)

2. RL is an abbreviation of reporting limit

3. Single sample hypotheses tests are not applicable to datasets having 100% censored data

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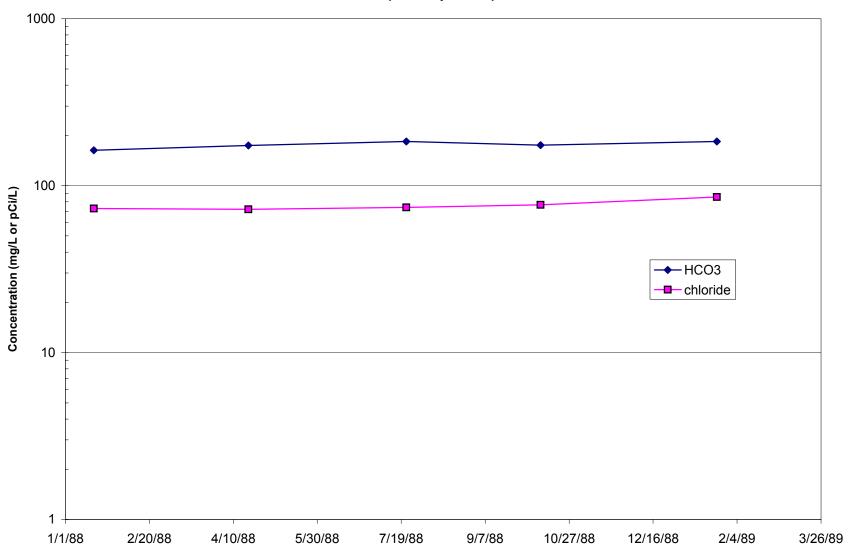
This Submittal Delivered by Email Only

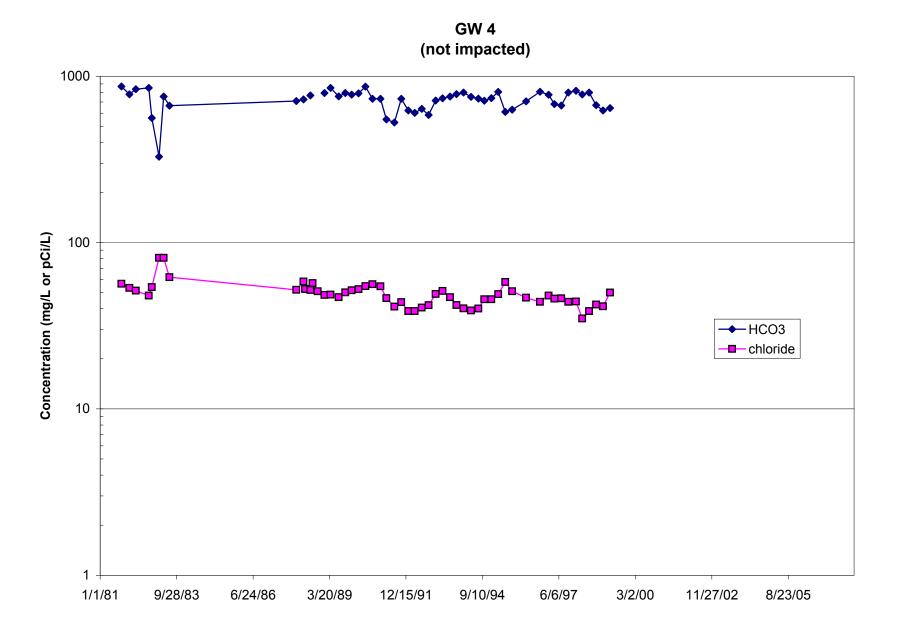
Appendix A Time Series Graphs of Indicator Parameters

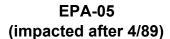


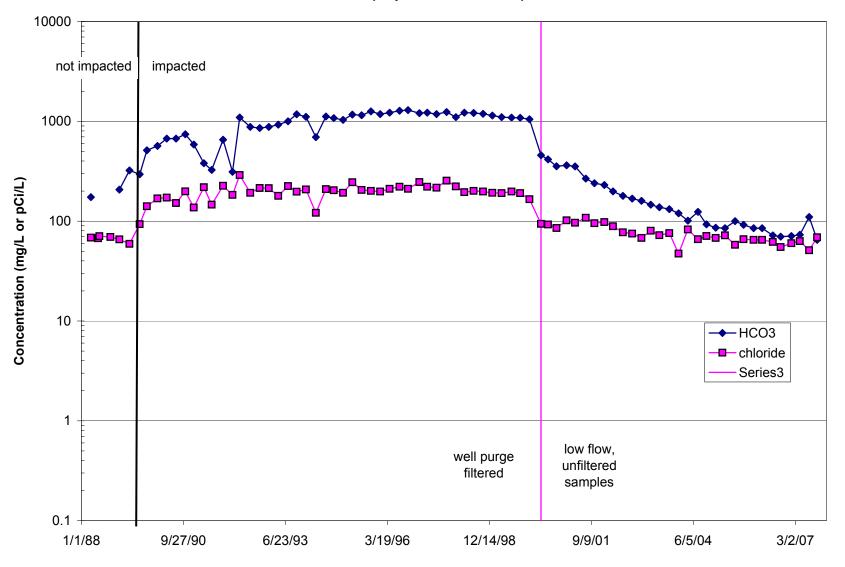
N.A. Water Systems, LLC Airside Business Park, 250 Airside Drive Moon Township, Pennsylvania 15108-2793 USA Tel: 412-809-6000 • Fax: 412-809-6075 Web site: www.nawatersystems.com

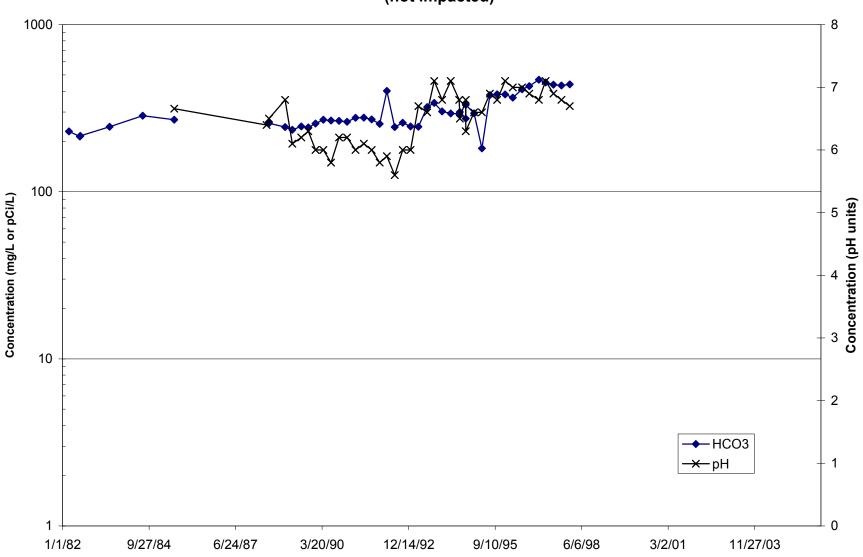






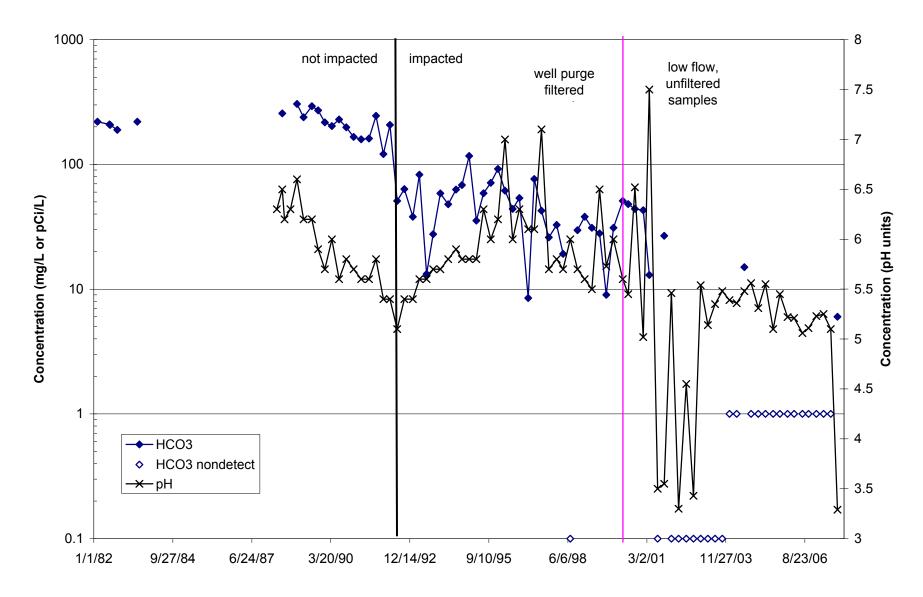


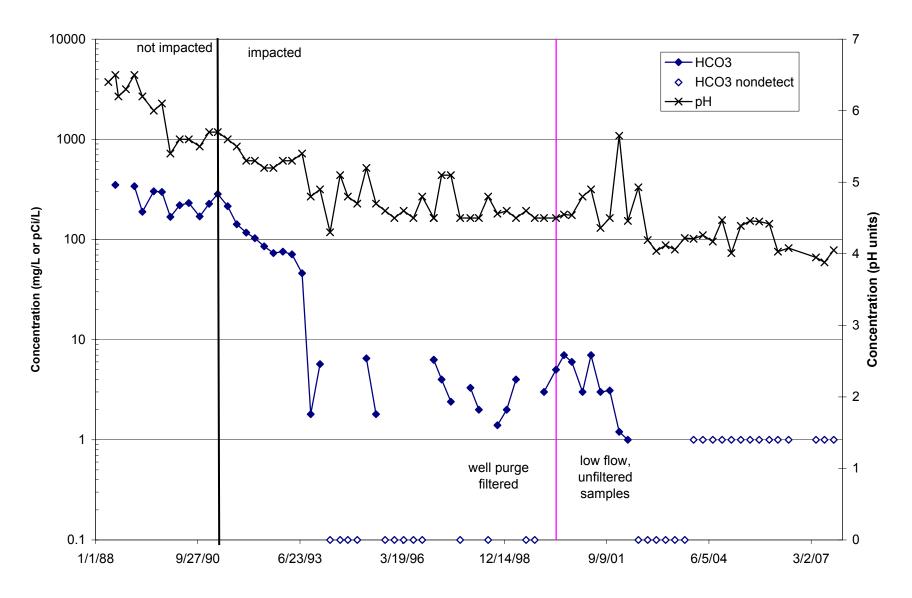


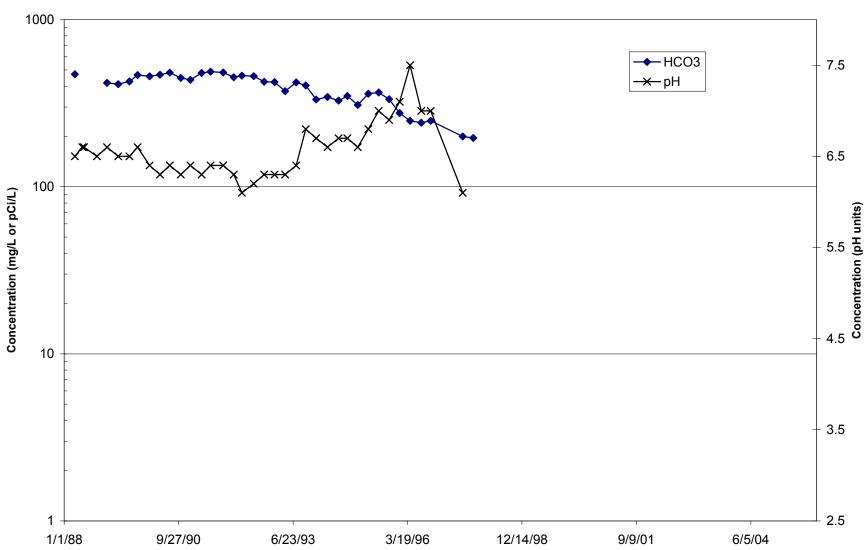


0411 (not impacted)

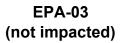


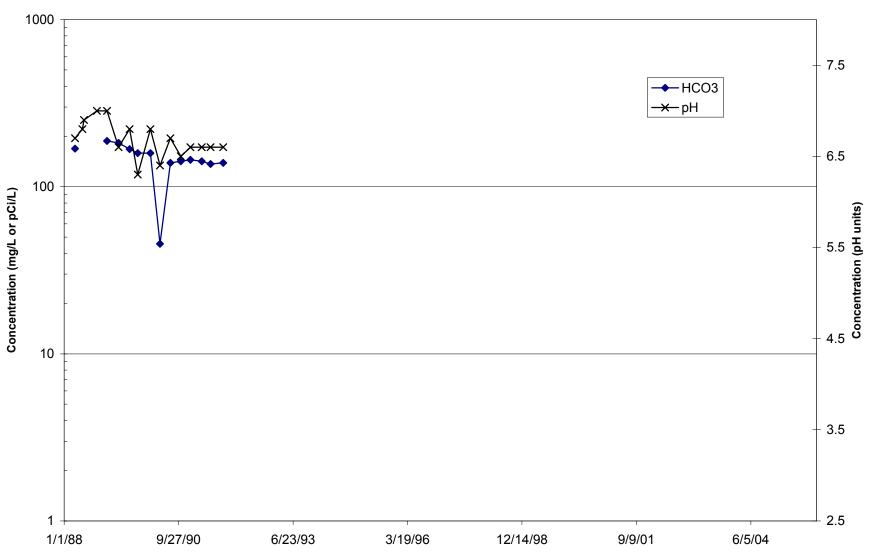




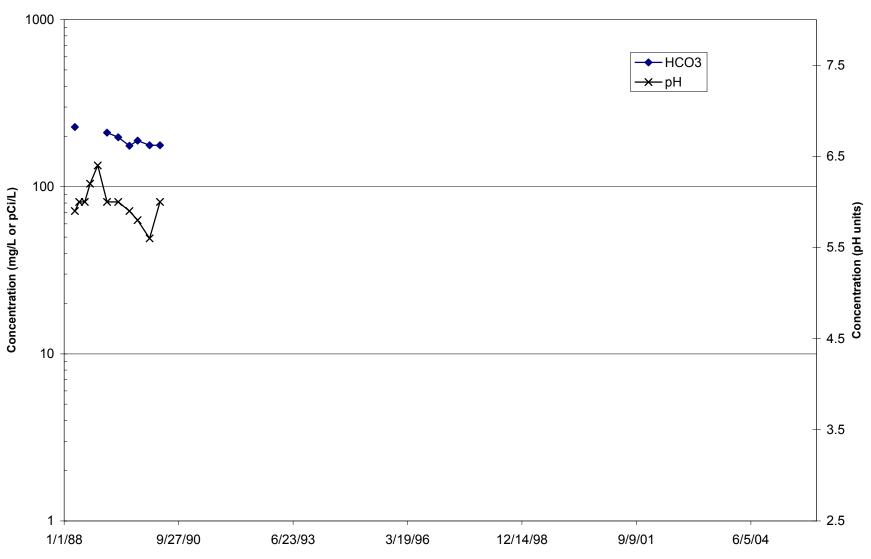


EPA-01 (not impacted)

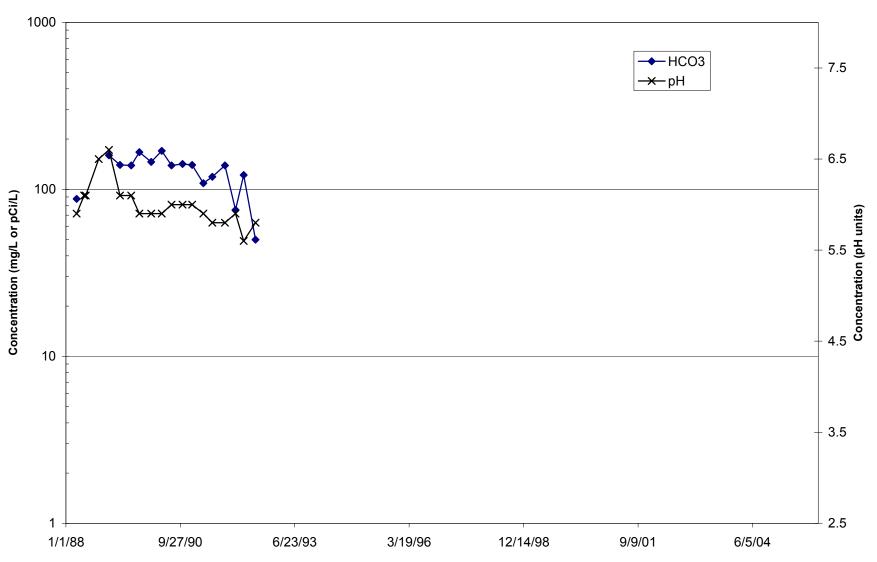




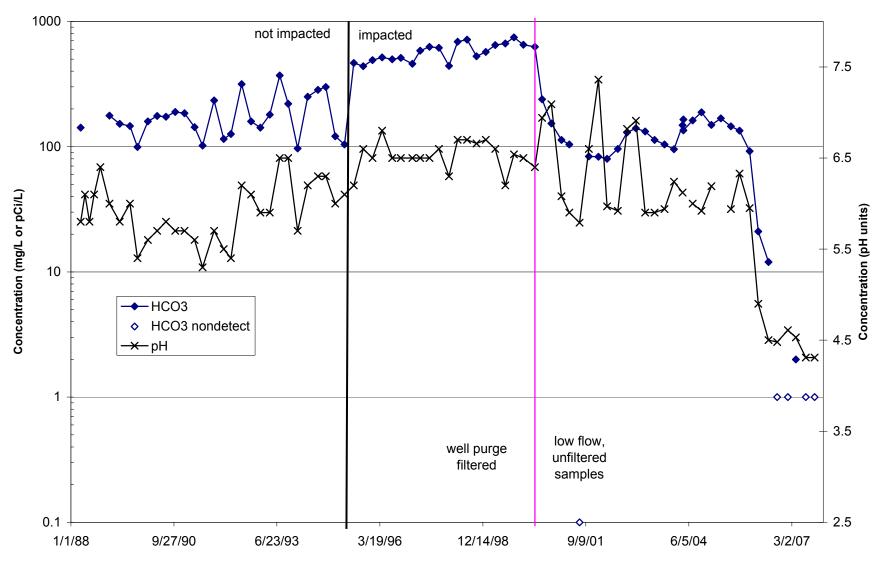




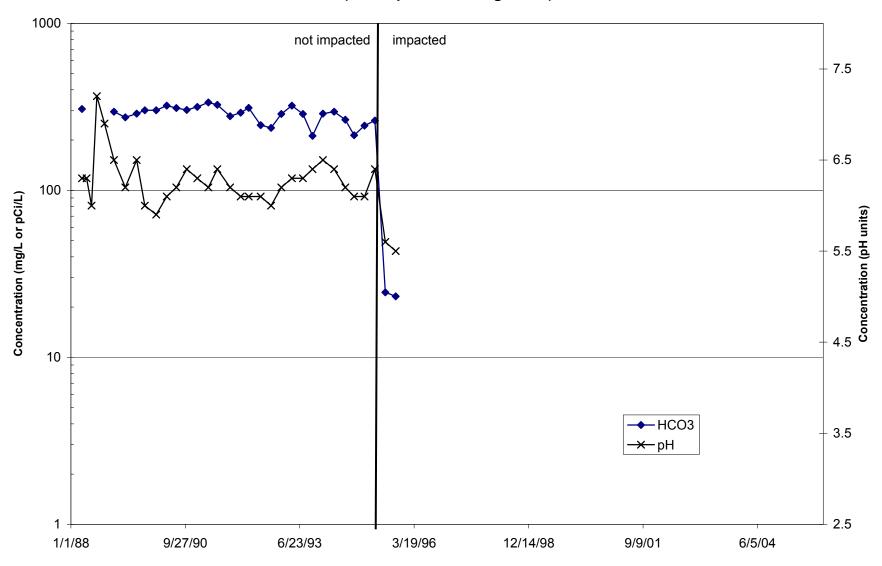




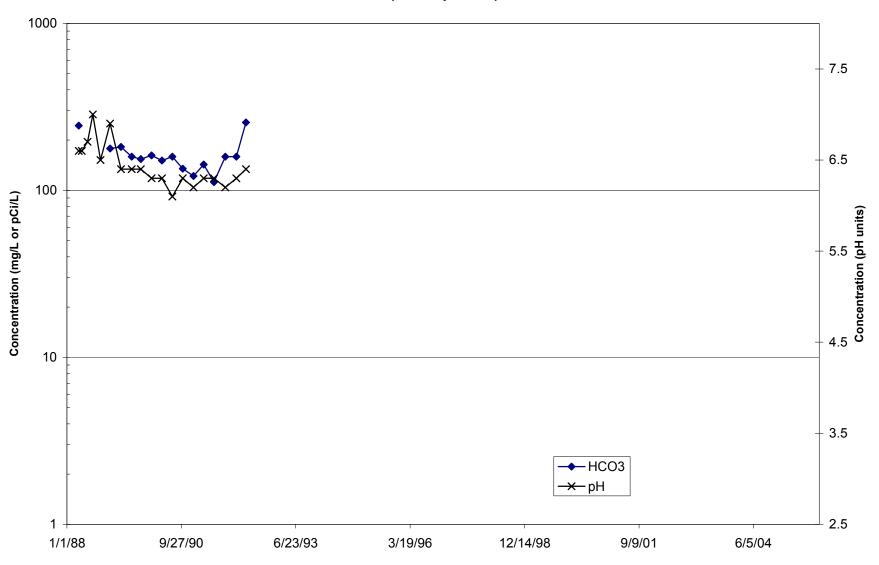
EPA-14 (background through 4/95, impacted after 5/2000)



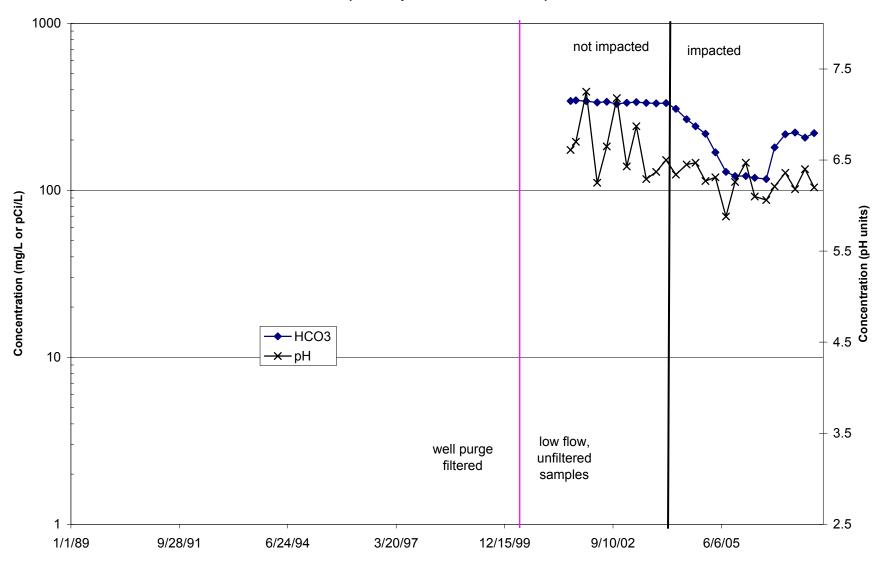
EPA-15 (not impacted through 4/95)







NBL-1 (not impacted until Jan 04)



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Appendix B Summary Tables of Output from ProUCL For UCL95 Statistics and Single Sample Hypothesis Tests



N.A. Water Systems, LLC Airside Business Park, 250 Airside Drive Moon Township, Pennsylvania 15108-2793 USA Tel: 412-809-6000 • Fax: 412-809-6075 Web site: www.nawatersystems.com

Nonparametric Background Statistics for Southwest Alluvium Data Sets with Non-Detects

	AI	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	v	CI	SO4	NO3_as_N	U	Chloroform	ab TDS	Rad-226	Rad-228	Rad totl	Th-230	Pb-210	Gross Alpha
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ua/L	ma/L	pCi/L	pCi/L	pCi/L	pCi/L	pCi/L	pCi/L
Total Number of Data	391	391	389	391	391	388	389	391	391	390	391	391	391	0	390		390		391	391	391	391	391
Number of Non-Detect Data	370	364		379	319	386	46	389	377	197	391	0	0	5	1	391	0	134	265		359	306	276
Number of Detected Data (or Distinct Obs. If	0.0			0.0	0.0				0						· · ·		•						2.0
zero nondetect)	21	27	0	12	72	2	343	2	14	193	0	262	306	386	389	0	307	257	126	292	32	85	115
Minimum Detected	0.1	0.001	N/A	0.006	0.01	0.05	0.01	0.03	0.05	0.001	N/A	9.8	605		0.001		1310	0.2	1	0.2	0.2	1	0.4
Maximum Detected	0.6	0.01	N/A	0.07	0.06	0.07	3.35	0.03	0.17	0.195	N/A	169	5830		0.367	N/A	10530	9.4	7	12	14.3	14.2	17.8
Percent Non-Detects	94.6%	93.1%	100.0%	96.9%	81.6%	99.5%	11.83%	99.5%	96.4%	50.51%	100.0%	0.0%	0.0%		0.26%	100.0%	0.0%	34.27%	67.77%		91.8%	78.3%	70.6%
Minimum Non-detect	0.1	0.001	1001070	0.005	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.01	0.0003	0.001	N/A	0.02	1	0.2	0.2	1	1
Maximum Non-detect	0.1	0.001		0.01	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.1	0.0003	1	N/A	0.2	1	0.2	0.2	1	1
Mean of Detected Data	0.182	0.00237		0.0255	0.0186	0.06	0.339	0	0.08	0.00708	0	74.82	2401		0.0419		4630	0.979	2.55	-	2.841	2.845	3.35
Median of Detected Data	0.14	0.001		0.01	0.02	0.06	0.13		0.08	0.003		67.8	2420		0.031		4795	0.6	2.2		1.6	2.2	2.1
SD of Detected Data	0.123	0.00259		0.0242	0.0104	0.0141	0.536		0.0291	0.0168		40.39	808		0.0448		1377	1.149	1.369		2.912	2.07	2.927
Mean of Log-Transformed Detected Data	-1.848	-6.402		-4.04	-4.106	-2.827	-1.973		-2.572	-5.705		4.115	7.72		-3.431		8.386	-0.354	0.813	0.23	0.58	0.871	0.911
SD of Log-Transformed Detected Data	0.51	0.757		0.868	0.481	0.238	1.392		0.299	1.067		0.705	0.385		0.659		0.358	0.73	0.487	0.931	1.02	0.558	0.757
Discernable Distribution (0.05)	none	none	N/A	none	none	none	none	N/A	none	none	N/A	none	none			N/A	none	none	none	none	Gamma	Lognormal	none
Skewness of Detected Data	2.38	2.06		1.24	1.99	N/A	2.931	N/A	2.45	8.339	N/A	0.14	1.04		4.177		0.40	3.919	1.342		2.87	2.21	2.09
Kaplan-Meier (KM) Method														<u> </u>									t
Mean	0.104	0.00109	<u> </u>	0.0066	0.0116	0.0501	0.3		0.0511	0.00401		74.82	2401	98.27	0.0418		4630	0.712	1.499	1.47	0.416	1.401	1.494
SD	0.0334	7.53E-04		0.00527	0.00554	0.00101	0.514		0.00769	0.0122		40.34	806.9	123.6	0.0448		1375	1	1.06		1.094	1.225	1.99
Standard Error of Mean	0.00173	3.88E-05	<u> </u>	2.78E-04	2.82E-04	7.28E-05	0.0261		4.04E-04	6.18E-04		2.043	40.86		0.00227		69.71	0.0507	0.0538		0.0562	0.0623	0.119
95% KM (t) UCL	0.107	0.00116	<u> </u>	0.00706	0.0121	0.0502	0.343		0.0517	0.00503		78.18	2468		0.0455		4744	0.796	1.588	1.612	0.509	1.504	1.689
95% KM (z) UCL	0.107	0.00116		0.00706	0.0121	0.0502	0.343		0.0517	0.00503		78.18	2468		0.0455		4744	0.795	1.588		0.509	1.503	1.689
95% KM (BCA) UCL	0.107	0.00117		0.0109	0.012	N/A	0.347		0.0706	0.00516		77.94	2465		0.0459		4750	0.798	1.624	1.621	0.812	1.532	1.687
95% KM (Percentile Bootstrap) UCL	0.107	0.00116		0.0108	0.0121	N/A	0.343		0.0613	0.00511		78.25	2470		0.0456		4740	0.797	1.611	1.61	0.71	1.513	1.693
95% KM (Chebyshev) UCL	0.112	0.00126		0.00781	0.0128	0.0504	0.414		0.0528	0.0067		83.72	2579		0.0517		4933	0.933	1.734		0.661	1.673	2.011
97.5% KM (Chebyshev) UCL	0.115	0.00134		0.00834	0.0133	0.0505	0.463		0.0536	0.00787		87.57	2656		0.056		5065	1.029	1.836	2.01	0.767	1.79	2.234
99% KM (Chebyshev) UCL	0.122	0.00148		0.00937	0.0144	0.0508	0.56		0.0551	0.0102		95.14	2807		0.0644		5323	1.216	2.035		0.975	2.021	2.674
Datasets without Nondetects																							
Student's-t UCL												78.18	2468				4744						
95% UCLs (Adjusted for Skewness)																							
95% Adjusted-CLT UCL												78.19	2470				4746						
95% Modified-t UCL												78.19	2468				4745						
Non-Parametric UCLs																							
95% CLT UCL												78.18	2468				4744						
95% Jackknife UCL												78.18	2468				4744						
95% Standard Bootstrap UCL												78.21	2468				4743						
95% Bootstrap-t UCL												78.3	2474				4743						
95% Hall's Bootstrap UCL												78.15	2470				4743						
95% Percentile Bootstrap UCL	1						1					78.25	2467				4739						
95% BCA Bootstrap UCL	1						1					78.23	2469				4751						
95% Chebyshev(Mean, Sd) UCL												83.72	2579				4933						
97.5% Chebyshev(Mean, Sd) UCL	1						1					87.57	2656				5065						
99% Chebyshev(Mean, Sd) UCL												95.14	2807				5323						
Potential UCL to Use																							
95% KM (t) UCL	0.107	0.00116		0.00706	0.0121	0.0502			0.0517										1.588		0.509	1.504	1.689
95% KM (z) UCL							İ							T T									
95% KM (BCA) UCL										0.00516					0.0459			0.798		1.621			
95% KM (Percentile Bootstrap) UCL	0.107	0.00116		0.0108	0.0121		İ		0.0613					T T					1.611			1.513	1.693
95% KM (Chebyshev) UCL							0.414																
97.5% KM (Chebyshev) UCL							İ							137.4									
99% KM (Chebyshev) ÚCL							İ							T T									
95% Student's-t UCL							1						2468				4744						
95% Modified-t UCL							1						2468				4745						
95% Chebyshev(Mean, Sd) UCL												83.72											
Notes			4	1		2		3			4			1		4		1					
			· · · · · · · · · · · · · · · · · · ·	-		_		-			-												

Note:

1. Data have multiple DLs - Use of KM Method is recommended

2. There may not be adequate detected values to compute meaningful and reliable test statistics and estimates

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

3. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values

determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).

4. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs!

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

	Sb	Ва	Cr	Cu	Fe	Hg	Ag	TI	Zn
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
Total Number of Data	0	26	37	13	Ŭ	8	21	0	25
Number of Non-Detect Data		26	36	11	15	8	21		10
Number of Detected Data (or Distinct Obs. If									
zero nondetect)		0	1	2	4	0	0		15
Minimum Detected					0.06				0.02
Maximum Detected					1.4				0.429
Percent Non-Detects		100.00%	97.30%	84.60%	78.95%	100.00%	100.00%		40.00%
Minimum Non-detect					0.05				0.01
Maximum Non-detect					0.1				0.1
Mean of Detected Data					0.418				0.0891
Median of Detected Data					0.105				0.05
SD of Detected Data					0.656				0.109
Mean of Log-Transformed Detected Data					-1.761				-2.889
SD of Log-Transformed Detected Data					1.434				0.936
Discernable Distribution (0.05)					Gamma				Gamma
Skewness of Detected Data					1.988				2.519
Kaplan-Meier (KM) Method for Datasets with									
Nondetects									
Mean					0.138				0.0636
SD					0.138				0.0030
Standard Error of Mean					0.298				0.0879
95% KM (t) UCL					0.079				0.0183
95% KM (z) UCL					0.268				0.0949
95% KM (BCA) UCL					1.4				0.0937
95% KM (Percentile Bootstrap) UCL					0.342				0.0951
95% KM (Chebyshev) UCL					0.482				0.0331
97.5% KM (Chebyshev) UCL					0.631				0.143
99% KM (Chebyshev) UCL					0.001				24.60%
					0.524				24.0070
Datasets without Nondetects									
Student's-t UCL									
95% UCLs (Adjusted for Skewness)									
95% Adjusted-CLT UCL									
95% Modified-t UCL									
Non-Parametric UCLs									
95% CLT UCL									
95% Jackknife UCL									
95% Standard Bootstrap UCL									
95% Bootstrap-t UCL									
95% Hall's Bootstrap UCL									
95% Percentile Bootstrap UCL									
95% BCA Bootstrap UCL									
95% Chebyshev(Mean, Sd) UCL									
97.5% Chebyshev(Mean, Sd) UCL									
99% Chebyshev(Mean, Sd) UCL									
Potential UCL to Use									0 00 10
95% KM (t) UCL 95% KM (z) UCL					0.275				0.0949
95% KM (BCA) UCL		i					i	i	i
95% KM (Percentile Bootstrap) UCL 95% KM (Chebyshev) UCL									
97.5% KM (Chebyshev) UCL 99% KM (Chebyshev) UCL									
95% Student's-t UCL									
95% Modified-t UCL									
Notes		5	4	4	1, 3	5	5		1
110100		3	4	4	1, 5	5	5		1

Note:

1. Data have multiple DLs - Use of KM Method is recommended

2. UCL recommendation based on guidance (table 16). None provided by software.

3. There may not be adequate detected values to compute meaningful and reliable test statistics and estimates

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

4. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).

5. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs!

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

6. Insufficent Number of Observations to produce Meaningful Statistics.

Comparisons of Background Sample Distributions to Comparison Values for Contaminants of Concern in the Southwest Alluvium Single Sample Sign Test

		onigie o	Sample Sig	11 1631																										
	AI	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	V	CI	SO4	NO3_as_N	U	Chloroform	Lab_TDS		Th-230	Pb-210	iross_Alph	Sb	Ва	Cr	Cu	Fe	Hg	Ag	TI	Zn
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ug/L	mg/L	pCi/L	pCi/L	pCi/L	pCi/L		mg/L	mg/L	mg/L	mg/L	mg/L	mg/L		mg/L
Raw Statistics																														
Number of Valid Data	391	391	389	391	391	388	389	391	391	390	391		391	391	390	391	390	391	391	391			26	37	13	19	8	21		25
Number of Distinct Data	8	6	3	7	5	2	90	2	6	28	1	262	306	279	189	3	307	61	24	38	59		3	4	2	6	2	2		12
Number of Non-Detect Data	370	364	389	379	319		46	389	377	197	391	0	0.00%	5	1	391	0	99	359	306	276		26	36	11	15	8	21		10
Number of Detected Data	21	27	0	12	72	2	343	2	14	193	0	391	391	386	389	-	390	292	32	85	115		0	1	2	4	0	0		15
Percent Non-Detects	94.63%	93.09%	100.00%	96.93%	81.59%	99.48%	11.83%	99.49%	96.42%	50.51%	100.00%	0.00%	0.00%	1.28%	0.26%	100.00%	0.00%	25.32%	91.82%	78.26%	70.59%		100%	97.30%	84.62%	78.95%	100.00%	100.00%		40.00%
Minimum Non-detect	0.1	0.001	0.01	0.005	0.01		0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.01	0.0003	0.001	N/A	0.2	0.2	1	1		0.001	0.001	0.01	0.05	0.0002	0.01		0.01
Maximum Non-detect	0.1	0.001	0.1	0.01	0.01		0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.1	0.0003	1	N/A	0.2	÷.=	1	1		0.1	0.05	0.02	0.1	0.001	0.05		0.1
Minimum Detected	0.1	0.001	N/A	0.006	0.01	0.05	0.01	0.03	0.05	0.001	N/A	9.8	605	0.09	0.001	N/A	1310	0.2	0.2	1	0.4		N/A	0.029	0.01	0.06	N/A	N/A		0.02
Maximum Detected	0.6	0.01	N/A	0.07	0.06		3.35	0.03	0.17	0.195		169	5830	1225	0.367		10530	12		14.2	17.8		N/A	0.029	0.01	1.4	14/73	N/A		0.429
Mean of Detected Data	0.182	0.00237	N/A	0.0255	0.0186		0.339	0.03	0.08	0.00708	-	74.82	2401	99.54	0.0419		4630	1.9	2.841	2.845	3.35		N/A	0.029	0.01	0.418		N/A		0.0891
Median of Detected Data	0.14	0.001	N/A	0.01	0.02		0.13	0.03	0.08	0.003	-	67.8	2420	74.1	0.031		4795	1.3	1.0	2.2	2.1		N/A	0.029	0.01	0.105		N/A		0.05
SD of Detected Data	0.123	0.00259	N/A	0.0242	0.0104	0.0141	0.536	0	0.0291	1.68%	N/A	40.39	808	124.1	0.0448	N/A	1377	1.783	2.912	2.07	2.927		N/A	N/A	0	0.656	N/A	N/A		0.109
Number Above Limit	0	0	0	12	2	1	144	0	0	2	0	0	292	322	197	0	351	19	6	82	1		0	0	0	1	0	0		0
Number Equal Limit	0	1	0	36	0	387	3	0	0	0	391	0	0	1	3	0	0	0	0	309	0		0	26	0	0	0	17		0
Number Below Limit	391	390	0	0	389	0	242	391	391	388	0	391	99	68	190	391	39	372	385	0	390		26	11	13	18	8	4		25
Number Observations Discarded			389	379		387					391	0	0				0			309				26				17		
H0: Site Median >= Comparison Value																														
(Form 2)	5	0.01	0.004	0.005	0.05	0.05	0.2	1	0.2	0.05	0.1	250	2125	30	0.03	80	3170	5	5	1	15	0.006	2	0.05	1	1	0.002	0.05		10
Test Value	-19.77	-19.75	0	12	-19.57	· 1	-4.988	-19.77	-19.77	-19.55	0	-19.77	9.76	12.86	0.356	-19.77	15.8	-17.85	-19.17	9.055	-19.67		0	0	0	1	0	0		0
Lower Critical Value (0.05)	-1.645	-1.645	-1	2	-1.645	-1	-1.645	-1.645	-1.645	-1.645	-1	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645		8	2	3	5	1	-1		7
P-Value	0	0	1	1	0	1	3.05E-07	0	0	0	1	0	1	1	0.639	0	1	0	0	1	0		1.49E-08	4.88E-04	1.22E-04	3.81E-05	0.00391	0.0625		2.98E-08
L																														
Obsevations below max detection limit =																														
non-detects ¹				0.01		_		0.1						0.1							1			0.05	0.02	0.1				0.1
Conclusion with Alpha = 0.05	Reject	Reject	Do not Reject	Do not Reject	Reject	Do not Reject	Reject	Reject	Reject	Reject	Do not Reject	Reject	Do not Reject	Do not Reject	Do not Reject	Reject	Do not Reject	Reject	Reject	Do not Reject	Reject	no data	Reject	Reject	Reject	Reject	Reject	Do not Reject	no data	Reject
1				I																										

Values lower than listed maximum nondetect value are treated by ProUCL as nondetects in single sample hypotheses tests
 All detection limits equal or higher than ARAR, all data rejected
 Nondetects equal or higher than ARAR rejected
 All values nondetects
 No data

6. Erroneous conclusion by ProUCL

Comparisons of Background Sample Distributions to Comparison Values for Contaminants of Concern in the Southwest Alluvium Single Sample Wilcoxon Signed Rank Test

	Single	Sample v	Vilcoxon Sig																											
	AI	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	V	CI	SO4	NO3_as_N	U	Chloroform	Lab_TDS	Rad_totl	Th-230	Pb-210	iross_Alph	Sb	Ва	Cr	Cu	Fe	Hg	Ag	TI	Zn
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ug/L	mg/L	pCi/L	pCi/L	pCi/L	pCi/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L		mg/L
Raw Statistics																														
Number of Valid Data	391	391	389	391	391	388	389	391	391	390	391	391	391	391	390	391	390) 391	391	391	391	0	26	37	' 13	19	8	21	0	25
Number of Distinct Data	8	6	3	7	5	2	90	2	6	28	1	262	306	279	189	3	307	61					3	4	2	6	2	2 2		12
Number of Non-Detect Data	370	364	389	379	319	386	46	389	377	197	391	0	0	5	1	391	C) 99	359	306	6 276		26	36	5 11	15	8	21		10
Number of Detected Data	21	27	0	12	72	2	343	2	14	193	0	391	391	386	389	0	390	292	32	85	5 115		0	1	2	4	0	0		15
Percent Non-Detects	94.63%	93.09%	100.00%	96.93%	81.59%	99.48%	11.83%	99.49%	96.42%	50.51%	100.00%	0.00%	0.00%	1.28%	0.26%	100.00%	0.00%	25.32%	91.82%	78.26%	70.59%		100.00%	97.30%	84.62%	78.95%	100.00%	100.00%		40.00%
Minimum Non-detect	0.1	0.001	0.01	0.005	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.01	0.0003	0.001	N/A	0.2	. 0.2	1	1		0.001	0.001	0.01	0.05	0.0002	0.01		0.01
Maximum Non-detect	0.1	0.001	0.1	0.01	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.1	0.0003	1	N/A	0.2	. 0.2	1	1		0.1	0.05	0.02	0.1	0.001	0.05		0.1
Minimum Detected	0.1	0.001	N/A	0.006	0.01	0.05	0.01	0.03	0.05	0.001	N/A	9.8	605	0.09	0.001	N/A	1310	0.2	0.2	: 1	0.4		N/A	0.029	0.01	0.06	N/A	N/A		0.02
Maximum Detected	0.6	0.01	N/A	0.07	0.06	0.07	3.35	0.03	0.17	0.195	N/A	169	5830	1225	0.367	N/A	10530) 12	14.3	14.2	2 17.8		N/A	0.029	0.01	1.4	N/A	N/A		0.429
Mean of Detected Data	0.182	0.00237	N/A	0.0255	0.0186	0.06	0.339	0.03	0.08	0.00708	N/A	74.82	2401	99.54	0.0419	N/A	4630) 1.9	2.841	2.845	5 3.35		N/A	0.029	0.01	0.418	N/A	N/A		0.0891
Median of Detected Data	0.14	0.001	N/A	0.01	0.02	0.06	0.13	0.03	0.08	0.003	N/A	67.8	2420	74.1	0.031	N/A	4795	5 1.3	1.6	2.2	2.1		N/A	0.029	0.01	0.105	N/A	N/A		0.05
SD of Detected Data	12.30%	0.26%	N/A	0.0242	0.0104	0.0141	0.536	0	0.0291	0.0168	N/A	40.39	808	124.1	0.0448	N/A	1377	1.783	2.912	2.07	2.927		N/A	N/A	0	0.656	N/A	N/A		0.109
Number Above Limit	0	0	389	11	2	1	144	0	0	2	0	0	292	322	197	0	351	19	6	82	2 1		0	0	0 0	1	0	0 0		0
Number Equal Limit	0	1	0	343	0	1	3	0	0	0	0	0	0	1	3	0	0 0) 0	0	3	3 0		0	0	0 0	0	0	0		0
Number Below Limit	391	390	0	37	389	386	242	391	391	388	391	391	99	68	190	391	39	372	385	306	390		26	37	' 13	18	8	21		25
T-plus	0	0	75855	473	3	1	32515	0	0	458	0	0	56328	67525	45502	0	70816	6 1241	424.5	23905	5 2		0	0	0 0	1	0	0 0		0
T-minus	76636	76245	0	703	76633	75077	42177	76636	76636	75787	76636	76636	20309	8721	29577	76636	5429	75396	76212	51561	76634		351	703	91	189	36	231		325
H0: Site Median >= Comparison Value																														
(Form 2)	5	0.01	0.004	0.005	0.05	0.05	0.2	1	0.2	0.05	0.1	250	2125	30	0.03	80	3170	5	5	1	15	0.006	2	0.05	1	1	0.002	0.05	0.002	10
· ·																														
Large Sample z-Test Value	-19.29	-19.26	18.47	-409.8	-18.43	-19.75	-2.468	-19.72	-19.44	-17.19	-19.77	-17.14	8.054	13.11	3.351	-18.98	14.68	-16.62	-18.86	-6.968	-17.94		-4.751	-5.541	0	1	0	-4.272		-4.38
Critical Value (0.05)	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645		-1.645	-1.645	5 21	53	5	-1.645		-1.645
P-Value	0	0	1	0	0	0	0.00679	0	0	0	0	0	1	1	1	0) 1	0	0	1.60E-12	2 0		1.01E-06	1.51E-08	3 1.00E-04	0	0.0039	9.67E-06		5.95E-06
·																					1				1					
Obsevations below max detection limit =																														
non-detects ¹			0.1	0.01				0.1						0.1							1			0.05	0.02	0.1				0.1
	1		Do not	1									Do not	Do not	Do not		Do not				1				1 1			1		
Conclusion with Alpha = 0.05	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	no data	Reject	Reject	Reject	Reject	Reject	Reject	no data	Reject
• • • • • • •																														
Notes			1,2,4	1, 2, 3, 6		3		1			3, 4			1		4				3	1	5	4	1, 3	1	1	2, 4	3, 4	5	1
				, , -, -							- /							+					•		•					

Values lower than listed maximum nondetect value are treated by ProUCL as nondetects in single sample hypotheses tests
 Detection limit higher than ARAR

Detection limit rights that and the
 Detection limit equals ARAR
 All data nondetects

5. No data
 6. Erroneous conclusion by ProUCL

Nonparametric Background Statistics for Zone 1 Data Sets with Non-Detects

		A -	De l	04	0.	Dh	Ma	Ma	NI:	0.			004 N	02 as N	<u>г п</u> т	<u> </u>		Ded 000	Ded 000	Ded 4a4	Th 020	Dh 040	Overes Alasha
Concentration Units	AI mg/L	As mg/L	Be mg/L	Cd mg/L	Co mg/L	Pb mg/L	Mn mg/L	Mo mg/L	Ni mg/L	Se mg/L	V CI mg/L mg/l		SO4 N mg/L	O3_as_N mg/L	U mg/L	ug/L	mg/L	pCi/L	pCi/L	Rad_totl pCi/L	pCi/L	pCi/L	Gross_Alpha pCi/L
Total Number of Data	234	234	234	234	111g/L 234	234	234	111g/L 234	230	U		- 234	234	233			234	233			234	234	234
Number of Non-Detect Data	203			234	234	234	204	234	230		234	234	234	<u></u> 167		234	234	233	234		234	189	82
Number of Detected Data (or Distinct Obs. If	203	190	234	231	210	233		229	221	224	234	0	0	107		200	0	4	70	2	215	109	02
· ·	31	20		2	24	1	222	F	2	10		156	107	66	194	1	196	220	164	222	19	45	150
zero nondetect)	0.1	38 0.001		0.005	24	1	233 0.66	0.03	0.00	0.001		156 9.4	187	66	-	0.91	186 2490	229 0.2	164	232 0.2	0.2	45	152
Minimum Detected	0.1	0.001		0.005	0.01		4.15	0.03	0.06	0.001		9.4 252	1410 3882	0.01			2490 5610	5.4			0.2 4.9	1.1	0.9 14
Maximum Detected			400.00/			00.00/						252				0.91	0100	-				9.1	
Percent Non-Detects	86.75%	83.76%	100.0%	98.72%	89.74%	99.6%	0.43%	97.86%	98.70%	95.73%	100.0%	•	0	71.67%		99.57%	0	1.72%	29.91%		91.88%	80.77%	35.04%
Minimum Non-detect	0.1	0.001	0.01	0.005	0.01	0.05	0.01	0.01	0.05	0.001	0.1 NA		NA I	0.01			NA	0.2	1	0.2	0.02	1	1
Maximum Non-detect	0.1	0.001	0.05	0.01	0.01	0.05	0.01	0.1	0.05	0.001	0.1 NA		NA ataa	0.1		1	NA	0.2	1	0.2	0.2	1	1
Mean of Detected Data	0.185	0.00174		0.00733	0.0171		2.434	0.12	0.0667	0.0019	-	7.13	2703	1.767			4225	1.269			0.974	2.58	2.757
Median of Detected Data	0.14	0.002		0.007	0.01		2.65	0.13	0.07	0.0015		37.9	2952	0.16			4569	1.2			0.7	2.1	2
SD of Detected Data	0.107	7.60E-04		0.00252	0.0112		0.814	0.099	0.00577	0.0011		5.53	649.3	7.961	0.0701		868.8	0.612			1.114	1.594	2.092
Mean of Log-Transformed Detected Data	-1.804	-6.446		-4.955	-4.208		0.824	-2.466	-2.711	-6.411		567	7.868	-1.745			8.324	0.132	1.111	1.044	-0.388	0.822	0.82
SD of Log-Transformed Detected Data	0.458	0.429	,	0.347	0.497		0.38	0.992	0.089			286	0.275	1.586			0.23	0.481	0.501	0.76	0.801	0.472	0.579
Discernable Distribution (0.05)	none	none	۲ ا		none		none		none	none	none			one	none		none		Gamma		ognormal		none
Skewness of Detected Data	2.288	0.882		0.586	2.65		-0.224	0.827	-1.732	0.863	9.4	472	-0.673	5.661	13.74		-0.723	2.292	2.015	1.103	2.897	2.492	2.318
Kaplan-Meier (KM) Method for Datasets with																							
Nondetects																							
Mean	0.111	0.00112		0.00504	0.0107		2.426	0.0319	0.0601	0.00104				0.524				1.25			0.263	1.385	2.106
SD	0.0479	4.06E-04		3.82E-04	0.00412		0.819		9.28E-04					4.277	0.0638		ļ	0.621	1.976	2.338	0.374	0.904	1.9
Standard Error of Mean	0.00318	2.69E-05		3.59E-05	2.75E-04		0.0537		7.50E-05					0.282				0.0408	0.13		0.0251	0.0598	0.125
95% KM (t) UCL	0.116	0.00116		0.0051	0.0112		2.515	0.0341	0.0602	0.00107				0.99				1.318			0.304	1.483	2.312
95% KM (z) UCL	0.116	0.00116		0.0051	0.0112		2.514		6.02E-02					0.988	0.0141			1.317	2.935		0.304	1.483	2.311
95% KM (BCA) UCL	0.118	0.00116		N/A	0.0112		2.519	0.142	N/A	0.00107				1.129				1.314	2.962		0.462	1.622	2.361
95% KM (Percentile Bootstrap) UCL	0.117	0.00117		N/A	0.0112		2.512	0.132	N/A	0.00107				1.043	0.0155			1.32	2.946	3.849	0.403	1.579	2.321
95% KM (Chebyshev) UCL	0.125	0.00124		0.0052	0.0119		2.66	0.0378	0.0604	0.00112				1.754	0.0255			1.428	3.287	4.257	0.372	1.645	2.649
97.5% KM (Chebyshev) UCL	0.131	0.00129		0.00527	0.0124		2.761	0.0403	0.0606	0.00116				2.287	0.0334			1.505	3.531	4.546	0.42	1.758	2.884
99% KM (Chebyshev) UCL	0.143	0.00139		0.0054	0.0135		2.96	0.0453	0.0608	0.00123				3.333	0.049			1.656	4.011	5.113	0.513	1.979	3.346
Datasets without Nondetects																							
Student's-t UCL											39	.92	2773				4319						
95% UCLs (Adjusted for Skewness)	-	-								+ +	30	0.32	2113		ł ł		4313					+	
95% Adjusted-CLT UCL	-	-								+ +	20	9.62	2770		ł ł		4316					+	
95% Modified-t UCL												.02 .03	2772				4318						
Non-Parametric UCLs													2112				4310					 	
95% CLT UCL											39	3.91	2772				4319					 	
95% Jackknife UCL												3.92	2773				4319					 	
95% Standard Bootstrap UCL												3.84	2773				4318					 	
95% Bootstrap-t UCL).09	2771				4317					 	
95% Hall's Bootstrap UCL											-	7.29	2771				4318						
95% Percentile Bootstrap UCL												.23	2772				4319					 	
95% BCA Bootstrap UCL	1		├							+ +		9.23	2772		+		4319			├		 	
95% Chebyshev(Mean, Sd) UCL	-		<u>├</u>							+ +			2888		 		4317			├		 	
97.5% Chebyshev(Mean, Sd) UCL	+		<u>├</u>							+ +		.84 3.88	2000		┟──┼		4473			├		 	
99% Chebyshev(Mean, Sd) UCL												7.88	3125				4790						
			├							+ +	47	.00	5120		<u> </u>		4190			├		 	
Potential UCL to Use	-		<u>├</u>							+ +					 					├		 	
95% KM (t) UCL	0.116	0.00116	<u>├</u>	0.0051	0.0112			0.0341	0.0602	0.00107					 					├	0.304	1.483	
95% KM (z) UCL	0.110	0.00110		0.0031	0.0112			0.0341	0.0002	0.00107											0.004	1.403	
95% KM (BCA) UCL							2.519											1.314		3.841			2.361
95% KM (Percentile Bootstrap) UCL	0.117	0.00117			0.0112		2.515	0.132		0.00107								1.514	2.946		0.403	1.579	2.301
95% KM (Chebyshev) UCL	0.117	0.00117	<u>├</u>		0.0112			J. 132		0.00107				1.754	0.0255				2.340	├	0.403	1.019	
97.5% KM (Chebyshev) UCL	+		<u>├</u>							╂───┼				1.734	0.0200					<u>├</u>		 	
	+		<u>├</u> ───┤							+					<u>├</u> ───┤					├			
99% KM (Chebyshev) UCL 95% Student's-t UCL			<u>├</u>					├		┼───┼		.92	2773		<u>├</u> ───┤		4319			├		ł	
95% Student's-t UCL 95% Modified-t UCL			├ ────┼							╂───┼		-			├ ───┤		4319			├ ────┤		_	
55% Moamea-LUCL			├ ────┼							╂────╁	35	.03	2772		├ ───┤		4318			├		_	
			├ ────┼							╂────╁					├ ───┤					├		_	
	+														<u>├</u>							 	
Notes		3	5 1	3	3	4		1, 3	3	3	5	+		1	1	Δ			ļ			 	
	1	5	51	, •	5	-		., •	J	. .	v		1		· ·			I I	l	1 L		L	

Note:

1. Data have multiple DLs - Use of KM Method is recommended

2. UCL recommendation based on guidance (table 16). None provided by software.

3. There may not be adequate detected values to compute meaningful and reliable test statistics and estimates

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

4. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).

5. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs! The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV). Nonparametric Background Statistics for Zone 1 Trace Metal Data Sets with Non-Detects

	01	D.	0	0		11	A	T 1	7
Concentration Units	Sb mg/L	Ba mg/L	Cr mg/L	Cu mg/L	Fe mg/L	Hg ma/L	Ag mg/L	TI mg/L	Zn mg/L
Total Number of Data	mg/L 0	, ,	U		mg/∟ 12	mg/∟ 0	,	mg/∟ 0	
Number of Non-Detect Data	0	14 11	11 11	3	12	0	11 11	0	16 9
Number of Detected Data (or Distinct Obs. If zero		11		3	<u>'</u>				9
nondetect)		3	0	1	11		0		7
Minimum Detected		0.079	0	0.026	0.25		0		0.01
Maximum Detected		0.079		0.026	14				5
Percent Non-Detects		78.57%	100.00%	75.0%	8.33%		100.00%		56.25%
Minimum Non-detect		0.05	0.01	0.02	0.33 //		0.01		0.01
Maximum Non-detect		0.05	0.01	0.02	0.1		0.01		0.01
Mean of Detected Data		0.1	0.05	0.02	6.386		0.05		0.1
Median of Detected Data		0.084			6.2				0.784
SD of Detected Data		0.00603			5.381				1.862
Mean of Log-Transformed Detected Data SD of Log-Transformed Detected Data		-2.471			1.169				-2.488
		0.0709			1.495				2.093
Discernable Distribution (0.05)		Normal			Normal				Lognormal
Skewness of Detected Data		0.492			0.0576				2.626
Kanlan Malar (KM) Mathed for Detects with		ļ							
Kaplan-Meier (KM) Method for Datasets with									
Nondetects		0.0000			5.075				0.051
Mean		0.0833			5.875				0.351
SD		0.00492			5.197				1.203
Standard Error of Mean		0.00301			1.573				0.325
95% KM (t) UCL		0.0886			8.701				0.921
95% KM (z) UCL		0.0882			8.463				0.886
95% KM (BCA) UCL		0.091			8.418				0.995
95% KM (Percentile Bootstrap) UCL		0.091			8.346				0.968
95% KM (Chebyshev) UCL		0.0964			12.73				1.767
97.5% KM (Chebyshev) UCL		0.102			15.7				2.379
99% KM (Chebyshev) UCL		0.113			21.53				3.583
Datasets without Nondetects									
Student's-t UCL									ļ
95% UCLs (Adjusted for Skewness)									ļ
95% Adjusted-CLT UCL									
95% Modified-t UCL									
Non-Parametric UCLs									
95% CLT UCL									
95% Jackknife UCL									
95% Standard Bootstrap UCL									
95% Bootstrap-t UCL									
95% Hall's Bootstrap UCL									
95% Percentile Bootstrap UCL									L
95% BCA Bootstrap UCL									
95% Chebyshev(Mean, Sd) UCL									
97.5% Chebyshev(Mean, Sd) UCL									
99% Chebyshev(Mean, Sd) UCL									
Potential UCL to Use									
95% KM (t) UCL		0.0886			8.701				
95% KM (z) UCL									
95% KM (BCA) UCL									
95% KM (Percentile Bootstrap) UCL		0.091			8.367				
95% KM (Chebyshev) UCL									
97.5% KM (Chebyshev) UCL									
99% KM (Chebyshev) UCL									3.583
95% Student's-t UCL									
95% Modified-t UCL									
	1								
									L

Note:

1. Data have multiple DLs - Use of KM Method is recommended

2. UCL recommendation based on guidance (table 16). None provided by software.

3. There may not be adequate detected values to compute meaningful and reliable test statistics and estimates

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV). 4. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values

determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).

5. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs!

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV). 6. Insufficent Number of Observations to produce Meaningful Statistics.

Comparisons of Background Sample Distributions to Comparison Values for Contaminants of Concern in Zone 1 Single Sample Sign Test

		Single	Sample Sig	n lest																									
	AI	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	v	CI	SO4	NO3_as_N	U	Chloroform	Lab_TDS	Rad_totl		Pb-210 G	ross_Alpha	Sb	Ba	Cr	Cu	Fe	Hg	Ag Ti	Zn
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ug/L	mg/L	pCi/L	pCi/L	pCi/L	pCi/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L mg/L	mg/L
Raw Statistics																													
Number of Valid Data	234	234	234	234	234	234	234	234	230	234	234	234	234	233	233	234	234	234	234	234	234	0	14	11	4	12	0	11	0 16
Number of Distinct Data	15	4	- 2	3	4	1	152	6	3	4	1	156	187	35	56	3	186	72	13	27	49		5	2	2 2	12		2	9
Number of Non-Detect Data	203	196	234	231	210	233	1	229	227	224	234	0	0	167	39	233	0	2	215	189	82		11	11	3	1		11	9
Number of Detected Data	31	38	0	3	24	1	233	5	3	10	0	234	234	66	194	1	234	232	19	45	152		3	0) 1	11		0	7
Percent Non-Detects	86.75%	83.76%	100.00%	98.72%	89.74%	99.57%	0.43%	97.86%	98.70%	95.73%	100.00%	0.00%	0.00%	71.67%	16.74%	99.57%	0.00%	0.85%	91.88%	80.77%	35.04%		78.57%	100.00%	75.00%	8.33%		100.00%	56.25%
Minimum Non-detect	0.1	0.001	0.01	0.005	0.01	0.05	0.01	0.01	0.05	0.001	0.1	N/A	N/A	0.01	0.0003	0.5	N/A	0.2	0.02	1	1		0.05	0.01		0.1		0.01	0.01
Maximum Non-detect	0.1	0.001	0.05	0.01	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.1	0.0004	1	N/A	0.2	0.2	1	1		0.1	0.05	0.02	0.1		0.05	0.1
Minimum Detected	0.1	0.001	N/A	0.005	0.01	0.05	0.66	0.03	0.06	0.001	N/A	19.4	1410	0.01	0.0004	0.91	2490	0.2	0.2	1.1	0.9		0.079	N/A	0.026	0.25		N/A	0.01
Maximum Detected	0.6	0.004	N/A	0.01	0.06	0.05	4.15	0.27	0.07	0.004	N/A	252	3882	51.8	0.975	0.91	5610	14.8	4.9	9.1	14		0.091	N/A	0.026	14		N/A	5
Mean of Detected Data	0.185	0.00174	N/A	0.00733	0.0171	0.05	2.434	0.12	0.0667	0.0019	N/A	37.13	2703	1.767	0.00862	0.91	4225	3.618	0.974	2.58	2.757		0.0847	N/A	0.026	6.386		N/A	0.784
Median of Detected Data	0.14	0.002	N/A	0.007	0.01	0.05	2.65	0.13	0.07	0.0015	N/A	37.9	2952	0.16	0.0013	0.91	4569	3.35	0.7	2.1	2		0.084	N/A	0.026	6.2		N/A	0.046
SD of Detected Data	0.107	7.60E-04	N/A	0.00252	0.0112	N/A	0.814	0.099	0.00577	0.0011	N/A	16.53	649.3	7.961	0.0701	N/A	868.8	2.332	1.114	1.594	2.092		0.00603	N/A	N/A	5.381		N/A	1.862
Number Above Limit	0	0	0	2	1	0	233	0	0	0	0	1	178	2	3	0	183	54	0	45	0		0	0	0 0	7		0	0
Number Equal Limit	0	0	0	97	0	234	0	0	0	0	234	0	0	0	0	0	0	5	0	189	0		0	7	0	0		10	0
Number Below Limit	234	234	. 0	0	233	0	1	234	230	234	0	233	56	231	230	234	51	175	234	0	234		14	4	4	5		1	16
Number Observations Discarded	0	0	234	232	0	234	0	0		0	234	0	C			0	0		0	189				7	r			10	
H0: Site Median >= Comparison Value																													
(Form 2)	5	0.01	0.004	0.005	0.05	0.05	0.2	1	0.2	0.05	0.1	250	2125	30	0.03	80	3170	5	5	1	15	0.006	2	0.05	1	1	0.002	0.05 0.00	2 10
Test Value	-15.3	-15.3	0	2	-15.17	0	15.17	-15.3	-15.17	-15.3	0	-15.17	7.975	-15	-14.87	-15.3	8.629	-7.996	-15.3	6.708	-15.3		0	0	0 0	7		0	0
Lower Critical Value (0.05)	-1.645	-1.645	-1	-1	-1.645	-1	-1.645	-1.645	-1.645	-1.645	-1	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645		3	-1	-1	2		-1	4
P-Value	0	0	1	1	0	1	1	0	0	0	1	0	1	0	0	0	1	6.43E-16	0	1	0		6.10E-05	0.0625	0.0625	0.806		0.5	1.53E-05
			Do not	Do not			Do not				Do not		Do not				Do not			Do not				Do not	Do not	Do not		Do not	
Conclusion with Alpha = 0.05	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject		Reject	Reject	Reject	Reject		Reject	Reject
Obsevations below max detection limit =																													
non-detects ¹			0.05	0.01				0.1																					
Notes			1, 2	1,3		2		1			2									3				3,4	4			3,4	

Values lower than listed maximum nondetect value are treated by ProUCL as nondetects in single sample hypotheses tests
 All detection limits equal or higher than ARAR, all data rejected
 Nondetects equal or higher than ARAR rejected
 Erroneous conclusion by ProUCL

Comparisons of Background Sample Distributions to Comparison Values for Contaminants of Concern in Zone 1 Single Sample Wilcoxon Signed Rank Test

Image mgl ></th> <th>Single S</th> <th>sample w</th> <th>licoxon Sig</th> <th>gned Rank</th> <th>lest</th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th> <th></th>		Single S	sample w	licoxon Sig	gned Rank	lest																									
Rev Cal > <th>AI</th> <th>As</th> <th>Be</th> <th>Cd</th> <th>Co</th> <th>Pb</th> <th>Mn</th> <th>Мо</th> <th>Ni</th> <th>Se</th> <th>V</th> <th>CI</th> <th>SO4</th> <th>NO3_as_N</th> <th>U</th> <th>Chloroform</th> <th>Lab_TDS</th> <th>Rad_totl</th> <th>Th-230</th> <th>Pb-210</th> <th>iross_Alph</th> <th>Sb</th> <th>Ва</th> <th>Cr</th> <th>Cu</th> <th>Fe</th> <th>Hg</th> <th>Ag</th> <th>TI</th> <th>Zn</th>		AI	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	V	CI	SO4	NO3_as_N	U	Chloroform	Lab_TDS	Rad_totl	Th-230	Pb-210	iross_Alph	Sb	Ва	Cr	Cu	Fe	Hg	Ag	TI	Zn
Number of Valid Data 254	Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ug/L	mg/L					mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
Number of District Data 16 4 1 152 6 3 16 175 13 27 13 27 140 5 2 2 12 13 14 13 13 13 13 13 14 13 13 13 13 14 13 13 13 14 13 14 13 14 <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td>																															
Number of Non-Oeteic Data 200 101 231 210 233 212 232 224 234 234 235 235 235 235 235 235 235 235 235 235 235 235 235 30 0 1 1 1 1 <	Number of Valid Data	234	234	234	234	234	234	234	234	230	234	234	== :		233	233	234	234	234	234	234	234	0	14	11	4	12	0	11	0	16
Number of Detected Data 31 38 0 3 10 0 2.34 66 194 1 2.36 6.376 10.005 7.5006 5.3767 10.005 7.5006 5.3767 10.005 7.5006 5.3767 10.005 0.0	Number of Distinct Data	15	4	2	3	4	. 1	152	6	3	4	1	156	187	35	56	3	186	5 72	13	27	49		5	2	2	12		2		9
Percent Nam-Detects 86.7% 83.76% 100.00% 97.8%	Number of Non-Detect Data	203	196	234	231	210	233			227	224	234	0	0	167	39	233	0) 2	215	189			11	11	3	1		11		9
Minimum Non-detect 0.1 0.001 0.001 0.001 0.001 0.003 0.5 N/A 0.02 0.02 1 0.05 0.01 0.001 0.001 0.003 0.01 0.02 0.02 1 1 0.05 0.01 0.01 0.001 0.01 0.003 0.01 0.003 0.01 0.003 0.01 0.003 0.01 0.02 0.02 1 1 0.05 0.01 0.03 0.01 0.01 0.003 0.01 0.003 0.01 0.003 0.01 0.003 0.01 0.003 0.01 0.003 0.01 0.003 0.01 0.00 0.004 N/A 0.01 0.003 0.01 0.03 0.01 0.03 0.01 0.03 0.01 0.03 0.01 0.03 0.01 0.03 0.01 0.03 0.01 0.01 0.02 0.01 0.01 0.02 0.01 0.01 0.02 0.01 0.01 0.02 0.01 0.01 0.01	Number of Detected Data	31	38	0	3	24	. 1	233	5	3	10	0	234	234	66	194	1	234	232	19	45	152		3	0	1	11		0		7
Maximum Non-detect 0.1 0.01 0.01 0.05 0.07 0.1 0.05 0.07 0.1 0.05 0.07 0.1 0.05 0.07 0.1 0.05 0.07 0.1 0.06 0.07 0.1 0.06 0.07 0.07 0.07 0.08 0.001 N.A 0.026 0.22 0.2 1.1 0.1 0.07 N.A 0.026 0.21 0.07 N.A 0.028 0.28 0.07 N.A 0.028 0.28 0.028 0.28 0.07 N.A 0.028 0.28 0.028 0.28 0.27 0.01 N.A 0.028 0.038 0.028 0.038 0.038 0.031 4.25 3.25 0.11 0.038 0.028 0.038 0.038 0.039 0.0397 N.A 0.031 0.028 0.038 0.038 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 0.031 <td>Percent Non-Detects</td> <td>86.75%</td> <td>83.76%</td> <td>100.00%</td> <td>98.72%</td> <td>89.74%</td> <td>99.57%</td> <td>0.43%</td> <td>97.86%</td> <td>98.70%</td> <td>95.73%</td> <td>100.00%</td> <td>0.00%</td> <td>0.00%</td> <td>71.67%</td> <td>16.74%</td> <td>99.57%</td> <td>0.00%</td> <td>0.85%</td> <td>91.88%</td> <td>80.77%</td> <td>35.04%</td> <td></td> <td>78.57%</td> <td>100.00%</td> <td>75.00%</td> <td>8.33%</td> <td></td> <td>100.00%</td> <td></td> <td>56.25%</td>	Percent Non-Detects	86.75%	83.76%	100.00%	98.72%	89.74%	99.57%	0.43%	97.86%	98.70%	95.73%	100.00%	0.00%	0.00%	71.67%	16.74%	99.57%	0.00%	0.85%	91.88%	80.77%	35.04%		78.57%	100.00%	75.00%	8.33%		100.00%		56.25%
Minimum Detected 0.01 NA 0.006 0.011 0.06 0.001 NA 10.0 0.001 0.01 0.006 0.011 NA 11.0 0.001 0.00 0.026 0.027 NA 0.026 0.026 0.027 NA 0.026 0.026 0.026 0.027 NA 0.007 NA 0.026 0.037 0.001 NA 0.007 0.01 NA 0.007 0.01 NA 0.007 0.011 NA 0.007 0.011 NA 0.007 0.011 NA 0.007 0.011 NA 0.0082 0.91 0.016 0.017 NA 0.026 0.91 NA 0.0082 0.91 0.026 0.91 NA 0.0082 0.91 0.026 0.91 NA 0.0082 0.91 NA 7.92 0.92 0.004 NA 0.026 0.026 0.026 0.026 0.026 0.026 0.026 0.026 0.026 0.026 0.026 0.026 0.026	Minimum Non-detect	0.1	0.001	0.01	0.005	0.01	0.05	0.01	0.01	0.05	0.001	0.1	N/A	N/A	0.01	0.0003	0.5	N/A	0.2	0.02	1	1		0.05	0.01	0.02	0.1		0.01		0.01
Maximum Detected 0.66 0.004 NA 0.007 0.001 NA 3.79 2882 0.16 0.001 0.04 NA 0.026 6.38 NA 0.004 NA 0.026 6.38 NA 0.007 NA 0.026 0.007 NA 0.007 NA 0.006 0.008 NA 0.006 0.008 NA 0.006 0.008 NA 0.006 0.00	Maximum Non-detect	0.1	0.001	0.05	0.01	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.1	0.0004	1	N/A	0.2	0.2	1	1		0.1	0.05	0.02	0.1		0.05		0.1
Mean of Detected Data 0.185 0.0071/2 NA 0.0073 0.017 0.05 2.434 0.12 0.068 0.0013 0.014 0.002 0.00 0.001 0.0015 NA 37.13 27.03 1.767 0.0084 NA 0.026 6.386 NA 0.026 Median of Detected Data 0.007 7.605 0.010 0.05 2.434 0.007 0.0115 NA 37.13 27.03 1.767 0.0084 NA 0.026 6.386 NA 0.047 NA 0.026 0.011 </td <td>Minimum Detected</td> <td>0.1</td> <td>0.001</td> <td>N/A</td> <td>0.005</td> <td>0.01</td> <td>0.05</td> <td>0.66</td> <td>0.03</td> <td>0.06</td> <td>0.001</td> <td>N/A</td> <td>19.4</td> <td>1410</td> <td>0.01</td> <td>0.0004</td> <td>0.91</td> <td>2490</td> <td>0.2</td> <td>0.2</td> <td>1.1</td> <td>0.9</td> <td></td> <td>0.079</td> <td>N/A</td> <td>0.026</td> <td>0.25</td> <td></td> <td>N/A</td> <td></td> <td>0.01</td>	Minimum Detected	0.1	0.001	N/A	0.005	0.01	0.05	0.66	0.03	0.06	0.001	N/A	19.4	1410	0.01	0.0004	0.91	2490	0.2	0.2	1.1	0.9		0.079	N/A	0.026	0.25		N/A		0.01
Interface of Detected Data 0.14 0.002 N/A 0.007 0.001 0.007 0.0015 N/A 379 29252 0.16 0.0031 0.91 4569 3.33 2.07 2.1 2 0.0084 N/A 0.026 6.2 N/A 0.046 SD of Detected Data 0.107 7.605-04 N/A 0.0022 0.011 N/A 16.53 64.03 7.91 0.01 N/A 86.83 2.33 2.11 1.24 D.0063 N/A N/A 5.81 N/A 1.046 Number Above Limit 0 0 0.0011 N/A 16.53 64.03 7.91 1.03 6.0 0 0 1.03 6.0 0 0 0 1.03 0.011 N/A 86.83 7.91 1.14 1.54 0.00 0	Maximum Detected	0.6	0.004	N/A	0.01	0.06	0.05	4.15	0.27	0.07	0.004	N/A	252	3882	51.8	0.975	0.91	5610	14.8	4.9	9.1	14		0.091	N/A	0.026	14		N/A		5
SD Optiened Data 0.007 7.062-04 NA 0.0022 0.017 NA 0.888 0.070 NA 0.888 5.332 1.114 1.504 2.022 0.00603 NA NA 5.331 NA 1.862 Number Apose Limit 0 0 135 0 2.33 0 0 1 10 2.33 0 0 1 178 2.332 1.114 1.504 2.092 0 0 0 0 0 1 188 5.41 0 45 0 0 0 0 1 178 2.332 1.114 1.504 2.092 0 0 0 0 0 0 0 0 0 1 1.862 2.332 1.114 1.504 2.092 0.0060 NA NA 5.331 NA 1.862 1.862 1.862 1.862 1.862 1.862 2.332 1.114 1.504 2.092 0.06 0 0 0 0 0 0 0 0 0 0 0 0 0 0<	Mean of Detected Data	0.185	0.00174	N/A	0.00733	0.0171	0.05	2.434	0.12	0.0667	0.0019	N/A	37.13	2703	1.767	0.00862	0.91	4225	3.618	0.974	2.58	2.757		0.0847	N/A	0.026	6.386		N/A		
Number Above Limit 0 0 233 0 0 0 1 178 2 3 0 153 64 0 0 0 7 0 0 0 Number Equal Limit 0 0 1 233 0 135 0 1 178 2 3 0 135 64 0 <	Median of Detected Data	0.14	0.002	N/A	0.007	0.01	0.05	2.65	0.13	0.07	0.0015	N/A	37.9	2952	0.16	0.0013	0.91	4569				-		0.084	N/A	0.026			N/A		
Number Equal Limit 0 0 1 0		0.107	7.60E-04	N/A	0.00252	0.0112	N/A			0.00577	0.0011	N/A	16.53	649.3	7.961	0.0701	N/A	868.8	3 2.332	1.114	1.594	2.092		0.00603	N/A	N/A	5.381		N/A		1.862
Number Below Limit 234 234 0 98 233 233 1 234 233 234 <	Number Above Limit	0	0	234	1	1	0	233	0	C	0	0	1	178	2	3	0	183	3 54	0	45	0		0	0	0	7		0		0
T-plus 0 0 27495 99 1 0 27494 0 0 1 24822 3 470 0 25771 4650 0 9028 0 0 0 63 0 0 0 T-minus 27495 27495 27494 27261 1 27495 27494 27495 27495 27494 27495 27495 27495 27495 27495 27495 27495 1724 21686 27495 19467 27495 19467 27495 105 66 10 15 66 136 HO: Site Median > 0.004 0.005 0.05 0.05 0.01 20.05 0.11 20.05 21.125 30 0.03 80 3170 5 5 1 15 0.06 2 0.05 0.01 0.002 0.05 0.002 0.05 0.002 0.05 0.002 0.05 0.002 0.05 0.002 0.00 0.00	Number Equal Limit	0	0	0	135	0	1	0	0	C	0	0	0	0	0	0	0	0) 5	0	0	0		0	0	0	0		0		0
T-minus 27495 27495 0 4851 27494 27261 1 27495 27495 27495 105 66 10 15 66 136 H0: Site Media >= Comparison Value (Form 2) 5 0.01 0.004 0.005 0.05 0.05 0.01 0.005 0.05 0.01 0.000 0.005 0.001 0.005 0.01 0.005 0.01 0.005 0.01 0.005 0.01 0.005 0.01 0.005 0.01 0.006 0.01 0.006 0.01 0.006 0.01 0.006 0.01 0.006 0.01 0.00 0.01 0.00 0.01 <td>Number Below Limit</td> <td>234</td> <td>234</td> <td>0</td> <td>30</td> <td>233</td> <td>233</td> <td></td> <td>234</td> <td>230</td> <td>234</td> <td>234</td> <td>233</td> <td></td> <td>231</td> <td>230</td> <td>234</td> <td></td> <td></td> <td></td> <td>189</td> <td>234</td> <td></td> <td>14</td> <td>11</td> <td>4</td> <td>5</td> <td></td> <td>11</td> <td></td> <td>16</td>	Number Below Limit	234	234	0	30	233	233		234	230	234	234	233		231	230	234				189	234		14	11	4	5		11		16
H0: Site Median >= Comparison Value (Form 2) Do not Reject <		0	0	27495	99	1	0	27494	0	C	0	0	1	24822	3		0					0		0	0	0	63		0		0
(Form 2) 5 0.01 0.004 0.005 0.05 0.02 1 0.00 3 0.00 1 1 0.002 0.005 0.002 0.005 0.002 0.005 0.002 <t< td=""><td></td><td>27495</td><td>27495</td><td>0</td><td>4851</td><td>27494</td><td>27261</td><td>1</td><td>27495</td><td>26565</td><td>27495</td><td>27495</td><td>27494</td><td>2673</td><td>27258</td><td>26791</td><td>27495</td><td>1724</td><td>21686</td><td>27495</td><td>19467</td><td>27495</td><td></td><td>105</td><td>66</td><td>10</td><td>15</td><td></td><td>66</td><td></td><td>136</td></t<>		27495	27495	0	4851	27494	27261	1	27495	26565	27495	27495	27494	2673	27258	26791	27495	1724	21686	27495	19467	27495		105	66	10	15		66		136
Image: constraint of the second sec	H0: Site Median >= Comparison Value																														
Critical Value (0.05) -1.645	(Form 2)	5	0.01	0.004	0.005	0.05	0.05	0.2	1	0.2	0.05	0.1	250	2125	30	0.03	80	3170	5	5	1	15	0.006	2	0.05	1	1	0.002	0.05	0.002	10
Critical Value (0.05) -1.645																															
Critical Value (0.05) -1.645																															
P-Value 0 0 1 0 0 1 0 0 0 0 0 0 1 0 0 0 0 1 0 1 0 0 0 0 0 1 0 0 0 0 1 0 0 0 0 1 0 0 0 0 1 0 0 0 0 1 0 </td <td>Large Sample z-Test Value</td> <td>-</td> <td></td> <td></td> <td></td> <td>1 110 1</td> <td></td> <td></td> <td></td> <td></td> <td>-</td> <td>1010</td> <td></td> <td></td> <td></td> <td>-</td> <td></td> <td></td> <td></td> <td></td> <td>0.000</td> <td></td> <td></td> <td>0</td> <td>0</td> <td>0</td> <td>63</td> <td></td> <td>0</td> <td></td> <td>0</td>	Large Sample z-Test Value	-				1 110 1					-	1010				-					0.000			0	0	0	63		0		0
Image: Description of the secting of the secting o		-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645		-1.645		-1.645			13	-1					35
Conclusion with Alpha = 0.05 Reject no data Reject Reject no data ""><td>P-Value</td><td>0</td><td>0</td><td>1</td><td>0</td><td>0</td><td>0</td><td>1</td><td>0</td><td>C</td><td>0</td><td>0</td><td>0</td><td>1</td><td>0</td><td>0</td><td>0</td><td>1</td><td>6.21E-20</td><td>0</td><td>1.50E-09</td><td>0</td><td></td><td>1.00E-04</td><td>5.00E-04</td><td>N/A</td><td>0.968</td><td></td><td>5.00E-04</td><td></td><td>0</td></th<>	P-Value	0	0	1	0	0	0	1	0	C	0	0	0	1	0	0	0	1	6.21E-20	0	1.50E-09	0		1.00E-04	5.00E-04	N/A	0.968		5.00E-04		0
Conclusion with Alpha = 0.05 Reject > <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td>																															
Obsevations below max detection limit =				Do not				Do not						Do not				Do not								Do not	Do not				
	Conclusion with Alpha = 0.05	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	no data	Reject	Reject	Reject	Reject n	o data	Reject r	no data	Reject
non-detects ¹ 0.05 0.01 0.1 0.5 1 0.03 1 1 0.2 1 0.1 0.05 0.05 0.1	Obsevations below max detection limit =																														
	non-detects ¹			0.05	0.01				0.1	0.5					1	0.003	1		1	0.2		1		0.1	0.05				0.05		0.1
																											İ				
Notes 1,2 1 3 1 1 3 1 1 1 1 3 1 4 1 1,3 5 1,3 1	Notes			1, 2	1		3		1	1		3			1	1	1		1	1	3	1	4	1	1,3	5			1,3		1

Values lower than listed maximum nondetect value are treated by ProUCL as nondetects in single sample hypotheses tests
 Detection limits higher than ARAR

Detection limits equal ARAR
 No data

5. Erroneous conclusion by ProUCL

Background Statistics for Zone 3 Data Sets with Non-Detects

	AI	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	v	CI	SO4	NO3 as N	U I	Chloroform	Lah TDS	Pad 226	Rad-228	Rad totl	Th-230	Pb-210	Gross Alpha
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	ma/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ua/L	mg/L	pCi/L	pCi/L	pCi/L	pCi/L	pCi/L	pCi/L
Total Number of Data	186	186	0	186	186	185	186	184	186	0	186	186	186	186	0	186	186	186	185	185	186	186	186
Number of Non-Detect Data	127	50			17		100	26				100	100	32		185	100	22	54	18	167	129	
Number of Detected Data (or Distinct Obs. If zero		50	100	1//	17	101	1	20	73	144	100	0	0	52	2	105	0	22	54	10	107	123	23
nondetect)	, 59	136		٩	169	4	185	158	113	42		138	175	154	184	1	178	164	131	167	19	57	157
Minimum Detected	0.1	0.001		0.01	0.01	0.05	0.42	0.02	0.05	0.001		150	1319	0.01	0.0007	1.1	2244	0.2	131	0.2	0.2		1.57
Maximum Detected	1.68	1.01		0.01	0.53	0.03	7.5	75	0.03	0.001		66	4674	61	0.0007	1.1	6930	23.7	22.3	40.9	57	11	69
Percent Non-Detects	68.28%	26.88%	100.00%	95.16%	9.14%		0.54%	14.13%	39.25%	77.42%	100.00%	0%	4074	17.20%	1.08%	99.46%	0930	11.83%	29.19%	9.73%	89.78%	69.35%	15.59%
Minimum Non-detect	00.2078	0.001	100.00 /8	0.005	0.01	0.05	0.04 /8	0.1	0.05	0.001	100.0076	078	078	0.01	0.0003	35.4078	0 /0	0.2	29.1970	9.7378	0.2	09.0076	10.0978
Maximum Non-detect	0.1	0.001		0.003	0.01	0.05	0.01	0.1	0.05	0.001				0.01	0.0003	1		0.2	1	0.2	0.2	1	1
	0.1	0.001		0.01	0.01	0.05	0.01	0.1	0.05	0.001				0.1	0.0003	1		0.2	1	0.2	0.2	1	1
Mean of Detected Data	0.422	0.121		0.02	0.0835	0.065	3.25	11.88	0.173	0.0026		31.62	2588	11.34	0.0791		4115	5.01	5.34	9.099	6.705	2.549	8.191
Median of Detected Data	0.422	0.0235		0.02	0.0000	0.065	3.3		0.173			30.85	2651	4.785	0.039		4237	4.5	4.3	7.9	2.3	2.343	5.4
Variance of Detected Data	0.01	0.0233		0.0007	0.00803	0.000167	2.724	266.4	0.012			72.59	504851	200.6	0.00792		1049301	17.13	14.78	55.58	170.2	2.889	91.4
SD of Detected Data	0.332	0.211		0.0265	0.0896	0.000107	1.65	16.32	0.141	0.00334		8.52	710.5	14.16	0.00732		1043301	4.139	3.845	7.455	13.05	1.7	9.56
CV of Detected Data	0.332	1.744		1.323	1.073	0.199	0.508	1.374	0.813	1.285		0.269	0.275	1.249	1.125		0.249	0.826	0.72	0.819	1.946	0.667	1.167
Skewness of Detected Data	1.636	2.409		2.916	2.85	0.135	0.313	1.631	1.75	2.646		1.052	0.557	1.16	1.255		0.243	2.032	1.985	1.86	3.549	2.725	3.849
Mean of Log-Transformed Detected Data	-1.125	-3.512		-4.284	-2.866	-2.749	1.008	1.176	-2.013	-6.38		3.42	7.821	0.277	-3.329		8.292	1.28	1.464	1.835	0.85	0.784	1.712
SD of Log-Transformed Detected Data	0.726	1.816		0.74	0.86	0.202	0.648	1.934	0.692			0.26	0.28	2.849	1.416		0.25	0.892	0.652	1.000	1.465	0.525	0.851
Discernable Distribution (0.05) of Detections	Gamma	lognormal			lognormal	normal r	normal	none	none	none		gamma	none	none	none		none	none			amma	gamma	lognormal
Discernable Distribution (0.00) of Detections	Gamina	lognonnar			lognonnai	nonnai i	Ionnai	none	none	none		gamma	none	none	none	1	none	none	lognonnai		jamma	gamma	lognonnai
Kaplan-Meier (KM) Method for Datasets with						<u> </u>									<u> </u>								
Nondetects																							
Mean	0.202	0.0886		0.0105	0.0768	0.0503	3.234	10.21	0.125	0.00136				9.391	0.0782			4.441	4.073	8.234	0.865	1.475	7.07
SD	0.238	0.187		0.00589	0.0878	0.00273	1.655	15.63	0.124					13.54	0.0886			4.175	3.779	7.538	4.512	1.175	9.136
Standard Error of Mean	0.0176	0.0138		0.000458	0.00646	0.000232	0.122	1.156	0.00916					0.996	0.00652			0.307	0.279	0.556	0.34	8.69%	0.672
95% KM (t) UCL	0.231	0.111		0.0112	0.0875	0.0507	3.435	12.12	0.14					11.04	0.089			4.949	4.534	9.153	1.426	1.618	8.181
95% KM (z) UCL	0.231	0.111		0.0112	0.0874	0.0507	3.434	12.11	0.14					11.03	0.089			4.946	4.532	9.148	1.424	1.618	8.175
95% KM (BCA) UCL	0.238	0.113		0.0113	0.0877	0.0701	3.445	12.19	0.14	0.00159				11.14	0.0908			4.996	4.509	9.196	1.881	1.697	8.217
95% KM (Percentile Bootstrap) UCL	0.234	0.112		0.0113	0.0876	0.0701	3.436	12.06	0.14	0.00159				10.92	0.0895			4.98	4.537	9.158	1.625	1.655	8.186
95% KM (Chebyshev) UCL	0.279	0.149		0.0125	0.105	0.0513	3.765	15.25	0.165	0.00191				13.73	0.107			5.78	5.289	10.66	2.346	1.854	9.999
97.5% KM (Chebyshev) UCL	0.312	0.175		0.0133	0.117	0.0518	3.994	17.43	0.182	0.00215				15.61	0.119			6.359	5.815	11.71	2.987	2.017	11.27
99% KM (Chebyshev) UCL	0.378	0.226		0.015	0.141	0.0526	4.445	21.71	0.216					19.3	0.143			7.496	6.848	13.76	4.246	2.339	13.76
Datasets without Nondetects																							
Student's-t UCL												32.66	2674				4239						
95% UCLs (Adjusted for Skewness)																							
95% Adjusted-CLT UCL												32.7	2676				4242						
95% Modified-t UCL												32.66	2674				4239						
Non-Parametric UCLs																							
95% CLT UCL												32.65	2674				4238						
95% Jackknife UCL				l l								32.66	2674				4239						
95% Standard Bootstrap UCL												32.68	2673				4237						
95% Bootstrap-t UCL												32.75	2675				4249						
95% Hall's Bootstrap UCL	1					† †						32.66	2676		†		4252						
95% Percentile Bootstrap UCL	1	1				†						32.71	2670				4237						
95% BCA Bootstrap UCL										1		32.65	2673			1	4235			I			
95% Chebyshev(Mean, Sd) UCL												34.35					4442			1			
97.5% Chebyshev(Mean, Sd) UCL												35.53					4584			1			
99% Chebyshev(Mean, Sd) UCL						l l						37.84					4862						
<u> </u>																							
Potential UCL to Use																							
95% KM (t) UCL	0.231			0.0112		0.0507	3.435			0.00157											1.426	1.618	
95% KM (z) UCL																							
95% KM (BCA) UCL					0.0877				0.14									4.996	4.509				8.217
95% KM (Percentile Bootstrap) UCL				0.0113		0.0701	3.436			0.00159													
95% KM (Chebyshev) UCL															0.107					10.66			
97.5% KM (Chebyshev) UCL		0.175						17.43						15.61									
99% KM (Chebyshev) UCL																							
95% Student's-t UCL													2674				4239						
95% Modified-t UCL													2674				4239			1			
95% Approximate Gamma UCL												32.65								1			
																				ĺ			
																				ĺ			
Notes			5	1, 3		3					5			1		4				İ			
			-	· · · · · · · · · · · · · · · · · · ·		· · · · · · · · · · · · · · · · · · ·					-												

Note:

Data have multiple DLs - Use of KM Method is recommended
 UCL recommendation based on guidance (table 16). None provided by software.

3. There may not be adequate distinct detected values to compute meaningful and reliable test statistics and estimates The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

4. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values

determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).

5. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs!

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

	Sb	Ва	Cr	Cu	Fe	Hg	Ag	TI	Zn
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
Total Number of Data	1	36	37	13	23	4	29	0	31
Number of Non-Detect Data	1	34	37	10	9	4	29		6
Number of Detected Data (or Distinct Obs. If									
zero nondetect)	0	2	0	3	14	0	0		25
Minimum Detected		0.54		0.028	0.03				0.02
Maximum Detected		0.54		0.06	67				6.859
Percent Non-Detects	100.0%	94.40%	0.00%	76.92%	39.13%	100.00%	100.00%		19.35%
Minimum Non-detect				0.01	0.1				0.02
Maximum Non-detect				0.02	0.1				0.1
Mean of Detected Data				0.042	9.682				0.766
Median of Detected Data				0.038	1.45				0.193
SD of Detected Data				0.0164	17.88				1.787
Mean of Log-Transformed Detected Data				-3.22	0.439				-1.622
SD of Log-Transformed Detected Data				0.384	2.354				1.511
Discernable Distribution (0.05)				Normal	Gamma				Lognormal
Skewness of Detected Data				1.034	2.879				3.177
Kaplan-Meier (KM) Method for Datasets with									
Nondetects									
Mean				0.0312	5.905				0.622
SD				0.00872					1.599
Standard Error of Mean				0.00296					0.293
95% KM (t) UCL				0.00290	11.2				1.12
95% KM (z) UCL				0.0361	10.97				1.12
95% KM (BCA) UCL				0.001	11.63				1.245
95% KM (Percentile Bootstrap) UCL				0.00	11.44				1.1243
95% KM (Chebyshev) UCL				0.0441	19.34				1.123
97.5% KM (Chebyshev) UCL				0.0441	25.15				2.453
99% KM (Chebyshev) UCL				0.0497	36.57				3.539
				0.0007	30.57				3.559
Datasets without Nondetects									
Student's-t UCL									
95% UCLs (Adjusted for Skewness)									
95% Adjusted-CLT UCL									
95% Modified-t UCL									
Non-Parametric UCLs									
95% CLT UCL									
95% Jackknife UCL									
95% Standard Bootstrap UCL									
95% Bootstrap-t UCL									
95% Hall's Bootstrap UCL									<u> </u>
95% Percentile Bootstrap UCL									<u> </u>
95% BCA Bootstrap UCL									
95% Chebyshev(Mean, Sd) UCL 97.5% Chebyshev(Mean, Sd) UCL									
99% Chebyshev(Mean, Sd) UCL									
Patantial UCL to Upa									
Potential UCL to Use				0.0007					
95% KM (t) UCL				0.0365					
95% KM (z) UCL					10.10				
95% KM (BCA) UCL					12.16				
95% KM (Percentile Bootstrap) UCL				0.06					
95% KM (Chebyshev) UCL									
97.5% KM (Chebyshev) UCL									
99% KM (Chebyshev) UCL									3.539
95% Student's-t UCL									ļ
95% Modified-t UCL									
Notes	6	4	5	1, 3		6	5		1

Note:

1. Data have multiple DLs - Use of KM Method is recommended

2. UCL recommendation based on guidance (table 16). None provided by software.

3. There may not be adequate detected values to compute meaningful and reliable test statistics and estimates

The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

4. Warning: Only one distinct data value was detected! It is suggested to use alternative site specific values

determined by the Project Team to estimate environmental parameters (e.g., EPC, BTV).

5. Warning: All observations are Non-Detects (NDs), therefore all statistics and estimates should also be NDs! The Project Team may decide to use alternative site specific values to estimate environmental parameters (e.g., EPC, BTV).

6. Insufficent Number of Observations to produce Meaningful Statistics.

Comparisons of Background Sample Distributions to Comparison Values for Contaminants of Concern in Zone 3 Single Sample Sign Test

			Sample Sig																											
	AI	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	v	CI		NO3_as_N	-	Chloroform							Ва	Cr	Cu	Fe	Hg	Ag	TI	Zn
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ug/L	mg/L	pCi/L	pCi/L	pCi/L	pCi/L m	ng/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
Raw Statistics																														1
Number of Valid Data	186		6 186	186	186	185	186	184	186	186	186	186	186	186		186	186	6 185	186		186	1	36	37	13	23	4	29	0	31
Number of Distinct Data	38	0	4 3	4	29	4	148	132	34	9	2	138	175	102		2	178	3 113	17	30	92	1	3	3	5	13	1	3		22
Number of Non-Detect Data	127		100	177	17	181	1	26	73	144	186	0	0	32	-	185	0	18	167	129	29	1	34	37	10	9	4	29		6
Number of Detected Data	59	136	6 0	9	169	4	185	158	113	42	0	186	186	154	184	1	186	6 167	19	57	157	0	2	0	3	14	0	0		25
Percent Non-Detects	68.28%	26.88%	6 100.00%	95.16%	9.14%	97.84%	0.54%	14.13%	39.25%	77.42%	100.00%	0.00%	0.00%	17.20%	1.08%	99.46%	0.00%	9.73%	89.78%	69.35%	15.59% 100	0.00%	94.44%	100.00%	76.92%	39.13%	100.00%	100.00%		19.35%
Minimum Non-detect	0.1	0.001	0.001	0.005	0.01	0.05	0.01	0.1	0.05	0.001	0.01	N/A	N/A	0.01	0.0003	1	N/A	0.2	0.2	1	1	0.05	0.05	0.01	0.01	0.1	2.00E-04	0.005		0.02
Maximum Non-detect	0.1	0.001	1 0.05	0.01	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.1	0.0003	1	N/A	0.2	0.2	1	1	0.05	0.1	0.1	0.02	0.1	2.00E-04	0.05		0.1
Minimum Detected	0.1	0.001	1 N/A	0.01	0.01	0.05	0.42	0.02	0.05	0.001	N/A	15	1319	0.01	0.0007	1.1	2244	0.2	0.2	1	1 N/	/A	0.054	N/A	0.028	0.03	N/A	N/A		0.02
Maximum Detected	1.68	1.01	1 N/A	0.09	0.53	0.08	7.5	75	0.67	0.015	N/A	66	4674	61	0.38	1.1	6930	40.9	57	11	69 N/	/A	0.054	N/A	0.06	67	N/A	N/A		6.859
Mean of Detected Data	0.422	0.121	1 N/A	0.02	0.0835	0.065	3.25	11.88	0.173	0.0026	N/A	31.62	2588	11.34	0.0791	1.1	4115	9.099	6.705	2.549	8.191 N/	/A	0.054	N/A	0.042	9.682	N/A	N/A		0.766
Median of Detected Data	0.31	0.0235	5 N/A	0.01	0.06	0.065	3.3	3.76	0.12	0.001	N/A	30.85	2651	4.785	0.039	1.1	4237	7.9	2.3	2	5.4 N/	/A	0.054	N/A	0.038	1.45	N/A	N/A		0.193
SD of Detected Data	0.332	0.211	1 N/A	0.0265	0.0896	0.0129	1.65	16.32	0.141	0.00334	N/A	8.52	710.5	14.16	0.089	N/A	1024	7.455	13.05	1.7	9.56 N/	/A	0	N/A	0.0164	17.88	N/A	N/A		1.787
Number Above Limit	0	97	7 0	9	87	3	185	112	30	0	0	0	135	18	104	0	151	121	5	56	19	0	0	0	0	7	0	0		0
Number Equal Limit	0	3	3 0	12	19	182	0	0	3	0	185	0	0	0	1	0	C) 1	0	130	0	0	0	21	0	1	0	23		0
Number Below Limit	186	86	6 1	0	80	0	1	72	153	186	1	186	51	168	81	186	35	63	181	0	167	0	36	13	13	15	4	6		31
Number Observations Discarded			185	177		182					185	0	0				C)		130		1		24				23		<u> </u>
H0: Site Median >= Comparison Value																														
(Form 2)	5	0.01	0.004	0.005	0.05	0.05	0.2	1	0.2	0.05	0.1	250	2125	30	0.03	80	3170	5	5	1	15	0.006	2	0.05	1	1	0.002	0.05	0.002	10
Test Value	-13.64	0.813	3 0	9	0.542	3	13.49	2.949	-9.092	-13.64	0	-13.64	6.159	-11	1.691	-13.64	8.506	6 4.276	-12.9	7.483	-10.85	0	0	0	0	7	0	0		0
Lower Critical Value (0.05)	-1.645	-1.645	5 -1	1	-1.645	-1	-1.645	-1.645	-1.645	-1.645	-1	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1.645	-1	12	3	3	6	-1	0		10
P-Value	0	0.792	2 0.5	1	0.706	1	1	0.998	4.84E-20	0	0.5	0	1	1.94E-28	0.955	0	1	1	0	1	9.77E-28	1 1	.46E-11	1.22E-04	1.22E-04	0.0669	0.0625	0.0156		4.66E-10
Obsevations below max detection limit =																														
																							0.4							
non-detects ¹		De N. C	De Nu f	De Not	Della	De Net	De Net	0.1			De Net		De Not	0.1	De Net		De Net	Do Not		De Net	-		0.1			0.1	Denet			0.1
Conclusion with Alpha = 0.05	Reject	Do Not Reject	Do Not Reject	Do Not Reject	Do Not Reject	Do Not Reject	Do Not Reject	Do Not Reject	Reject	Reject	Do Not Reject	Reject	Do Not Reject	Reject	Do Not Reject	Reject	Do Not Reject	Do Not Reject	Reject	Do Not Reject	Do n Reject Reje		Reject	Reject	Reject	Do not Reject	Do not Reject	Reject	no data	Reject
·						•				-						-		-	•											
Notes		1	3, 4	3		3		1			3, 4			1						3		2, 4	1			1, 4	4			1

Values lower than listed maximum nondetect value are treated by ProUCL as nondetects in single sample hypotheses tests
 All detection limits equal or higher than ARAR, all data rejected
 Nondetects equal or higher than ARAR rejected
 Erroneous conclusion by ProUCL

Comparisons of Background Sample Distributions to Comparison Values for Contaminants of Concern in Zone 3 Single Sample Wilcoxon Signed Rank Test

				igned Rank						1	1							1												
	Al	As	Be	Cd	Co	Pb	Mn	Мо	Ni	Se	v	CI		NO3_as_N	-	Chloroform			Th-230			Sb	Ва	Cr	Cu	Fe	Hg	Ag	TI	Zn
Concentration Units	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	ug/L	mg/L	pCi/L	pCi/L	pCi/L	pCi/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
Raw Statistics																														
Number of Valid Data	186	186	5 186	5 186	186	185	186	184	186	186	186	186	186	186	186	186	186	6 185	186	186	186	1	36	37	13	23	4	29	0	31
Number of Distinct Data	38	84	4 3	8 4	29	4	148	132	34	9	2	138	175	102	141	2	178	3 113	17	30	92	1	3	3	5	13	1	3		22
Number of Non-Detect Data	127	50	186	5 177	17	181	1	26	73	144	186	0	0	32	2	185	0	18	167	129	29	1	34	37	10	9	4	29		6
Number of Detected Data	59	136	6 0	9	169	4	185	158	113	42	0	186	186	154	184	1	186	6 167	19	57	101	0	2	0	3	14	0	0		25
Percent Non-Detects	68.28%	26.88%	100.00%	95.16%	9.14%	97.84%	0.54%	14.13%	39.25%	77.42%	100.00%	0.00%	0.00%	17.20%	1.08%	99.46%	0.00%	9.73%	89.78%	69.35%	15.59%	100.00%	94.44%	100.00%	76.92%	39.13%	100.00%	100.00%		19.35%
Minimum Non-detect	0.1	0.001	0.001	0.005	0.01	0.05	0.01	0.1	0.05	0.001	0.01	N/A	N/A	0.01	0.0003	1	N/A	0.2	0.2	1	1	0.05	0.05	0.01	0.01	0.1	2.00E-04	0.005		0.02
Maximum Non-detect	0.1	0.001	0.05	i 0.01	0.01	0.05	0.01	0.1	0.05	0.001	0.1	N/A	N/A	0.1	0.0003	1	N/A	0.2	0.2	1	1	0.05	0.1	0.1	0.02	0.1	2.00E-04	0.05		0.1
Minimum Detected	0.1	0.001	N/A	0.01	0.01	0.05	0.42	0.02	0.05	0.001	N/A	15	1319	0.01	0.0007	1.1	2244	1 0.2	0.2	1	1	N/A	0.054	N/A	0.028	0.03	N/A	N/A		0.02
Maximum Detected	1.68	1.01	N/A	0.09	0.53	0.08	7.5	75	0.67	0.015	N/A	66	4674	61	0.38	1.1	6930	0 40.9	57	11	69	N/A	0.054	N/A	0.06	67	N/A	N/A		6.859
Mean of Detected Data	0.422	0.121	N/A	0.02	0.0835	0.065	3.25	11.88	0.173	0.0026	N/A	31.62	2588	11.34	0.0791	1.1	4115	5 9.099	6.705	2.549	8.191	N/A	0.054	N/A	4.20%	9.682	N/A	N/A		0.766
Median of Detected Data	0.31	0.0235	5 N/A	0.01	0.06	0.065	3.3	3.76	0.12	0.001	N/A	30.85	2651	4.785	0.039	1.1	4237	7.9	2.3	2	5.4	N/A	0.054	N/A	0.038	1.45	N/A	N/A		0.193
SD of Detected Data	0.332	0.211	N/A	0.0265	0.0896	0.0129	1.65	16.32	0.141	0.00334	N/A	8.52	710.5	14.16	0.089	N/A	1024	1 7.455	13.05	1.7	9.56	N/A	0	N/A	0.0164	17.88	N/A	N/A		1.787
Number Above Limit	0	97	185	5 9	87	3	185	112	30	0	0	0	135	18	104	0	151	121	5	56	19		0	0	0	7	0	0		0
Number Equal Limit	0	3	s 0	165	19	1	0	0	3	0	0	0	0	0	1	0	0) 1	0	1	0		0	3	0	1	0	0		0
Number Below Limit	186	86	6 1	12	80	181	1	72	153	186	186	186	51	168	81	186	35	5 63	181	129	167		36	34	13	15	4	29		31
T-plus	0	11532	17264	153	8396	187	17390	13885	2603	0	0	0	14409	675.5	12127	0	15898	3 12007	740	6821	1331		0	0	0	121	0	0		0
T-minus	17391	5304	127	78	5632	16833	1	3135	14234	17391	17391	17391	2982	16716	5079	17391	1493	3 5013	16651	10385	16060		666	595	91	132	10	435		496
H0: Site Median >= Comparison Value																														
(Form 2)	5	0.01	0.004	0.005	0.05	0.05	0.2	1	0.2	0.05	0.1	250	2125	30	0.03	0.05	3170) 5	5	1	15	0.006	2	0.05	1	1	0.002	0.05	0.002	10
											-																			
Large Sample z-Test Value	-12.32	3.964	12.18	-304.9	-0.478	-13.31	11.83	7.433	-8.559	-12.58	-13.6	-11.83	7.772	-10.91	4,705	-13.6	9.797	4.707	-11.94	-2.688	-10.02		-5.549	-6.221	0	-0.54	0	-4.999		-4.855
Critical Value (0.05)	-1.645				-1.645	-1.645	-1.645		-1.645	-1.645					-1.645	-1.645			-1.645		-1.645		-1.645	-1.645	21	-1.645		-1.645		-1.645
P-Value	3.42E-35	1	1	0	0.316	0	1	1	5.69E-18	1.43E-36	0	1.43E-32	1	5.12E-28	1	0	1	1	3.463E-33	3.59E-03	6.16E-24		1.44E-08	2.47E-10	1.00E-04	0.295	N/A	2.88E-07		6.02E-07
			-	-		Ţ					-		-			-														
		Do Not	Do Not		Do Not		Do Not	Do Not					Do Not		Do Not		Do Not	Do Not								Do not	Do not			
Conclusion with Alpha = 0.05	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject		Reject	Reject	Reject	Reject	Reject	Reject		Reject
	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Rejeor	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject	Reject		Reject	Reject	Reject	Reject	Reject	110,001		Reject
Obsevations below max detection limit =																		1										 		
non-detects ¹								0.1						0.1				1					0.1					, I		
	<u> </u>			1				0.1						0.1				+					0.1		<u> </u>			·	+	
Notes			2.4	23		2		1			34			4						0		4	4	2						
NOLES			∠, 4	∠, 3		3		1			3, 4			1				1		3	Z	, 4	1	2				3	5	

Values lower than listed maximum nondetect value are treated by ProUCL as nondetects in single sample hypotheses tests
 Detection limits higher than ARAR
 Detection limits equal ARAR
 All data are below detection limits

5. No data

APPENDIX D

Annual Land Use Report for 2009



UNC CHURCH ROCK

UNITED NUCLEAR CORPORATION



P.O. Box 3077 Gallup, New Mexico 87305-3077 Telephone: (505) 905-6651 Fax: (505) 905-6654

March 17, 2010

REPT 07 COAG '08

Mr. Keith I. McConnell, Deputy Director Decommissioning and Uranium Recovery Licensing Directorate Division of Waste Management and Environmental Protection Office of Federal and State Materials and Environmental Management Programs U.S. Nuclear Regulatory Commission 11545 Rockville Pike #2 White Flint, Mail Stop T7 E-18 Rockville, MD 20852-2738

Re: Annual Land Use Report for 2009

Dear Mr. McConnell:

The above report is submitted, pursuant to our NRC Source Materials License No. SUA-1475, Condition 31.

Sincerely,

Map Chinchill

Max Chischilly, Jr. Radiation Safety Officer

DY:

Enclosure

Cc: Jack E. Whitten, NRC Region IV Steve Hill, GE Roy Blickwedel, GE Yolande Norman, NRC



Page 1 of 5

ANNUAL LAND USE SURVEY REPORT FOR 2009 UNITED NUCLEAR CORPORATION LICENSE NO. SUA – 1475

CONDITION NO. 31

MARCH 16, 2010



SURVEY OF LAND OWNERSHIP AND USE WITHIN TWO-MILE RADIUS OF MILL SITE

UNITED NUCLEAR CORPORATION License No. SUA-1475 Condition No. 31

1.0 Introduction

This report has been prepared pursuant to License Condition 31 of United Nuclear Corporation's License No. SUA – 1475. The information submitted in this report was acquired from the master title plate published by the Bureau of Land Management. United Nuclear Corporation maintains the surface ownership records. The map is a copy of the USGS Quadrangle of Hardground Flats, Oak Springs, and Churchrock, the photo revised in 1979. United Nuclear Corporation's Radiation Safety Officer performed the land use survey.

2.0 Area Ownership and Use

Reference the attached map (figure 1) for location in regards to Mill Site.

Area:	Owner:	Usage:
Section 1	Navajo Tribe	Grazing and well monitoring
Section 2*	United Nuclear	Mill & Tailings site, one employee homesite and well monitoring.
Section 3	Navajo Tribe	Grazing and well monitoring
Section 4	Indian Allotted	Grazing
Section 6	Indian Allotted	Grazing
Section 9	Navajo Tribe	Grazing
Section 10	Indian Allotted	Grazing, eleven homesite, and well monitoring
Section 11	Navajo Tribe	Grazing
Section 12	BLM, A, Etah, Etal	Grazing
Section 13	Navajo Tribe	Grazing
Section 14	Indian Allotted	Grazing
Section 15	Navajo Tribe	Grazing

Page 3 of 5



Area: Section 33	Owner: Navajo Tribe	Usage: Grazing
Section 34	BLM (western portion) United Nuclear (eastern portion)	Grazing
Section 35	Indian Trust for Navajo Tribe	Grazing and UNC'S Reclaimed Northeast Churchrock Mine Site is located in this section.
Section 36	United Nuclear	Unauthorized grazing and well monitoring.
Map Northern Portion of 2 mile radius	Navajo Reservation	Grazing, 24 home sites, and Quivira's reclaimed Church- Rock Mine site is located in this area.
Section 31	Indian Allotted	Grazing

*Additional Note for Section 2:

The Mill has been decommissioned and has been cleaned to meet releasable standards for unrestricted use. Final reclamation activity on the tailings area was completed in 1996 with the exception of evaporation pond area, currently used for ground water remediation. Final cover radon flux test result was reported in the "Report On Radon Emanation Testing Of Final Radon Cover Over UNC'S Church Rock Tailings Site, Docket No. 40-80907" submitted on January 3, 1997.

3.0 Current ongoing groundwater tailings seepage remediation activity

- Sample/monitor wells on Sec. 2 and 36 (UNC), Sec. 1 and 3 (Navajo Tribe) and Sec. 10 (Indian Allotted) on a quarterly basis.
- 2) Continual pumping/extraction of wells RW-11, RW-16, RW-17, RW-A, PB-2, NW-1 NW-2, NW-3 (pumping since Feb. 2009 and turned off on Nov. 2009) and NW-4 (pumping started on Nov. 2009). These wells are on UNC's Sec. 36 to enhance the remedy for cut off and containment of the migrating Zone 3 seepage impacted water.
- 3) Monthly monitoring (i.e. measure field parameters, in-house bicarbonate and chloride titration test) of wells NBL-1, NBL-2, PB-2, PB-3, PB-4, RW-A and NW series 1 thru 5 to track the northern most migration of the seepage impacted water in Zone 3, Sec. 36.



4.0 Well ID, Use, Location and Formation

Well ID: United Nuclear	Use: Domestic	Location: Sec. 2-Mill Site	Formation: Westwater
Circle Wash	Domestic/ Livestock	Section 14	Alluvium
Unknown ID Abandoned	No Known Use	Section 11	Alluvium
J.E. Soper#1 Abandoned	No Known Use	Section 1	Two Wells- Members Mancos
BLM – 2	Monitor	Section 12	Alluvium
14T – 586 (Friendship well #	Livestock	N.I.R. N Part of map	Lower Gallup
NR – 1	Monitor/ Inactive	N.I.R. N Part of map	Alluvium
15K – 303	Livestock	N.I.R. NE Part of Map	Upper Gallup

5.0 Significant changes or events which took place in 2009 are as follows:

- 1) No change under item 2.0 (Area ownership and Use) and item 4.0 (Well ID, Use, Location and Formation). Total current homesite is thirty-six within the two mile radius of mill site.
- 2) Pumping started during February 2009 for wells NW-1, NW-2 and NW-3 also a new pumping regime was implemented during the second week of November 2009, whereby NW-3 was turned off and pumping started for NW-4 in Sec. 36. All the pumped ground water is discharged into the tailings North Evaporation Pond in Sec. 2.



- 3) On August 18, 2009 remedial construction activity (i.e. Interim Removal Action) involving various agencies and contractors (i.e. USEPA, NNEPA, MWH, MACTEC, AVM and UNC) started on UNC's inactive Northeast Church Rock Mine Site in Sec. 35 (Indian Trust Land) and the nearby adjacent Navajo Reservation Land. The work ceased on January 11, 2010 due to winter weather conditions and will resume during the spring of 2010. Also, three households were temporarily relocated to an off site housing complex for about five months due to the residents in close proximity to or within the construction activity area.
- 4) As requested by Chester Engineers, an injection test was performed during October 30, 2009 and November 4, 2009 in well NBL-2 (background monitor well in Zone 3, Sec. 36) by MACTEC and UNC as part of a potential remedial option for the on going SWSFS (site – wide supplemental feasibility study). The objective and result of this test is reported on pg. 29 of Chester Engineers 2009 Ground water Corrective Action Annual Review Report for UNC.

