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1. PURPOSE

The Monitored Geologic Repository (MGR) Waste Package Operations (WPO) of the Civilian Radioactive Waste Management System Management and Operating Contractor (CRWMS M&O) performed calculations to provide input for disposal of Pu-ceramic waste forms. The Puceramic (Refs. 1 and 2) is designed to immobilize excess plutonium from weapons production, and has been considered for disposal at the potential Yucca Mountain site. Because of the high content of fissile material in the ceramic, the waste package (WP) design requires special consideration of the amount and placement of neutron absorbers, and the possible loss of absorbers and fissile materials over geologic time. For some WPs, the corrosion-allowance material (CAM) and the corrosion-resistant material (CRM) may breach (Ref. 3, Section 10.5.1.2), allowing the influx of water. Water in the WP will moderate neutrons, increasing the likelihood of a criticality within the WP; and the water may, in time, gradually leach the fissile components and neutron absorbers out of the WP, further affecting the neutronics of the system.

This study presents calculations of the long-term geochemical behavior of WPs containing Puceramic disks arranged according to the "can-in-canister" concept. The cans containing Puceramic disks are embedded in canisters filled with high-level waste (HLW) glass (Ref. 1). The objectives of this calculation were to determine:

- The extent to which criticality control material, suggested for this WP design, will remain in the WP after corrosion/dissolution of the initial WP configuration (such that it can be effective in preventing criticality).
- The extent to which fissile plutonium and uranium will be carried out of the degraded WP by infiltrating water (such that internal criticality is no longer possible, but the possibility of external criticality may be enhanced).
- * The nominal chemical composition for the criticality evaluations of the WP design, and to suggest the range of parametric variations for additional evaluations.

The chemical compositions (and subsequent criticality evaluations) for some of the simulations are calculated for time periods to $\sim 6.10^5$ years. The longer time is based on the analysis in Reference 4 (Figure C-13), which suggests that 5% of the WPs may remain flooded for longer than $2\cdot10^5$ years. However, it is important to note that after 10^5 years of flooding, most of the materials of interest (fissile and absorber materials) will have either been removed from the WP, reached a steady state, or been transmuted.

The calculation included elements with high neutron absorption cross sections, notably gadolinium (Gd), as well as the fissile materials. The results of this calculation will be used to ensure that the type and amount of criticality control material used in the WP design will prevent criticality.

A previous calculation of the degradation of Pu-ceramic WPs found maximum losses of -15% of

the original Gd content of the Pu-ceramic WP (Ref. 5. Table 5.3-1). However, the water and HLW glass compositions, and the ceramic degradation rates, used in that prior study, have been superceded by new information. In addition, the development of a faster and more flexible version of the reaction-path code has made it possible to calculate degradation of the WPs under a wider range of rate combinations. Consequently, the current study provides a greater understanding of the effects of composition and rate uncertainty on the potential losses of Gd and fissile materials.

This calculation was prepared under procedure AP-3.12Q, Revision 0, ICN 0.

2. METHOD

The method used for this calculation involves the following steps:

- Use of the qualified and modified version of the EQ3/6 reaction-path code (software package, Section 4.1) for tracing the degradation of the WP. The software estimates the concentrations remaining in the aqueous solution and the composition of the precipitated solids. (EQ3 is used to determine a starting fluid composition for EQ6 calculations; it does not simulate reaction progress.)
- * Evaluation of available data for degradation rates of package materials, to be used as EQ6 reaction rates (rkl in the EQ6 "6i" files).
- Use of "solid-centered flow-through" mode (SCFT) in EQ6; in this mode, an increment of aqueous "feed" solution is added continuously to the WP system, and a like volume of the existing solution is removed, simulating a continuously-stirred tank reactor. This mode is discussed in Section 4.
- Determination of fissile material concentrations in solution as a function of time (from the output of EQ6 simulated reaction times up to $6.10⁵$ years).
- Calculation of the amount of fissile material released from the WP as a function of time (fissile material loss reduces the chance of criticality within the WP).
- * Determination of concentrations of neutron absorbers, such as Gd and B, in solution as a function of time (from the output of EQ6 over times up to $6.10⁵$ years).
- * Calculation of the amount of neutron absorbers retained within the WP as a function of time.
- Determination of composition and amounts of solids (precipitated minerals or corrosion products, and unreacted package materials).

Detailed description of each step is available in Section 5 of this calculation.

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3. ASSUMPTIONS

All assumptions are for preliminary design; these assumptions will require verification before this calculation can be used to support procurement, fabrication, or construction activities. All assumptions are used throughout Section 5 and Section 6.

- 3.1 It is assumed that an aqueous solution fills all voids within WPs, and that the solutions that drip into the WP will have a composition approximating that of the J-13 well water (as given in Ref. 6, Tables 4.1 and 4.2; this composition is given in Table *5-4* of the current document) for $\sim 10^6$ years. (The J-13-like compositions are hereafter referred to as "incoming water", rather than "J-13 well water", to distinguish the idealized compositions from actual well water samples.) The basis for the first part of this assumption is that it provides the maximum degradation rate with the potential for the fastest flushing of the neutron absorber from the WP, and is thereby conservative. The basis for the second part of the assumption is that the groundwater composition is controlled largely by transport through the host rock, over pathways of hundreds of meters, and the host rock composition is not expected to change substantially over $10⁶$ years. For a few thousand years after waste emplacement, the composition may differ because of perturbations resulting from reactions with engineered materials and from the thermal pulse. These are not taken into account in this calculation because the corrosion allowance barrier and CRM are not expected to breach until after that perturbed period. Therefore, the early perturbation is not relevant to the calculations reported in this document. See Assumption 3.3.
- 3.2 It is assumed that the density of the incoming water is 1.0 g/cm³. The basis for this assumption is that for dilute solutions, the density is extremely close to that for pure water, and that any differences are insignificant in respect to other uncertainties in the data and calculations. Moreover, this value is used only initially in EQ316 to convert concentrations of dissolved substances from parts per million to molalities.
- 3.3 The assumption that the water entering the WP can be approximated by the J-13 well water implicitly assumes: (1) that the incoming water will have only a minimal contact, if any at all, with undegraded metal in the corrosion allowance barrier, and (2) that any effects of contact with the drift liner will be minimal after a few thousand years. The basis for the first part of this assumption is that the water will move sufficiently rapidly through openings in the WP barriers such that its residence time in the corroded barrier will be too short for significant reaction to occur, and the corrosion products lining the cracks should consist primarily of inert Fe oxides. The second part of this assumption is justified by the following: (A) the drift liner at the top of the drift is expected to collapse with the roof support well before 1000 years; and (B) the water flowing through the concrete liner, dominantly along fractures, will be in contact with the degradation products of the liner, which will have come close to equilibrium with the water moving through the rock above the repository (most recent drift designs do not contain significant amounts of concrete); and (C) recent evaluations of codisposal WPs show that

degradation of the WP materials (specifically, HLW glass and steel) overwhelms the native chemistry of the incoming water (Figures 5-2 through 5-20 of Ref. 7 show pH variations of 3 to 10 in WP). Thus, even though the chemistry of the infiltrating water may vary substantially, the effects of the variations will likely be insignificant in a WP that undergoes significant alteration. An evaluation of the effects of varying incoming water chemistry is given in Section 5.3.2 of the current document.

- 3.4 It is assumed that water may circulate sufficiently freely in the partially degraded WP that all degraded solid products may react with each other through the aqueous solution medium. The basis for this assumption is that this provides one bound for the extent of chemical interactions within the WP.
- 3.5 It is assumed that 25 \degree C thermodynamic data can be used for the calculations. The bases of this assumption are two-fold. First, the initial breach and filling of a WP is unlikely to occur before $10⁴$ years (Ref. 4, Figure C-12), when the WP contents have cooled to \leq 50 °C (Ref. 8, Figures 3-22 and 3-24). Second, the assumption is conservative, with respect to loss of the Gd, the criticality control material. Gd carbonates and phosphates are likely to be the solubility-limiting solids for the Gd. Since the solubilities of solid carbonates and $GdPO_4\text{-}H_2O$ are temperature-independent or retrograde (Ref. 9, Tables IV and V), use of the lower-temperature database is likely to be conservative.
- 3.6 In most calculations, it is assumed that chromium and molybdenum will oxidize fully to chromate (or dichromnate) and molybdate, respectively. The first basis of the assumption is the body of available thermodynamic data (the dataO file in the accompanying electronic media, Ref. 10), which indicates that in the presence of air the chromium and molybdenum would both oxidize to the VI valence state. Laboratory observation of the corrosion of Cr- and Mo-containing steels and alloys, however, indicates that any such oxidation would be extremely slow. In fact, oxidation to the VI state may not occur at a significant rate with respect to the time frame of interest, or there may exist stable Cr(III) or Cr(VI) solids (not present in the EQ3/6 thermodynamic database) that substantially lower aqueous Cr concentration. For the present analyses, the assumption is made that over the times of concern the oxidation will occur. The second basis of the assumption is that it is conservative with respect to solubility of $GdOHCO₃$ (the expected solubilitycontrolling phase for Gd); extreme acidification of the water will enhance solubility (Ref. 5, Section 5.3) and transport of Gd out of the WP, thereby separating it preferentially from fissile material.
- 3.7 It is assumed that the CRM (the inner barrier) of the WP will react so slowly with the infiltrating water (and the water already in the WP) as to have negligible effect on the chemistry. The bases for this assumption consist of the facts that the CRM is fabricated from Alloy 22 (see nomenclature in Section 5.1.1), which corrodes very slowly compared (1) to other reactants in the WP and (2) to the rate at which soluble corrosion products will likely be flushed out of the WP.

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- 3.8 In most calculations, is assumed that gases in the solution in the WP will remain in equilibrium with the ambient atmosphere outside the WP. In other words, it is assumed that there is sufficient contact with the gas phase in the repository to maintain equilibrium with the $CO₂$ and $O₂$ present, whether or not this be the normal atmosphere in open air or rock gas that seeps out of the adjacent tuff. Under these conditions, the partial pressure of CO₂ exerts important controls on the pH and carbonate concentration in the solution and hence on the solubility of uranium, gadolinium and other elements. The basis of this assumption is that it is consistent with the approach taken in the viability assessment (VA) (Ref.. 11, Figure 4-27), is simple, and generally causes higher losses of gadolinium at high pH, and is thus conservative. A sensitivity study on the effects of $fCO₂$ is given in Section 5.3.2, and calculations with variations in oxygen fugacity are presented in Sections 6.1 and 6.2.
- 3.9 It is assumed that precipitated solids that are deposited remain in place, and are not mechanically eroded or entrained as colloids in the advected water. The basis for this assumption is that it conservatively maximizes the size of potential deposits of fissile material inside the WP.
- 3.10 It is assumed that the corrosion rates used in this calculation encompass rates for microbially assisted degradation, and that the degradation rates will not be controlled principally by bacteria. The bases for this assumption are (1) steel corrosion rates measured under environmental conditions inherently include exposure to bacteria, and (2) the lack of organic nutrients available for bacterial corrosion will limit the involvement of bacteria. It is assumed that bacteria act as catalysts, particularly for processes such as the reduction of sulfate, but this catalytic effect is not expected to change significantly the types of solids formed in the WP.
- 3.11 It is assumed that sufficient decay heat is retained within the WP over times of interest to cause convective circulation and mixing of the water inside the WP. The basis for this assumption is the analysis in Reference 12 (Att. VI).
- 3.12 It is assumed that the rate of entry of water into, as well as the rate of egress from, a WP is equal to the rate at which water drips onto the WP. The basis for this assumption is that for most of the time frame of interest, i.e., long after the corrosion barriers become largely degraded, it is more reasonable to assume that all or most of the drip will enter the degraded WP than to assume that a significant portion will instead be diverted around the remains. However, the calculations include scenarios with very low drip rates, which effectively simulate diversion of the bulk of the water striking the WP.
- 3.13 It is assumed that the most insoluble solids for a fissile radionuclide will form, i.e., equilibrium will be reached. The basis for this assumption is conservatism; the highest chance of internal criticality occurs when the fissile solids have lowest solubility, and are thus retained in the WP.

- 3.14 A number of minor assumptions have been made about the geometry of the Pu-ceramic WP. The bases of these assumptions are outlined and referenced in the spreadsheet Puceram.xls (Ref. 10), and are also discussed in Section 5.1. The assumptions about WP geometry are always intended to obtain the greatest accuracy in the representation, and where inadequate information is available to choose among possible WP designs, the choice that appears to lead to greatest conservatism is made.
- *3.15* For any WP components that were described as "304" stainless steel, without indication of the carbon grade, the alloy was assumed to be the low-carbon equivalent (see Section 5.1.1 for nomenclature). The basis of this assumption is that, in general, the carbon in the steel is totally insignificant compared to the carbon supplied by the fixed $CO₂$ fugacity of the EQ3/6 calculation, and to the constant influx of carbonate via the incoming water.
- 3.16 It is assumed that the thermodynamic behavior of hafnium (Hf) can be treated as if it were zirconium (Zr). The basis of this assumption is the extreme similarity of the chemical behaviors of the two elements (Ref. 13, p. 272). Thermodynamic data for many important Hf solids and aqueous species are lacking, thus Zr was substituted for Hf in the calculation.
- 3.17 It is assumed that the decay of 239 Pu to 235 U can be conservatively approximated by a simple exponential correction to the reported amounts of solids in the EO6 runs, after completion of each run. EQ6 currently has no built-in capability to handle radioactive decay. The basis of this assumption is that it is conservative for internal criticality since Pu solids are generally less soluble than U solids. The assumption causes an overestimate of the amount of 235U remaining in the WP with time. The exponential correction is not made in the tables and figures presented in this document, but is left up to the user of the calculation results. For the calculations with significant Gd loss, the loss generally occur within two half lives of 239 Pu, so that the correction is less than a factor of four.

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4. USE OF COMPUTER SOFTWARE AND MODELS

This section describes the computer software used to carry out the calculation.

EQ3/6 Software Package-The EQ3/6 software package originated in the mid-1970s at Northwestern University (Ref. 14). Since 1978, Lawrence Livermore National Laboratory (LLNL) has been responsible for maintenance of EQ3/6. The software has most recently been maintained under the sponsorship of the Civilian Radioactive Waste Management Program of the United States Department of Energy (DOE). The major components of the EQ3/6 package include: EQ3NR, a speciation-solubility code; EQ6, a reaction-path code, which simulates water/rock interaction or fluid mixing in either a pure reaction progress mode or a time mode; EQPT, a data file preprocessor; EQLIB, a supporting software library; and various supporting thermodynamic data files. The software deals with the concepts of the thermodynamic equilibrium, thermodynamic disequilibrium, and reaction kinetics. The supporting data files contain both standard state and activity coefficient-related data. Most of the data files support the use of the Davies or B-dot equations for the activity coefficients; two others support the use of Pitzer's equations. The temperature range of the thermodynamic data on the data files varies from 25 °C only for some species to a full range of 0-300 °C for others. EQPT takes a formatted data file (a "data0" file) and writes an unformatted "data1" file, which is actually the form read by EQ3NR and EQ6. EQ3NR is useful for analyzing groundwater chemistry data, calculating solubility limits, and determining whether certain reactions are in states of partial equilibrium or disequilibrium. EQ3NR is also required to initialize an EQ6 calculation..

EQ6 simulates the consequences of interactions among an aqueous solution and a set of degrading reactants. This code operates both in a pure reaction progress frame and in a time frame. In a time frame calculation, the user specifies rate laws for the progress of the irreversible reactions. Otherwise, only relative rates are specified. EQ3NR and EQ6 use a hybrid Newton-Raphson technique to make thermodynamic calculations. This method is supported by a set of algorithms that create and optimize starting values. EQ6 uses an ordinary differential equation (ODE) integration algorithm to solve rate equations in time mode. The codes in the EQ3/6 package are written in FORTRAN 77 and have been developed to run on personal computers (PCs) running Microsoft DOS or Windows, and under the UNIX operating system. Further information on the codes of the EQ3/6 package is provided in References 14, 15, 16, and 17.

Solid-Centered Flow-through Mode-EQ6 Version 7.2b, as distributed by LLNL, does not contain an SCFT mode. To add this mode, it is necessary to change the eq6.for source code, and recompile the source. However, by using a variant of the "special reactant" type built into EQ6, it is possible to add the functionality of SCFT mode in a very simple and straightforward manner. This mode was added to EQ6 per Software Change Request (SCR) LSCR198 (Ref. 18), and the Software Qualification Report (SQR) for Media Number 30084-M04-001.

The new mode is induced with a "special-special" reactant. The EQ6 input file nomenclature for this new mode is jcode=5; in the Daveler format, it is indicated by the reactant type

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DISPLACER. The jcode=5 is immediately trapped and converted to jcode=2, and a flag is set to indicate the existence of the DISPLACER reactant. Apart from the input trapping, the distinction between the DISPLACER and SPECIAL reactants is seen only in one 9-line block of the EQ6 FORTRAN source code (in the reacts subroutine), where the total moles of elements in the rock plus water system (mte array) is adjusted by adding in the DISPLACER reactant, and subtracting out a commensurate amount of the total aqueous elements (mteaq array).

4.1 SOFTWARE APPROVED FOR QUALITY ASSURANCE (QA) WORK

The addendum to EQ6 was approved for QA work by the M&O (M&O identifier CSCI: UCRL-MA-1 10662 V 7.2b; final approval in Ref. 18). An installation and test report (Ref. 19) was written and submitted to Software Configuration Management (SCM), and the proper installation was verified before the runs described in this calculation were made. The implementation of the SCFT mode is covered by SCR LSCRl98, and the SQR for Media Number 30084-M04-001. The SCFT addendum was installed on three of the Central Processing Units (CPUs), and the installation and test reports were filed and returned to SCM before the calculations were run. All calculations were run under the Windows 95 operating system. In this study *EQ316* was used to provide the following:

- * A general overview of the expected chemical reactions.
- * The degradation products from corrosion of the waste forms and canisters.
- An indication of the minerals, and their amounts, likely to precipitate within the WP.

The programs have not been used outside the range of parameters for which they have been verified. Some runs with high HLW glass degradation rates and low flush rates had temporary excursions to high ionic strength, above the range for which the B-Dot ionic strength correction has been validated. However, the excursions were temporary, and they occurred in the early part of the calculation, which principally sets up the mass balance for the completion of the run. Results from these high-ionic strength runs are included for conservatism since such runs likely overestimate Gd loss. The *EQ3/6* calculations reported in this document used Version 7.2b of the code, which is appropriate for the application, and were executed on Pentium series (including "Pentium II") PCs.

The EQ316 package has been verified by its present custodian, LLNL. The source codes were obtained from SCM in accordance with the Office of Civilian Radioactive Waste Management AP-SI.1Q Revision 1, ICN 0 procedure. The code was installed on the Pentium and Pentium II PCs according to an M&O-approved Installation and Test procedure (Ref. 19).

4.2 SOFTWARE ROUTINES

Spreadsheet analyses were performed with the Office 97 version of Microsoft Excel, installed on a PC running Microsoft Windows 95. The specific spreadsheets, used for results reported in this document, are included in the electronic media (Ref. 10).

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4.3 MODELS

None used.

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S. CALCULATION

The calculations begin with selection of data for compositions, amounts, surface areas, and reaction rates of the various components of Pu-ceramic WPs. These quantities are recalculated to the form required for entry into EQ6. For example, weight percentages of elements or component oxides are converted to mole fractions of elements; degradation rates in pm/y are converted to mole/ (cm^2-s) , etc. Spreadsheets (Ref. 10) provide details of these calculations. The final part of the input to EQ6 consists of the composition of incoming water together with a rate of influx to the WP that corresponds to suitably chosen percolation rates into a drift and drip rate into a WP (Section 5.1.1.3). The EQ6 output provides the results of the chemical degradation calculations for the WP, or components thereof. In selected cases, the degradation of the WP is divided into phases, e.g., degradation of HLW glass before breach of the stainless steel cans, and subsequent exposure of the Pu-ceramic to water. The results include the compositions and amounts of solid products and of elements dissolved in solution. Details of the results are presented in Section 6.

In all tables from this document, the number of digits reported does not necessarily reflect the accuracy or precision of the calculation. In most tables, two to four digits after the decimal place have been retained, to prevent round-off errors in subsequent calculations.

Some "preliminary" and "prototype" drawings (Refs. 22 and 23) have been used to calculate geometric surface areas and masses of WP components, as inputs for EQ6. However, the results of the EQ6 runs should change little due to changes in package design. The surface areas are used only as part of the rate equation, such that the total rate is: (surface area).(EQ6 rkl parameter). The variation in rkl is far larger than the variation in component surface areas that may results from design differences. Geometric surface areas of HLW glass, ceramic, and stainless steel are not expected to vary by more than 50% among WP designs (as judged by the comparison of areas used in this calculation, versus Ref. 5). In contrast, the rkl are highly uncertain and are varied, as part of this calculation, over 1 to 3 orders of magnitude. Therefore, any variations in the geometric surface areas are trivial and are subsumed in the parametric rate study built into this calculation. Similar arguments can be made about the compositions and masses of HLW glass and steel; Section 5.3.3 (Table 5-9) shows that large variations in HLW glass composition have relatively small effects.

5.1 CALCULATION INPUTS

5.1.1 WP Materials and Performance Parameters

This section provides a brief overview of the physical and chemical characteristics of Pu-ceramic WPs, and describes how the WP is represented in the EQ6 inputs. The conversion of the WP physical description, into parameters suitable for the EQ6 input files, is performed by the spreadsheet Pu-ceram.xls (Ref. 10). Additional details of the description may be found in References 1, 2, 20, 21, 22, and 23, and the references cited therein.

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Material nomenclature used throughout this document includes: SB-575 N06022 (hereafter referred to as Alloy 22), UNS N06625 and SA-240 S30403 (hereafter referred to as 304L), and SA-516 (hereafter referred to as A516).

5.1.1.1 Physical and Chemical Form of Pu-Ceramic **WPs**

It is convenient to consider the Pu-ceramic WP as several structural components, as illustrated in Figures *5-1* and 5-2 (the latter from Ref. 2); specifically:

- (1) The outer shell, consisting of the CAM and the inner Alloy 22 CRM.
- (2) The "outer web", a carbon steel (A516) structural basket designed to hold the HLW glass pour canisters (GPCs) in place.
- (3) The 5 GPCs, which consist of 304L canisters filled with solidified HLW glass.
- (4) The 28 cans that contain the Pu-ceramic disks; the 7 magazines that hold the cans; and the rack that holds the 7 magazines in place within the GPC. These are all constructed of 304L stainless steel. After the rack-and-magazine assembly is emplaced within the GPC, molten HLW glass will be poured into the GPC, encasing the internal components.
- (5) The Pu-ceramic disks. Each can holds 20 ceramic disks, for a total of 560 = (20 disks).(4cans).(7 magazines) disks per GPC, or 2800 disks in all 5 GPCs.

The details of each component are given in spreadsheet Pu-ceram.xls (Ref. 10).

Figure 5-1. Cross-section of Pu-ceramic Waste Package

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Figure 5-2. Lengthwise Section of GPC, with Magazine and Rack Assembly

Table 5-1 provides a summary of the compositions of the principal alloys used in the calculations, along with reasonable high and average degradation rates. For a comparable specific surface area, the carbon steel is expected to degrade much more rapidly than the stainless steels (304L). In addition, the stainless steels contain significant amounts of Cr and Mo, and under the assumption of complete oxidation (Assumption 3.6), should produce more acid, per unit volume, than the carbon steel.

NOTES: ^aCompositions from Reference 24, pages 10 and 17 (Sections 5.2 and 5.4). Details given in Pu-ceram.xls, sheet 'Compositions'.

Degradation rates from Reference 25, pages 11-13 (304L), and Reference 26, pages 5-44 to 5-55, Figures 5.4-3, 5.4-4, and 5.4-5 (carbon steel). Conversions performed in Pu-ceram.xls, sheet "Rates". For 1 mole \approx 100 g.

Table 5-2 gives the molar composition of the HLW glass used in the calculations (Ref. 27, Att I, p. I-7). The actual HLW glass composition used in the GPCs may vary significantly from these values since the sources of the HLW glass and melting processes are not currently fixed. For example, compositions proposed for Savannah River HLW glass vary by a factor of \sim 6 in U₃O₈ content, from 0.53 to 3.16 wt% (Ref. 28, Table 3.3.8). The silica and alkali contents (Na, Li, and K) of the HLW glass have perhaps the most significant bearing on EQ6 calculations. The amount of silica in the HLW glass strongly controls the amount of clay that forms in the WP, and the silica activity controls the presence of insoluble uranium phases, such as soddyite $((UO₂)₂SiO₄·2H₂O)$. The alkali content can induce pH to rise in the early stages of the EO6 run,

as HLW glass degrades. The Si and alkali contents in Table 5-2 are typical for proposed DOE HLW glasses (Ref. 28, Table 3.3.8). The effects of such compositional variations are examined in the sensitivity studies discussed in Section 5.3 of this document.

Rates for HLW glass degradation were taken from Reference 26 (Figure 6.2-5) and normalized in spreadsheet Pu-ceram.xls, sheet "Rates" (Ref. 10). The high rate corresponds approximately to pH 9 at 70 °C, and the low rate to pH 8 at 25 °C. In the transmuted composition of the HLW glass, the content of an element in the HLW glass is the sum of its isotopes listed in HLW glass composition in Table 5-2.

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Table 5-2. HLW Glass Composition and Rates

NOTES: $*^{238}$ Pu decayed to $*^{234}U$; $*^{41}$ Pu decayed to $*^{237}$ Np; Ag, Cs, Zn, and Th removed because of low quantity, chemical insignificance, and/or lack of reliable thermodynamic data.

Pasic Reference 27, Attachment I, page I-7.

 $^{\circ}$ Normalized to 1 mole \equiv 100 g. Conversions performed in spreadsheet Pu-ceram.xls, sheet "Compositions."

Reference 26, page 6-5 and Figure 6.2-5; fracture factor = 30, conversions performed in Pu-ceram.xls, sheet "Rates".

Table 5-3 summarizes the composition of the "baseline" ceramic used for all the calculations reported herein. The three rates chosen for the calculations are based on Slide 32 of Reference 21; these rate data were the best available at the time the study began. Very recently, LLNL has provided updated rates that are a factor of 3 to 10 lower for the same conditions (Figure 6.1 of

Ref. 2). The new LLNL rate data are shown in Figure 5-3. However, LLNL has also determined that the specific surface areas of the Pu-ceramic samples were \sim 3 to 10 times higher than previously thought, due in part to the sample porosity. Since the total reaction rate for the ceramic is the product of surface area and the fundamental rate constant, the two new observations by LLNL (of lower rates and higher specific surface area) tend to cancel out. Hence this study retains the original range of rates, given in Table 5-3, spanning nearly three orders of magnitude.

Table 5-3. Composition^a and Degradation Rates^b of Pu-ceramic

NOTES: ^eReference 1, Table 3.1.

Reference 21, slide 32.

'Replaced by Zr In EQ6 runs, then converted back to Hf for mass calculations. $\overline{\text{``For 1 mole speedal reactant}} = 100 \text{ g}.$

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Figure 5-3. Normalized Rates for Single-pass Flow Test

5.1.1.2 Chemical Composition of Incoming Water

It was assumed that the incoming water would have a composition approximating that of the J-13 well water (Assumptions 3.1 and 3.3). Table *5-4* contains the EQ3NR constraints for the incoming water (based on EQ3NR input file jl3noc3O.3i), and the EQ6 elemental molal composition used for this calculation. The basic composition is taken from Harrar et al. (Ref. 6, Tables 4.1 and 4.2) and represents the best estimates of a committee assembled specifically to review the composition of J-13 well water. For those elements listed as "trace" constituents in Table 5-4, an arbitrary small molality (10^{-16}) was added to assure numerical stability. Section 5.3.2 of the current document contains a discussion on the effect of varying the incoming water compositions.

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Table 5-4. EQ3NR and EQ6 Compositions of Incoming Water

NOTE: ^{*}Aqueous

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5.1.1.3 Drip Rate of Incoming Water

It is assumed (Assumption 3.12) that the drip rate onto a WP is the same as the rate at which water flows through the WP. Four drip rates were used: 0.0015 , 0.015 , 0.15 , and 0.5 m³/year per WP. The justification for these rates is given in Section 5.1.1.3 of Reference 7.

5.1.1.4 Densities and Molecular Weights of Solids

The current SCFT version of EQ6 (Section 4.2 of this document) automatically calculates total mineral volumes, using the molar volume values (VOPrTr) embedded in the EQ3/6 data0.nuc.R8d file (included in the electronic media, Ref. 10). Molar volumes are not available for many solids that are in the EQ6 database; such solids are flagged in the database with fictitious molar volumes of "500.000." However, the dataO file contains valid molar volume entries for the solids that comprise the vast bulk of the volume in the current study. The new SCFT version of EQ6 automatically calculates the total volume of solids, and outputs the volumes in the *.elem min.txt and *.elem tot.txt files (where the asterisk denotes the root file name in Table 6-1); the placeholder "500.000" entries are ignored.

5.1.1.5 Atomic Weights

Atomic weights were taken from References 29 and 30, and are listed in spreadsheet Pu-ceram.xls, sheet "AtomWts" (Ref. 10).

5.2 DATA CONVERSION

The data presented in Section 5.1 are transformed into EQ3/6 format by converting mass fractions into mole fractions; normalizing surface areas, volumes, and moles to I liter reactive water in the system; and converting rates to mole/ cm^2 -s). Most of these conversions are straightforward and are perforned in the spreadsheets, which are included in the electronic media for this document (Ref. 10). Spreadsheet Pu-ceram.xls (1) calculates volumes and areas of WP components (sheets "Magazine,Rack,Can,Disk" and "GPC & Outer Web"); (2) converts raw percent compositions to normalized molar compositions (sheet "Compositions"); (3) converts experimental reaction rates (in units such as μ m/y) to EQ6 units (mole/(cm²·s)), and converts drip rates in m^3 /y to normalized mole/(cm² \cdot s) for the EQ6 system (sheet "Rates"); and (4) calculates the system void volume, normalizes the WP component volumes and areas to the EQ6 system containing I liter void space, and calculates the EQ6 moles of each component from the normalized volumes, densities, and molecular weights (sheet "Void&Norm").

In Section 6, the "moles" in plots are for the sub-sampled (I liter fluid) EQ6 system; all EQ6 solid special reactants (i.e., WP steels, HLW glass, and ceramic) are converted to 100 g/mole. To obtain the actual kg mass of a reactant in the WP, multiply *plotted* moles by 459.3965 = (0.1 kg/mole "special reactant")-(4593.965 liters void space in initial system). To obtain the actual mass of a mineral in the WP, multiply *plotted* moles by (mineral molecular weight)-(0.001 kg/g).(4593.965 liters void space in initial system). The mineral molecular weights are given in

the EQ3/6 database (dataO.nuc.R8d), which is included in the electronic file distribution for this document (Ref. 10). (Some of the older calculations used the data0.nuc.R8a (included in the Ref. 10) and dat0.nuc.R8p databases. The "R8a" version is identical to R8d, except the latter contains a molar volume for rhabdophane; since the latter phase is a trivial fraction of the total solids volume, this difference is insignificant. The "R8p" is identical to Rga, except is contains data for a hypothetical Pu-phosphate, exactly as described in Reference 7 (p. 51). Since the hypothetical Pu-phosphate solid never precipitated in these runs, there was no functional difference among the three databases.) Note that 4593.965 liters is from the variable WOIDS in sheet "Void&Norm" of spreadsheet Pu-ceram.xls (VVOIDS is given in $cm³$, and must be divided by 1000 $cm³/l$ to obtain 4593.965).

-5.3 PREPARATORY STUDIES: UPDATED COMPOSITIONS AND THERMOCHEMICAL DATA

Since the issue of the original study of Pu-ceramic degradation geochemistry (Ref. 5), new information became available on three topics: (1) the thermochemistry of the Gd carbonate system; (2) the projected long-term chemistry of the repository waters; and (3) the' chemistry of the HLW glass produced by Savannah River Site (SRS). Before the matrix of calculations described in Section *5.4* was designed, three sensitivity studies had been performed'to determine how the new data might change results of EQ6 calculations. These sensitivity studies used the input files for Runs 4 and 6 from Reference 5 (Table 5.3-1); Run 4 produced the highest Gd loss in that prior study. All input and spreadsheet files referred to in this section can be found on the electronic media accompanying this calculation (Ref. 10).

5.3.1 Effects of Gd Carbonate Thermochemistry

The previous Pu-ceramic calculations (Ref. 5) used a custom version of the Swedish Nuclear Fuel and Waste Management Co. (SKB) thermodynamic database (Ref. 31), prepared by LLNL. The Pu-ceramic waste forms contain Gd, but little or no phosphate; the encasing HLW glass contains some phosphate, but not enough to precipitate all the Gd in the ceramic. Thus, when these waste forms degrade in water, Gd phosphate may not be the solubility-limiting phase. When the SKB database was used in the previous study, the calculations indicated that $GdOHCO₃$ would be the solubility-limiting phase for Gd (Ref. 5, Figures 5.3.4-2 and 5.3.4-5). However, the thermodynamic data for this phase were originally estimated from the data for $NdOHCO₃$, and there was substantial uncertainty in the SKB data for the hydroxyl and carbonate aqueous complexes of Gd.

Recently, Weger et al. (Ref. 32) performed experiments to determine the solubility of three solid Gd carbonates and one hydroxide, and estimated the formation constants for several dissolved Gd carbonate and hydroxyl complexes. A version of the SKB database was customized to include the new Weger et al. data; this new database is hereafter referred to as the Weger database. Runs 4 and 6 reported in Reference 5 (Table 5.3-1) were repeated using this new database. In brief, the new Weger et al. data appear to have no significant impact and produce approximately the same calculated Gd losses, as were obtained with the original SKB database.

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53.1.1 Entry and Testing of the New Data

When thermodynamic data are added to the EQ316 databases, it is important to achieve the highest possible consistency. Large errors can occur if the logarithms of stability constants *(log K)* are derived from reaction free energies (ΔG^{mn}) reported in the literature, without correcting for the differences in standard state free energies G_f^0 . In most experiments, the measured quantities are actually closer to the *log Ks*. While $G_f^p/(RT)$ values are used in the formal statement of the statistical sum $f(\chi)$ (e.g., Equation 8 in Ref. 32), quantities of the form:

$$
-\ln K = (G_f^0 / RT)_{\text{NaCd(CO}_3)_{2} \cdot 6H_2O} - (G_f^0 / RT)_{\text{Na}^+} - (G_f^0 / RT)_{\text{Ga}^+}
$$
\n
$$
-2 \cdot (G_f^0 / RT)_{\text{CO}^2} - 6 \cdot (G_f^0 / RT)_{\text{H},O}
$$
\n(Eq. 1)

are actually determined in the minimization of $f(\gamma)$.

For the current calculations, the $G_f^0/(RT)$ values reported in Table 2 of Reference 32 were used to calculate *log K*'s in terms of the basis species used in EQ3/6 (H⁺, H₂O, HCO₃⁻, Na⁺, and Gd^{3+}).

The spreadsheet GdCO3 Weger Rai 2 EQ6.xls was used to perform thermochemical conversions. As a cross-check, the $log K$ s for formation of Gd(OH)₃⁰ (aq), Gd₂(CO₃)₃, and $NaGd(CO₃)₂·6H₂O$ were calculated in terms of the basis species used in Reference 32, and compared with the values in Table 4 of Reference 32. The results agree to the number of significant digits reported by Weger et al. Table *5-5,* derived from GdCO3_Weger_Rai_2_EQ6.xls, details the conversion of the Weger et al. data into EQ3/6 format. The value of the gas constant, 1.987216 cal/(mole-K), was taken from Reference 30 (p. 57). The Weger et al. speciation does not use all the hydroxyl and carbonate complexes included in the SKB database; these additional SKB complexes were removed from the Weger database for the purposes of testing.

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Table 5-5. Conversion of Weger et al. Data to EQ3/6 Basis

NOTES: "Aqueous

To test the Weger database, EQ6 runs were used to simulate the solubility experiments for Gd₂(CO₃)₃ (Ref. 32, Figure 3 and Table A.3). The EQ3NR and EQ6 input files are included in the electronic attachment (Ref. 10), as files weger.3i, weger2s.6i, and weger3s.6i. Figure 5-4 shows the results of the comparison (calculated via weg fig3.xls). The discrete symbols represent the experimental data (Ref. 32, Table A.3). The agreement between the EQ6 calculation and the experiments is good, given the spread in experimental results and the inherent uncertainty in the EQ6 "B-Dot" ionic strength corrections. The B-Dot method is normally considered accurate only for ionic strength ≤ 1 , and the Weger et al. experiments reached ionic strengths $>$ 3.

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Figure 5-4. Comparison of EQ6 Calculations with Weger et al. Experiments

5.3.1.2 Comparison of Pu-Ceramic Runs with and without New Data

To test the consequences of using the Weger database, versus the SKB database, run numbers 4 and 6 from Reference 5 (Table 5.3-1) were repeated with both datasets. To ensure a uniform basis for comparison, both calculations were performed with the new addendum to EQ3/6 (described in Section 4.2 of this calculation).

Table 5-6 summarizes the net Gd losses for the two runs, using both databases. Figures 5-5 and 5-6 plot results for runs 4 and 6, respectively (the "moles" on the right are for the normalized EQ6 system, as discussed in Reference 5 (p. 24).

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Table 5-6. Percent Loss of Gadolinium for Entire WP

NOTE: Runs described in Reference 5, Table 5.3-1.

For old run 4 (Figure 5-5), the CO₂ fugacity is buffered to $\sim 10^{-2.5}$ bar, and the pH stays moderate (< 8); consequently, $Gd_2(CO_3)$ is stabilized relative to $GdOHCO_3$ for the calculation with the Weger database. For run 6 (Figure 5-6), the $CO₂$ pressure is lower (~10^{-3.5} bar), the average pH is higher; consequently, both databases yield GdOHCO₃ as the solubility determining solid.

5.3.1.3 Summary of Results for Thermodynamic Data Sensitivity Study

It is apparent from Table 5-6 and Figures 5-5 and 5-6 that the Weger et al. data do not significantly alter the calculation results, and use of the SKB database is slightly conservative for systems where the solid Gd phases are carbonates.

NOTE: Runs described in Reference 5, Table 5.3-1.

5.3.2 **Chemistry of Incoming Water**

There were several motivations for this study. First, the incoming water composition used in previous WPO studies (Ref. 5, Assumptions 3.8 and 3.12) was derived by a somewhat involved process, and resulted in log_{10} O₂ and CO₂ fugacities of -0.6778 and -2.5390 , respectively. These specific and high-precision values made comparison against other studies somewhat difficult (in the literature and the VA, it is more common to use "round" values like -0.7 and

 -3). Second, the previously assumed CO₂ fugacity is not consistent with the value log_{10} fCO₂ = -3.0 assumed in the VA calculations (Ref. 11, Figure 4-27). The source of this discrepancy is subtle: the author of Reference 33 (Assumption 3.8) attempted to obtain the highest consistency between log_{10} fCO₂ and reported analyses for J-13 well water (Ref. 6, Tables 4.1 and 4.2). However, the J-13 well water is drawn from below the water table, under a substantial head; thus true J-13 well water is probably in equilibrium with a higher $CO₂$ fugacity than the fracture and pore waters at the repository level. Since the solubility of actinides and gadolinium can depend very strongly on the effective $CO₂$ pressure, it is appropriate to use the VA value as the recommended mean. Third, to make the reported J-13 well water analyses consistent, the analysis used in Reference 33 (Assumption 3.8) required adjustments to the reported Na⁺ to maintain charge balance; these adjustments are difficult to justify since Na is generally analyzed to high accuracy. Fourth, a slight error was discovered in the original data reduction for the J-13 well water analyses, used for Reference 33; the concentration of Fe in J-13 water was accidentally assigned to the concentration of F. This very minor error has a small effect on pH, but for the sake of consistency, it was desirable to recalculate the water compositions. Fifth, it is desirable to have standard water compositions with trace elements added in sufficient quantity to prevent numerical instabilities. It was recently discovered that some species (e.g., MoO₄⁻) require minimal concentrations, $\sim 10^{-16}$ molal, compared to $\sim 10^{-22}$ molal used previously.

The sensitivity analysis for incoming water composition consisted of several steps:

(1) Six initial EQ3NR input files were created (jl3noc25.3i, jl3noc3O.3i, jl3noc35.3i, jl3cal25.3i, jl3cal30.3i, and jl3cal35.3i). The files either held Ca concentrations at saturation with calcite (those files have "cal", for calcite, as the $4th$, $5th$, and $6th$ letters), or allowed Ca⁺⁺ to be fixed at the average value reported in Table 4.2 of Reference 6 (those files have "noc" for "no calcite" as the middle letters). The log_{10} fCO₂ values were set at -2.5 , -3.0 , or -3.5 (corresponding to 25, 30, and 35, as the last two digits of the file names), to encompass the VA mean value, the approximate value suggested in Reference 33, and the normal ambient value (Ref. 34, p. F-210). The concentrations of Fe^{2+} , Al^{3+} , and Mn^{2+} were set to equilibrium with goethite, diaspore, and pyrolusite, respectively.

(2) Six EQ6 files, analogous to Cerd2N0 51.6i in run 4 of Reference 5 (p. 23, Table 5.3-1), were created by substituting in the water compositions from the 3p pickup files created in step (1) above; these files were named r4noc25.6i, r4noc30.6i, r4noc35.6i, r4cal25.6i, r4cal3O.6i, and r4cal35.6i. These files cover "stage I" of the run, in which a high drip rate of $0.5 \text{ m}^3/\text{day}$ is assumed to enter the WP, reacting with the steels and HLW glass, but not the ceramic. The runs terminate after -3800 years, when the pH has stabilized and all the HLW glass (but not all the steel) is degraded. These files were formatted to use the new EQ6 SCFT mode described in Section 4.2. A seventh file, r4base.6i, was created by reformatting Cerd2N0_5I.6i to be compatible with the new SCFT mode, and was used for base-case comparisons. The new SCFT method dynamically adjusts step size, and produces results slightly different from the allpost/nxtinput method described in Reference 5, pp. 11-14.

(3) Six EQ6 files, analogous to Cerd2W0 0015I.6i in run 4 of Reference 5 (p. 23, Table 5.3-1), were created by picking up the results of the runs described in step (2), adjusting the ending time, adding in the Pu-ceramic as a reactant, and adjusting the drip rate to 0.0015 m^3 /y. These runs constitute the second stage of the calculation, out to -6.10^5 years. The file names are r4noc25 .6i, r4noc30 .6i, r4noc35 .6i, r4cal25 .6i, r4cal30 .6i, and r4cal35 .6i, with the obvious correspondence to the files described in step (2). A seventh file, r4base .6i, was made from Cerd2W0 0015I.6i, formatted for consistency with the new SCFT method, to act as a base case (unchanged J-13 well water composition).

(4) The Gd, Pu, and U % loss were calculated from the elem tot.txt files generated by step (3).

Table 5-7 shows the comparison of Gd, Pu, and U losses for the 7 runs. Overall, the differences are small. It is notable that the highest Gd losses occur in the samples with the lowest overall $CO₂$ fugacity, and that these runs have the highest average pH. However, the runs with log_{10} $fCO_2 = -3.5$ have lower dissolved H₂CO₃ and HCO₃⁻, and are less able to buffer the system pH during the degradation of metals.

Table 5-7. Gd, Pu, and U % Losses for Run 4, Using Varied Incoming Water Compositions

5.3.3 Effects of HLW Glass Composition Variations

A sensitivity study, investigating effects of HLW glass composition variations, was undertaken for two reasons. First, the previous study of Pu-ceramic degradation used a single composition to represent "typical" HLW glasses produced at the SRS (Ref. 5, Table 5.1.1.1-2, p. 16). A very similar, standardized HLW glass composition was recently developed in Reference 27 (Att. I, p. I-7). Ultimately, these two compositions trace to a 1987 report, issued before SRS had finalized its HLW glass production process. In addition, other reports suggested large composition ranges for SRS HLW glass (Ref. 28, Table 3.3.8). Second, in the previous calculation, all phosphorus was eliminated from the idealized HLW glass composition. The phosphorus was removed from the previous calculation, to prevent EQ6 from precipitating low-solubility Gd phosphates; the removal was conservative since there was uncertainty about the accuracy of HLW glass phosphorus analyses, and elimination of phosphorus would lead to greater Gd losses from the WP.

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There is a mechanism to obtain current HLW glass compositions, and these current compositions show phosphorous is a real and predictable component of the SRS HLW glasses. The "Waste Acceptance Product Specifications for Vitrified High-Level Waste Forms", or WAPS, specifies (Ref. 35, p. 7) that the producer shall project the composition in the WQR (Waste Qualification Report). Reference 36 (Tables 1 through 3) is such a report, and gives seven extreme compositions expected for SRS HLW glass. The most widely varying compositions from that report are given in Table *5-8* (converted into moles), along with the "base" composition used for HLW glass in previous Pu-ceramic calculations (Ref. 5, Table 5.1.1.1-2).

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Table 5-8. Elements (in moles) in HLW Glass, Normalized to 100 g/mole

NOTES: References for HLW glass: Spreadsheet: glass2sensitivity.xis; Base from Reference 5, Table 5.1.1.1-2, and associated spreadsheets; all others from Reference 36, Tables 1 to 3.

The recent glass analysis given by Bibler et al. (Ref. 37, Table 1) is akin to batch1 in Table 5-8 above. However, the analyses in Reference 37 are for tank 51, which is the oldest tank at SRS; this tank was filled with residue from an early process, and may not be typical of the overall SRS production.

To test the effects of the variability in HLW glass composition, Run 4 from Reference 5 (Table 5.3-1) was redone with each of the HLW glass compositions in Table 5-8. This run had produced the highest Gd loss in the previous study. The stagel calculation involved exposure of just the steel components and HLW glass at a high drip rate $(0.5 \text{ m}^3/\text{year})$, and the stage2 part of the calculation added the Pu-ceramic as a reactant and dropped the drip rate (water entry into the

WP) to 0.0015 m³/year. The new SCFT mode (Section 4.2) was used. (Note: when the new mode is used on a second-stage input file generated by the older mode described in Reference *5,* pages 11-13, the new mode calculates the same Gd loss [-14.8%], as found in Table 5.3-1 of Reference 5.) However, for the results in Table 5-9 of this document, it was necessary to redo the first-stage of the calculation as well, which yields a slightly lower overall loss for the Base case (-13.9%). The first stage calculation involves very rapid and dramatic changes in chemical composition, and the newer method provides a slightly more accurate solution to the differential equations in such circumstances.

From Table 5-9, it is apparent that the new HLW glass compositions from Reference 36 ('Blend" through "Purex') generate slightly lower Gd losses than does the Base case. The principal reason for the lower loss appears to be the small phosphate content for the newer compositions. The phosphate converts some of the Gd into insoluble $GdPO_4\cdot H_2O$, which lowers overall Gd loss in the system. Overall, the wide variations in the new HLW glass compositions appear to have little effect on the total calculated Gd loss. Other elements (particularly Pb) compete with Gd for the phosphate, so there is no simple correlation between phosphate content and Gd loss.

Table 5-9. Gadolinium Losses for Run 4, Using Different HLW Glass Compositions

NOTE: References for glass: Spreadsheets case4 tot.xls, glass2sensitivity.xls; Reference 36, Tables 1-3. Reference for Run 4: Reference 5, Table 5.3-1.

5.4 EQ6 CALCULATIONS

5.4.1 Scenarios Considered

The rationale for selection of EQ6 simulation scenarios is to provide conservative assessments of solubility and transport of criticality control materials (Gd and Hf, the neutron absorbers) and fissile materials (i.e., U and Pu compounds) in the WP. The proposed criticality control materials are incorporated into the ceramic. Upon degradation of the ceramic, the Hf is expected to form insoluble Hf oxides, which will remain in the WP. Degradation of the ceramic is also expected to yield varied amounts of (1) aqueous (dissolved) Gd; (2) a solid Gd carbonate (most likely GdOHCO₃); and (3) possibly GdPO₄·H₂O (Ref. 5, Section 5.3; and Ref. 7, Section 5.3.1). Gadolinium phosphate is sparingly soluble in neutral solutions, though the solubility does increase at low and high pH (Ref. 7, Section 5.3.1). Formation of $GdPQ₄·H₂O$ will be limited by

the availability of phosphate from degrading HLW glass. Dissolution of solid Gd carbonates and phosphates, at high pH, is also enhanced by dissolved carbonate. Uranium and plutonium are also quite soluble in the alkaline, carbonate-rich solutions produced when HLW glass degrades. Thus the matrix of EQ6 calculations should include scenarios that may yield high pH, particularly high pH with high dissolved carbonate.

However, solid GdOHCO₃ also dissolves readily under acid conditions (Ref. 5, Section 5.3). Low-pH conditions might occur if steel degrades separately from the HLW glass. To obtain sustained, low-pH conditions, it is generally necessary to break the degradation process into two stages. The first stage involves an early breach of the 304L stainless steel canisters holding the HLW glass, followed by fast degradation of the HLW glass and removal of the alkaline components during a period of relatively high drip rate. In the second stage, the 304L cans holding the ceramic are allowed to breach, exposing some portion of the Pu-U-Gd-Hf-ceramic to acid conditions. To keep the pH low, the drip rate must be reduced for the second stage. One difficulty with these two-stage runs is that they require very high HLW glass degradation rates, to ensure the HLW glass is degraded and leached before exposure of the ceramic. The two-stage runs are somewhat unrealistic, in that the 304L cans are thin, and unlikely to last for long periods after breach of the HLW glass canisters. A few two-stage runs are included in the matrix, just to test the sensitivity of the system to sustained, low-pH conditions capable of dissolving GdOHCO₃.

Thus, the reaction scenarios can be divided into two general categories. The first category comprises single-stage cases, in which all reactants (steels, HLW glass, and fissile materials) are *exposed* simultaneously to the water in the WP. Because the reaction rates of the materials in the WP may vary greatly, all materials do not necessarily *degrade* simultaneously. The second category comprises two-stage runs. In the first stage, the A516 outer web (basket) and the GPCs (HLW glass and 304L steel) are first exposed to water, until the HLW glass is completely degraded and its alkalinity largely flushed out of the system. The first stage is actually run twice; once up to approximately three times as long as is required to degrade the HLW glass, in order to locate the true pH minimum; and it is run a second time, just to the commencement of the lowpH plateau, to create an EQ6 pickup file for the second stage. This repetitive process ensures that the maximum acidity will be achieved in the second stage. In the second stage, the 304L cans, magazines, rack, and Pu-ceramic disks are added as reactants. The aim of the two-stage runs is to force a "conservative" condition of high acidity, by degrading the HLW glass rapidly, before all the acid-producing steel is degraded. The early HLW glass degradation and flushing requires very high HLW glass degradation rates; the total effective rate of the HLW glass is further increased by considering cracks as part of the total surface area. These high HLW glass degradation rates, if used with a slow flush rate, can produce unreasonably high ionic strengths (> 1); such high ionic strengths are beyond the applicability of EQ316's "B-Dot" corrections.

The primary cases for this study included 20 single-stage and 5 two-stage runs, with varied combinations of steel, HLW glass, and ceramic degradation rates and different water fluxes. These 25 cases used ambient gas fugacities of 0.2 bar for O_2 , and 10^{-3} bar for CO_2 . The CO_2 pressure used is consistent with the long-term ambient assumed for the VA (Ref. 11, Figure 4-

27). Tables 5-10 and 5-11 present a summary of EQ6 runs and corresponding degradation configurations for the ambient gas fugacities.

Table 5-10. Summary of Single-stage EQ6 Cases for Pu-ceramic WP
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Table 5-11. Summary of Multiple-stage EQ6 Cases for Pu-ceramic WP

In addition to the 25 cases reacted under the ambient gas-phase conditions, a number of cases were done under low-oxygen fugacity constraints to test sensitivity of reaction-paths to $fO₂$ (oxygen fugacity, or idealized partial pressure of oxygen). Table 5-12 summarizes the configurations for the sensitivity tests. Cases were selected on the basis of whether Gd loss was larger than 1% at the end of the reaction-path. In the test, the $log_{10}(f(O_2))$ is changed from -0.7 to -10 (Cases 26 through 30) and -15 bar (Cases 31 through 35). As will be shown later, the range of f02 used here is reasonable.

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Table 5-12. Summary of Sensitivity Tests for Selected EQ6 Cases for Pu-ceramic WP

An additional test case (Case 36 in Table 5-12) was conducted to determine the effect of reducing the highest Pu-ceramic rate by a factor of 3. This test (pr0 2131, where the "r" stands for "rate") is a slight modification of pO02131, the case with the highest total Gd loss (78%). The motivation for this case was LLNL's discovery that their previously reported rates had been overestimated by a factor of at least 3, due to an underestimate of the experimental surface area (Ref. 2, Table 6.2). Substitution of this lower rate roughly halved the Gd loss.

5.42 EQ6 Run Conditions and Nomenclature

Cases 1 through 20 in Table 5-10 were single-stage runs. Cases 21 through 25 of Table 5-11 were multiple-stage runs.

Tables 5-10 through 5-12, under "Case ID" column, give the root-file names used to describe the runs. The EQ6 input files corresponding to these runs end with the extension ".6i" (e.g., pO0111 1.6i is the EQ6 input file name for Case 1); these input files are included in the electronic media accompanying this calculation (Ref. 10). Each EQ6 run has associated tabdelimited text files, also included in the electronic media (e.g., p00_1111.elem_aqu.txt for Case 1). Most of the important run conditions could be inferred from the root-file name. For most cases, the Case ID is evaluated from left to right, as follows:

- The first letter "p" indicates Pu-ceramic.
- The second and third characters (first and second digits after "p") indicate sensitivity studies and revisions of the input file, with the third digit ranging from 0 to 9 (for the current document, the third digit never exceeded 2).
- The fourth character indicates the form of Fe oxide in the reaction products in multiple-

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stage runs (e.g., "h" for hematite and "g" for goethite, respectively). In single-stage runs, the default third character is an underscore ("["]"), and hematite is always allowed to form. Hematite and goethite are the primary Fe(lIl) iron oxide minerals observed to form in rust. Hematite is thermodynamically more stable, and its stability increases with temperature.

- The fifth digit is 1 or 2, indicating the average or high rate of steel corrosion in Table 5-1.
- The sixth digit in this block is 1, 2, or 0, with 1 and 2 indicating the low and high HLW glass corrosion rates listed in Table 5-2, respectively, and 0 indicating that no undegraded HLW glass is present in that stage of the EQ6 run.
- The seventh digit in the block is 1, 2, 3, or 0, with 1, 2, and 3 indicating the rates of Pu ceramic dissolution of low, average, and high, respsctively, as defined in Table 5-3, and 0 indicating no Pu-ceramic is corroding in that stage of the EQ6 run.
- The last digit in the block indicates the choice of water flush rate, with 1, 2, 3, and 4 indicating 0.0015 m³/y, 0.015 m³/y, 0.15 m³/y, and 0.5 m³/y, respectively.

An example is Case 21 in Table 5-11, which lists root file names pOh2204/pO2h2022 for the two input files. The first root name, pOlh2204, covers the time period before the breach of the 304L cans and exposure of the 304L magazines and rack, and represents stage number one, 01 (second and third characters) of a multiple-stage run. The fourth character, h, indicates that hematite is allowed to form (is not suppressed) in this case. The fifth character, 2, indicates the selection of the faster degradation rates for the A516 and 304L steels in the WP (Table 5-1). The sixth character, 2, indicates the faster HLW glass degradation rate in Table 5-2. The seventh character, 0, indicates the absence of Pu-ceramic in the EQ6 system (since the run is pre-breach, the ceramic is not yet exposed to chemical corrosion). The last character is 4, indicating the highest water flush rate of $0.5 \text{ m}^3/\text{year}$. The second root-file name, p02h2022, covers the time after breach of the cans and magazines containing Pu-ceramic, and is Case 21, stage two (second and third characters). As always, the last block of characters, 2022, gives the rates. The fifth character again indicates the faster steel corrosion rates. The sixth character is 0 since no HLW glass remains in the EQ6 system (though the HLW glass corrosion products are carried through the calculation). The seventh character indicates the Pu-ceramic corrosion has the "average" rate in Table 5-3, and the last character corresponds to a water drip rate of 0.015 m^3/year . In general, the first stage of a multistage run is comparatively short $(-10^3$ to -10^4 years), and the second stage of the run is carried out to at least $10⁵$ years, and up to $6.10⁵$ years. While the first stage is important in setting up the chemical conditions, the second stage is generally of greater interest for neutronics calculations since the corrosion product compositions can vary widely in the first stage, but achieve a quasi-steady state at longer times.

The case IDs in Table 5-12 are for the sensitivity test cases. The second character indicates the type of test: "s" for fO_2 sensitivity, or "r" for ceramic rate sensitivity. For the fO_2 sensitivity cases, the third character is 0 or 2, representing $fQ_2 = 10^{-10}$ bar or 10⁻¹⁵ bar, respectively.

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 Example 20 Finally resource to the Calculation

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6. RESULTS

6.1 SUMMARY OF RESULTS

Table 6-1 summarizes the total percentage Gd, Pu, and U lost at the end of the EQ6 runs. The complete output tables (aqueous, mineral, and total moles) for all the cases are included in the electronic media, as tab-delimited text files. A summary of the files included in the electronic media is given in Attachment II.

Table 6-1. Summary of Gd, Pu, and U Losses for all EQ6 Cases

In the following sections, Cases 3, 8, 13, 14, 18, 22, 25, and 26 will be discussed in more detail. Cases 3, 8, 13, 14, and 18 were selected for detailed discussion because they represent the runs with the highest net Gd, Pu, or U loss. Case 3 is discussed with Case 26 (its lower $fO₂$) analogue), to illustrate the indirect effect of $fO₂$ on Gd loss. Cases 22 and 25 were selected as representative of the two-stage scenarios; Case 25 is closest to the cases that produced highest Gd loss in the prior study of Pu-ceramic degradation (Ref. 5, Table 5.3-1).

Table 6-2 summarizes several aspects of the selected cases, focussing on conditions tied to Gd loss.

Case	Number оf Stages	Corrosion Rates [®]	Water Drip Rate (m^3y)	Years/10 ⁶	pH min / max	Peak Gd Conc. (molal)	Width Gd Peak ^b (y)	%Gd Loss
3		HLW: Low Steel: Average Ceramic: High	0.0015	6.3	5.8/9.4	$7.2E-3;$ $1.4E-3$	3.6E3; 3.3E4	20.45
8	1	HLW: High Steel: Average Ceramic: High	0.0015	6.3	7.3/10.1	4.9E-6; 5E-7	4.6E3; 1E4	0.02
13	1	HLW: Low Steel: High Ceramic: High	0.0015	6.3	5.3/9.3	5E-2	5.4E3	77.92
14	\ddagger	HLW: Low Steel: High Ceramic: High	0.15	0.43	5.8/8.2	$9.7E-4.$ 3.9E-4	1.1E3: 1.4E3	49.01
18	1	HLW: High Steel: High Ceramic: High	0.0015	6.3	7.9/10.0	$3.9E-6.$ $2.6E - 7:$ 7.3E-8	2.2E3; 1.6E5; 3.3E5	0.02
22	$\overline{2}$	HLW: High Steel: High Ceramic: \bullet Medium	0.5/ 0.015	0.031/5.3	5.2/8.6	6E-8	4.7E5	0.07
25	$\overline{2}$	HLW: High Steel: High Ceramic: High	0.15/ 0.0015	0.035/6.3	5.0/8.1	$4.4E-2$	1.2E3	13.89
26 ^c	1	HLW: Low Steel: Average Ceramic: High	0.0015	6.3	6.3/9.3	8.3E-5	5.7E4	1.30

Table **6-2.** Gd Loss Characteristics and pH of Selected Cases

NOTES: "See Tables 5-1 through 5-3 for rates.

bMost cases have multiple "peaks." Peaks and widths approximated from plots in Gd_peaks_Pu-ceram_083099.doc, included in the electronic media for this document (Ref. 10). ^eAnalogous to Case 3, except log₁₀(fO₂) = -10 (versus -0.7).

 12.280

6.2 CASES 3 AND 26 (p00_1131 and ps0_1131)

The results of Case 3 (p00 1131) and its corresponding sensitivity test, Case 26 (ps0 1131), are plotted in Figures 6-1 through 6-7 to demonstrate the general consequences of WP component degradation. The consequences include: changes in pH; variations in the dissolved Gd, Pu, and U content of the water flushed from the WP; and the formation of solubility-controlling and space-filling minerals and solid solutions. Figures 6-1 through 6-7 also demonstrate the indirect effects of oxygen fugacity on pH and Gd loss.

Figure 6-1. Effect of fO_2 on pH: Cases 3 (p00_1131) and 26 (ps00_1131)

Figure 6-1 illustrates the tie between materials degradation and pH. The pH in the WP decreases with steel degradation (especially with 304L steels) until the steels are exhausted. Since this example uses a low HLW glass degradation rate, the alkalinity from HLW glass is insufficient to neutralize the pH until all the steel is consumed. The Mg and Ca carbonate minerals formed from

HLW glass dissolution do not fully neutralize the acid produced from steel degradation. The Cr contained in the steels is oxidized when it is released into water, and lowers the pH:

$$
Cr + H2O + 1.5O2 = CrO4- + 2H+
$$
 (Eq. 2)

After the steel is consumed and its associated acidity is flushed from the system, the pH climbs (at -50,000 years) due to HLW glass dissolution. Once the HLW glass is consumed, the pH gradually drops to the ambient for the incoming water at $fCO₂ = 10⁻³$ bar. However, in the sensitivity test, the lower oxygen fugacity $(fQ_2 = 10^{-10})$ in the WP prevents the pH from increasing until $\sim 10^{4.81}$ years, which is $\sim 10^4$ years later than that in the case with 0.2 bar fO₂; the rapid pH increase occurs when the Pu ceramic and "Out Web AS16" (the carbon steel support structure) have been exhausted and while the GPC 304L and Cans 304L continue to degrade. The lower oxygen fugacity in the system limits chromate oxidation, and thus lowers the production of acid via Equation 2.

The concentrations of aqueous Gd, U, and Pu as functions of time are shown in Figure 6-2 and 6-3. The solubility-limiting phase for Gd is GdOHCO₃; like most carbonates, this phase is more soluble under acid conditions, thus the times of highest aqueous Gd correspond to times of low pH. Thus, the case with lower fO₂ (Case 26) also yields lower Gd loss (1.3%, versus 20.45% for Case 3; Table 6-1).

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NOTES: Compare with Figure 6-2. The lower fO₂ results in slightly lower dissolved Cr and a less extreme pH minimum. Dissolved Pu is lower than that for Case 26 because aqueous Pu(VI) carbonate complexes are insignificant at this lower fO2.

Figure 6-3. Case 26 (fO₂ = 10⁻¹⁰ bar): pH and Total Aqueous Cr, Gd, Pu, and U

In Figures 6-4 and 6-5, selected minerals that control the solubility of chromium are plotted together with pH as a function of time for $fO_2 = 0.2$ and 10^{-10} bar, respectively. At the higher fO_2 , the only Cr solids predicted to form are BaCrO₄ and crocoite (PbCrO₄), and there is insufficient Ba and Pb in the system to precipitate all the chromate produced by steel oxidation. The formation of solid, mixed-valence state CrO₂, in Case 26 (fO₂ = 10^{-10}) results in a less dramatic drop in pH; compare Figure 6-5 to Figure 6-4 (Case 3; $fO_2 = 0.2$). In summary, the formation of Cr-bearing solids affects the pH in the system and, consequently, the losses of Gd, Pu, and U from the WP. Reducing the uncertainty in the stability of Cr(III) and Cr(IV)containing solids in the EQ3/6 thermodynamic database in EQ3/6 would significantly reduce uncertainty in simulating the degradation of Pu-ceramic.

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0 **^I**- **^I**~~~~~~~~~~~9.5 i 1 *I* 9'5 \mathbb{P} -2 \uparrow 2 \uparrow 2 \uparrow 2 \uparrow 2 \uparrow 2 \uparrow 5.5 -3

<u>Waste Package Operation</u>

Waste Package Operations - The Calculation Calculation of Pu-Ceramic Waste Packages: Effects of Calculation for Chemical Degradation of Pu-Ceramic Waste Packages: Effects of Updated Materials Composition and Rates

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NOTE: Formation of mixed valence state CrO₂ causes a lesser drop in pH, compared to Figure 6-4.

Figure 6-5. Case 26 (fO₂ = 10⁻¹⁰ bar): pH and Minerals Controlling Cr Solubility

The above results demonstrate the significant effect of oxygen fugacity on calculated pH and losses in Pu, Gd, and U from the WP. The fO₂ used in most Cases (0.2 bar) is probably higher than the average fO_2 experienced in a corroding WP. A lower fO_2 of 10^{-10} bar dramatically reduces losses of *Gd*, Pu, and U. This lower $fO₂$ value is not unreasonable. The common observation of coexisting magnetite and hematitefFeOOH in rust on cars implies local, low *fO2* during corrosion of steels.

The fO₂ of the magnetite/goethite and magnetite/hematite buffers are easily calculated from the log_{10} K values for the magnetite, hematite, and $Fe³⁺$ decomposition reactions, as given in the EQ3/6 dataO.nuc.R8d thermodynamic constants file. Equations 3 through 6 below give the dataO reactions necessary to carry out the calculations.

 $Fe₂O₃ + 6H⁺ = 2Fe⁺⁺⁺ + 3H₂O$ log₁₀K = 0.1086 at 25 °C (Eq. 5)

$$
\text{Fe}^{++} + 0.5 \cdot \text{H}_2\text{O} = 0.25 \cdot \text{O}_{2 \text{ (g)}} + \text{H}^+ + \text{Fe}^{++} \quad \log_{10} \text{K} = -7.7654 \text{ at } 25 \text{ °C} \tag{Eq. 6}
$$

Combining equations algebraically $[(Eq. 4) - (Eq. 6) - 3 \cdot (Eq. 3)]$ yields:

$$
Fe3O4 + 0.25 O2(g) + 1.5·H2O = 3FeOOH log10K = 16.6343
$$
 (Eq. 7)

At equilibrium:

$$
log_{10}K = 3 \cdot log_{10}[Fe_3O_4] - 1.5 \cdot log_{10}[H_2O] - 0.25 \cdot log_{10}(fO_2) - log_{10}[Fe_3O_4]
$$
 (Eq. 8)

Where the \iint notation here indicates activity (~concentration). The activity of solid Fe₃O₄ in Fe₃O₄ is 1, the activity of water relative to a standard state of pure water is \sim 1, and the activity of solid FeOOH in FeOOH is 1, so Equation 8 reduces to:

$$
log_{10}(fO_2) = -4 \cdot log_{10}K = -66.5372
$$
 (Eq. 9)

Similarly, equations $(Eq. 4) - (Eq. 6) - 1.5 (Eq. 4)$ can be combined to yield:

$$
Fe3O4 + 0.25·O2 (g) = 1.5·Fe2O3 \t\t log10(fO2) = -72.2996 \t\t (Eq. 10)
$$

Thus the formal fO₂ is $\sim 10^{-67}$ to 10⁻⁷² bar for these commonly observed rust assemblages. These low values do not imply that only a tiny fraction of a molecule of $O₂$ can coexist with magnetite/Fe(lII)-oxide rusts; rather, they imply that corrosion can produce locally reduced assemblages, even in the presence or air. Other redox-active couples in natural waters imply more moderate fO₂. For example, commonly observed nitrogen redox couples (e.g., nitratenitrite in pore waters of Chihuahuan deserts) imply $fO₂ < 10^{-10}$ bar (Ref. 38, pp. 632-634).

Figures 6-6 and 6-7 show some of the principal solubility-controlling solids predicted by the EQ6 runs for Cases 3 and 26. The hematite abundance plateaus rapidly, and results principally from degradation of the A516 outer web support structure. Because of the low HLW glass degradation rate, smectite clays form slowly. However, a "mole" of smectite clay $(((Ca, Mg)_{0.165}, (K, Na)_{0.33})Fe₂Al_{0.33} Si_{3.67}H₂O₁₂) has a much larger molar volume than a "mole" of$ hematite (424.293 versus 30.274 cm³, respectively; from the V0PrTr entries in data0.nuc.R8d). $GdOHCO₃$ is the principal solubility-controlling phase for Gd, though some $GdPO₄·H₂O$ forms with the phosphate present in HLW glass and steels.

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NOTE: The phosphate to form GdPO4:H₂O originates as a trace constituent of steels and waste glass; phosphate content of incoming water (< 1 ppm) is comparatively insignificant. Compare to Figure 6-7.

Figure 6-6. Case 3 (fO₂=0.2 bar): pH and Some Solubility-Controlling Minerals

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Figure 6-7. Case 26 (fO₂=10⁻¹⁰ bar): pH and Some Solubility-Controlling Minerals

Tables 6-3 and 6-4 give the elemental composition for the water in the WP (aqueous phase) and for the degradation products (in mole%) for Case 3 (p00_1131) at selected reaction times.

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Table 6-3. Solution Composition in Molality in Selected Years for Case 3 (p00_1131)

NOTES: ^{*}Hf was converted to Zr for EQ6 Calculations (Assumption 3.16).

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Table 6-4. Composition of Corrosion Products (mole%) and Density in Selected Years for Case 3 $(p00_11131)$

NOTES: ^aHf was converted to Zr for EQ6, then converted back to Hf for mass and density calculations.
⁹For EQ6 system (1 liter aqueous fiuld). To obtain total grams in WP, multiply by total system volume of 4593.965 liters.

6.3 CASE 8 (p00 1231)

In Case 8, the high HLW glass and high ceramic degradation rates raise the system pH over 10 in less than $5 \cdot 10^3$ years (Figure 6-8), in the process elevating the solubilities of U and Pu (Figure 6-9). In the higher pH periods, the dissolved U and Pu are almost entirely as $UO_2(CO_3)$ ⁴⁻ and $PuO₂(CO₃)₃⁴$. Consequently, there are significant losses of U (89.2%) and Pu (29.9%). Because there is no protracted period of low pH, only 0.02% of Gd is lost (Table 6-2). The combination of fast ceramic degradation, high pH, and fixed fCO₂ leads to formation of calcite (with Ca principally from ceramic and HLW glass) and GdOHCO₃, the latter as the solubility-controlling phase for Gd (Figure 6-10). These solid carbonates also serve as buffers against rapid drops in pH.

Figure 6-8. Case 8 (p00_1231): pH and Consumption of Package Materials

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Figure 6-9. Case 8: pH and Total Aqueous Gd, Pu, and U

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NOTE: GdOHCO₃ controls Gd solubility after ~400 years.

Tables 6-5 and 6-6 give the aqueous phase composition and the composition of the solid corrosion products in the WP for Case 8.

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Table 6-5. Solution Composition in Molality in Selected Years for Case 8 (p00_1231)

NOTES: ^aHf was converted to Zr for EQ6 Calculations (Assumption 3.16).

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Table 6-6. Composition of Corrosion Products (mole%) and Density in Selected Years for Case 8 (pO0_1231)

Years Element	4360	9767	30506	126670	633860
O	5.39E+01	5.44E+01	5.81E+01	5.98E+01	5.99E+01
ĀĪ	8.81E-01	8.76E-01	$1.02E + 00$	1.08E+00	1.08E+00
в	2.74E+00	2.53E+00	1.94E+00	$6.43E-16$	2.70E-16
Ba	8.39E-03	8.34E-03	9.68E-03	1.03E-02	1.03E-02
Ca	4.81E-01	4.78E-01	5.55E-01	4.57E-01	4.78E-01
CĪ	0.00E+00	$2.16E - 18$	0.00E+00	$0.00E + 00$	0.00E+00
$\overline{\mathbf{C}}$ r	$0.00E + 00$	2.98E-03	3.46E-03	1.38E-03	0.00E+00
Cu	2.41E-02	2.39E-02	2.78E-02	2.95E-02	2.94E-02
F	0.00E+00	5.40E-19	0.00E+00	$0.00E + 00$	0.00E+00
Fe	1.12E+01	1.20E+01	1.60E+01	1.88E+01	1.87E+01
Gd	7.63E-02	7.68E-02	8.80E-02	9.35E-02	9.33E-02
Ή	1.76E+01	1.66E+01	8.69E+00	5.55E+00	5.60E+00
ਟ	8.25E-01	8.03E-01	3.29E-01	6.03E-01	5.47E-01
F	1.12E-02	1.22E-02	1.68E-02	1.99E-02	2.00E-02
κ	4.44E-01	4.39E-01	4.49E-01	2.59E-01	1.32E-01
π	1.92E-01	1.19E-01	0.00E+00	$0.00E + 00$	$0.00E + 00$
Mg	3.46E-01	3.44E-01	3.97E-01	3.44E-01	3.30E-01
Mn	4.09E-01	4.32E-01	5.66E-01	6.52E-01	6.52E-01
Mo	$0.00E + 00$	$0.00E + 00$	2.14E-31	$0.00E + 00$	0.00E+00
Ñ	$0.00E + 00$	$0.00E + 00$	$0.00E + 00$	0.00E+00	$0.00E + 00$
Na	1.61E+00	1.54E+00	4.57E-01	2.33E-02	3.20E-02
Ni	2.25E-01	3.45E-01	7.02E-01	9.87E-01	9.86E-01
Np	$0.00E + 00$	$0.00E + 00$	8.67E-17	0.00E+00	$0.00E + 00$
Pb	3.00E-03	2.98E-03	3.46E-03	3.67E-03	3.67E-03
$\overline{\mathsf{Pu}}$	5.21E-02	5.34E-02	6.20E-02	6.59E-02	6.58E-02
ड	0.00E+00	2.43E-16	$0.00E + 00$	$0.00E + 00$	$0.00E + 00$
$\overline{\mathsf{si}}$	8.01E+00	7.98E+00	$9.32E + 00$	9.96E+00	1.01E+01
Tc	0.00E+00	1.23E-31	$0.00E + 00$	1.51E-31	1.42E-31
π	9.14E-01	9.08E-01	$1.05E + 00$	$1.12E + 00$	1.12E+00
Ū	0.00E+00	2.81E-02	2.91E-02	3.12E-02	3.10E-02
Hf(Zr)	8.82E-02	8.76E-02	1.02E-01	1.08E-01	1.08E-01
Tota! (%)	100	100	100	100	100
Total (g)	3.19E+03	3.31E+03	$3.31E + 03$	3.35E+03	$3.35E + 03$
Density (g/cm ⁺)	3.28	3.37	3.84	4.09	4.10

NOTES: 'Hf was converted to Zr for EQ6. then converted back to Hf for mass and density calculations. bFor EQ6 system (1 Iter aqueous fluid). To obtain total grams In WP. multiply by total system volume of 4593.965 liters.

6.4 CASE 13 (p00_2131)

Case 13 combines fast steel and ceramic degradation with slow HLW glass degradation and a very low drip rate. These conditions allow early degradation of the ceramic, and exposure of the Gd solids to sustained, low pH. Consequently, this case experienced the highest Gd loss, ~79%. Figure 6-11 shows the relationship of WP materials degradation to pH, and Figure 6-12 shows the total dissolved Gd (Gd [aq]), and the total Gd in solids, per liter of void space. It is apparent that for a substantial period, the dissolved Gd is more significant than the Gd present in solids. The dissolved Gd is largely lost from the system by the flushing action. Because of the low flush rate, the pH eventually reaches high values (due to HLW glass dissolution); consequently, this system loses much of the U to aqueous carbonate complexes.

Figure 6-11. Case 13 (p00_2131): pH and Package Materials Remaining

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NOTES: Moles solids calculated per liter of void space. This case had the highest Gd loss. Gd (aq) is the total dissolved (aqueous) Gd concentration.

Figure 6-12. Case 13: pH, Moles Aqueous Gd, and Moles Gd Solids

Tables 6-7 and 6-E, respectively, give the concentrations of dissolved elements in the aqueous phase, and the composition and density of the corrosion products for Case 13.

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Table 6-7. Solution Composition in Molality in Selected Years for Case 13 (p00_2131 and pl0_2131)

NOTES: ⁴Hf was converted to Zr for EQ6 Calculations (Assumption 3.16).

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Table 6-8. Composition of Corrosion Products (mole%) and Density at Selected Times for Case 13 (p00_2131 and pl0_2131)

NOTES: ^aHf was converted to Zr for EQ6, then converted back to Hf for mass and density calculations.
⁹For EQ6 system (1 liter aqueous fluid). To obtain total grams in WP, multiply by total system volume of 4593.965 liters.

6.5 CASE 14 (p00 2133)

In Case p00_2133, the combination of rapid steel degradation and low HLW degradation causes low pH (< 6) for $\sim 6.10^3$ years (Figure 6-13). The HLW glass remains largely intact after most of the steel has corroded (Figure 6-13). Once the steel is exhausted, HLW glass dissolution raises the pH in the system. However, because the drip rate is comparatively high, the pH swings are less extreme than those in Case 13 (Figures 6-11 and 6-12). The concentration of soluble Gd (primarily Gd^{3+} or $Gd(CO_3)_2$) is high (~10⁻³ mole/l) for several thousand years (Figure 6-14). Because of the high drip rate, the peak Gd concentration is less extreme than that in Case 13, but the net Gd fluence (concentration x drip rate) is sufficient to remove a significant portion of Gd from the WP. This Case has the second highest Gd loss (49%). However, the lack of very high pH lessens formation of U and Pu carbonate complexes; consequently, U and Pu losses, 0.22 and 0.2%, respectively, are low (Table 6-1). In the sensitivity tests, Cases 30 and 35, Gd loss is decreased to 37.5, and 0.10%, respectively, by changing $fO₂$ from 0.2 to 10⁻¹⁰, and from 0.2 to 10^{-15} bar, respectively (Table 6-1).

NOTES: This case shows the effects of combining high steel dissolution rates with a low HLW glass rate. Corrosion of steels decreases pH until the steels are exhausted; glass dissolution then increases pH.

Figure 6-13. Case 14 (p00_2133): pH and Package Materials Remaining

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NOTES: The degradation of steel, and subsequent oxidation of Cr(O) to Cr(VI), lowers pH until the steel is exhausted. In the low pH period, Gd occurs mainly as Gd³⁺. After exhaustion of steel (Figure 6-13), Gd solubility decreases and $Gd(CO₃)₂$ becomes the primary dissolved Gd species.

Figure 6-14. Case 14: pH, Gd Species, and Total Aqueous Gd and Cr

Tables 6-9 and 6-10, respectively, give the concentrations of dissolved elements in the aqueous phase, and the composition and density of the corrosion products for Case 13.

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Table 6-9. Solution Composition in Molality in Selected Years for Case 14 (p00_2133)

NOTES: ^aHf was converted to Zr for EQ6 Calculations (Assumption 3.16).

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Table 6-10. Composition of Corrosion Products (mole%) and Density in Selected Years for Case 14 $(p00_2133)$

NOTES: "Hf was converted to Zr for EQ6, then converted back to Hf for mass and density calculations.
"For EQ6 system (1 liter aqueous fiuld). To obtain total grams in WP, multiply by total system volume of 4593.965 liters.

6.6 CASE 18 (p00_2231)

Case 18 (p00_2231) and Case 8 (p00_1231) are similar in that the relatively high HLW glass dissolution rate, and low groundwater drip rate, keep pH above 7 (Figure 6-15). The loss of Gd in Case 18 is low $($ < 0.02%) and losses of Pu and U are significant, 13.99 and 72.95%, respectively (Table 6-1). The minimum pH for Case 18 (Figure 6-15) is higher than that for Case 8 because, in Case 18, the steel is consumed early when there is ample alkalinity from the ceramic and HLW glass to consume H^+ produced via Equation 2.

Tables 6-11 and 6-12 give the aqueous phase composition and the composition of the solid corrosion products in the WP for Case 18.

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Table 6-11. Solution Composition in Molality in Selected Years for Case 18 (p00_2231)

NOTES: ^eHf was converted to Zr for EQ6 Calculations (Assumption 3.16).

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Table 6-12. Composition of Corrosion Products (mole%) and Density in Selected Years for Case 18 $1000 - 2231$

NOTES: "Hf was converted to Zr for EQ6, then converted back to Hf for mass and density calculations.
⁹For EQ6 system (1 liter aqueous fluid). To obtain total grams in WP, multiply by total system volume of 4593.965 liters.

6.7 TWO-STAGE CASES 22 AND 25 (pO1g2204IpO2g2O22 and pOlg2203IpO2g2O31)

For the five two-stage tests, U contained in the HLW glass is totally lost in the first stage. At the second stage, loss of U is less than 1% in all of the cases, and losses of Pu are also not significant (Table 6-1). The high steel and HLW glass dissolution rates and high groundwater flushing rate remove most of the alkalinity from the HLW glass in the first stage. Since the Puceramic is exposed only in the second stage, the potential for Gd loss occurs only in the second stage. The second stage begins with a period of low pH (Figures 6-16 and 6-17), due to corrosion and oxidation of Cr in the remaining GPC stainless steel, and in the 304L cans, magazines, and racks. Uranium solubility is not dramatic under such pH conditions (Tables 6-1 and 6-13), and U retention in the second stage is higher than in the first. Because of the lower flush rate for Case 25, the low pH persists for longer time. Of the two-stage runs, only Case 25 achieves significant Gd loss because the Gd is degraded completely in the period of low pH. However, the ceramic itself is alkaline, and its degradation counteracts acidity produced by steel degradation. Case 25 is most analogous to the cases that produced high $(-15%)$ Gd loss in the prior study of ceramic WP degradation (Ref. 5, Table 5.3-1).

The "stair step" pH pattern, shown in Figures 6-17 and 6-18 is related to the appearance and disappearance of minerals (corrosion products). Figure 6-19 illustrates one cause of the "stair step". After all the steel has corroded, the only solid Cr phase is BaCrO4. As the system is flushed with Cr-poor water, the BaCrO₄ transforms into $BaCO₃$, dominantly by the reaction:

$$
BaCrO4 + HCO3- = BaCO3 + CrO4- + H+.
$$
 (Eq. 11)

Thus, until the BaCrO₄ disappears completely, the pH drops. Once the BaCrO₄ is exhausted, pH returns nearly to the ambient of incoming water at 10^{-3} bars fCO₂. There is one last small stair step, related to the disappearance of crocoite (PbCrO4).

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Figure 6-17. Case 25 (p02g2031, 2nd Stage): pH and Package Materials Remaining

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NOTES: Gd (aq) is the total dissolved (aqueous) Gd concentration. High Gd (aq) exists only as a brief, sharp peak before pH climbs and GdOHCO₃ precipitates.

Figure 6-18. Case 25 (2nd Stage): pH, Solid Corrosion Products, and Gd (aq)
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Figure 6-19. Case 25 (2nd Stage): Minerals Causing pH "Stair Step"

Tables 6-13 through 6-16 give the elemental composition of the aqueous phase and the precipitated solids (corrosion products) for Cases 22 and 25.

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Table 6-13. Solution Composition in Molality in Selected Years for Case 22 (p01g2204/p02g2022)

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^aHf was converted to Zr for EQ6, then converted back to Hf for mass and density calculations.
PFor EQ6 system (1 liter aqueous fluid). To obtain total grams in WP, multiply by total system volume **NOTES:** of 4593.965 liters.

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Table 6-15. Solution Composition in Molality in Selected Years for Case 25 (p01g2203/p02g2031)

NOTES: ^{*}Hf was converted to Zr for EQ6 Calculations (Assumption 3.16).

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Table 6-16. Composition of Corrosion Products (mole%) and Density in Selected Years for Case 25 (p01g2203/p02g2031)

NOTES: "Hf was converted to Zr for EQ6, then converted back to Hf for mass and density calculations.
"For EQ6 system (1 liter aqueous fluid). To obtain total grams in WP, multiply by total system volume of 4593.965 liters.

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7. ATTACHMENTS

Attachment I. Document Input Reference Sheets (8 pages)

Attachment II. Listing of Files on Electronic Media (7 pages)

I

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Referenced in the Electronic Media:

Roberts, W.L.; Rapp, G.R., Jr.; and Weber, J. 1974. *Encyclopedia of Minerals.* New York, New York: van Nostrand Reinhold Company. TIC: 241917.

Attachment I. Document Input Reference Sheet

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Attachment **II.** Listing of Files on Electronic Media

This attachment contains the MS-DOS directory for files placed on the electronic media (Ref. 10). The files are of nine types:

- 1) Excel files (extension = xis), called out in the text and tables. The file Pu-ceram.xls is the principal source of the *.6i input files used for sections 5.4.2 and 6.
- 2) EQ3/6 input files (extensions = 3i or 6i), as discussed in Sections 5.3 and 5.4.2. The 6i input files described in Section 5.4 and 6 are formatted for the SCFT mode, and have 8-character names p???sgfw.6i, where the s is a steel rate index (1 or 2), the g is a HLW glass rate index (0, 1 or 2; 0 used when no undegraded HLW glass is present), the f is the fissile (ceramic) degradation rate index (0, 1, 2, or 3), and w is the incoming water drip rate index (1, 2, 3, 4). The ??? indicate run-specific nomenclature described in section 5.4.2.
- 3) EQ6 output files (text, extension $= 60$).
- 4) Tab-delimited text files (extension = txt), with names p???????.elem????.txt. as discussed in Section 5.4.2; these contain total aqueous moles (*.elem_aqu.txt), total moles in minerals and aqueous phase ($*$ -elem_m_a.txt), total moles in minerals, aqueous phase, and remain special reactants (*.elem tot.txt), and the total moles in minerals alone (*.elem min.txt). The *.elem_tot.txt and *.elem_min.txt also have the volume in cm³ of the minerals and total solids (including special reactants) in the system.
- *5)* FORTRAN source files (extension = for) for the version of EQ6 used in the calculations.
- 6) MS-DOS/Win95/Win98 executables (extension $=$ exe) for the version of EQ6 and runeq6 used in the calculations, and the autoexec.bat file that sets up the environment.
- 7) EQ6 data files used for the calculations, with the text file dataO.nuc.R8d, and the binary version datal.nuc.
- 8) Selected binary files (bin extension) used for plots in this document; the format of these files is specified in the EQ6 addendum source (Ref. 18). The binary files are not necessary to use the output of the calculation, but are provided as a convenience for plotting.
- 9) A copy of this calculation document, Pu-ceramic_312hws*.doc, as of 09-12-1999.

Below are listed the contents of the DOS directories within the electronic attachment. The file sizes are given in bytes.

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