Chapter 5 Streams, Reservoirs, and the Savannah River

The southeastern United States has abundant freshwater resources, which provide diverse ecosystems. There are five main drainage basins on SRS. The five streams that originate on, or pass through SRS before entering the Savannah River are Upper Three Runs, Beaver Dam Creek, Fourmile Branch, Steel Creek, and Lower Three Runs. A sixth stream, Pen Branch, does not flow directly into the Savannah River but joins Steel Creek in the Savannah River floodplain swamp. The upper reaches of Lower Three Runs were impounded in 1958 to form Par Pond. L Lake was formed in 1985 by damming Steel Creek. The purpose of Chapter 5 is to describe and define the physical and biological characteristics of these systems and provide a synopsis from the many aquatic studies conducted through the years at SRS.

To open any section of chapter 5, click on its "bookmark" listed in the column on the left.

5.1 Upper Three Runs

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Drainage Description and Surface Hydrology

General Description

Upper Three Runs is a large, cool (annual maximum temperature of 26.1°C [79°F]), black water stream in the northern part of SRS (Figure 5-1). With headwaters arising offsite, near Aiken, Upper Three Runs drains an area of approximately 545 km^2 (209 mi²) and discharges directly into the Savannah River. Upper Three Runs is approximately 40 km (25 mi) long, with the lower 28 km (17 mi) within the boundaries of SRS. Upper Three Runs receives more water from underground sources (Dublin-Midville aquifer system) than the other SRS streams; hence, it has low conductivity, low hardness, and low pH values (Specht 1987).

Effluent Contribution

Upper Three Runs is the only major tributary on SRS that never has received major thermal discharges. Its two significant tributaries are Tinker Creek, the largest, and Tims Branch. Above its confluence with Tims Branch, Upper Three Runs is relatively unimpacted. Begin ning in late 1988, Upper Three Runs began receiving effluents from the F-/H-Area Effluent Treatment Facility, which discharges into the creek just downstream of the Road C bridge. Tims Branch receives industrial wastes from the fuel fabrication facilities (300-M Area) and the Savannah River Technology Center (700-A Area) including nonprocess cooling water, steam condensates, process effluents, and treated groundwater effluents. Three smaller trib utaries of Upper Three Runs originating from the 200-F and the 200-H Separations Areas

Figure **5-1.** Location of Upper Three Runs on SRS

and S-Area carry ambient temperature cooling water, steam condensates, powerhouse wash down waters, and ash disposal basin effluents.

Upper Three Runs, either directly or via tributaries, receives the following NPDES-permit ted discharges: cooling water, blowdown, stormwater, lab drains, air stripper discharge, steam condensate, M-Area wastes, process-water, Consolidated Incineration Facility waste water, neutralization wastewater, and F/H Effluent Treatment Facility wastewater.

Flow Measurements

The U.S. Geological Survey measures flow at several locations on Upper Three Runs. Flow statistics of Upper Three Runs at SRS Road A, above SRS Road C, and near New Ellenton (Highway 278) (Figure 5-2) are summarized in Table 5-1. Records for the most downstream station (Upper Three Runs at Road A) date back to June 1974 (no records from February 1978 through September 1978). In water year 1995, the mean flow of Upper Three Runs at Road A was 8.0 m³/s (284 ft³/s). Over the period of record (water years 1974-1995) at Road A, the mean flow was 6.9 m³/s (245 ft³/s), the 7-day low flow was 2.4 m³/s (84 ft³/s), and the 7Q10 was 2.8 m³/s (100 ft³/s). The maximum, minimum, and mean daily flows from October 1982 to September 1995 at the three stations are shown in Figure 5-3 through Fig ure 5-5.

Figure 5-2. Flow Measurement Sampling Stations for Upper Three Runs

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Table **5-1.** Flow Summary for Upper Three Runs

cms = cubic meters per second.

cfs = cubic feet per second.

Figure 5-3. Maximum, Mean, and Minimum Flow Measurements for Upper Three Runs at U.S. Highway 278, Octo ber 1982-September 1995.

Figure 5-4. Maximum, Mean, and Minimum Flow Measurements for Upper Three Runs above Road C on SRS, Octo ber 1982-September 1995.

Figure 5-5. Maximum, Mean, and Minimum Flow Measurements for Upper Three Runs at Road A at SRS, October 1982-September 1995.

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Section 5.1-Upper Three Runs
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Water Chemistry and Quality

Studies and Monitoring

Water-Quality Monitoring

Westinghouse Savannah River Company Environmental Monitoring Section has monitored the water quality of Upper Three Runs since 1973. During routine monitoring, three sam pling locations have been established on Upper Three Runs. These locations - Upper Three Runs at U.S. Highway 278, Tims Branch 5, and Upper Three Runs at Road A - are sampled monthly for physical and biological water quality parameters and for metals (Fig ure 5-6). Samples also are collected annually and analyzed for pesticides, herbicides, and PCBs. In 1990, the Upper Three Runs at Highway 278 sampling location was moved approximately 1.6 km (1 mi) downstream and closer to the SRS boundary to provide a bet ter indication of contaminants originating above SRS; this new sampling location is desig nated Upper Three Runs **IA.**

All routine water quality monitoring data reported in the following sections can be found in the annual SRS Environmental Reports.

Figure 5-6. Water Chemistry and Quality Sampling Stations for Upper Three Runs

Comprehensive Cooling Water Study

From 1983 to 1985, Upper Three Runs and Tims Branch were studied as part of the Com prehensive Cooling Water Study (CCWS). This study was designed to assess the impact of current and proposed SRS activities on water quality. The CCWS included the following four locations on Upper Three Runs and Tims Branch (Figure 5-6):

- * Upper Three Runs at Road 8-1 upstream of SRS point source effluents and upstream of Tinker Creek (02 on map)
- * Upper Three Runs at Road C primarily consists of headwater conditions, but also reflects point-source effluents to tributaries of Upper Three Runs from 200-F and 200-H Separations Areas, as well as construction impacts from 200-S Area (13 on map).
- * Upper Three Runs upstream of confluence with Savannah River consists of all upstream point-source and runoff conditions, as well as temporal flooding from high water conditions of the Savannah River (18 on map).
- Tims Branch cumulative effects of point-source effluents discharged to this stream from 300-M Area, the Savannah River Technology Center, and the Savannah River Ecol ogy Laboratory (12 on map).

Comprehensive results and discussion of CCWS data can be found in Newman et al. (1986) and Lower (1987). Gladden et al. (1985) present a synopsis of Upper Three Runs and Tims Branch water quality data prior to the CCWS.

Priority Pollutants Survey

In 1984, an instream survey of priority pollutants was conducted to determine the levels of volatile, acid, base and neutral compounds in Upper Three Runs. This study sampled four locations: Upper Three Runs at Highway 278 (01), Tims Branch near Road C (12), Upper Three Runs upstream of Road C (04), and Upper Three Runs at Road A (17). Potential sources for organics in Upper Three Runs included process sewer outfalls, local outcrop ping of groundwater, and the M-Area settling basin. The results of this study are presented in the sections that follow and in Lower (1987).

Chemical Assessment Studies

From 1985 to 1993, four studies were conducted on the waters of Tims Branch and Upper Three Runs to determine impacts from specific SRS operations. In addition, a toxicity study was done on Upper Three Runs waters upstream of all SRS discharges. These studies, which are summarized later in this chapter included:

- Effect of changes in M-Area effluent on Tims Branch (June 1985-December 1986)
- * Effect of M-Area Liquid Effluent Treatment Facility on Tims Branch (September 1988-March 1990)
- Effect of F/H-Area Effluent Treatment Facility on Upper Three Runs (July 1987-February 1990)
- Effect of contaminants from Mixed Waste Management Facility on Upper Three Runs wetlands (1993)
- "* Effect of Upper Three Runs water on *Ceriodaphnia dubia* (1994)

Field Data

Water Temperature

The mean water temperature at sampling locations on Tims Branch and Upper Three Runs ranged from 15.2 to 17.3°C (59 to 63°F) during the CCWS and routine monitoring program (Table 5-2). These temperatures are reflective of the nonthermal nature of this stream.

pH Measurements

The mean pH in the Upper Three Runs watershed is slightly acidic, with means ranging from 6.03 near Road 8-1 to 6.65 at the stream confluence with the Savannah River (Table **⁵** 2). Throughout the CCWS and during routine monitoring, waters at all Upper Three Runs and tributary sites displayed wide pH variations (3.1-8.3).

Physical Characteristics and General Chemistry

Dissolved Oxygen

Mean dissolved oxygen concentrations ranged from 7.80 to 8.10 mg/l throughout Upper Three Runs during the CCWS (Table 5-3). Dissolved oxygen concentrations measured since the CCWS have been slightly higher, with ranges from 5.0 to 15.1 mg/l. Dissolved oxygen in McQueen Branch and Crouch Branch fluctuates from 1.2 to 13.5 mg/l (Table 5-3).

Suspended Solids and Turbidity

Mean concentrations of suspended solids and turbidity were highest and most variable dur ing the CCWS (Table 5-3). The range in maximum suspended solids concentrations varied considerably (46.8 mg/I - 892 mg/i) throughout the watershed. Maximum turbidities ranged from 32 to 352 **NTU** (Nephelometric Turbidity Units). Suspended solids and turbidity con centrations have been lower since the CCWS. From 1987 to 1991, maximum suspended sol ids concentrations ranged from 6.0 to 97 mg/l, and turbidities have ranged from 3.9 to 120 NTU. From 1992 to 1995, maximum suspended solids concentrations ranged from 31 to 89 mg/I and turbidities ranged from 1.02 to 200 NTU in Upper Three Runs and its tributaries (Table 5-3).

Water quality in tributaries to Upper Three Runs that drain the Defense Waste Processing Facility (DWPF) site was monitored from 1983 to 1995 to determine the effect of construc tion activities on the tributaries and Upper Three Runs. In both McQueen Branch and Crouch Branch, total suspended solids and turbidity increased over baseline and reference stream concentrations. During the period of construction mean total suspended solids ranged from 18.26 to 182.33 mg/l in McQueen Branch and from 50.12 to 99.62 mg/1 in Crouch Branch. After construction means ranged from 5.63 to 135.83 mg/l and 26.13 to 156.99 mg/l in McQueen and Crouch Branch, respectively (Bodie and Scott 1995).

Turbidity ranged from 25.62 to 236.13 NTUs in McQueen Branch during construction and from 1.86 to 158.89 NTUs since construction. In Crouch Branch the values were 98.67 to 187.33 NTUs during construction and 30.62 to 237.44 NTUs after construction.

The higher suspended solid loads and turbidity values occurred after periods of heavy rain and may be the result of poorly functioning sedimentation basins (Bodie and Scott 1995).

Table 5-2. Upper Three Runs Field Data

NA = Not analyzed.

Blank Spaces = Mean not calculated due to insufficient data in report. **aCCWS =** Comprehensive Cooling Water Study (Newman et al. 1986).

61987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

C19 ⁹ ² -19 ⁹ ⁵ = Data taken from Amett et al. 1993, 1994, 1995, and 1996.

Table 5-3. Upper Three Runs Physical Characteristics and General Chemistry

 ${}^{a}NTU = Nephelometric$ Turbidity Units.

 $c_{1987-1991}$ = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

d1992-1995 = Data taken from Arnett 1993, 1994, 1995, and 1996.

bCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

Statistically there appears to have been no significant adverse impact to Upper Three Runs from the construction activities.

Conductivity

During the CCWS, specific conductance in Upper Three Runs and Tims Branch was low compared to other onsite streams and the Savannah River (Newman et al. 1986). Specific conductance ranged from 1.4 to 20.1μ S/cm at the headmost waters of Upper Three Runs and increased in range from 20.6 to 68.9 μ S/cm near the confluence with the Savannah River (Table 5-3). Specific conductances measured from 1987 to 1991 were highest at Tims Branch 5 (mean 51.2 μ S/cm) and lowest at the upstream location, Upper Three Runs **IA** (mean 21.5 μ S/cm). Between 1992 and 1995, mean conductivity at Tims Branch was 64.5 μ S/cm, 76.75 μ S/cm at Crouch Branch, 45.5 μ S/cm at McQueen Branch, and 26.25 μ S/cm and 24 at the two Upper Three Runs locations (Table 5-3).

Conductivity values were monitored in McQueen Branch, Crouch Branch, and Upper Three Runs during and since construction of DWPF to determine the impact of construc tion activities on the tributaries' water quality and the effects of tributary water quality on Upper Three Runs water quality. During construction, mean conductivity ranged from 55.93 to 66.80 μ S/cm in McQueen Branch and from 90.02 to 108.27 μ S/cm in Crouch Branch. After construction conductivity ranged from 28.31 to 60.55 μ S/cm in McQueen Branch and from 87.27 to 95.02 µS/cm in Crouch Branch (Bodie and Scott 1995).

Major Anions and Cations

Alkalinity, Chloride, and Sulfate

Alkalinity, chloride, and sulfate concentrations in Upper Three Runs increased with dis tance downstream during the CCWS (Table 5-4). The waters near the confluence with the Savannah River had mean alkalinity and sulfate concentrations of 5.09 mg CaCO₃/l, chloride concentrations of 2.05 mg/I, and sulfate concentrations of 1.18 mg/l. Except for alka linity and sulfate in Tims Branch, similar concentrations of alkalinity, chloride, and sulfate have been measured in Upper Three Runs in the years following the CCWS. Alkalinity in Tims Branch has averaged 14.3 mg CaCO₃/l since the study.

Calcium, Magnesium, Sodium, and Potassium

Concentrations of total calcium, magnesium, sodium, and potassium in Upper Three Runs waters during the CCWS and as part of routine monitoring are given in Table 5-5. The CCWS determined that most of the calcium, magnesium, and sodium was in the dissolved phase (Newman et al. 1986). Mean potassium concentrations were below detection limits during CCWS and were not measured during routine monitoring. All monitoring data indi cate that concentrations of these cations are low, but increase with distance downstream. Tims Branch tended to have slightly higher concentrations of these cations relative to the mainstream Upper Three Runs.

Aluminum, Iron, and Manganese

Table 5-5 gives concentrations of total aluminum, manganese, and iron measured during the CCWS and routine monitoring. The highest concentrations of these cations were mea

Table 5-4. Upper Three Runs Major Anions

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

 $b_{1987-1991}$ = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

 $c_{1992-1995}$ = Data taken from Arnett 1993, 1994, 1995, and 1996.

Table 5-5. Upper Three Runs Major Cations (Total)

NA = Not analyzed.

Blank Spaces = Mean not calculated due to insufficient data in report. 3

CCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

 c 1992-1995 = Data taken from Arnett 1993, 1994, 1995, and 1996.

Nutrients

Table 5-6. Upper Three Runs Nutrients

NA = Not analyzed.

ND = None detected.

Blank Spaces = Mean not calculated due to insufficient data in report.

 a CCWS = Comprehensive Cooling Water Study (Newman et al. 1986).
 b 1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

'1992-1995 = Data taken from Arnett 1993, 1994, 1995, 1996.

 $\rm ^d$ All nondetectable quantities were excluded from the calculation of means.

e1992-1993.

f1993-1995.

91994-1995.

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Table 5-7. Upper Three Runs Trace Elements (Total)

 $NA = Not analyzed$.

ND = None detected.

Blank Spaces = Mean not calculated due to insufficient data in report.

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b_{1987-1991 = Data taken from Amett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.}

c1992-1995 = Data taken from Arnett 1993, 1994, 1995, 1996.

dAll nondetectable quantities were excluded from the calculation of means.

Priority Pollutants

Concentrations of all 88 tested volatile, acid, and base/neutral organics were below detec tion limits in waters in Upper Three Runs at each of the sampling locations. The 1984 study indicated that the natural quality of these streams was low in detectable synthetic organics and that potential groundwater outcropping of chlorinated hydrocarbons from the 300-M Area to the Upper Three Runs watershed was not occurring in detectable quantities at the time of the study (Lower 1987).

Pesticides, Herbicides, PCBs, and Volatile Organic Compounds

Water samples also are collected annually from Tims Branch and Upper Three Runs during routine monitoring and analyzed for pesticides, herbicides, PCBs, and volatile organic com pounds (VOCs). From 1987 to 1994, no analytes were detected in Upper Three Runs or Tims Branch. In 1995, pesticides and VOCs were detected in Tims Branch. VOCs and her bicides also were detected in Upper Three Runs (Table 5-8) (Arnett and Mamatey 1996).

Lower (1987) reports the results of analyses for pesticides, herbicides, and PCBs from 1982 to 1985, and results from 1967 to 1981 can be found in Gladden et al. (1985). During these periods, concentrations were also near or below detection limits at all locations.

Chemical, Including Radionuclide, and Toxicity Assessment Studies

Changes in M-Area Effluent

A chemical and biological study was conducted in Tims Branch from September 1988 to March 1990 to determine if changes to the discharges of the A-014 and A-015 outfalls had an adverse effect on communities in Tims Branch. Seven locations (Figure 5-6; 05 and 06) were sampled semimonthly for chemical parameters. ECS (1990) documented the results. Table 5-9 summarizes the data obtained in this study.

Table 5-8. Pesticides, Herbicides, and Volatile Organic Compounds Detected in Water Samples from SRS Streams in 1995. All Other Pesticides, Herbicides and Volatile Organic Compounds at all Other Sample Locations were Not Detected

Source: Arnett 1996.

^a See Figure 5-6.

During the study, a noticeable difference in water chemistry was observed along Tims Branch at TB-1 and TB-11. With the exception of total Kjeldahl nitrogen at Site 1 and tem perature at Site 3, all other chemical parameters at the sites were significantly higher than the control site. The A-015 discharge increased the phosphate content of Tims Branch; how ever, the phosphate value for Site 4 had recovered to the level of control Site 2. The A-014 discharge was the primary source for nitrates in Tims Branch. The most noticeable effect of nitrates was at Site 4, which was significantly different from the control site only during release of the effluent.

Overall, the discharge from M-Area did appear to impact the water quality of Tims Branch. A comparison of preoperational and postoperational data for Tims Branch and the outfall channels indicated that total suspended solids, phosphate, dissolved oxygen, and tempera ture were significantly higher during 1988-1990. The nitrate values for Site 1 and Site 3 were lower than in 1985-1986, due to the modifications of effluents discharged to A-014 and A-015. Post-operational phosphate values increased for all sites along Tims Branch.

M-Area Liquid Effluent Treatment Facility

A chemical and biological assessment was conducted to address the effects of the M-Area Liquid Effluent Treatment Facility on the chemistry and biota of Tims Branch and Upper Three Runs. The discharge point and eight locations (Figure 5-6) along Tims Branch and Upper Three Runs were sampled monthly for chemical parameters. The results of this study are documented in ECS (1987) and are summarized in Table 5-10.

The study indicated that a beaver pond immediately upstream greatly influenced the water chemistry at site TB-2 (Figure 5-6). High organic-matter concentrations in the waters of upstream sites TB-2, TB-0, and TB-1 created low dissolved oxygen conditions. Sites TB-3, TB-4, TB-5, and TB-6 had much higher mean dissolved oxygen concentrations.

Nitrate concentrations were highest at M-Out (maximum 47 mg/1), and decreased as the dis charge peak moved downstream to TB-4 (maximum 2.7 mg/1). Background nitrate concen trations at Tims Branch sites increased by a factor of three between 1985 and 1986. Nitrite concentrations had trends similar to the nitrate, but were one or two orders of magnitude lower. Ammonium and total Kjeldahl nitrogen levels did not appear to be related to the M-Area Liquid Effluent Treatment Facility discharges.

Uranium was detected two times in water samples at M-Out and at some upstream Tims Branch sites, and once in Upper Three Runs (TB-6). Zinc levels originally were elevated at M-Out and **TB-1.**

Traces of trichloroethylene were observed frequently at M-Out, with an overall mean con centration of 2.9 μ g/l. No detectable levels of tetrachloroethene or 1,1,1-trichloroethylene were measured.

F-/H-Area Effluent Treatment Facility

In June 1987, a 4-year biological study was initiated in anticipation of the fall 1988 startup of effluent discharges into Upper Three Runs by SRS F-/H-Area Effluent Treatment Facility (ETF). Using Hester-Dendy multiplate samplers, six sites were sampled quarterly for mac roinvertebrates. Sampling locations included one site upstream of the Road C bridge, four sites downstream of the Road C bridge, and one site on Tims Branch (Figure 5-6). The dis-

NA = Not analyzed.

^aSee Figure 5-6.

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Table **5-10.** (cont)

NA = Not analyzed.

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Table **5-10.** (cont)

NA **=** Not analyzed.

charge point for the **ETF** effluent (H-016) is at Road C. The results of this study are docu mented in Enwright Environmental (1990) and Chem-Nuclear Laboratory Services (1991) and Table 5-11.

Thirty-one parameters were analyzed monthly on water samples collected from the five sites on Upper Three Runs. Six parameters were analyzed monthly on water samples from the mouth of Tims Branch. Of the 31 water chemistry parameters analyzed, only two parameters were measured at higher levels at downstream sites than at the control site: nitrate-nitrogen and uranium. Nitrate-nitrogen levels also were found to be elevated in Tims Branch, which suggests that tributary, as being the source of elevated nitrate at downstream Upper Three Runs sites, not the F-/H-Area ETE Uranium was not sampled in Tims Branch. All other measured parameters exhibited similar concentrations upstream and downstream, with no observed impact by the F-/H-Area ETF.

Table 5-11. Water Quality Parameters for Stream Assessments for F-/H-Area ETF Effluent Discharge, July 1987 February 1990

NA = Not analyzed.

aSee Figure *5-6.*

Table **5-11.** (cont)

 $NA = Not analyzed$.

^aSee Figure 5-6.

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Table 5-11. (cont)

 $\bar{\alpha}$

NA = Not analyzed.

^aSee Figure 5-6.

 $\bar{\beta}$

Impact of the Mixed Waste Management Facility Groundwater Contamination

Algae

Although the type locality for several diatom species have their origin in Upper Three Runs (Patrick and Reimer 1966), the algae of this stream have not been quantitatively sampled except for three sampling stations that were sampled for periphyton ash-free dry weight and chlorophyll content between 1983 and 1985 (Specht 1987). These measurements were pri marily used for pooled data comparisons of thermal versus nonthermal streams on SRS. When extracted from this context, the data revealed limited descriptive information about the periphyton community of Upper Three Runs.

Macrophytes

Introduction

Aquatic macrophytes provide stream structure, substrate for periphyton development, cover and substrate for smaller animals, and a source of carbon for the stream system. Although aquatic macrophytes are an important component to the function of many aquatic systems, they tend to be less important in flowing waters. Streams such as Upper Three Runs tend to have little macrophyte colonization because of the closed canopy above the stream channel. The reduction in available sunlight falling directly upon the water limits macrophyte coloni zation to the few open areas of the channel such as those found at road bridges and power line crossings.

Comprehensive Cooling Water Study

The only data dealing with aquatic macrophytes in Upper Three Runs comes from the Com prehensive Cooling Water Study (Specht 1987) and was collected during 1984-1985. The data are from one sample at one station in the creek near a bridge and show limited macro phyte development. A survey of the entire stream to document the extent of this important component of the lower food chain would provide valuable information on the structure and condition of this SRS watershed.

Zooplankton

No studies have been conducted on the zooplankton in Upper Three Runs.

Macroinvertebrates

Sampling Locations and Methods

Aquatic Insects Survey

From September 1976 through August 1977, Morse et al. (1980) conducted an extensive sampling program for aquatic insects at six locations adjacent in the Upper Three Runs drainage basin. Sampling was conducted at Upper Three Runs at Roads 8-1 and F, Tinker Creek at Road 8-1, Mill Creek at Road E-2, Mill Creek at an unnamed fork 2.75 km (1.75 mi) upstream from Road E-2, and Boggy Gut at U.S. Forest Service at Road 781-4 (Figure 5-7). At each station, aquatic insects were collected biweekly from natural sub strates and by light trapping. Details of sampling methods can be found in Morse et al. (1980).

Comprehensive Cooling Water Study

During the Comprehensive Cooling Water Study, macroinvertebrates were collected monthly from November 1984 through September 1985 at two locations in Upper Three Runs (just above Road C and in the creek mouth) using Hester-Dendy multiplate samplers and drift nets (Chimney and Cody 1986; Firth et al. 1986). In the creek mouth, macroinver tebrates also were sampled quarterly from March 1982 through September 1984 using drift nets and Hester-Dendy multiplate samplers (Specht et al. 1984; O'Hop et al. 1985). Details of sampling methods can be found in Specht et al. (1984) and Firth et al. (1986). These results are summarized in Specht (1987).

Figure 5-7. Macroinvertebrate Sampling Stations for Upper Three Runs

F-/H-Area Effluent Treatment Facility Monitoring Program

Macroinvertebrates were sampled quarterly from July 1987 through July 1991 using Hester Dendy multiplate samplers and semiannually by qualitative sampling of natural substrates at five locations in Upper Three Runs in the vicinity of Road C. These samples were taken to determine if the F-/H-Area Effluent Treatment Facility effluent was adversely impacting the macroinvertebrate community of the stream. Beginning in March 1990 and continuing through July 1991, this sampling program was continued at three of the five stations.

Tims Branch Monitoring Program

Macroinvertebrates were sampled from June 1985 through December 1986 at five stations in the Tims Branch tributary of Upper Three Runs and at two locations in Upper Three Runs to determine if effluents from M Area and the A-/M-Area sanitary wastewater treatment plant were adversely affecting the Tims Branch macroinvertebrate community. Additional sam pling was conducted at four locations in Tims Branch from September 1988 through March 1990 using Hester-Dendy multiplate samplers.

Other Sampling Programs

Macroinvertebrates were sampled during the summer of 1993 at six locations in Upper Three Runs and in eight tributaries: Cedar Creek, Boggy Gut, Tinker Creek, Reedy Branch, Mill Creek, McQueen Branch, Crouch Branch, and Tims Branch. This one-month sample identified streams with great variability in populations among locations as a first step in developing a long-term sampling program. Hester-Dendy muliplate samplers were used to collect macroinvertebrates. Parameters measured included total number of taxa collected at each station, mean number of taxa per sampler, density of organisms, biomass, relative abun dance of functional feeding groups, and relative abundance of major taxonomic groups (Specht 1994b).

Macroinvertebrates also were sampled in September 1994 to develop a biotic index for southeastern streams using Hester-Dendy multiplate samplers. While not specifically designed to characterize SRS streams, these data contribute to a better understanding of the streams. Upper Three Runs at Road C and its tributaries, Tinker Creek and Mill Creek, were 3 of the 16 sampled locations (Specht and Paller 1995).

Past investigators collected from SRS locations and described "peculiar" species of a bivalve in the genus *Elliptio* (Britton and Fuller 1979). In 1989, *Elliptio* specimens were collected from seven locations on the SRS: the Savannah River, Upper Three Runs drainage (four sites), Pen Branch, and Lower Three Runs. The clams' morphology was described and com pared with known specimens. In addition, electrophoresis was used to elicit differences in enzyme structures.

Results

Aquatic Insects Survey

A total of 34,206 aquatic insects were collected during the one-year study (Morse et al. 1980, 1983). The study identified at least 551 species of aquatic insects, including at least 52 species and 2 genera new to science. Several other species were collected from Upper Three Runs that are either rarely collected or were not reported in South Carolina before this study. The species list that Morse et al. (1980, 1983) compiled for Upper Three Runs contains more species than have been reported for any other North American streams of compara ble size.

Morse et al. (1980) concluded: "It is clear that the aquatic insect fauna of the Upper Three Runs drainage is unusual. It not only includes many rare species, but also, because of its special combination of ecological characteristics, contains species not often found living together in the same freshwater system. This spring-fed stream is colder and generally clearer than most surface waters at its low elevation, reminiscent of unpolluted streams in northern North America or high in the Appalachian Mountains. On the other hand, its shifting sandy bottom and somewhat tea-colored water are visibly indistinguishable from those of other lowland southeastern streams. As a result, many typically northern and mountain species co-exist here with southern lowland species. In consideration of the many rare insect species in the Upper Three Runs drainage and of the several "typically northern" species co-existing here with an endemic southern fauna, it is probably safe to say that this stream has persisted for a very long time without appreciable alteration. It is therefore appropriate to regard the system as an outstanding example of an unpolluted, spring-fed, Sandhills waterway."

More recent studies of the aquatic insects in Upper Three Runs identified more than 650 species, including 104 species of caddisflies (Trichoptera) (Floyd et al. 1993). Ninety three species of caddisflies, representing 14 families, were identified during biweekly col lections over a one-year period at two locations on Upper Three Runs downstream of all SRS discharges. Three species *(Oxyethira setosa, Triaenodes smithi, and Nyctiophylax seratus)* are new distributional records for South Carolina. Two species of *Triaenodes* are new to science. Other species considered endemic to the Upper Three Runs drainage, rare outside the drainage, or of limited distribution included *Cheumatopsyche richardsoni, Oxyethira dunbartonensis, Protoptila morettii, Hydrophysche elissoma, Triaenodes ochraceus, Neotrichiafalca, Oecetis morsei, and Pycnopsyche virginica* (Floyd et al. 1993).

Comprehensive Cooling Water Study

Introduction

Macroinvertebrate drift and Hester-Dendy multiplate data are summarized in this section. More detailed information can be found in Specht et al. (1984), O'Hop et al. (1985), Firth et al. (1986), Chimney and Cody (1986), and Specht (1987).

Taxa Collected

Sixty-two macroinvertebrate taxa (Table 5-12) were collected from multiplate samplers near Road C during the 1984-1985 program. However, these data are not directly compa rable to the data collected by Morse et al. (1980, 1983) or to data from the F-/H-Area Effluent Treatment Facility monitoring program since there were differences in the level of taxonomic resolution among the studies.

Table **5-12.** Summary Statistics for Macroinvertebrate Data from Hester Dendy Multiplate Samplers, Upper Three Runs, October 1984-September **1985**

$AFDW =$ ash-free dry weight.

Densities

The mean density of organisms on the multiplate samplers near Road C was 582.7 organ isms/ $m²$ in 1984-1985, while densities in the creek mouth were much higher, averaging 2839.5 organisms/m² (Table 5-12). However, biomass was higher at Road C (0.500 g/m²) than in the creek mouth (0.190 g/m^2 ; Table 5-12) The number of taxa collected per sampler was similar at the two stations, averaging 15.1 taxa/sampler at Road C and 16.1 taxa/sampler in the creek mouth (Table 5-12). The macroinvertebrate collections from multiplate samplers at Road C were overwhelmingly dominated by tanytarsine chirono mids (70.1%; Firth et al. 1986). Other abundant taxa included unidentified chironomids (21.6%) and the mayfly, *Stenonema modestum* (1.6%). Dominant taxa in the creek mouth were orthoclad chironomids (28.0%), chironomini chironomids (16.7%), tanytarsine chi ronomids (8.5%), brachycentrid and hydropsychid caddisflies (3.1% for each taxa; Chim ney and Cody 1986).

Macroinvertebrate drift densities were much higher at Road C (6848.9 organisms/1000 m^3 ; Firth et al. 1986) than in the creek mouth (377.6 organisms/1000 m³; Chimney and Cody 1986). Abundant taxa in the drift at Road C included chironomids (21.0%), hydrop sychid caddisflies (10.5%), baetid mayflies (8.7%), blackfly larvae (6.7%), and the stone fly *Taeniopteryx longicera* (6.5%; Firth et al. 1986). The composition of drift in the creek mouth was dominated more by chironomids (52.9%) but also included relatively high abundances of baetid mayflies (12.3%), hydropsychid caddisflies (3.4%) and blackfly lar vae (7.4%; Chimney and Cody 1986).

Effluent Treatment Facility Monitoring Program

Taxa Collected

During the five-year monitoring program, 292 macroinvertebrate taxa were collected from Upper Three Runs from quantitative sampling (Hester-Dendy multiplate samplers) and sampling of natural substrates combined (Table 5-13). Of these taxa, 96 were collected exclusively from natural substrates.

Dominant Taxa

Introduction

Table 5-14 presents the relative abundance of major taxonomic groups of macroinverte brates at each sampling station.

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Table 5-13. (cont)

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Table 5-13. (cont)

Table **5-13.** (cont)

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Table **5-13.** (cont)

Table **5-13.** (cont)

Table 5-13. (cont)

aSpecimen not identifiable to genus.

Table 5-14. Relative Abundance (Percent Composition) of Major Taxonomic Groups of Macroinvertebrates Collected from Hester-Dendy Multiplate Samplers, Upper Three Runs, July 1987-July 1991

Chironomids

At all stations, chironomids were the most common group of macroinvertebrates collected, ranging from 49.9 to 55.3% of the total number of organisms collected from the multiplate samplers. Most of the chironomids were in the subfamilies Orthocladiinae and Chironomi nae. Dominant species of chironomids collected from the multiplate samplers included *Rheocricotopus robacki, Polypedilum convictum, Tvetenia discoloripes,* and *Corynoneura* sp.. *Rheotanytarsus* sp., *Conchopelopia* sp., *Cladotanytarsus* sp., and *Rheosmittia* sp., and *Parametriocnemus lunbecki* also were abundant on natural substrates. The most common nonchironomid species of dipteran was the blackfly, *Simulium jonesi.*

Trichoptera

Trichoptera (caddisflies) composed 11.0-13.4% of the organisms collected from the multi plate samplers. The most abundant caddisflies on the multiplates were *Hydropsyche elis soma* and *Cheumatopsyche* sp. These two taxa and *Brachycentrus numerosus* were the most commonly collected caddisflies on natural substrates.

Plecoptera

Plecoptera (stoneflies) were much more common in Upper Three Runs than in any other SRS streams, where they made up 11.4-13.2% of the macroinvertebrates collected from the multiplate samplers. Stoneflies prefer cool water, which is provided in Upper Three Runs by naturally occurring flows of groundwater into the stream. The most common stoneflies were *Perlesta placida, Acroneuria abnormis,* and *Isoperla* sp.

Ephemeroptera and Coleoptera

Ephemeroptera (mayflies; 7.2-9.4%) and Coleoptera (beetles; 5.6-10.7%) accounted for most of the remaining organisms collected from the multiplate samplers. *Stenonema modestumlsmithae, Heptageniaflavescens,* and *Paraleptophlebia guttata* were the domi nant mayflies, while *Stenelmis crenata* and several other elmid beetles were the most com mon beetles in the stream.

Mean Number of Taxa and Density of Organisms

The mean number of taxa collected from multiplate samplers at a given station (all multi plates combined) ranged from 27.0 to 45.0 during the five year study (Table 5-15), while the mean number of taxa per sampler ranged from 12.3 to 22.9 (Table 5-16). Mean densi ties of organisms ranged from 213.9 to 799.2 organisms/ $m²$ (Table 5-17), which is lower than the densities reported for most other SRS streams. However, macroinvertebrate biom ass (ash-free dry weight) was comparable to most other streams, ranging from 0.1399 to 0.6860 g/m² (Table 5-18). Distinct seasonal variations in biomass were observed, with biomass peaking in the spring as organisms reached their maximum size and declining over summer, just after the period of peak emergence.

Macroinvertebrate diversity, as calculated by the Shannon-Weaver diversity index was high at all stations, ranging from 3.90 to 4.59 (Table 5-19). Generally, diversity values above 3 are indicative of unimpacted streams.

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Table 5-15. Mean Number of Taxa Collected During Each Sampling Period from Hester-Dendy Multiplate Samplers, Upper Three Runs, July 1987-July 1991

Table 5-16. Mean Number of Taxa Collected Per Hester-Dendy Multiplate Sampler, Upper Three Runs, July 1987 July 1991

Table 5-17. Mean Density of Macroinvertebrates (no./m²) Collected from Hester-Dendy Multiplate Samplers, Upper Three Runs, July 1987-July 1991

Table 5-18. Mean Biomass (g/m^2) of Macroinvertebrates Collected from Hester-Dendy Multiplate Samplers, Upper Three Runs, July 1987-July 1991

Table **5-19.** Shannon-Weaver Diversity of Macroinvertebrates Collected from Hester-Dendy Multiplate Samplers, Upper Three Runs, July 1987-July 1991

	Station									
Year	1(05)	2(07)	3(08)	4 (09)	5(10)					
1987	4.22	4.42	4.50	4.56	4.48					
1988	4.38	4.59	4.56	4.46	4.38					
1989	4.42	4.35	4.08	4.15	4.22					
1990	4.16	4.31	4.31	3.90	4.28					
1991	4.08	4.17			4.12					

Temporal Trends

During the course of the five-year study, declines in the mean number of taxa per station, mean number of taxa per sampler, mean density of organisms, and diversity were observed at all stations (Table 5-15, Table 5-16, Table 5-17, and Table 5-19). The decline in density of organisms was the most dramatic, with mean density for all stations combined declining from 667.5 organisms/m² in 1987 to 273.8 organisms/m² in 1990. In 1991, densities increased somewhat to an average of 427.0 organisms/m². Taxa that exhibited the greatest declines in abundance were primarily chironomids, such as the orthoclads, *Rheocricotopus robacki, Thienemanniella* spp., *Tvetenia discoloripes, Orthocladius* spp., and *Cricotopus* spp.; the Chironomini midge, *Polypedilum convictum;* and the tanytarsine midge, *Rheot anytarsus* spp. Species richness among Ephemeroptera, Plecoptera, and Trichoptera, which are generally considered to be pollution-intolerant organisms, did not decrease signifi cantly, and the relative abundance of Plecoptera and Trichoptera actually increased, although overall densities of these orders remained relatively constant. Although the densi ties declined, the total biomass of the macroinvertebrate community did not decrease over time (Table 5-18), which indicates that larger organisms were being collected.

The observed changes in the macroinvertebrate community are not attributable to effects from the F-/H-Area Effluent Treatment Facility effluent since they occurred at the upstream reference station (Station 1 [05]) as well as the stations located downstream from the F-/H Area Effluent Treatment Facility. At this time, there is no explanation for the observed changes, but data from 1991 indicate that the stream appears to be recovering somewhat from the unknown perturbation. Possible perturbations included increased siltation from the Defense Waste Processing Facility construction area and/or toxic inputs to the stream from areas upstream from SRS.

Tims Branch Monitoring Program

1985-1986 Sampling Program

The 1985-1986 sampling program collected 147 macroinvertebrate taxa in Tims Branch, including 97 taxa from multiplate samplers and 50 taxa collected exclusively from natural substrates during qualitative sampling efforts (Table 5-20). The fewest total number of taxa (23) was collected from multiplate samplers at Station TB-2, which was upstream from the M-Area effluent discharge but just downstream from a stagnant pond. Outflow from the

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Table **5-20.** (cont)

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Table **5-20.** (cont)

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Table 5-20. (cont)

WSRC-TR-97-0223-0223 5-48 WSRC-TR-97-0223 pond often had no detectable levels of dissolved oxygen, and the macroinvertebrate community of this station was typical of streams with high organic loading. The mean num ber of taxa collected per sampler was also lowest at Station TB-2 (5 taxa/sampler; Table 5 21) and increased in a downstream direction to 12.8 taxa per sampler in the mouth of the creek (Station TB-4). Dominant groups of macroinvertebrates at Station TB-2 included oli gochaetes (66.4%) and chironomids from the tribe Chironomini (23.1%; Table 5-22). The macroinvertebrate community downstream from the M-Area effluent discharge gradually improved to a community more typical of less impacted streams, with the oligochaetes and Chironomini being replaced by orthoclad and tanytarsine chironomids, as well as Coleoptera (beetles), Ephemeroptera (mayflies), Plecoptera (stoneflies), and Trichoptera (caddisflies). At Station TB-4, 57 taxa were collected from multiplate samplers. Mayflies composed 19.5% of the total organisms collected; stoneflies 3.6%; caddisflies, 4.8%; bee tles, 5.5%; orthoclads, 30.9%; and tanytarsine chironomids 16.6% (Table 5-22).

The total density of organisms in 1985-1986 was highest at Station TB-2 (3679.1 organ isms/m²) and generally declined in a downstream direction to 584.0 organisms/m² at Station TB-4 (Table 5-21). Total biomass was also highest at Station TB-2 (0.253 g/m²), declined to 0.072 g/m² at Station TB-1, and then increased to 0.123 and 0.207 g/m² at Stations TB-3 and TB-4, respectively (Table 5-21).

1988-1990 Sampling Program

Macroinvertebrate identification during the 1988-1990 study generally was performed only to the family level; therefore direct comparisons of taxa richness with the 1985-1986 study are not possible. Also, no taxa list is presented for the 1988-1990 program since all of the families identified were present in the 1985-1986 study.

The macroinvertebrate community in Tims Branch in 1988-1990 differed somewhat from that of 1985-1986. Station TB-2 still was dominated by pollution-tolerant taxa, but had shifted from a community dominated by oligochaetes to one dominated by pollution-toler ant chironomids of the tribe Chironomini, which constituted 90.7% of the organisms col lected from multiplate samplers at this station (Table 5-23). At the most downstream station (Station TB-4), mayflies and caddisflies were less abundant than in 1985-1986, composing 14.5% and 1.2%, respectively, of the macroinvertebrates on the multiplate samplers, while stoneflies were more abundant, making up 14.3% of the organisms collected in 1988-1990 (Table 5-23), as compared to just 3.6% in 1985-1986.

The density of organisms was much lower at Station TB-2 in 1988-1990 (421.1 organ isms/m² as compared to 3679.1 in 1985-1986). Densities at Stations 1 and 3 were also somewhat lower, while mean densities at Station 4 were higher (763.5 organisms/ $m²$ as compared to 584.0 in 1985-1986; Table 5-21). At present, there is no explanation for the observed declines. The decline in density of organisms is of particular concern since densi ties were observed to decline steadily during the last year of the sampling program (Figure 5-8).

In contrast to the decline in macroinvertebrate densities, macroinvertebrate biomass was higher at three of the four sampling stations in 1988-1990 than in 1985-1986 (Table 5-21). The increased biomass with decreased densities indicates that the relative size of the organ isms has increased, possibly due to shifts in species composition.

Table 5-21. Summary Statistics for Macroinvertebrate Data from Hester-Dendy Multiplate Samplers, Tims Branch, **1985-1990**

Table 5-22. Relative Abundance (Percent Composition) of Major Taxonomic Groups of Macroinvertebrates, Tims Branch, June 1985-December 1986

a_{Includes} subtaxa values in parentheses.

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Table 5-23. Relative Abundance (Percent Composition) of Major Taxonomic Groups of Macroinvertebrates, Tim's Branch, September 1988-March 1990

^aIncludes subtaxa values in parentheses.

Figure **5-8.** Density of Macroinvertebrates in Tims Branch, **1985-1990**

1993-1994 Sampling Program

Upper Three Runs and three of its tributaries (Tims Branch, McQueen Branch, and Crouch Branch) received effluent from 25 National Pollutant Discharge Elimination System (NPDES) outfalls in A, B, F, H, M, and S Areas and the Flowing Streams Facility at Road C. The new NPDES permit issued in 1996, eliminated 10 of theses outfalls by deletion or consolidation of several outfalls into one. Upper Three Runs supports a reasonably rich macroinvertebrate community. Total number of taxa collected per station in 1993 ranged from 42 at Road F to 52 at Road C (Table 5-24). Thirty-six taxa were collected at Road C in 1994. The mean number of taxa per sampler in 1993 ranged from 20.0 at Banks Mill Road (offsite) to 26.8 at Road C. In 1994, the mean number of taxa per sampler at Road C was 20.8. Density ranged from 444.7 organisms/m² at Road F to 1378.8 organisms/m² at the railroad bridge upstream of Road C in the 1993 study and was 765.4 organisms/ $m²$ in 1994 at Road C. Biomass in 1993 ranged from 0.0296 g/m^2 at Road A to 0.8470 g/m^2 at Road 8-1. Biomass was 0.0390 g/m^2 at Road C in 1994 (Table 5-24). The most abundant group of macroinvertebrates collected at all of the stations in 1993 were the orthoclad chironomids (33.7-57.0%). In 1994, the most abundant group was the Chironomini (31.4%). Other abun dant groups were Tanytarsini chironomids, Chironomini chironomids, Trichoptera, and Coleoptera in 1993 and Orthocladiinae, Tanytarsini, other Diptera, and Ephemeroptera in 1994 (Table 5-25). Collector-gatherers was the most abundant functional feeding group at any station in either year (Table 5-26). No longitudinal trends were observed for any mea sured population parameter.

Sampled tributaries of Upper Three Runs included Cedar Creek and Boggy Gut (both upstream of the SRS boundary), Tinker Creek, Reedy Branch, Mill Creek, McQueen Branch, Tims Branch, and Crouch Branch. The data indicate that the three tributaries that receive NPDES discharges (Tims Branch, McQueen Branch, and Crouch Branch) are per turbed to some extent as evidenced by fewer total taxa, fewer mean taxa per sampler, and lower densities (Table 5-27).

Crouch Branch was the most disturbed of the tributaries sampled in 1993 with only 19 taxa collected, none of which was the more sensitive Ephemeroptera-Plecoptera-Trichorptera taxa. Chironomini chironomids dominated its community (Table 5-28), generally an indica tion of perturbed conditions. Crouch Branch had relatively low oxygen concentrations, which may be responsible, at least in part, for the observed perturbation. In addition, Crouch and McQueen Branches had high suspended solids loads during the construction of the Defense Waste Processing Facility which also may have had some effect on the streams' macroinvertebrate communities. Crouch Branch was also the most disturbed stream sam pled in 1994. It had 25 taxa and was dominated by Chironomini and oligochaetes (Table 5 27 and Table 5-28).

The remainder of the sampled tributaries showed no evidence of perturbation in 1993. Mill Creek had the highest number of taxa (63), and Tinker Creek had the fewest (46). Dominant taxa in most streams included the orthoclad chironomids, the Tanytarsine chironomids, a Tanypodinae chironomid, and a Chironomini chironomid (Table 5-27 and Table 5-28). As in Upper Three Runs proper, the most abundant functional group was the collector-gatherers (Table 5-29). Based on the Hester-Dendy data from 1994, Tims Branch showed evidence of some disturbance (Specht 1994b; Specht and Paller 1995).

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Table 5-24. Mean Number of Taxa, Density, and Biomass For Macroinvertebrates collected on Hester-Dendy Multi plate Samplers in Upper Three Runs, July-August 1993 and September 1994

Source: Specht 1994b; Specht and Paller 1995.

Table 5-25. Percent Composition of Macroinvertebrate Taxa Collected on Hester-Dendy Multiplate Samplers in Upper Three Runs, July-August 1993 and September 1994

Table **5-26.** Relative Abundance (Percent Composition) of Macroinvertebrate Functional Groups in Upper Three Runs, July-August **1993** and September 1994.

Source: Specht 1994b; Specht and Paller 1995.

Table 5-27. Mean Number of Taxa, Density, and Biomass For Macroinvertebrates collected on Hester-Dendy Multi plate Samplers in Tributaries of Upper Three Runs, July-August 1993 and September 1994).

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Table 5-28. Percent Composition of Macroinvertebrate Taxa Collected on Hester-Dendy Multiplate Samplers in Tributaries of Upper Three Runs, July-August 1993 and September 1994.

	1993										1994				
	Cedar Creek	Boggy Gut	Tinker Creek at Kennedy Pond Road	Tinker Creek at 2-1	Reedy Branch	Mill Creek at Road E-2	McQueen Branch at Road F	N at Road Tims Branch	Tims Branch at Road C	4 Crouch Branch at Road	Tinker Creek at Kennedy Pond Road	$E-2$ Mill Creek at Road	Tims Branch at Road 2	4 Crouch Branch at Road	
Hydra	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.06	0.00	0.00	
Ephemeroptera	3.22	2.76	13.42	2.91	4.9	6.46	0.68	0.00	4.57	0.00	19.95	9.95	0.39	2.24	
Plecoptera	0.31	3.37	0.00	3.4	0.75	0.00	7.19	0.00	3.05	0.00	0.00	0.00	0.33	0.00	
Trichoptera	9.03	2.14	6.28	0.97	4.26	3.99	1.71	8.05	2.49	0.00	2.99	2.20	0.07	0.00	
Coleoptera	1.35	12.25	2.71	18.84	6.08	1.08	1.02	17.30	11.91	1.10	1.00	0.52	2.08	0.00	
Odonata	0.21	0.31	0.11	0.00	0.21	0.38	3.08	0.18	0.28	1.10	0.31	0.13	0.13	3.59	
Megaloptera	0.10	1.07	1.19	0.10	0.21	0.00	0.68	0.00	0.69	0.00	0.06	0.13	0.00	0.00	
Hemiptera	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.14	0.55	0.00	0.00	0.00	0.00	
Other Diptera	5.19	3.68	0.76	4.66	1.71	2.22	4.11	2.78	1.25	1.66	0.50	0.19	1.69	0.90	
Tanypodinae	1.35	5.36	1.52	4.76	9.49	12.29	18.84	6.57	9.28	2.21	2.00	4.85	3.65	9.87	
Tanytarsini	20.67	24.96	24.89	10.10	42.96	33.02	24.66	31.36	17.31	5.52	51.81	46.86	57.20	4.48	
Orthocladiinae	48.29	23.58	19.81	46.21	6.50	17.87	8.56	26.83	38.23	0.55	4.74	7.24	17.72	0.45	
Chironomini	7.68	20.37	28.03	6.99	21.54	22.56	29.10	4.81	10.66	86.19	11.97	23.79	11.92	56.05	
Pseudochiromomini	0.00	0.00	0.00	0.10	0.11	0.00	0.00	0.00	0.00	0.55	0.00	0.00	0.00	0.00	
Oligochaeta	2.60	0.15	0.22	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.94	3.88	3.65	17.94	
Gastropoda	0.00	0.00	0.22	0.00	1.07	0.06	0.34	0.00	0.00	0.00	0.19	0.00	0.07	1.35	
Pelecypoda	0.00	0.00	0.22	0.00	0.11	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	
Amphipoda	0.00	0.00	0.00	0.00	0.11	0.06	0.00	2.04	0.00	0.00	0.00	0.00	0.00	0.00	
Hydracarina	0.00	0.00	0.54	0.97	0.00	0.00	0.00	0.09	0.14	0.55	0.00	0.00	0.00	0.00	
Decapoda	0.00	0.00	0.11	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.06	0.06	0.06	0.45	
Nemertea	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.06	0.00	0.13	0.00	
Turbellaria	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	3.12	0.06	0.00	1.79	
Bivalva	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.06	0.00	0.00	0.00	
Hydracarina	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.19	0.06	0.00	0.98	
Total	100.00	100.00	100.03 100.01		100.01	99.99	99.97		100.01 100.00	99.98	99.95	99.98	99.03	100.09	

Table 5-29. Relative Abundance (Percent Composition) of Macroinvertebrate Functional Groups in Tributaries of Upper Three Runs, July-August 1993 and September 1994

	1993											1994				
	Creek Cedar ⁽	$\ddot{5}$ Boggy	Tinker Creek at Kennedy Pond Road	$2-1$ Creek at Road Tinker	Reedy Branch	E.2 Creek at Road Eik	McQueen Branch at Road F	\mathbf{a} Tims Branch at Road	ပ Tims Branch at Road	4 at Road Crouch Branch	Creek at Kennedy Pond Road Tinker	Creek at Road E-2 Mill	Tims Branch at Road 2	₩ Crouch Branch at Road		
Collector- Gatherers	65.94	75.19	70.88	82.62	76.33	72.43	54.45	79.93	72.58	93.37	43.20	78.54	62.93	78.48		
Collector-Fil- terers	24.09	7.04	14.07	4.95	4.26	9.19	0.34	11.19	8.17	1.10	37.66	9.31	25.54	1.35		
Predators	2.28	10.26	4.33	9.13	12.05	15.15	25.34	8.33	12.60	5.52	6.17	5.43	6.71	16.59		
Scrapers	1.45	1.84	9.85	0.97	3.52	1.96	2.40	0.00	2.22	0.00	11.91	4.91	0.33	2.69		
Shredders	4.57	5.67	0.87	2.33	3.73	1.08	17.47	0.56	4.43	0.00	0.81	1.36	4.50	0.00		
Piercer-Herbi- vores	1.66	0.00	0.00	0.00	0.11	0.19	0.00	0.00	0.00	0.00	0.25	0.45	0.00	0.90		
Total	99.99	100.00	100.00	100.00	100.00	100.00	100.00	100.01	100.00	99.99	100.00	100.00	100.01	100.01		

Bivalve Mollusks

Mill Creek is a tributary of Upper Three Runs. Recent research indicates that the taxonomy of the bivalve population in the Upper Three Runs' drainage is more complex than was pre viously suspected.

Based on conchology and electrophoresis it is apparent that while mollusks in the Savannah River are *Elloptio icterina,* the traditional classification scheme for Upper Three Runs *Elliptio* needs to be modified. The Mill Creek *Elliptio* population, formerly considered E. *icterina,* contains two discrete species, *E. compeanata* and one whose shells match the holotype of *E. hepatica* (Davis and Mulvey 1993).

Fish

Introduction

Upper Three Runs was sampled in 1984-1985, 1992, and 1993. The 1984-1985 samples were taken as part of the CCWS. These samples included ichthyoplankton collections from the lower reaches of the stream. The 1992 samples were taken as a dual effort to character ize stream fish assemblages on SRS and assess possible impacts from the outcropping of contaminated groundwater from F and H Areas into Upper Three Runs.

Comprehensive Cooling Water Study

Introduction

Most of the sampling in Upper Three Runs during the CCWS was devoted to the collection of ichthyoplankton from sample stations at Road C, Road A, and in the creek mouth (Figure 5-9). Samples were collected with 0.505-mm mesh nets weekly during daylight from Feb ruary through July during 1984 and 1985 (Paller 1985; Paller et al. 1986).

Ichthyoplankton Taxa Collected

Paller (1985) collected 358 ichthyoplanktons from Upper Three Runs during 1984, and Paller et al. (1986) collected 217 during 1985 (Table 5-30 and Table 5-31). The dominant taxa during both years was spotted sucker *(Minytrema melanops).* Other relatively abundant taxa were darters *(Etheostoma* spp.), crappie *(Pomoxis nigromaculatus),* minnows (prima rily *Notropis* spp.), blueback herring *(Alosa aestivalis),* and American shad *(Alosa sapidis sima).* The occurrence of American shad in Upper Three Runs reflects the relatively large size and substantial flow of this creek that make it a suitable spawning area for this species. Perhaps the most unusual feature of the ichthyoplankton assemblage in Upper Three Runs was the relatively high abundance of spotted sucker larvae. Larvae of this species were not

□ Ichthyoplankton Stations A Area

- Figure 5-9. Fish Sampling Stations in Upper Three Runs
- Table 5-30. Number and Percent Composition of Ichthyoplankton Collected from Upper Three Runs, March 14-June 3, 1984

found in such numbers in any of the other SRS creeks, suggesting that Upper Three Runs is an important spawning site for spotted sucker.

Ichthyoplankton Densities

Mean ichthyoplankton densities at the three sample stations in Upper Three Runs ranged from 29/1000 m^3 to 69/1000 m^3 during 1984 (Figure 5-10). Mean densities during 1985 ranged from $21/1000 \text{ m}^3$ to $41/1000 \text{ m}^3$. These densities were low to moderate compared with densities in the other unimpacted creeks sampled during the CCWS (Paller 1985; Paller et al. 1986). There were few consistent differences among sample stations except that the sample station in the creek mouth supported more taxa, partly because of the accessibility of this sample station to riverine taxa such as the anadromous American shad and blueback herring.

Figure 5-10. Mean Ichthyoplankton Density at Three Stations in Upper Three Runs, 1984 and 1985

Studies to Assess Impacts from F and H Areas

Introduction

Paller (1993) electrofished sample sites on Upper Three Runs, several tributary streams, and a control stream, Hollow Creek, to assess possible impacts of the F- and H- Area seepage basins on fish assemblages in Upper Three Runs. The tributaries include two third-order streams: Cedar Creek, north of the SRS; and Tinker Creek, which joins Upper Three Runs upstream from the F- and H- Area seepage basins. The third tributary was Crouch Branch, a small first-order tributary of Upper Three Runs that drains F and H Areas. The control stream, Hollow Creek, was selected because of its proximity to Upper Three Runs and its generally similar size, bottom substrate, channel morphometry, and water quality.

There were eight sample stations on Upper Three Runs (Figure 5-9), four (Stations 4-8) downstream from and four upstream from F and H Areas. There was one sample station on Cedar Creek (Station 9), one sample station on Tinker Creek (Station 10), one sample station on Crouch Branch (Station 11), and three sample stations (Stations 12-14) on Hollow Creek. Each sample station consisted of three 100-m stream reaches. Repeated electrofishing passes (usually seven) were made through each reach until no or few fish were collected. Habitat data also were collected, consisting of stream width, and depth, current velocity, number of logs and stumps, substrate type, dissolved oxygen concentration, pH, temperature, and con ductivity. Percentages of stream bottom area covered or overhung by submerged debris (leaves, twigs, and wood fragments), submerged brush/snags, submerged root masses, and aquatic macrophytes. Low-growing riparian vegetation were estimated visually.

Comparison to Other SRS Streams

In general, the fish assemblages in Upper Three Runs were typical of those in third- and higher-order streams on SRS (Paller 1992). Shiners (Cyprinidae) and sunfishes (Cen trarchidae) numerically dominated the fish assemblages at most Upper Three Runs sample stations. Larger predatory and benthic insectivorous species were present, as is typically the case in larger streams. The smaller tributary sample stations were strongly dominated by shiners, followed by pirate perch, madtoms, and darters. This pattern, too, is typical of unim pacted streams on the SRS (Paller 1992).

Differences Among Sample Stations

Differences in number of species and number of individuals among sample stations did not indicate degradation of the fish assemblages downstream from the waste management areas. Three of the five highest mean species numbers were observed at sample stations that were downstream from F and H Areas (Table 5-32). The lowest mean species numbers were observed upstream from the confluence of Upper Three Runs and Tinker Creek, possibly reflecting smaller stream size, other habitat factors, or offsite impacts. There were no signif icant differences in total number of fish per sample station; densities at the stations down stream from F and H Areas were within the range or higher than numbers at the stations upstream of F and H Areas.

Canonical discriminant analysis was used to assess differences among sample stations based on the numbers of the 10 most abundant species. Sample stations were divided into three groups: Upper Three Runs upstream from the waste management areas, Upper Three Runs downstream from the waste management areas, and Hollow Creek. Differences among the three station groups were significant at P<0.05. Upper Three Runs stations downstream from Environmental Information Document-SRS Ecology *Section 5.1-Upper Three Runs* Eish Fish Fish Exerces From The Savannah River **Fish** *Fish* Fish *Fish Chapter 5*—Streams, Reservoirs, and the Savannah River *Fish Chapter Disco Pines* Hallow Creak and Savaral

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the waste management areas differed from Upper Three Runs sample stations upstream from the waste management areas in that they had higher numbers of many species and from Hollow Creek sample stations in that they had lower densities of coastal shiner *(Not ropis petersoni),* dusky shiner *(Notropis cummingsae)* and redbreast sunfish *(Lepomis auri tis).*

Differences Among Different Habitats

Canonical discriminant analysis was used to assess habitat differences that could be responsible for the observed differences in fish assemblage structure among station groups. Habitat differences among the three station groups were significant at P<0.05. Upper Three Runs stations downstream from the waste management area were wider, deeper, and had faster currents than both Upper Three Runs sample stations upstream from the waste management area and Hollow Creek sample stations. The sample station groups also dif fered in macrophyte coverage, coverage by brush, and other factors. These and related hab itat differences can exert strong influences on fish assemblage structure and fish-capture probability and were probably responsible for the observed differences among station groups.

Effects of F- and H-Area Operations

Paller (1993) observed that natural changes in species composition resulting from increases in stream size and other longitudinal habitat changes would be expected to be gradual and continuous. In contrast, changes caused by the introduction of toxicants from a point source could be abrupt, with the greatest change occurring near the discharge point. However, indices of similarity in species composition between adjacent station pairs indi cated less differences between the sample station just upstream from the F and H Areas and the sample station just downstream from F and H Areas than between the putatively unimpacted sample stations farther upstream. Thus, this analysis also supported the con clusion of no measurable community level impacts associated with contaminants outcrop ping from F and H Areas. Further conclusions about the impacts of F and H Areas on fish in Upper Three Runs must remain tentative until additional data are available, including the analyses of contaminant levels in fish from potentially polluted sites.

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5.2 Beaver Dam Creek

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Drainage Description and Surface Hydrology

General Description

Beaver Dam Creek, which is approximately 5 km (3 mi) long and drains an area of approxi mately 2.2 km^2 (0.85 mi²), originates at the effluent outfall canal of D Area and flows south, parallel to Fourmile Branch, to the Savannah River (Figure 5-11). The discharge of Beaver Dam Creek mixes with a portion of Fourmile Branch discharge in the Savannah River flood plain swamp before entering the Savannah River. Beaver Dam Creek is deep, narrow, and channelized in its headwaters and near its mouth and opens into a slough-like channel in its midreaches (Firth et al. 1986).

Effluents Contribution

Before SRS operations, Beaver Dam Creek probably had only intermittent or very low flow. Beaver Dam Creek has received thermal effluents since 1952 as a result of cooling water discharges from combined heavy water production (shut down in 1982) and operation of a coal-fired power plant in 400-D Area. In fall 1988, extended reactor outages decreased site electrical and steam demands, thus reducing the operations and thermal effluent of 400-D Area. Currently, Beaver Dam Creek receives condenser cooling water from the coal-fired power plant, neutralization waste water, sanitary waste water, ash basin effluent waters, and various laboratory waste waters.

Figure 5-11. Location of Beaver Dam Creek on SRS

Flow Measurements

The U. S. Geological Survey measures flow at 400-D (Figure 5-12). Table 5-33 summarizes flow statistics of Beaver Dam Creek at 400-D. Records for the station at 400-D begin in June 1974. In water year 1995, the mean flow of Beaver Dam Creek at 400-D was $2.0 \text{ m}^3/\text{s}$ (92 ft³/s). Over the period of record (water years 1974-1995) at 400-D, the mean flow was 2.3 m³/s (82 ft³/s), the 7-day low flow was 0.45 m³/s (16 ft³/s), and the 7Q10 was 0.0003 $m³/s$ (0.01 ft³/s). The maximum, minimum, and mean daily flows from October 1982 to September 1995 at 400-D are shown in Figure 5-13.

Figure 5-12. Flow Sampling Station on Beaver Dam Creek

Table 5-33. Flow Summary for Beaver Dam Creek

 $a_{cms} = Cubic meters per second.$

 b cfs = Cubic feet per second.</sup>

Figure 5-13. Maximum, Mean, and Minimum Flow Measurements from October 1982 to September 1995 for Beaver Dam Creek at 400-D, October 1982-September 1995

Water Chemistry and Quality

Studies and Monitoring

Water-Quality Monitoring

Westinghouse Savannah River Company Environmental Monitoring Section (EMS) con ducted routine water-quality monitoring of Beaver Dam Creek from 1973 to 1986. EMS monitored one location on Beaver Dam Creek near the Savannah River Swamp (Figure 5 14; location 08 on map) monthly for physical and biological water quality indicators and quarterly for metals. Since 1986, water-quality measurements consisting of water tempera ture, pH, dissolved oxygen concentration, conductivity, and oxidation/reduction potential have been taken hourly downstream of D-Area discharge. Lower (1987) summarizes water quality monitoring results for Beaver Dam Creek from 1983-1985. All routine water quality monitoring data reported in the following sections can be found in the annual SRS Environ mental Reports.

Comprehensive Cooling-Water Study

From 1983 to 1985, three locations on Beaver Dam Creek were studied as part of the Com prehensive Cooling Water Study (CCWS). This study was designed to assess the impact of current and proposed SRS activities on water quality. The Beaver Dam Creek sampling locations, which are shown in Figure 5-14, included:

- Beaver Dam Creek at 400-D outfall prior to ash basin effluent (01 on map)
- Beaver Dam Creek downstream of 400-D ash basin effluent (02 on map)
- Beaver Dam Creek near confluence with Savannah River (09 on map)

Figure 5-14 Water Chemistry and Quality Sampling Stations for Beaver Dam Creek

A synopsis of historical water quality monitoring of Beaver Dam Creek prior to the CCWS can be found in Gladden et al. (1985).

Comprehensive results and discussion of CCWS data can be found in Lower (1987).

Priority Pollutants Survey

In 1984, a special instream survey of priority pollutants was conducted to determine the lev els of volatile, acid, and base/neutral organic compounds in Beaver Dam Creek. One loca tion in Beaver Dam Creek, downstream of 400-D effluent, was sampled (Figure 5-14; 02 on map). The "Comprehensive Cooling Water Study Final Report, Volume II: Water Quality" (Lower 1987) documented these results. This document also reports the results of analyses for pesticides and PCBs in Beaver Dam Creek.

Chemical Assessment Studies

A Clean Water Act Section 316(a) Demonstration special study was conducted in 1988 on the waters of Beaver Dam Creek to determine impacts from specific SRS operations. The data collected during the 316(a) Demonstration are reflective of the thermal flow regimes that exist when D Area is operating at a reduced power level and are most reflective of cur rent conditions in Beaver Dam Creek. This study, documented in "Compliance of the Savannah River Site D-Area Cooling System with Environmental Regulations" (Specht et al. 1990), is discussed later in this chapter. Stations 03 to 09 on Figure 5-14 correspond to the stations sampled during this monitoring program (1A-6).

Field Data

Water Temperature

Stream water temperatures at the mainstream Beaver Dam Creek sites during the CCWS ranged from 11.2 to 34°C (52.2 to 93.2°F), with an average temperature of 25.5°C (Table 5-34). These temperatures reflected the moderate thermal effluent associated with the stream. In Beaver Dam Creek near the confluence with the Savannah River, the temperatures were slightly lower than at the upstream sites. However, the range of temperatures was greater, reflecting river water inputs during high water conditions in the winter and the input of ther mal effluents from Four Mile Creek during the spring and summer (Lower 1987). Routine monitoring on Beaver Dam Creek near the Savannah River Swamp from 1987 to 1991 indi cated temperatures were comparable to those measured during the CCWS at the same loca tion (Table 5-35). Hourly temperature measurements ranged from 5 to 32° C (41 to 89.6 $^{\circ}$ F), with a mean temperature of $22^{\circ}C$ (71.6 $^{\circ}F$).

pH Measurements

The pH measurements at all locations during the CCWS ranged from 6.0 to 7.7, with an average near 7.0 (Table 5-34). These values reflected Savannah River source water concen trations. From 1987 to 1991, hourly pH measurements ranged from 4.9 to 10.9 (Table 5-35).
Table 5-34. Beaver Dam Creek Field Data

 \overline{PCCWS} = Comprehensive Cooling Water Study, Newman et al. (1986).

Blank Spaces = Mean not calculated due to insufficient data in report.

Data taken from Arnett et al. 1992; Cummins et al. 1991, 1990; Davis et al. 1989; Mikol et al. 1988.

Physical Characteristics and General Chemistry

Dissolved Oxygen

The mean dissolved oxygen concentration in the mainstream waters of Beaver Dam Creek was 7.25 mg O_2/l , with ranges of 5.40 to 10.0 mg O_2/l during the CCWS (Table 5-36). Near the confluence with the Savannah River (swamp), the mean dissolved oxygen concentration was much lower (5.53 mg O_2/l). The lower dissolved oxygen concentration in the swamp was likely the result of thermal input from Fourmile Branch. From 1987 to 1991, concentra tions of dissolved oxygen (mean 7.85 mg O_2/l) have continued to reflect concentrations measured during the CCWS (Table 5-35).

Suspended Solids and Turbidity

Mean turbidity and total suspended solids concentrations measured during the CCWS were elevated compared to Savannah River source water concentrations (Newman et al. 1986). Newman et al. (1986) documented that the source water contributions and the mainstream velocity regime contributed to the conditions in Beaver Dam Creek. The routine monitoring program does not measure suspended solids and turbidity.

Conductivity

Mean specific conductance in Beaver Dam Creek ranged from 79.1 to 91.1 μ S/cm during the CCWS (Table 5-36). These values are reflective of the specific conductance of Savannah River source water. Hourly specific conductance measurements from 1987 to 1991 ranged from 44 to 302 μ S/cm, with a mean of 107 μ S/cm (Table 5-36).

Table **5-36.** Beaver Dam Creek Physical Characteristics and General Chemistry, 1987-1991

 ${}^{a}NTU =$ Nephelometric turbidity units.

^bCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

Major Anions and Cations

Alkalinity, Chloride and Sulfate

Table 5-37 gives alkalinity, chloride, and sulfate concentrations in Beaver Dam Creek dur ing the CCWS. Total alkalinity and chloride concentrations measured during the CCWS reflected concentrations of Savannah River water. Mean total alkalinity concentrations in Beaver Dam Creek ranged from 16.7 to 18.1 mg CaCO₃/l; Savannah River concentrations ranged from 16.5 to 19.6 mg CaCO₃/I (Lower 1987). Mean chloride concentrations in Beaver Dam Creek ranged from 5.70 to 6.38 mg/l; while Savannah River concentrations ranged from 5.53 to 6.29.mg/1 (Lower 1987). Sulfate concentrations during the CCWS were higher downstream (mean 11.3 mg/1) than at the upstream location (mean 6.77 mg/I). Sulfate con centrations in the Beaver Dam Creek swamp waters (mean 8.40 mg/1) were slightly reduced from the downstream location, but were elevated compared to the upstream location and were attributed to ash basin effluent entering Beaver Dam Creek. The routine monitoring program does not measure total alkalinity, chloride, and sulfate.

Calcium, Magnesium, Sodium, and Potassium

Concentrations of total calcium, magnesium, sodium, and potassium in Beaver Dam Creek waters during the CCWS are given in Table 5-38. Data collected during the CCWS indi cated that the transport of these cations in Beaver Dam Creek was almost entirely in the dis solved phase (Newman et al. 1986). Calcium, magnesium, sodium, and potassium are not measured during routine monitoring.

Aluminum, Iron, and Manganese

Table 5-38 gives concentrations of total aluminum, manganese, and iron measured during the CCWS. Approximately 94% of the aluminum was associated with the solid phase, 86% of the iron was in the solid phase, and about 25-40% of the manganese was in the solid phase (Newman et al. 1986). The concentrations of these elements were reflective of Savan nah River source water. The routine monitoring program does not measure aluminum, man ganese, and iron.

Table 5-37. Beaver Dam Creek Major Anions

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

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Table **5-38.** Beaver Dam Creek Major Cations (Total)

 ${}^{\text{a}}$ CCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

Nutrients

Phosphorus

During the CCWS, all phosphorus species in Beaver Dam Creek waters were found in con centrations similar to those of its source water, the Savannah River (Table 5-39). Mean con centrations of total phosphorus and total orthophosphate in the mainstream waters during the CCWS were 0.13 mg P/I and 0.10 mg P/l, respectively. In the Beaver Dam Creek swamp waters, mean concentrations of total phosphorus (0.092 mg P/I) and total orthophosphate (0.064 mg P/1) were slightly lower than at the mainstream locations. The routine monitoring program does not measure phosphorus species.

Nitrogen

During the CCWS, all nitrogen species in Beaver Dam Creek waters were found in concen trations similar to those of its source water, the Savannah River (Table 5-39). Concentra tions of nitrite and ammonia were lowest at the Beaver Dam Creek swamp location. Mean concentrations of nitrate were similar at all three locations, with a range of 0.310-0.325mg/i (Table 5-39). Newman et al. (1986) documented that at the mainstream locations, mean per centages of total nitrogen and nitrate, ammonia, and organic nitrogen were 40%, 20%, and 40%, respectively. The routine monitoring program does not measure nitrogen species.

Trace Elements

Total trace elemental concentrations measured in Beaver Dam Creek during the CCWS largely reflected Savannah River source water concentrations (Table 5-40). During the CCWS, mean total arsenic concentrations ranged from 2.7 to 3.7 μ g/l; total cadmium ranged from 0.55 to 0.77 μ g/l; total chromium ranged from 7.7 to 11.8 μ g/l; total copper ranged from 2.8 to 5.4 µg/l; total lead ranged from 2.1 to 3.1 µg/l; and total zinc ranged from 4.8 to 5.3 **gg/l.**

aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

Table 5-40. Beaver Dam Creek Trace Elements (Total)

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

Total nickel concentrations sometimes were elevated during the CCWS, with ranges from 3.2 to 4.6 μ g/l. The elevated nickel concentrations may have been the result of the 400-D Area powerhouse and ash sluicing operations (Lower 1987).

Total mercury and uranium concentrations were consistently below the analytical detec tion limits for those elements during the CCWS. The routine monitoring program does not measure trace elements.

Organic Carbon

Total organic carbon concentrations in Beaver Dam Creek were similar to concentrations in the Savannah River. Mean total organic carbon concentrations ranged from 5.65 to 6.75 mg/1 in Beaver Dam Creek, while the Savannah River average was 6.10 mg/l. Approximately 72-84% of the organic carbon was present in the dissolved phase (New man et al. 1986).

Priority Pollutants

Concentrations of all 88 volatile, acid, and base/neutral organics tested in Beaver Dam Creek during the 1984 instream survey were below the associated analytical detection limits (Lower 1987). These results confirmed the 1981 point source data for outfall D-001 and indicated that the input of Savannah River water for various operations in 400-D Area had no adverse impact on levels of organics in Beaver Dam Creek.

Pesticides, Herbicides, and PCBs

Pesticides, herbicides, and PCBs are not measured in Beaver Dam Creek during routine water quality monitoring.

Lower (1987) reports the results of analyses for pesticides, herbicides, and PCBs from 1982 to 1985; results from 1967 to 1981 can be found in Gladden et al. (1985). During these periods, concentrations were also near or below detection limits at all locations.

Chemical, Including Radionuclide, and Toxicity Assessment Studies

In 1988, a Clean Water Act Section 316(a) Demonstration was initiated to determine whether Beaver Dam Creek could support a balanced indigenous biological community because water temperature during D-Area operations exceeded limits set forth in the National Pollutant Discharge Elimination System (NPDES) permit. Seven sampling loca tions were selected on Beaver Dam Creek to represent different habitats of the stream (Figure 5-14; locations 03-09).

Forty water quality parameters were assessed monthly at the sampling stations in Beaver Dam Creek (Table 5-41). The data show that 400-D Area operations greatly influence the water quality of Beaver Dam Creek. Generally, water quality parameters at Station **1A** (03) were consistently different from all other stations on Beaver Dam Creek. Station **1A** was in the original intermittent stream channel of the creek, upstream of the confluence of

Table 5-41. Mean Values for Water Chemistry Parameters at Stations Sampled in Beaver Dam Creek, 1988-1990

Source: Specht et al. 1990.

^aFive sampling dates; all others were sampled monthly for 18 months.

the 400-D Area discharge canal. However, this station was usually dry, except after periods of heavy rainfall, and the water collected was reflective of rainwater runoff.

Station **1A** had the lowest mean values for pH, conductivity, dissolved oxygen, tempera ture, orthophosphorus, sulfate, nitrate, total phosphorus, and dissolved and total strontium. Means for total hardness, total suspended solids, turbidity, dissolved and total aluminum, dissolved and total barium, dissolved and total iron, total lead, dissolved and total manga nese, and total zinc were higher at Station **IA** than at the other stations. Excluding Station 1A, means for the study period were generally similar among stations on Beaver Dam Creek (Table 5-41).

This study indicated that there were no characteristics of the Beaver Dam Creek water qual ity measurements that represented obvious indicators of stress to the biological communi ties. Water temperatures never exceeded the maximum NPDES limit of 32.2°C (90'F). However, the water temperatures below the thermal discharge from the D-Area power plant were often more than 2.8°C (5°F) higher than the water temperatures at the 5G pumphouse at the Savannah River or the upstream station (1A).

No radionuclide or toxicity data have been collected from Beaver Dam Creek.

Algae

There have been no studies of the algae of Beaver Dam Creek with the exception of three stations that were sampled for periphyton from 1983 to 1985 (Specht 1987) as part of the CCWS. Periphyton biomass values for the Beaver Dam Creek sampling stations were sta tistically similar to each other and to the values of the other SRS streams. There was no evi dence that pollutants discharged to Beaver Dam Creek via D-Area effluents had an adverse impact on the periphyton community of this stream system.

Macrophytes

Introduction

Aquatic macrophytes provide stream structure, substrate for periphyton development, cover and substrate for smaller animals, and a source of carbon for the stream system. Although aquatic macrophytes are an important component of the function of many aquatic systems, they tend to be less important in flowing than still waters.

Comprehensive Cooling Water Study

The CCWS examined the Beaver Dam Creek aquatic macrophyte community (Specht 1987). Data from the single station sampled for macrophytes in 1983-1984 show a total of 28 taxa; of the 28 taxa, only 4 were found growing in the stream channel, and the remain der were found growing in the riparian area. Macrophyte data from a different single sta tion were collected from October 1984-September 1985. Eight taxa were found in measurable quantities (Table 5-42). Differences were found between spring-summer values and those of fall-winter sampling periods. Most aquatic macrophyte growth was in back waters and along the creek margins (Specht 1987). The number of taxa and population parameters compare favorably with those for at least one nonthermal stream location sur veyed during the same sampling period.

Effects of Reduced Power Plant Operation

Normal successional patterns and development of macrophyte beds would be expected to continue in suitable sections of Beaver Dam Creek with the reduced operation of the D-Area power plant.

Table 5-42. The Annual Mean Macrophyte Area, Volume, Biomass, Percent Cover and Standard Deviation (s), Octo ber 1984-September 1985

Source: Firth et al. 1986.

aTwenty reaches were sampled.

Zooplankton

Comprehensive Cooling Water Study

Chimney and Cody (1986a) studied the zooplankton in Beaver Dam Creek for regulatory compliance issues covered by the Clean Water Act Section 316(a). The study documented the temporal and spatial characteristics of zooplankton species based on quarterly sampling from December 1984 to August 1985. Surface-water grab samples were collected at two sta tions approximately one-third of the distance down the channel. Species richness comprised 7 Protozoa, 15 Rotifera, 14 Cladocera, 4 Copepoda, and 1 Ostracoda; Steel Creek was the only SRS stream with a higher species richness. Figure 5-15 indicates that the greatest den sities occurred during March 1985 and May 1985. During March 1985, Cladocera domi nated the assemblage; during May 1985, Protozoa and Copepods were most abundant.

Regulatory Compliance Study

Specht et al. (1990) performed a study in Beaver Dam Creek from September 1988 to Feb ruary 1990 to meet regulatory compliance of the D-Area cooling system. Results were based on seven sampling stations distributed throughout the length of the stream at approxi mately equal intervals. Species richness during these years averaged 26 Protozoa, 27 Rotifera, 7 Cladocera, 8 Copepoda, and 1 Ostracoda. Figure 5-16 illustrates the mean monthly zooplankton densities. Seasonal patterns were different from the results in 1985, probably due to sampling differences and large natural variability in abundances. Large and typical standard errors are associated with this variation. These results are representative of most zooplankton populations in flowing water habitats.

Figure 5-15. Mean Total Density of Zooplankton in Beaver Dam Creek, 1984-1985

Figure 5-16. Mean Total Density of Zooplankton in Beaver Dam Creek, 1988-1990

Macroinvertebrates

Sampling Locations and Methods

Comprehensive Cooling Water Study

From September 1982 through August 1983, macroinvertebrates were sampled quarterly in the mouth of Beaver Dam Creek using Hester-Dendy multiplate samplers and drift nets (Specht and Painter 1983).

From November 1983 through September 1985, as part of the CCWS, macroinvertebrates were sampled monthly at five stations in Beaver Dam Creek (Figure 5-17) using Hester Dendy multiplate samplers, leaf bags, and drift nets; at three of these five stations, macroin vertebrates also were collected from aquatic macrophytes (Kondratieff and Kondratieff 1984, 1985; Firth et al. 1986). The CCWS macroinvertebrate data were summarized by Specht (1987).

Clean Water Act Compliance Study

In support of a Clean Water Act Section 316(a) Demonstration, macroinvertebrate sampling also was conducted monthly at six stations in Beaver Dam Creek (Figure 5-17) from Sep tember 1988 through February 1990 using Hester-Dendy multiplate samplers and quarterly using drift nets and dip nets (Specht et al. 1990). Detailed sampling methods for the 1988 1990 program can be found in Nagle et al. (1990). Stream temperatures in Beaver Dam Creek averaged several degrees cooler during the 1988-1990 study than during the 1983 1985 study, due to reduced power production at the D-Area power house.

Figure 5-17. Macroinvertebrate Sampling Stations for Beaver Dam Creek

Additional Study

Results

Introduction

A program to develop a biotic index for southeastern streams using Hester-Dendy multi plate samplers sampled macroinvertebrates in the upstream portion of Beaver Dam Creek in September 1994. While not specifically designed to characterize SRS streams, these data contribute to a better understanding of the streams (Specht and Paller 1995).

The thermal regime in 1988-1990 is similar to present (1997) conditions; therefore, the macroinvertebrate data from 1988-1990 should be representative of present conditions. Most of the data presented and discussed in this section are Hester-Dendy data from the 1984-1985 and 1988-1990 sampling programs because these are the most comparable data sets, with respect to sampling methods and stations.

Number of Taxa

In 1984-1985, the total number of macroinvertebrate taxa collected on Hester-Dendy mul tiplate samplers at each of the five sampling stations ranged from 36 to 61, while in 1988 1990, excluding chironomid genera, the number of taxa collected at the six sampling sta tions was slightly higher, ranging from 41 to 64 (Table 5-43). In both sampling programs, the most upstream station had the fewest taxa, while the slough-like area (Station 7 in 1984-1985; Station 04 in 1988-1990) had the most taxa. Total number of taxa collected in September 1994 at one station was 27 (Table 5-44).

The 1988-1990 sampling program including all sampling methods, collected 163 macroin vertebrate taxa from Beaver Dam Creek (Table 5-45). Of these taxa, 124 were collected on the multiplate samplers, and 39 were collected exclusively by the supplemental sampling methods. Taxa that were most likely to be missed or under-represented by the multiplate samplers included some species of mollusks (Gastropoda and Pelecypoda), dragonflies and damselflies (Odonata), most aquatic bugs (Hemiptera), and many aquatic beetles (Coleoptera). Many of these taxa are exclusively benthic and do not readily colonize artifi cial substrates that are above the stream bottom. Taxa that are likely to be over-represented on the multiplate samplers include many species of chironomids, at least one of which *(Stenochironomus)* actively burrowed into the plates of the multiplate samplers in large numbers but was rarely collected on natural substrates in Beaver Dam Creek.

Densities

The mean density of macroinvertebrates collected on multiplate samplers ranged from 921.2 to 1776.5 organisms/m² in 1984-1985 and from 773.7 to 2348.0 organisms/m² in 1988-1990 (Table 5-46). Mean density in September 1994 at Station 5 was 502.8 organ isms/ $m²$. Densities at upstream stations (01, 02, and 03) were generally higher than at downstream stations (04, 05, and 06) throughout the sampling programs. Densities were fairly comparable between studies for any given station. The mean biomass of macroinver tebrates ranged from 0.076 to 0.220 g ash-free dry-weight (AFDW)/ m^2 in 1984-1985, while biomass in 1988-1990 was substantially higher at most stations, ranging from 0.185 to 0.394 g AFDW/m² (Table 5-46). Biomass in September 1994 was 0.442 g AFDW/m².

 a Specht 1987.

b_{Specht et al. 1990.}

CSpecht and Paller 1995.

^dStation 5 was approximately 0.8 km (0.5 mi) upstream from Station 01 (1988-1990), in the D-Area discharge canal.

echimney and Cody 1986b.

Table 5-44. Mean Number of Taxa, Density and Biomass of Macroinvertebrate Taxa and of Percent Composition of Functional Groups Collected on Hester-Dendy Multiplate Samplers in Beaver Dam Creek, September 1994

Source: Specht and Paller 1995.

Table 5-45. Macroinvertebrate Taxa Found in Beaver Dam Creek, All Sampling Methods Combined. "X" Denotes Quantitative (Multiplate) Sampling and "*" Denotes Qualitative Sampling (Dip Nets and Drift) Only, Sep tember 1988-February 1990

 $\ddot{}$

Table 5-45. (cont)

Table 5-45. (cont)

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Table 5-45. (cont)

 $\hat{\mathcal{A}}$

Table 5-45. (coat)

 \overline{a}

Source: Specht et al. 1990.

^aHigher level taxa not included in count if a lower level taxon was present.

Sources: ^aSpecht 1987; ^bSpecht et al. 1990; ^cSpecht and Paller 1995.

dStation numbers used in the 1984-1985 study differ from those used in the 1988-1990 study.

eStation 5 was approximately 0.8 km (0.5 mi) upstream from Station 01 (1988-1990) in the D-Area discharge canal.

 $AFDW =$ ash-free dry weight.

Dominant Species

As a group, Chironomidae were by far the most common macroinvertebrates in Beaver Dam Creek in 1988-1990, comprising from 51.8% (Station 05) to 75.9% (Station 01) of the organisms collected from the multiplate samplers (Table 5-47). At Stations 01 through 04, the chironomid subfamily Orthocladiinae was the most abundant macroinvertebrate taxon, accounting for 22.4-29.6% of the organisms collected, while at Stations 05 and 06, chironomids of the tribe Chironomini were most abundant (20.6 and 25.0%, respectively). Tanytarsini chironomids were also abundant at all stations, comprising 8.4-19.9% of the organisms collected. Other groups of macroinvertebrates that contributed at least 5% to the total density at one or more stations included mayflies (Ephemeroptera; 0.8-26.5%), caddisflies (Trichoptera; 2.1-11.8%), beetles (Coleoptera; 1.0-11.9%), oligochaete worms (1.4- 6.6%), and non-chironomid dipterans (0.8-5.6%; Table 5-47). Groups that contrib uted at least 5% to the total density at Station 5 in 1994 were Ephemeroptera (73%) and Trichoptera (12%).

Differences Among Sampling Stations

In **1988-1990,** Station 01 had fewer macroinvertebrate taxa and slightly lower biomass than the other sampling stations (Table 5-43 and Table 5-46) but was similar to the other sampling stations with respect to densities of organisms and mean number of taxa (Table 5-46 and Table 5-48). There were some indications of longitudinal changes in the relative abundances of macroinvertebrate taxonomic groups in Beaver Dam Creek. Station 01 had the lowest relative abundance of mayflies (0.8%) and the highest relative abundance of chironomids (75.9%; Table 5-47) of all of the stations. In general, the relative abundance of mayflies increased in a downstream direction and beetles (Coleoptera) were much more abundant at Stations 05 and 06 (11.9% and 5.1%, respectively) than at the four upstream stations (1.0- 2.1%; Table 5-47). Conversely, trichopterans were more abundant at the three upstream stations (5.9-11.8%) than at the three downstream stations (2.1-4.8%; Table 5-47).

Table 5-47. Mean Relative Abundance (Percent Composition) of Higher Order Taxonomic Groups of Macroinverte brates from Quantitative Sampling in Beaver Dam Creek, September 1988-February 1990 and September 1994

Source': Specht et al. 1990: Specht and Paller 1995.

^aSome totals may not equal 100.00 due to rounding.

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Table 5-48. Descriptive Statistics for Total Macroinvertebrate Densities (Organisms/m²) and Mean Number of Taxa from Quantitative (Multiplate) Sampling in Beaver Dam Creek, September 1988-February 1990

Source: Specht et al. 1990.

aAbbreviations:N=Total number of samples analyzed; Med=Median; SD=Standard Deviation; CV=Coefficient of Variation; Min=Minimum; Max=Maximum.

bOnly four samples were collected at Station 03 in January 1989.

^cSamples were not retrieved from Station 06 in February 1990 due to high water levels.

Taxonomic Composition Changes Since 1984-1985

The taxonomic composition of Beaver Dam Creek has changed substantially between 1984-1985 and 1988-1990, with most stations exhibiting increases in the relative abun dance of mayflies (Ephemeroptera), snails (Gastropoda), beetles (Coleoptera), caddisflies (Trichoptera) and Tanytarsini chironomids, and an overall decline in the relative abundance of Chironomini chironomids (Table 5-45, Table 5-47, and Table 5-49). Many species of Chironomini generally are considered to be pollution-tolerant (Beck 1977), while most spe cies of mayflies and caddisflies are intolerant of poor water quality (Hynes 1970). Although Ephemeroptera were common in 1994 (73% of the total number of macroinvertebrates), the group comprised almost exclusively *Baetis.* Some species of this genus have been reported to be very tolerant of heavy metals (Heliovaara and Vaisanen 1993), which are present in the coal ash, the ash basins, and the run-off to the creek from these areas. However, it appears that, with respect to taxonomic composition, the macroinvertebrate community of Beaver Dam Creek has undergone substantial improvement since 1984-1985.

Table 5-49. Mean Relative Abundance (Percent Composition) of Higher Order Taxonomic Groups of Macroinverte brates Collected on Hester Dendy Multiplate Samplers, October 1984-September 1985

Source: Specht et al. 1990.

aTaxa were grouped in "Chironomidae."

^bSome totals may not equal 100.00 due to rounding.

^cStations 05, 06, 07, 08 calculated from Firth et al. 1986.

dStation 09 calculated from Chimney and Cody 1986b.

Fish

Introduction

Three fisheries sampling programs have been conducted on Beaver Dam Creek. The most extensive was the D-Area Clean Water Act Section 316(a) Demonstration, which was con ducted from September 1988 through February 1990 to evaluate the effects of heated cooling water discharged to Beaver Dam Creek from the D-Area power plant. Fisheries sampling in Beaver Dam Creek also was conducted in conjunction with the CCWS (1983 1985) and the Comprehensive Cooling Water Quarterly Study (1984-1985). The results of all three programs will be presented in this discussion, although the 316(a) study will be emphasized because it is more recent and comprehensive than the earlier studies.

D-Area 316(a) Demonstration

Introduction

Six stations from the upper reaches to the mouth of Beaver Dam Creek were electrofished monthly from September 1988 through February 1990 (Figure 5-18) (Specht et al. 1990). These stations encompassed a range of habitats from wide, deep, slow-flowing areas near the creek mouth (Station 06) to narrower, more swiftly flowing channels (Stations 01, 02, and 03). Some portions of the creek were heavily overgrown with aquatic vegetation, par ticularly the slough-like habitat represented by Station 04. Initially, all stations also were sampled monthly by hoopnetting. However, hoopnet sampling was discontinued at Sta tions 02, 03, and 04 to prevent the accidental trapping of alligators. Hoopnet sampling also

Figure 5-18. Fish Sampling Stations for Beaver Dam Creek.

was suspended at all sample stations during April and May, the months of maximum alliga tor movement.

Adult Fish

Forty-five species were collected by electrofishing and hoopnetting during the 18-month study (Specht et al. 1990). The number of fish collected by electrofishing ranged from 561 at Station 02 and 532 at Station 05 to 167 at Station 01 (Table 5-50). The number of species collected from each station was less variable, ranging from 25-32. Electrofishing catches varied monthly, and long-term patterns were not evident.

Fewer fish were collected at Station 01 than at the other stations (Table 5-50). Station 01 was closest to the D-Area outfall and had higher flow rates than the other sample stations. A major portion of the sample area at Station 01 was strongly channelized with little vegeta tion or instream structure to serve as foraging or refuge areas for fish. Stations 02, 04, and 05, in contrast, generally had open canopies with abundant aquatic and shoreline vegetation.

The electrofishing data were converted to catch per unit effort (CPUE, expressed as number caught per 100 m [328 ft]) by dividing the number caught in each transect by the transect length. Total (all species) mean CPUE ranged from approximately 10.4 fish/100 m at Sta tion 02 to approximately 3.1 fish/100 m at Station 01 (Table 5-51). CPUE fluctuated widely monthly. Low CPUE occurred when Savannah River flood waters inundated Stations 03, 04, 05, and 06. Flooding greatly increased water depth at these stations and allowed fish to disperse into the surrounding floodplain.

Statistical testing indicated that CPUE was significantly lower at Station 01 than at the other stations. The low CPUE at Station 01 was probably a result of high current velocities and relatively poor habitat. Other differences were likely the result of local variations in instream structure, vegetation, substrate, current velocity, flooding, and other factors that influenced habitat quality and sampling efficiency.

The most abundant species (by number) collected by electrofishing were spotted sucker *(Minytrema melanops),* coastal shiner *(Notropis petersoni),* redbreast sunfish *(Lepomis auritus),* largemouth bass *(Micropterus salmoides),* and spotted sunfish *(Lepomis puncta tus)* (Table 5-50). Sunfish (Centrarchidae) were the most abundant family, composing nearly 40% of the fish collected from Beaver Dam Creek. Minnows and shiners (Cyprin idae) (26%) and suckers (Catostomidae) (17.3%) were also abundant (Table 5-52). Several taxa, including the bannerfin shiner *(Cyprinella leedsi),* blackbanded darter *(Percina nigro fasciata),* and channel catfish *(Ictalurus punctatus),* exhibited a distinct longitudinal zona tion, being more abundant either towards the headwaters or towards the stream mouth. Most of the more common species, however, were abundant throughout the stream.

Seventeen species were collected by hoopnetting (Table 5-53). The greatest number of fish (96) and greatest number of species (13) were collected from Station 06, at the mouth of Beaver Dam Creek. While Stations 02, 03, and 04 were sampled for only four months due to the high potential for alligator mortality by drowning in the hoopnets, they yielded a total of 11 species and 99 individuals, indicating that fish were abundant at these stations.

For further analysis, the hoopnetting data were converted to CPUE (expressed as number caught per net per day; Table 5-54). Mean CPUE was highly variable. However, catch rates

a_{Not} included in taxa count.

 \cdots

Table 5-51. Electrofishing CPUE (No./100 m) at Stations 01-06 in Beaver Dam Creek, September 1988-February 1990

Table 5-52. Relative Abundance (Percent Composition) of Species Collected by Electrofishing at Stations 01-06 in Beaver Dam Creek, September 1988-February **1990**

assumpling at these stations was discontinued after December due to the danger hoopnets presented to the alligator population. b_{Not} included in taxa count.

Table 5-54. Hoopnetting CPUE (No./Net Day) at Stations 01-06 in Beaver Dam Creek, September 1988-February $1990^{\rm a}$

aStations 02, 03, and 04 discontinued after four months. No nets were set in March or April 1989.

were generally higher during the warmer months. Zero catch rates occurred most often dur ing the fall and winter of 1988-1989 (Figure 5-19), especially at Station 01, probably because of seasonal reductions in fish activity and movement.

Channel catfish constituted 60% of the total hoopnetting catch, followed by redbreast sun fish (10.5%), bluegill *(Lepomis macrochirus) (5.1%),* and flat bullhead *(Ictalurus platycephalus)* (4.2%) (Table 5-55). In addition, white catfish *(Ictalurus catus)* were abun dant at Stations 05 and 06, perhaps because this species entered the lower reaches of Beaver Dam Creek from the Savannah River.

One objective of the 316 (a) Demonstration was to determine the health of the Beaver Dam Creek fish community. This was done by comparing fish communities in Beaver Dam Creek to fish communities in relatively unimpacted streams. These included several nonthermal streams on the SRS and several other southeastern streams described in the literature (sum marized in Paller et al. 1988) (Table 5-56).

Total number of species ranged from 21 in Upper Three Runs Creek to 59 in Steel Creek. The total in Beaver Dam Creek, 46, was near the maximum. The high species number in Beaver Dam Creek was partly a function of sampling effort, but it also reflects the habitat diversity of this stream, which contains swamp, slough, and stream environments.

Sunfishes and black bass dominated the fish assemblage in Beaver Dam Creek (38% of the collections), followed by minnows (23%), suckers (15%), and catfishes (11%) (Table 5-56). The relative abundance of minnows and sunfishes was within the range of the other south eastern streams, but the relative abundance of suckers and catfishes was greater. The high relative abundance of catfishes in Beaver Dam Creek was partly due to the intensive hoop-

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aStations 02, 03, and 04 discontinued after four months. No nets were set in March or April 1989.

Table 5-56. Species Number and Relative Abundance (Percent Composition) of Fishes in Nine Southeastern Streams. Relative Abundances are Based on the Number of Individual Fish in Each Family

aSpecht et al. 1990, electrofishing and hoopnetting combined.

^bKelly et al. 1981.

CBoschung and O'Neil 1981.

dMundy and Boschung 1981.

ePardue and Huish 1981.

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Table 5-56. (cont)

f_{Pardue} and Huish 1981.

gPaller et al. 1988.

h_{Paller} and Saul 1986.

Table 5-57. Total Number of Ichthyoplankton Collected from Stations 01-06 in Beaver Dam Creek, February-July **1989**

aDarter = *Percina* sp. or *Etheostoma* sp.

^bHigher taxon not counted in total if lower taxon present.

l,

Taxa	February March		April	May	June	July
darters	6	3	5	3	2	0
sunfish	0	0	31	11	3	4
minnow	0	0		0	2	0
sucker	0	0	11		0	0
other	0			2	0	0
total no.	6		55	17		

Table 5-58. Monthly Ichthyoplankton Collections from Beaver Dam Creek, February-July 1989
Table 5-59. Average Densities (Organisms/1000m³) of Ichthyoplankton Collected from Stations 01-06 in Beaver Dam Creek, February-July 1989

aDifferences are due to rounding error.

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Table 5-60. Number and Percent Composition of Ichthyoplankton Collected from Steel Creek, Upper Three Runs, and Lower Three Runs, **1984-1985**

Table 5-60. (cont)

Source: Specht (1987).

^aprincipally eggs.

^btotal incorrect in original table; actual total is 1,513.

^chigher taxon not counted if lower taxa were present.

Comprehensive Cooling Water Study

Adult and Juvenile fish

The CCWS involved two adult/juvenile fish sampling programs on Beaver Dam Creek: one quarterly and one reporting on the distribution of fish during the winter (Paller and Osteen 1985; Paller and Saul 1986). For the quarterly study, two hoopnets were placed in the mouth of Beaver Dam Creek and five sites were electrofished four times a year. Four of the electrofishing sites corresponded to Stations 03-06 of the 316(a) Demonstration. The overwintering study included one electrofishing sample station in the mouth of Bea ver Dam Creek during 1984 and three during 1985 in the mouth (Station 06), lower reaches of the floodplain swamp (roughly analogous to Station 05), and upper reaches of the floodplain (roughly analogous to Station 02). In addition, hoopnets were set in the mouth of the creek during the 1985 overwintering study.

Electrofishing CPUE near the mouth of Beaver Dam Creek was considerably lower dur ing the quarterly (1.6/100 m) and overwintering (1.4/100 m) studies than during the 316(a) Demonstration (7.1/100 m). Similar trends also were observed at Stations 02, 03, 04, 05, and 06. The studies used the same sampling methods and equipment, suggesting that differences in sampling efficiency were not responsible for the differences in CPUE.

Electrofishing data from the overwintering study indicated that the dominant species, in rank order, at Stations 02, 05, and 06 were spotted sunfish, redbreast sunfish, largemouth bass, bowfin, *(Amia calva)* gizzard shad *(Dorosoma cepedianum)* and bluegill (Table 5 61). Dominant species at the same stations during the 316(a) Demonstration were spotted sucker, redbreast sunfish, largemouth bass, spotted sunfish, striped mullet *(Mugil ceph alus),* and bluegill. Definite reasons for these differences are unknown, but interannual

Table 5-61. Dominant Species^a Collected from Beaver Dam Creek by Electrofishing at Stations 02, 05, and 06 during the Overwintering Study (1984-1985), and at Each Station during the 316(a) Demonstration

aExcluding minnows and other small fishes. bPaller and Saul 1986.

^cSpecht et al. 1990.

dBluegill ranked eighth.

variations in recruitment and changes in habitat between studies (e.g., aquatic plant growth) are possible causes for sampling variation.

Mean overall hoopnetting CPUE for the overwintering study (1.2 fish/net day) was approxi mately the same as for the 316(a) Demonstration (1.0 fish/net day); CPUE during both stud ies was higher than during the quarterly study (0.6 fish/net day). Channel catfish were dominant during the overwintering program, but flat bullhead *(Ameiurus platycephalus),* black crappie *(Pomoxis nigromaculatus),* redear sunfish *(Lepomis microlophus),* and blue back herring *(Alosa aestivalis)* constituted substantial proportions of the catch (Table 5-62). Channel catfish were also strongly dominant during the 316(a) Demonstration. In general, all these studies revealed the presence of diverse fish assemblages in Beaver Dam Creek over time.

Ichthyoplankton

Larval fish and fish eggs were collected from five sample stations in Beaver Dam Creek dur ing the 1984-1985 Comprehensive Cooling Water Study. Four of these sample stations were analogous to the stations sampled in the 316(a) demonstration in 1989 (Stations 03, 04, 05, and 06). Sampling methodology and effort was similar for both studies.

More ichthyoplankton were collected in 1984 (334 individuals) (Table 5-63) and 1985 (253 individuals) (Table 5-64) than in 1989 (97 individuals). This may be an effect of low water levels during the 1989 spawning season since many species spawn most successfully when floodwaters inundate terrestrial areas (Martin et al. 1981). Relative abundance also differed among years. While sunfishes (Centrarchidae) were dominant during 1984 and 1989 (Table 5-57 and Table 5-63), they constituted a comparatively small proportion of the total catch in 1985 (11.4%, Table 5-64). Similarly, the relative abundance of clupeids (herring and shad), suckers, darters, and brook silversides varied substantially among years.

Table 5-62. Dominant Species Collected from Hoopnets at Station 06 during the CCWS Overwintering Study (1984-1985) and the 316(a) Demonstration

apaller and Saul 1986.

bSpecht et al. 1990.

eBluegill ranked tenth in abundance for November 1984-April 1985 and sixth for September 1988-February 1990.

Table **5-63.** Number and Percent Composition of Ichthyoplankton Collected from Beaver Dam Creek, 1984

Source: Specht et al. 1987.

^aCalculated to compare with current program.

bPrincipally eggs.

Table 5-64. Number and Percent Composition of Ichthyoplankton Collected from Beaver Dam Creek, 1985

Source: Specht et al. 1987.

aCalculated to compare with current program.

b_{principally eggs.}

Mean ichthyoplankton densities in 1984, 1985, and 1989 were highest in the creek mouth and the stations (04 and 05) in the lower reaches of Beaver Dam Creek (1989 data in Table 5-59). The lowest densities occurred in the upper reaches of Beaver Dam Creek during all years.

The preceding comparisons indicate considerable annual variability in ichthyoplankton den sity and species composition. Some of this variability stems from sampling error, but some probably reflects real differences in ichthyoplankton abundance. Water level is known to strongly influence the attractiveness of tributary streams to spawning anadromous fish such as blueback herring (Frankensteen 1976) and the spawning success of other species that require coves, backwaters, shallows, and inundated vegetation where larvae and eggs are sheltered and protected from currents (Martin et al. 1981). A noteworthy trend that was con sistent across all years was greater densities and species richness in the lower reaches of Beaver Dam Creek reflecting the relative importance of this portion of the stream as a spawning area for some species.

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5.3 Fourmile Branch

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Drainage Description and Surface Hydrology

General Description

Fourmile Branch originates near the center of SRS and flows southwesterly for approxi mately 24 km (15 mi) (Figure 5-20). The watershed, which drains about 57 km² (22 mi²), includes several SRS facilities: C Area (reactor), F and H Areas (separations facilities, tank farms, and seepage basins), and the Solid Waste Disposal Facility (SWDF). At its headwa ters, Fourmile Branch is a small blackwater stream that is relatively unimpacted by SRS operations (Specht 1987).

In its lower reaches, Fourmile Branch broadens and flows through a delta that has been formed by the deposition of sediments. Although most of the flow through the delta is in one main channel, the delta has numerous standing dead trees, logs, stumps, and cypress trees, which provide structure and reduce water velocity in some areas. Downstream of the delta the creek flows in one main channel; the majority of the flow discharges into the Savannah River at river kilometer 244.7 (river mile 152.1), while a small portion of the creek flows west and enters Beaver Dam Creek. When the Savannah River floods, water from Fourmile Branch flows along the northern boundary of the floodplain swamp and joins with Pen Branch and Steel Creek instead of flowing directly into the river (Specht 1987).

Effluent Contributions

Fourmile Branch receives effluents from F, H, and C Areas. Before C-Reactor shutdown in 1985 the reactor discharged heated Savannah River water (as hot as 70'C [158°F]) into Fourmile Branch via Castor Creek. Resulting water temperatures were in excess of 60'C

Figure **5-20.** Location of Fourmile Branch on SRS

(140'F) in Fourmile Branch just downstream of its confluence with Castor Creek. With the contribution of C-Reactor cooling water, the flow in Fourmile Branch measured about 11.3 m³/s (400 ft³/s) (Murphy et al. 1991). The flows and temperatures associated with C Reactor operations no longer occur.

Prior to 1996, Fourmile Branch received effluents from 16 National Pollutant Discharge Elimination System (NPDES) outfalls in C, F, and H Areas, and Central Shops as well as groundwater from beneath F and H Areas. Until 1985, Fourmile Branch also received thermal effluents from C Reactor. With the new NPDES permit issued in 1996, outfalls were reduced from 16 to 5 due to deletions of waste streams and the consolidation of the outfalls. Effluent from the new 1.05-million gallon per day Centralized Sanitary Wastewa ter Treatment Facility began discharging to Fourmile Branch in 1995.

Fourmile Branch, either directly or via tributaries, receives the following NPDES-permit ted discharges: 186 basin overflows, cooling water, floor drains, steam condensate, pro cess wastewater, laundry effluent, stormwater, sanitary treatment wastewater, ash basin runoff, and lab drains.

Flow Measurements

The U.S. Geological Survey measures flow at several locations on Fourmile Branch. Table 5-65 summarizes flow statistics of Fourmile Branch at Site No. 7 and at SRS Road A-13.2 (Figure 5-21). Records for the most downstream station (Fourmile Branch at Road A 13.2) date back to November 1976. In water year 1995, the mean flow of Fourmile Branch at Road A-13.2 was 1.1 m³/s (37.3 ft³/s). Over the period of record (water years 1977-1995) at Road A-13.2, the mean flow was 3.2 m^3 /s (113 ft $\frac{3}{s}$), the 7-day low flow was 0.22m³/s (7.6 ft³/s), and the 7Q10 was 0.23 m³/s (8.2 ft³/s). Figure 5-22 and Figure 5-23 show the maximum, minimum, and mean daily flows from October 1982 to September 1995.

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Table 5-65. Flow Summary for Fourmile Branch

cms = cubic meters per second.

cfs = cubic feet per second.

Figure 5-21. Flow Measurement Sampling Stations for Fourmile Branch

Figure 5-22. Maximum, Mean, and Minimum Flow Measurements for Fourmile Branch at Site No. 7, October 1982 September 1995

Figure 5-23. Maximum, Mean, and Minimum Flow Measurements for Fourmile Branch at Road A-13.2, October 1982 September 1995

E Area H Area
F Area **A**

Water Chemistry and Quality

Studies and Monitoring

Water-Quality Monitoring

Westinghouse Savannah River Company's Environmental Monitoring Section has con ducted routine water-quality monitoring in Fourmile Branch since 1973. It samples loca tions on Fourmile Branch--near Road A-7 and near Road A-13.2 - monthly for physical and biological parameters and quarterly for metals (Figure 5-24; 08 and 10). The Environmental Monitoring Section also collects a sample annually from the same locations and analyzes it for pesticides, herbicides, and PCBs. All routine water quality monitoring data reported in the following sections can be found in the annual SRS Envi ronmental Reports.

Comprehensive Cooling Water Study

From 1983 to 1985, the Savannah River Ecology Laboratory (SREL) studied Fourmile Branch as part of the Comprehensive Cooling Water Study (CCWS). The study was to provide "an assessment of the potential effects of present and proposed SRS activities on the quality of waters used for cooling at the SRS" (Newman et al. 1986). The study included sampling the following four locations on Fourmile Branch (Figure 5-24):

- 1. Fourmile Branch at Road E-1 (02)
- 2. Fourmile Branch at Road 3, near Fourmile Branch at Road A-7 (08)
- 3. Fourmile Branch at Castor Creek (09)
- 4. Fourmile Branch at Road A-13.2, near Fourmile Branch at Road A **(11)**

	Station Name	ZU -01 ÓЗ Fourmile Branch 02 04 06				
01	Upstream $H(I)$	07 _c C Area				
02	Road E-1, CW3, HW7					
03	Road C1	۵۵۰				
04	Downstream H (III)	Castor Creek 09.				
05	Road C					
-06	Upstream F (IV), CW2 & HW2					
07	115 TransLine and Downstream	10°				
	F (VI) and CW1, FW4					
08	Road A-7					
09	Road 3					
10	Road A					
11	Road A-13.2	2 o				
		Scale in Miles				

Figure 5-24. Water Quality Monitoring Stations on Fourmile Branch

The data collected from the thermal portions of Fourmile Branch reflect impacts associ ated with reactor operation and are not relevant to the current conditions of Fourmile Branch. However, because a limited amount of monitoring data are available, portions of the CCWS data will be presented.

Gladden et al. (1985) presented a synopsis of historical water quality monitoring of the Fourmile Branch system prior to the CCWS. Comprehensive results and discussion of CCWS data can be found in Lower (1987).

Priority Pollutant Study

In 1984, a special instream survey of priority pollutants was conducted to determine the levels of volatile, acid, and base/neutral compounds in Fourmile Branch. The sample locations were Fourmile Branch at Road C (05 on map), Road A-7 (08 on map), and Road A (10 on map). Potential sources for organics into Fourmile Branch were discharges from F and H Areas, shallow groundwater outcropping, runoff from precipitation, and input of Savannah River water. The results of this study are in the following sections and in Lower (1987).

Chemical Assessment Studies

Since 1985, three special studies have been conducted on Fourmile Branch to determine the impacts of groundwater contamination on Fourmile Branch. The Savannah River Technology Center (SRTC) conducted two studies in 1988 and 1989 aimed at characteriz ing the shallow groundwater outcropping into Fourmile Branch and its associated seepline (Looney et al. 1988; Haselow et al. 1990). The third study, (Dixon et al. 1994) which began in 1991, is a continuation of the 1988 and 1989 studies and is aimed at deter mining whether annual rainfall and natural groundwater flow will dilute and flush the remaining contaminant plume out of the shallow groundwater and Fourmile Branch wet land. The Chemical Assessment Studies later in this section summarize the results of the SRTC studies.

Field Data

Water Temperature

When C-Reactor was operating (from the 1950s to 1985), temperatures in Castor Creek reached in excess of 70'C (158°F), while temperatures in the nonthermal portions of Fourmile Branch averaged 16.9°C (62.4°F) at Fourmile Branch near Road 3 (Table 5-66). After the thermal and nonthermal waters of this system blended, the mean temperature was 39.4° C (102.9°F) (Fourmile Branch at Road A-13.2), and then dropped to 28° C (82.4°F) near the confluence with the Savannah River.

Between C-Reactor's shutdown in 1985 and 1991, temperatures in Fourmile Branch at Road A ranged from 4.2° C to 31° C (39.5 to 87.8°F) with means ranging from 16.8 to 18.5 °C (62.2 to 65.3 °F) and averaged 18.5 °C (65.3 °F) (Table 5-66). Since 1992, the temperatures have ranged from 4.5° C to 25.5° C (40.1 to 77.9 $^{\circ}$ F) and averaged 17 $^{\circ}$ C (62.6 $^{\circ}$ F) (Table 5-66). The wide fluctuations in temperature reflect seasonal temperature differ ences. Temperatures upstream at Road A-7 reflect a similar range of 6.4° C to 27° C (43.5) to 80.6° F) and an average of 17° C (62° F).

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Table 5-66. Fourmile Branch Field Data

Source: Newman et al. 1986; Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988; Arnett 1993, 1994, 1995, 1996.

Blank spaces = mean not calculated due to insufficient data in report.

NA **=** not analyzed.

pH Measurements

Since 1987, the pH in Fourmile Branch has varied from 5.4 to **8.1** at Road A-7 and from 3.1 to 8.5 at Road A (Table 5-66). The pH measurements from the CCWS also fell within these ranges. The change in the ionization constant $(K_w = [H^+] [OH])$ associated with the change in water temperature is responsible, in part, for the changes in pH (Newman et al. 1986).

Physical Characteristics and General Chemistry

Dissolved Oxygen

Concentrations of dissolved oxygen are strongly correlated to water temperature. Because there was no thermal input near Road A-7, dissolved oxygen concentrations measured between 1987 and 1991 (mean 8.4 mg/1) were similar to those measured during the CCWS (mean 7.8 mg/1) (Table 5-67). During the CCWS, the mean dissolved oxygen concentration

Table **5-67.** Fourmile Branch Physical Characteristics and General Chemistry

Sources: Newman et al. 1986; Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988; Arnett 1993, 1994, 1995, 1996.

NTU = Nephelometric Turbidity Units.

upstream of Road 3 at the swampy headwater pool (Road E-1) was 6.8 mg/l. Dissolved oxy gen concentrations near Road A- 13.2 (mean 6.0 mg/I) were lower during the CCWS than during the period following cessation of reactor operations (mean 7.9 mg/l; at Road A), due to the thermal input to the system. Mean dissolved oxygen concentrations between 1992 and 1995 were 8.9 mg $O_2/1$ at Road A and 8.22 mg $O_2/1$ at Road A-7 (Table 5-67).

Turbidity and Suspended Solids

Turbidity and suspended solids have been lower in Fourmile Branch since the cessation of C Reactor operations (Table 5-67). Mean turbidities and suspended solids near Road A-7 between 1987 and 1991 were 8.2 NTU and 5.1 mg/l, respectively. However, during the CCWS, the mean turbidity near Road 3 was 20.8 NTU, and the mean suspended solids concentration was 7.82 mg/1. Between 1987 and 1991, mean turbidities and suspended solids have been 5.2 NTU and 3.1 mg/l, respectively, near Road A. During the CCWS, the

mean turbidity near Road A-13.2 was 18.5 NTU, and the mean suspended solids concentra tion was 9.31 mg/i. Between 1992 and 1995, mean turbidity was 4.5 NTU and Road A and 7.5 NTU at Road A-7 (Table 5-67).

Conductivity

Specific conductivity in Fourmile Branch between 1987 and 1991 averaged 56.5 µS/cm at Fourmile Branch near Road A-7 and 44.3 g.S/cm at Fourmile Branch near Road A (Table 5 67). Between 1992 and 1995, specific conductivity averaged 60 µS/cm at Road A and 70.75 μ S/cm at Road A-7 (Table 5-67). During the CCWS, mean conductivity in Fourmile Branch near Road A-13.2 was 66 µS/cm. Discharging Savannah River water to Fourmile Branch resulted in this higher conductivity.

Major Anions and Cations

Alkalinity, Chloride, and Sulfate

Monitoring data from 1987-1991 indicate that mean concentrations of total alkalinity are similar in waters near Road A-7 (9.7 mg CaCO₃/l) and near Road A (10.3 mg CaCO₃/l) (Table 5-68). These mean concentrations are also similar to those measured during the CCWS (8.52 mg CaCO₃/l at Road 3 and 14.14 mg CaCO₃/l near Road A-13.2). Mean chloride concentrations between 1987 and 1991 were higher at Road A-7 (5.7 mg/l) than at Road A (3.2 mg/1), which may reflect groundwater outcroppings from the F- and H-Area seepage basins. Mean sulfate concentrations were highest near Road A-7, also reflecting groundwater input from the F- and H-Area seepage basins. Higher sulfate and chloride con centrations also were observed during the CCWS. Between 1992 and 1995, mean chloride concentrations in Fourmile Branch ranged from 3.58 to 4.22 mg/l; mean sulfate concentra tions ranged from 4.75 to 6.5 mg/l; and mean alkalinity ranged from 8.25 to 11.75 mg $CaCO₃/1$ (Table 5-68).

Calcium, Magnesium, Sodium, and Potassium

Concentrations of calcium, magnesium, sodium, and potassium near Road A-7 and Road A remain similar to the concentrations measured during the CCWS (Table 5-69). As in other onsite aquatic systems, calcium, magnesium, and sodium are transported almost entirely in the dissolved phase throughout Fourmile Branch (Newman et al. 1986). Mean concentra tions of potassium were generally below the detection limits in the nonthermal portion of Fourmile Branch, but increased to about 1 mg/l in downstream thermal waters, reflecting Savannah River source waters. Potassium is not measured during routine SRS water-quality monitoring.

Aluminum, Iron, and Manganese

Concentrations of aluminum, manganese, and iron are in Table 5-69. During the CCWS, concentrations of aluminum were highest near Road A- 13.2, reflecting concentrations in the Savannah River. Concentrations of manganese and iron were highest at the upstream loca tion (Road E-1). The high concentrations of iron (mean 3.59 mg/1) measured at the upstream Fourmile Branch location were attributed to a large amount of iron oxide deposi tion in the marshy area.

Table 5-68. Fourmile Branch Major Anions

Source: Newman et al. 1986; Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988; Arnett 1993, 1994, 1995, 1996.

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Table *5-69.* Fourmile Branch Major Cations (Total)

Source: Newman et a]. 1986; Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988; Arnett **1993,** 1994, 1995, 1996.

Blank spaces = mean not calculated due to insufficient data in report.

NA = not analyzed.

From 1987-1991, the maximum aluminum (0.37 mg/1) and manganese (0.21 mg/I) con centrations were measured near Road A-7. The maximum iron concentration, 1.7 mg/l, was measured near Road A. From 1992 to 1995, the maximum aluminum (0.865 mg/1), iron (1.51 mg/I), and manganese (0.16 mg/1) concentrations were all at Road A-7 (Table 5-69).

Nutrients

Phosphorus concentrations are in Table 5-70. The mean concentrations of phosphorus species in Fourmile Branch waters during the CCWS were higher near Road-A 13.2 (0.089 mg/1) than the concentrations near Road 3 (0.023 mg/i). Analyses determined that the larger percentage of the phosphorus was dissolved orthophosphate. From 1987-1991, mean concentrations of phosphorus have been lower near Road A than near Road A-7 (0.075 mg/i near Road A-7 and 0.028 mg/l near Road A). Between 1992 and 1995, the concentrations were almost the same (mean of 0.038 mg/i at Road A and 0.03 mg/l at Road A-7).

Nitrogen

Organic nitrogen and nitrite concentrations during the CCWS were similar at all Fourmile Branch sites (Table 5-70). Nitrite is not measured in routine water-quality monitoring, and organic nitrogen was measured only near Road A-7 between 1987 and 1995. The concentrations of organic nitrogen measured near Road A-7 were similar to those mea sured in the CCWS. Ammonia concentrations were highest in the most upstream location (Road E-1) and decreased with distance downstream.

The CCWS measured extremely high concentrations of nitrate (2.3 mg $NO₃/l$) near Road 3 during the CCWS. These elevated concentrations can be attributed to the outcropping of water from the F-and H-Area seepage basins (Fenimore and Horton 1973). Nitric acid was added to the seepage basins during that time. Ammonia and nitrate concentrations measured since 1992 are similar to prior measurements (Table 5-70).

Trace Elements

SREL measured low levels of trace elements in Fourmile Branch during the CCWS. The detection limits reported for routine monitoring are higher than the concentrations mea sured during the CCWS; therefore, only the CCWS data is discussed below. All data are in Table 5-71. The CCWS measured the highest mean concentrations of arsenic (2.5 μg/l), cadmium (0.45 μg/l), copper (3.8 μg/l), and nickel (3.6 μg/l) in Fourmile Branch at Road A. Fourmile Branch near Road A-7 had the highest mean concentration of chro mium (9.1 μ g/l), while the most upstream location, Road E-1, had the highest mean concentrations of lead $(2.2 \mu g/l)$ and zinc $(8.2 \mu g/l)$.

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Table 5-70. Fourmile Branch Nutrients

Source: Newman et al. 1986; Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988; Arnett 1993, 1994, 1995, and 1996.

NA= not analyzed.

ND = none detected.

^aAll nondetectable quantities were excluded from the calculation of means.

 b 1992-1993; nitrate + nitrite.

 c 1993-1995.

d,994-1995.

Table 5-71. Fourmile Branch Trace Elements (Total)

Source: Newman et al. 1986, Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988; Arnett 1993, 1994, 1995, and 1996.

Blank spaces = mean not calculated due to insufficient data in re

NA = not analyzed.

ND = none detected.

^aAll nondetectable quantities were excluded from the calculation of means.

Organic Carbon

During the CCWS, mean total organic carbon concentrations in nonthermal waters ranged from 4.3 to 8.0 mg/l, while the thermal waters had total organic carbon concentrations of 6.4-7.3 mg/1. Between 1987 and 1991, total organic carbon concentrations ranged from 0.06 to 9.04 mg/1.

Priority Pollutants

Lower (1987) reports the results of an instream study to determine the levels of volatile, acid, and base/neutral organics in Fourmile Branch. Concentrations of all 88 tested organics were below the detection limits at each of the three sampling locations.

In 1984, an instream survey of priority pollutants was conducted to determine the levels of volatile, acid, and base/neutral compounds in Fourmile Branch. Three locations-Fourmile Branch at Road C, Road A-7, and Road A-were sampled for this study. Potential sources for organics into Fourmile Branch were discharges from F and H Areas, shallow groundwa ter outcropping, runoff from precipitation, and input of Savannah River water. The results of this study are in the sections that follow and in Lower (1987).

Pesticides, Herbicides, PCBs, and Volatile Organic Compounds

Pesticides, herbicides, PCBs, and volatile organic compounds also are measured during routine water-quality monitoring. No pesticides, herbicides, PCBs, or volatile organic com pounds have been detected in Fourmile Branch.

Lower (1987) reports the results of analyses for pesticides, herbicides, and PCBs from 1982 to 1985; results from 1967 to 1981 can be found in Gladden et al. (1985). During these peri ods, concentrations were also near or below detection limits at all locations.

Chemical, Including Radionuclide, and Toxicity Assessment Studies

Seepline Metals

In 1987, SRTC initiated a sampling and analysis program in the vicinity of the F- and H-Area seepage basins because observations of the basins suggested that water outcropping along the seepline was impacting a small area of the terrestrial/wetlands environment along Fourmile Branch. SRTC collected samples upstream and downstream from both F and H Areas (Figure 5-24) and analyzed for selected metals, inorganic constituents, pH, con ductivity, nitrate, and nitrite. The results of this study are presented in Table 5-72.

Results from the analysis of the water from Fourmile Branch suggested that the input of sodium nitrate from the seep areas or area outfalls was measurable in the stream. The remaining constituents did not measurably impact the stream (Looney et al. 1988).

Haselow et al. (1990) collected soil cores, stream, and seepline water samples from Four mile Branch in 1988 and 1989 as a follow-up to the Looney et al. (1988) study. These sam ples also were analyzed for selected metals, inorganic constituents, pH, conductivity, nitrate, and nitrite. The results of this study are in Table 5-72.

Metals analyses showed that all concentrations in Fourmile Branch were below the pro posed drinking water standards for those analytes. Aluminum and sodium concentrations were elevated, but no standards exist for these constituents. The high sodium concentrations are probably from the caustic discharged to the basins, and the aluminum is probably being leached from the soil matrix. Nitrate was elevated at both the F- and H-Area seeplines but

	88CW-1 ^a	88CW-2	88CW-3	88CW-4	89HW-9	89HW-4	89HW-7	89HW-8
pН	NA	NA	NA	NA.	6.30	7.21	6.37	5.34
Conductivity (mV/cm)	NA	NA	NA	NA	93.5	320	91.6	35.9
Alkalinity (mg/l)	NA	NA	NA	NA	ND	148	8.0	ND
Total Hardness	NA	NA	NA	NA	139	73.3	196	196
Chloride mg/l)	NA	NA	NA	NA	4.6	1.2	5.5	5.7
Nitrate (mg/l)	NA	NA	NA	NA	1.4	8.9	5.7	0.15
Silver (mg/l)	< 0.01	< 0.01	< 0.01	< 0.01	< 0.004	< 0.004	< 0.004	< 0.004
Aluminum (mg/l)	< 0.2	< 0.2	<0.2	< 0.2	0.05	0.08	0.47	0.07
$\text{Arsenic}(\text{mg/l})$	NA	NA	NA	NA	< 0.003	< 0.003	< 0.003	< 0.003
Cadmium (mg/l)	< 0.005	< 0.005	< 0.005	< 0.005	< 0.006	<0.006	< 0.006	< 0.006
Chromium (mg/l)	< 0.01	0.01	< 0.01	0.018	< 0.003	< 0.003	0.021	< 0.003
Copper (mg/l)	< 0.025	< 0.025	< 0.025	< 0.025	< 0.007	< 0.007	< 0.007	< 0.007
Mercury (mg/l)	< 0.0002	< 0.0002	0.0003	< 0.0002	< 0.0002	< 0.0002	< 0.0002	< 0.0002
Sodium (mg/l)	14	14	12	2.9	11.7	63.3	10.2	25.9
Nickel (mg/l)	< 0.04	< 0.04	< 0.04	< 0.04	< 0.034	< 0.034	< 0.034	< 0.034
Lead (mg/l)	< 0.005	< 0.005	< 0.005	< 0.005	< 0.003	< 0.003	< 0.003	< 0.003
Selenium (mg/l)	NA	NA	NA	NA	< 0.002	< 0.002	< 0.002	< 0.002
$\text{Zinc} \text{ (mg/l)}$	< 0.02	< 0.02	0.059	0.029	< 0.009	< 0.009	0.028	0.009

Table 5-72. Results from Seepline Sampling Study of Fourmile Branch

Source: Haselow et al. 1990.

 $NA = not analyzed$.

 $ND = not detected.$

^aRefer to Haselow et al. (1990) for description of these sampling locations.

was no more than half the concentration in the stream at all sampling locations. The concen tration of chloride apparently was unaffected by the basins because upstream samples were about the same as the concentrations along the seepline and in Fourmile Branch.

Discharges to the seepage basins were discontinued in 1988, and the basins were capped and sealed in 1990 to minimize the release of contaminants to the environment. Scientists hypothesized that after eliminating the contaminant source, annual rainfall and natural groundwater flow would dilute and flush the remaining contaminant plume out of the shal low groundwater and Fourmile Branch wetland system. After the contaminant plume in the shallow groundwater is diluted and flushed, the Fourmile Branch wetland systems below the basins will begin to recover.

To investigate this hypothesis and to monitor the postbasin-closure temporal changes in the contaminant levels outcropping along the Fourmile Branch seepline, a semiannual program was initiated to monitor Title 40 CFR Part 264 Appendix IX metals and various inorganics. Samples were collected from five seepline locations in F Area, five seepline locations in H Area, and three stream locations on Fourmile Branch. Analytical results for the seepline sampling locations are in Dixon and Rogers (1994), Chappell et al. (1995), Koch and Dixon (1997) and in Table 5-73 and Table 5-74. The sampling indicates that as of 1996 the seepline and Fourmile Branch still were influenced by contaminants migrating from the F- and H-Area seepage basins. However, when compared to 1989 concentrations (Haselow et al. 1990), the concentrations of most constituents have declined.

Table **5-73.** Average Tritium, Conductivity, and pH values for F- and H-Area Seeplines, March 1989-September 1996

Source: Koch and Dixon 1997.

Table 5-74. Average Tritium and Conductivity Values for Seepline below 643-E, March 1993-September 1996

Source: Koch and Dixon 1997.

Seepline Tritium

The Savannah River Technology Center established a monitoring program to track changes in tritium, pH, and specific conductivity at the Fourmile Branch seepline down gradient from the F- and H-Area seepage basins.

Measurements from the September 1993 survey indicated higher tritium concentrations, conductivity measurements, and pH values than in recent previous surveys, but lower than measurements recorded in 1990. Decreased rainfall prior to the sampling event contributed to the increased tritium concentrations and conductivity values during this survey.

September 1993 tritium values in the F-Area seepline ranged from 12 to 17,800 pCi/ml (Figure 5-25). Six of twenty-one locations sampled had above-background tritium activi ties that exceeded 1989 measurements by 10%. The tritium activity of 17,800 pCi/ml exceeded the maximum value recorded in 1989 (14,000 pCi/ml [Dixon et al. 1994])

Tritium values in the H-Area seepline ranged from 124 to 15,500 pCi/ml (Figure 5-26). Four of twenty-one sampling locations had above-background tritium activities that exceeded 1989 measurements by more than 10%. No tritium activity exceeded the maxi mum activity of 24,000 pCi/ml measured in 1989.

Below-average rainfall prior to the September 1993 sampling event caused the tritium plume to move upward through the soil profile, while the toe of the plume moved back toward the seepage basins and the seepline intercept. Therefore, tritium concentrations at most locations increased considerably. With increased rainfall, the plume should move deeper into the soil profile and outcrop closer to Fourmile Branch. Tritium concentrations at the seepline should return to levels typically measured in the past. It is important to note that total tritium fluxes to the wetlands and Fourmile Branch have steadily declined since basin closure. Overall results suggest that the tritium plume is flushing from the Fourmile Branch system (Dixon et al. 1994).

In September 1996, tritium concentrations ranged from 7-7150 pCi/ml at the F-Area seepline, from 20-1980 pCi/ml at the H-Area seepline and from 16-19, 100 pCi/ml in the drainage below 643-E (a decommissioned area of the Solid Waste Management Facility).

The following conclusions are based on trends in measured concentrations from 1986 to 1996:

- tritium concentrations measured at most locations in September 1996 are relatively unchanged compared to the previous sampling events, but are significantly lower than the 1986 baseline tritium concentration
- "• total tritium fluxes to Fourmile Branch and its associated wetlands have steadily declined since basin closure, supporting the hypothesis that the tritium plumes are being flushed from shallow groundwater
- * most of the tritium detected along the F-Area seepline is localized in two areas (8 of 22) sample locations)
- most of the tritium detected along the seepline below 643-E (a decommissioned area of the Solid Waste Management Facility) is on the west side of the drainage
- differences in tritium concentrations measured from one sampling event to the next represent seasonal variability and the effects of rainfall

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Figure 5-25. Comparison of March 1989, June 1993, and September 1993 Tritium Concentrations for Selected F-Area Seepline Locations

Toxicity

In 1994, a study was done to determine if the cladoceran *Ceriodaphnia dubia* is adversely affected by Fourmile Branch water that does not receive NPDES discharges, if *Ceriodaph nia* can be cultured for extended times in Fourmile Branch water, and if *Ceriodaphnia* cul tured in Fourmile Branch water are sensitive to a reference toxicant. SRS surface waters are extremely soft, with hardnesses ranging from approximately 2 to 30 mg/l. Waters this soft may not have adequate calcium or trace minerals to support long-term survival of *Ceri odaphnia.* Detailed results of this study are presented in Specht (1994a).

Fourmile Branch water was acutely toxic (measured as percent survival) to *Ceriodaphnia dubia* in three of five monthly tests and was chronically toxic (measured as reproductive success) in five of five monthly tests.The reference toxicant (sodium chloride) tests on Fourmile Branch water indicated test organisms in that water were severely stressed and extremely sensitive to the reference toxicant. Water from Fourmile Branch above any SRS NPDES outfall is not capable of sustaining cultures of *Ceriodaphnia dubia.*

In order to determine the source of the observed toxicity in Fourmile Branch, a Toxicity Identification Evaluation (TIE) was performed in 1995 on water collected from Fourmile Branch at Road F following U.S. Environmental Protection Agency protocols (Durhan et al. 1993; Norberg-King et al. 1991). The results of the TIE indicate that naturally occurring iron is responsible for the toxicity to *Ceriodaphnia dubia* (ETT Environmental 1995a, b).

Algae

Phytoplankton

The former C-Reactor cooling water effluent system, which includes Fourmile Branch and its principal tributary Castor Creek, is a low-potential impact area for phytoplankton. The food base throughout this system is composed of detrital material and attached algae (per iphyton), rather than phytoplankton, as is typical in lotic systems (Wetzel 1983). Primary producers in the Fourmile Branch system consist mainly of periphyton and macrophytes (Specht 1987).

Periphyton

The periphyton of Fourmile Branch were studied from 1983-1985 as part of the CCWS. Methods and sampling locations can be found in Specht (1987).

Macrophytes

Aquatic macrophytes provide stream structure, substrate for periphyton development, cover and substrate for smaller animals, and a source of carbon for the stream system. Although aquatic macrophytes are an important component of the function of many aquatic systems, they tend to be less important in flowing waters. Macrophytes could not colonize the channels of systems such as Fourmile Branch during the large thermal and flow impacts from reactor operations. Colonization of the stream has occurred since the cessation of C-Reactor operations in 1985; unfortunately, there are no data dealing with the invasion of submerged macrophytes in Fourmile Branch. However, there is some research, started in 1987 (Sharitz et al. 1993), addressing vegetational succession in the Fourmile Branch corridor that does deal with the emergent and wetland species colonizing the stream banks.

Comprehensive Cooling Water Study

The 1984-1985 data collected during the CCWS (Specht 1987) were the only data dealing with aquatic macrophytes in Fourmile Branch. The CCWS data are not representative of the current status of aquatic macrophytes in the Fourmile Branch system because they were collected at only two stations, both of which had been impacted by reactor opera tions. The data showed fewer total number of taxa than post-thermal and reference streams and had no taxa growing in the stream channel. Area, volume, biomass, and percent cover for aquatic macrophytes at the two stations are in Table 5-75.

Table 5-75. Annual Mean, N, and Standard Deviation for the Following Aquatic Vascular Plant Parameters at Each Mapped Station: Area, Volume, Biomass and Percent Cover, October 1984-September 1985

Source: Firth et al. (1986).

Expectations Since the Cessation of C-Reactor Operation

Normal successional patterns and development of macrophyte beds would likely have occurred in suitable sections of Fourmile Branch since the cessation of C-Reactor operation. This probability cannot, however, be supported or refuted because of the lack of observa tions and data. There may be adequate justification for either resampling of the areas stud ied during the CCWS, or a survey of the entire stream to document the invasion of this important component of the lower food chain.

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Zooplankton

Chimney and Cody (1986) performed the only systematic study of the zooplankton in Fourmile Branch which documented the temporal and spatial characteristics of zooplank ton species based on quarterly sampling from December 1984 to August 1985. Surface water grab samples were collected adjacent to macrophyte beds at a single station where the stream enters the Savannah River swamp. Due to thermal discharges, mean tempera ture was greater than 32°(89.6 F). Species richness consisted of 8 Protozoa, 21 Rotifera, 6 Cladocera, 2 Copepoda, and 1 Ostracoda.

The data in Figure 5-27 indicate that the greatest densities occurred during March 1985. With the exception of December 1984, Protozoa and Rotifera made up more than 80% of the monthly total densities. This result is typical of zooplankton populations, which are warm-water, summer species (Hutchinson 1967).

Figure 5-27. Mean Total Zooplankton Density in Fourmile Branch
Macroinvertebrates

Sampling Locations and Methods

Locations

Macroinvertebrates were sampled monthly from September 1982 to August 1983 in the mouth of Fourmile Branch (Specht et al. 1984), monthly from November 1983 through Sep tember 1984 at three locations in Fourmile Branch (Kondratieff and Kondratieff 1984, 1985; O'Hop et al. 1985), and monthly from October 1984 through September 1985 at six locations in Fourmile Branch (Figure 5-28) (Specht 1987). Because C Reactor, which discharged to Fourmile Branch, was placed on cold standby in June 1985, these data primarily represent the macroinvertebrate community that existed when C Reactor was operating at full power; however, the data also document early recovery of thermal stations between June and Sep tember 1985. After C Reactor shut down in June 1985, Lauritsen and Starkel (1989) col lected macroinvertebrate samples at one location in Fourmile Branch from June 1985 to September 1987 to document macroinvertebrate recolonization. During the recolonization study, they collected samples at time intervals ranging from one day to two weeks from June through August 1985 and monthly from October 1985 to September 1987.

	Station Name	E Area H Area F Area $\chi^{\text{Road 4}}$ Fourmile Branch -Road C Road E-1 Road C-4- C Area ◈ $01 - 6$		
01	12 (Road A-7)	Castor Creek		
$02\,$	14 (Road A-13.2)			
03	15			
04	16			
05	17			
06	$18\,$	$02 -$		
		03. $\overline{2}$ 0 1 Scale in Miles 06		

Figure 5-28. Location of the Fourmile Branch Macroinvertebrate Sampling Stations

Macroinvertebrates were sampled during the summer of 1993 at six locations in Fourmile Branch using Hester-Dendy multiplate macroinvertebrate samplers (Specht 1994b). Macro invertebrates also were sampled in September 1994 using Hester-Dendy multiplate sam plers to develop a biotic index for southeastern streams.While not specifically designed to characterize SRS streams, these data contribute to a better understanding of them. Fourmile Branch was sampled at Road C and Road A-13.2 (Table 5-76) (Specht and Paller 1995).

Methods

A majority of the Fourmile Branch macroinvertebrate sampling was conducted with Hester Dendy multiplate samplers. However, macroinvertebrates also were collected from leaf bags, artificial snags, macrophytes, natural substrates, and drift at some stations. Details of sampling methods can be found in Specht et al. (1984), Kondratieff and Kondratieff (1984, 1985), Firth et al. (1986), Lauritsen and Starkel (1989), Specht (1994b) and Specht and Paller (1995). Most of the data in this summary is from the 1984-1985 sampling period because fewer stations were sampled in the other years. Table 5-76 identifies the locations of all sampling stations and the sampling methods employed in each study.

Results

Introduction

From the onset of macroinvertebrate sampling in 1982 until C-Reactor was shut down in June 1985, Fourmile Branch was subject to severe thermal stress, with temperatures as high as 51.7'C (125°F) recorded at some macroinvertebrate sampling stations. However, during reactor outages, which sometimes lasted for several weeks or longer, stream temperatures were near ambient. Many species of macroinvertebrates were able to colonize the stream quickly during reactor outages. In addition, it is likely that some of the organisms collected in the severely thermal areas drifted downstream from portions of the stream where temper atures were ambient and that the organisms were no longer alive when collected. Therefore, the macroinvertebrate data collected from Fourmile Branch between 1982 and June 1985 must be interpreted cautiously to prevent erroneous conclusions.

Macroinvertebrate Community During Operation of C-Reactor

Nonthermal Areas

Station 12 was located at Road A-7 upstream from the discharge of thermal effluent into Fourmile Branch. Macroinvertebrate data collected from Hester-Dendy multiplate samplers (Table 5-77) indicate that this station was similar to other nonthermal SRS streams with respect to density of organisms, biomass, average number of taxa, and total number of taxa collected. Dominant taxa at Road A-7 included leeches (Hirudinea), the mayfly *Stenonema modestum,* blackflies, tanytarsine and orthoclad chironomids, and several species of hydropsychid caddisflies (Table 5-78 and Table 5-79). The taxonomic composition of the Station at Road A-7 differed from most other nonthermal stream stations sampled during the CCWS in that it was overwhelmingly dominated by filter-feeders (Table 5-80). The domination of filter-feeders makes Station 12 most similar to the macroinvertebrate com munities found downstream from L Lake and Par Pond. Fourmile Branch receives nutrient inputs from several small sanitary wastewater treatment plants and nitrogen inputs from the F -and H-Area seeplines. The creek also supports numerous beavers, which dam the creek.

Table 5-76. Locations of Macroinvertebrate Sampling Stations and Sampling Methods in Fourmile Branch, 1982-1987

HD = Hester-Dendy multiplate samplers.

 $D = Drift.$

LB =Leaf bags.

SN = Artificial snags.

SB = Substrate.

M = Macrophyte.

Table 5-77. Mean Number of Taxa, Density, and Biomass for Macroinvertebrates Collected on Hester-Dendy Multiplate Samplers in Fourmile Branch, October 1984-September 1985

Source: Firth et al. 1986.

Source: Firth et al. 1986.

^aIncludes subtaxa in parentheses.

bSubtaxa not differentiated.

Table **5-79.** (cont)

Table **5-79.** (cont)

Table **5-79.** (cont)

Table 5-79. (cont)

Source: Firth et al. 1986.

Table 5-80. Relative Abundance (Percent Composition) of Macroinvertebrate Functional Groups in Fourmile Branch, October 1984-September 1985

Source: Firth et al. 1986.

^aTotals not equalling 100.00% are due to rounding error.

It is likely that ponding from beaver activity occurred upstream from Station 12 and that the nutrient inputs from SRS sources stimulated algal production, which resulted in an abun dant source of food for filter-feeding macroinvertebrates.

Thermal Areas During reactor operation, few macroinvertebrates were found in the water column of Four mile Branch. Only a few species of oligochaetes, nematodes, and chironomids survived deep in the sediments. During reactor outages, the thermal portion of the stream quickly was colonized by several species of mayflies, caddisflies, and beetles. Gastropods also were collected at the thermal stations during reactor outages. The gastropods probably moved into the main channels of the creek from cooler side channels. Thus, the resident number of taxa, density of organisms, and biomass during reactor operation were much lower than suggested by the data in Table 5-77 and Table 5-79, which include data collected during reactor outages. Creek Mouth During reactor operation, water temperatures in the mouth of Fourmile Branch were always at least several degrees warmer than the Savannah River, except when the creek was flooded by high river levels. During reactor operation, water temperatures in the creek mouth exceeded 35 $^{\circ}$ C (95 $^{\circ}$ F) beginning in May and approached 40 $^{\circ}$ C (104 $^{\circ}$ F) in July. The macroinvertebrate community in the mouth of Fourmile Branch was far more diverse than that of the delta, but still exhibited strong evidence of thermal perturbation. The community was dominated by dipterans (primarily chironomids; Table 5-78). Some species, such as *Hyalella azteca* (Amphipoda) and *Caenis* spp. (Ephemeroptera), were abundant during the winter and spring. *Physella heterostropha* (Gastropoda) was abundant from January to April, especially on the multiplate samplers positioned near the creek bottom. From June to October, densities of all macroinvertebrates were low and the fauna consisted mostly of *Cheumatopsyche* spp. (Trichoptera), *Simulium* spp. (Diptera), and Nematoda. When river levels were high and the Savannah River swamp was flooded, the number of taxa and densi ties increased on multiplate samplers in the mouth of Fourmile Branch. When compared to the mouths of the other four creeks that drain SRS, Fourmile Branch had far lower mean densities, taxa richness, and biomass. Recovery of Macroinvertebrate Community After Shutdown of C Reactor Introduction A smaller mesh size (#106) was used in the recovery study than in the CCWS (#600 mesh size). Therefore, density estimates from the recovery study are not directly comparable to density estimates from the CCWS. The smaller mesh size was used to collect very early instars of macroinvertebrates that might otherwise be missed by conventional sampling. Recolonization Following the shutdown of C Reactor in June 1985, macroinvertebrates rapidly colonized the stream. Within two months of shutdown, macroinvertebrate taxa richness in Fourmile Branch (mean of 20 taxa/sampler) was comparable to that of nonthermal SRS streams sam pled during the CCWS (19.7-25.9 taxa/sampler). In subsequent months, the number of taxa

collected was variable, but increased slightly over time, to about 23 taxa/sampler when sam pling ended in September 1987 (Figure 5-29). Mean densities of macroinvertebrates were low during the first few weeks after reactor shutdown. However, in July 1985 densities increased sharply and peaked at 23,631 organisms/m2 (Figure 5-30), which was the highest density reported during the recolonization study.

Taxa Collected

Table 5-81 lists all macroinvertebrate taxa collected during the recolonization study. The dominant organisms found during the first months of recovery were Orthocladiinae, Tany tarsini, and other chironomid early instar larvae, which accounted for up to 98.8% of the macroinvertebrates collected. Chironomids are often among the first colonizers of aquatic habitats that have been severely stressed. These data indicate that recolonization of Four mile Branch following reactor shutdown was rapid for certain groups of aquatic insects. Other insects, such as caddisflies and mayflies, were somewhat slower to colonize, but still were present in relatively large numbers within two months of reactor shutdown. Mean macroinvertebrate biomass was similar to nonthermal streams within a month of reactor shutdown, and the distribution of biomass in functional groups was similar to most nonther mal streams at SRS.

Macroinvertebrate Community After Long-Term Shutdown of C Reactor

1993

Three of the six stations sampled in 1993, (at Roads 4, A-7, and A- 13.2) had relatively high numbers of taxa (ranging from 46 to 52) and high mean numbers of taxa per sampler (rang ing from 20.2 to 32.2; Table 5-82). Two of the remaining stations (at Road E-1 and C-4) had somewhat lower values for these parameters, while the macroinvertebrate commu nity at Road C was extremely depauperate with only five taxa collected. The macroinverte brate community was dominated by chironomids at most stations. Ephemeroptera (mayflies) were the most abundant group at the Road A- 13.2 station, which is downstream of all the outfalls and in the section of the stream that received thermal effluent (Table 5-83). Collector-filterers, which feed on suspended organic matter were common at both Road A 13.2 (29.7% of the functional groups) and Road A-7 (35.7%) stations (Table 5-83). The source of perturbation at Road C appears to be low dissolved oxygen concentrations. Dis solved oxygen levels at that station were extremely low at the time of sample collection (0.8mg/l) and are probably responsible for the poor macroinvertebrate community at this location. Contrast these data with data from 1994, when dissolved oxygen concentration at the time of collection was 4.7 mg/l, still the lowest from all the creeks sampled, but not so low as to preclude the survival of macroinvertebrates. The macroinvertebrate data from 1993 indicate that the quality of the stream appears to improve with downstream distance (Specht 1994b).

1994

In 1994, there were 55 taxa collected in Fourmile Branch at Road C and 48 taxa at Road A-13.2. Mean number of taxa per sampler was 27.6 at Road C and 30.6 at Road A-13.2 (Table 5-82). Biomass was 0.0904 g/m^2 at Road C and 0.3176 g/m^2 at Road A-13.2. Tanytarsii (41%), oligochaetes (13%), Tanypodinae (14%), and Chironomini (11%) dominated

Figure 5-29. Macroinvertebrate Taxa Richness in Fourmile Branch at Road A-13.2 following C-Reactor Shutdown (Scale on y-axes are different).

Figure 5-30. Total Macroinvertebrate Mean Density (no./m²) in Fourmile Branch at Road A-13.2 following C-Reactor Shutdown

Table **5-81.** Macroinvertebrate Taxa Found on Artificial Substrates in Fourmile Branch, June 1985-September 1987

Table **5-82.** Mean Number of Taxa, Density and Biomass for Macroinvertebrates Collected on Hester-Dendy Multiplate Samplers in Fourmile Branch, July - August **1993** and September 1994

Source: Specht 1994b; Specht and Paller 1995.

Station designations between the years are not comparable.

Table 5-83. Percent Composition of Macroinvertebrate Taxa Collected on Hester-Dendy Multiplate Samplers in Four mile Branch, July - August 1993 and September 1994

Source: Specht 1994b; Specht and Paller 1995.

'Note that station designations between the years are not comparable.

the macroinvertebrate community at Road C. At Road A-13.2, Ephemeroptera (45%), Tany tarsini (20%), and Trichoptera (13%) were the most common organisms (Table 5-83). Col lector-gatherer was the most common functional feeding group at both locations (Table 5 84). The macroinvertebrate community at Road C underwent substantial recovery between 1993 and 1994 (Specht and Paller 1995).

Table 5-84. Relative Abundance (Percent Composition) of Macroinvertebrate Functional Groups in Fourmile Branch, July - August 1993 and September 1994

Source: Specht 1994b; Specht and Paller 1995.

^aStation designations between the years are not comparable.

Fish

Introduction

Classifying Studies

Fish studies in Fourmile Branch can be classified temporally into three phases. The earliest phase (prior to 1980) was descriptive and nonquantitative. Results of this work are summa rized in a list of species and their relative abundances in Fourmile Branch and other riverine ecosystems (Bennett and McFarlane 1983), which is not discussed further.

Assessment of Thermal Impacts

The second and most extensive phase (1983-1985) concentrated on the assessment of impacts caused by the thermal discharges from C Reactor. Fourmile Branch was sampled at various locations, including the ambient temperature section of the stream above C Reactor, the section below C Reactor, and the section that flowed through the Savannah River flood plain swamp (Figure 5-31). These studies focused on distribution and relative abundance, reproduction, and habitat use.

Assessment of Potential Impacts from Contaminants

The third phase of work (performed in 1990) was undertaken to assess potential impacts to fish communities associated with contaminants outcropping from the waste disposal sites in F and H Areas. The latter work also documented the changes that have occurred in the Fourmile Branch fish community since C Reactor operations were terminated in 1985. The following discussion summarizes the second and third phases of fisheries work.

Figure 5-31. Fish Sampling Stations on Fourmile Branch

C-Reactor Operations

Fish Movements in Relation to Reactor Operation

Introduction

Figure 5-32. Variability in the Daily Mean Number of Individuals Entering Fourmile Branch at Different Reactor Cycles, **1983-1985**

Upstream Migration

A dam approximately 1 km (0.6 mi) above the nets limited upstream migration within Four mile Branch. Additionally, a nearby backwater served as a limited refuge area when water temperatures were high (McFarlane 1976). Observations during reactor outages revealed that the dam provides a barrier to the upstream migration of fish independent of reactor operation.

Species Collected

Totals of 39 and 29 species of fish were collected from the Fourmile Branch and Pen Branch channels (Aho et al. 1986), respectively. Centrarchids (sunfishes and basses) were the most abundant taxa, more than 30% of the species (Figure 5-33). Four species (spotted sunfish *[Lepomis punctatus],* lake chubsucker *[Erimyzon sucetta],* golden shiner *[Notemigonus cry soleucas],* and redbreast sunfish *[Lepomis auritus])* accounted for more than 50% of the fish

entering the stream channels. Individuals caught moving into the stream included both juve niles and adults. Cyprinids, principally the coastal shiner *(Notropis petersoni),* and mosqui tofish *(Gambusia affinis),* frequently were observed in the streams during periods of ambient conditions. Their absence from the collection (Figure 5-33) reflects a sampling bias of the nets against small fishes.

Effects of Reactor Restart

The number of fish killed when thermal conditions were reestablished was low compared to the cumulative number that moved upstream during ambient water temperature conditions. Visual observations suggested that fish responded to the increased flow rates and release of heated effluents associated with a reactor restart by moving downstream into the Savannah River floodplain swamp. Some potential refuge areas existed along the stream margin (e.g., marshes and isolated groundwater seepage zones), but were either ephemeral or too small to provide sufficient shelter for the number of fish that migrated upstream during reactor out ages.

Fish Assemblages at Thermal Swamp Sites

Introduction

Aho et al. (1986) also studied the fish assemblage structure within the Savannah River swamp system. Four sites (FM1, FM2, FM3, and FM4) were in or near the flow path of Fourmile Branch (original text has descriptions of locations). A total of 11,996 individuals representing 51 species were collected by electrofishing during the study. (Table 5-85 lists the species collected in Fourmile Branch.) The vast majority of species were categorized as year-round residents. However, two species (hickory shad *[Alosa mediocris]* and striped mullet *[Mugil cephalus])* were migratory. Centrarchids and cyprinids represented more than 40% of the taxa and individuals.

Faunal Groups

Although the more common species occurred throughout the Savannah River swamp sys tem, a cluster analysis based on relative abundance provided evidence of spatial variation (Figure 5-34). Eight major faunal group clusters were formed. The four sites within Four mile Branch grouped together, along with the three sites associated with Stave Island (downstream of Fourmile Branch in the Savannah River Swamp) and one site in Steel Creek. Temporal variability in assemblage composition was minor compared to spatial dif ferentiation of the groupings. In the majority of cases (>70%), the seasonal censuses from the same site (spring, summer, and fall) fell within the same cluster.

Dominant Species in Faunal Groups

While no cluster had a unique fish fauna, cross classifying the clusters with the dominant fish species (those contributing at least 5% of the numerical size of a faunal group) high lighted basic compositional differences between the different groups (Table 5-86). Large bodied species, such as the bowfin *(Amia calva),* gizzard shad *(Dorosoma cepedianum),* longnose gar *(Lepisosteus osseus),* and largemouth bass *(Micropterus salmoides)* were the dominant species of the depauperate Group A assemblage, which included the Fourmile Branch sites. Largemouth bass, brook silversides *(Labidesthes sicculus),* longnose gar, and cyprinids also were identified as common species (Figure 5-33). For the remaining faunal

Table 5-85. Fish Species Present at Four Fourmile Branch Study Sites within the Savannah River Swamp System

Source: Aho et al. (1986).

^aFewer than five individuals were collected.

Figure 5-34. Average Linkage Cluster Analysis of the Fish Assemblage Structure from Collections from Twelve Sites During Three Seasons in the Savannah River Swamp System

Table 5-86. Two-Way Table Comparing Faunal Clusters and Fish Species Associations Contributing at Least 5 Percent to the Abundance of a Specific Cluster within the Savannah River Swamp System.

Source: Aho et. al. 1986.

aFaunal clusters are as shown in Figure 5-34.

^bMean species richness is derived from the total number of species from each site-season collection; mean abundance is the mean number of fish per 50-m transect.

> clusters, the species assemblage had greater representation by minnows and centrarchids. Groups C through **E** represented a shift in general body shape from large-bodied to an assemblage dominated by small-bodied fishes.

Spatial Differentiation

The spatial differentiation in assemblage structure corresponded to two major habitat gra dients within the Savannah River swamp system (Table 5-87). Faunal cluster A was most closely linked with sites that had high water temperatures. Areas experiencing high levels of thermal loading had depauperate assemblages dominated by mosquitofish when water temperatures were elevated, but were rapidly reinvaded when ambient temperature condi tions returned. The second gradient involved different degrees of canopy closure. Faunal Cluster B was associated with sites characterized as having a predominantly closed over story of cypress and tupelo, limited emergent macrophyte growth, and low habitat com plexity; open water and standing tree buttresses were the dominant habitat features. The number of species was moderate compared to other groups, but abundance was low and dominated by taxa commonly associated with less structured habitats. Open canopy, braided water courses with periodically high flow rates, and high habitat diversity with extensive macrophyte growth and downed timber characterized faunal clusters C through H and included sites where the canopy had been lost as a result of prior disturbances. Spe cies associated with structurally complex habitats dominated these groups. Subtle eleva tions in water temperature at sites near Stave Island (S12, S13, and S14) did not appear to additionally influence assemblage structure for either open or closed canopy groups.

Table 5-87. Summary of the Site Characteristics for the Twelve Locations Studied for Fish Species Composition within the Savannah River Swamp System

Source: Aho et al. 1986.
^a No instream macrophytes or overhanging vegetation.

Effects of Canopy Cover

Several factors may have contributed to the difference in species richness and abundance between open- and closed-canopy areas. The occurrence of macrophyte growth can affect fish-assemblage dynamics by providing a refuge from predators (Werner and Hall 1979; Werner et al. 1983; Keast 1978). Macrophytes also may influence the availability of food resources by increasing the amount of substrate for secondary productivity of aquatic inver tebrates, the dominant prey category for most of the fishes in the Savannah River swamp system.

Fish Assemblages at Thermal Stream Sites

Upstream of C Reactor

Paller et al. (1986) electrofished Fourmile Branch above C Reactor (at SRS Road A-7) on six dates and identified 20 taxa. Minnows were the most numerically abundant taxon fol lowed by redbreast sunfish, spotted sunfish *(Lepomis punctatus),* and pirate perch *(Aphre doderus sayanus)* (Table 5-88). These taxa were also among the most numerous species in the undisturbed upper coastal plain streams of South Carolina as determined by Meffe and Sheldon (1988).

Table 5-88. Number of Fish Collected During Electrofishing at Fourmile Branch Near Road A-7

Source: Paller et al. 1986.

Downstream of C Reactor

No attempts were made to collect adult or juvenile fish in the portion of Fourmile Branch immediately below C Reactor. It is assumed that this area was devoid of fish when C Reac tor was operating (except for some possible short-term migration of fish into this area dur ing reactor outages) because average monthly temperatures during C-Reactor operation exceeded 50'C (122°F) throughout the year.

Paller and Saul (1986) studied fish communities in SRS streams between November 1984 and August 1985. Sampling stations for this study included sites near Road A and Road A- 13.2, where electrofishing was scheduled quarterly. Collection data from those sites are summarized in Table 5-89. The observations from the study by Paller and Saul (1986) indi cate that only mosquitofish inhabited the portion of Fourmile Branch below the C Reactor outfall during reactor operation. However, when C Reactor shut down, sunfish migrated into the area according to the findings of Aho et al. (1986) discussed earlier.

Effects of Cool Water

Paller and Saul (1986) collected additional samples in the lower reaches of Fourmile Branch (February 21, 1985). They sampled fish on one occasion (February 21, 1985) in the Fourmile Branch delta using a backpack electroshocker. Collections were taken from rela tively cool backwater areas during a period of reactor operation. Only one specimen, a brook silverside, was captured. Mosquitofish were observed in the cooler portions of the Fourmile Branch delta on numerous occasions when C Reactor was operating. Thus, mos quitofish were undoubtedly year-round residents of portions of the delta.

Table 5-89. Fish Collected from Fourmile Branch near Road A and Road A- 13.2

Source: Paller and Saul 1986.

Effects of Reactor Operation

Paller and Saul (1986) electrofished the mouth of Fourmile Branch on 15 occasions (Table 5-90). Eleven taxa were represented. They found that few or no fish were present when C Reactor was operating and water temperatures were elevated, but that fish rapidly reinvaded the area during reactor outages.

Ichthyoplankton Distribution

Introduction

Paller and Saul (1986) also collected ichthyoplankton from seven sampling stations on Fourmile Branch during 1984 and 1985. One station was situated near Road A-7 in the undisturbed headwaters upstream from the confluence with the reactor cooling water effluent. One station was at Road A, approximately 8 km (5 mi) downstream from C Reactor, one at the inflow into the delta, three in the thermal swamp downstream from the delta, and one in the creek mouth (Figure 5-3 1). The three thermal swamp stations were grouped together in the following analysis because they had similar habitats and temperatures.

Source: Paller and Saul 1986.

Effects of Reactor Operation

A total of 206 ichthyoplankton was collected from Fourmile Branch between March 14 and July 31, 1984. Centrarchids were the most abundant taxa, although brook silverside and blueback herring *(Alosa aestivalis)* were also well-represented (Table 5-91).

C Reactor was operating at full power throughout April and May 1984, and operation was intermittent during March. As a result, temperatures at Road A ranged from 33.9 to 40.1°C (93-104°F), and temperatures at the inflow into the delta ranged from 30.1 to 44.8°C (86.1-112.6°F) during the spring spawning season of 1984 (Figure 5-35). Ichthyoplankton were absent from these sites with the exception of some brook silverside eggs and unidentifiable eggs collected from the Road A sample site in May 1984. These eggs probably drifted into the channel of Fourmile Branch from cooler side-channel waters (Paller 1985).

Thermal Variation in Swamp

Temperatures in the Fourmile Branch thermal swamp and creek mouth ranged from 18 to 42°C (64.4 to 107.6°F), and were lower and much more variable than at the inflow into the Fourmile Branch delta and at the Road A sample stations. The temperature variability in the thermal swamp was due to the intermittent intrusion of relatively cool river water during periods of high water in the Savannah River. During these periods, the river water displaced the thermal plume and created suitable habitats for fishes in normally hot areas. Most of the larvae collected from the Fourmile Branch thermal swamp during April and May 1984 were spawned during periods of high river water levels when the swamp was inundated with cool river water. These larvae were principally centrarchids, but also included blueback herring and threadfin or gizzard shad *(Dorosoma* spp.). Some larvae also were collected when tem peratures were relatively high (37°C [98.6°F]) in mid-April 1984. These larvae may have drifted into the main swamp channels from the cooler backwater areas (Paller 1985).

Table **5-91.** Number and Percent Composition of Ichthyoplankton Collected from Fourmile Branch, March 14-July 31, 1984

Source: Paller 1985. ^aPrimarily eggs.

Figure 5-35. Average Ichthyoplankton Densities and Water Temperatures at Four Sampling Stations in Fourmile Branch March-July 1984

A total of 174 ichthyoplankton was collected from Fourmile Branch between February and July 1985 (Table 5-92). Unidentified ichthyoplankton (primarily eggs) were most common. Mean densities of ichthyoplankton upstream from C Reactor were generally low (<15/ 1000 **m³)** (Figure 5-36), and most of the organisms collected were minnows or cen trarchids. Throughout the sampling, ichthyoplankton were largely absent from the sample station near Road A, where water temperatures sometimes exceeded 40'C (104'F) when C Reactor was operating. Farther downstream in the delta, some cooling had occurred, but temperatures still remained near 40'C (104'F) during much of the sampling.

Mouth of Fourmile Branch

No larvae or eggs were collected from the mouth of Fourmile Branch except in February, March, and May 1984. The ichthyoplankton densities in the mouth of Fourmile Branch were greatest in May. All of the ichthyoplankton collected during May were unidentifiable eggs. They were taken on a single sample date when C Reactor was briefly shutdown and the water temperature was 27°C (80.6°F). These data suggest that fish began spawning in the creek mouth as soon as temperatures became tolerable (Paller 1985).

Comparison of 1984 and 1985 Data

Except for the densities at the creek mouth, ichthyoplankton were less abundant in Four mile Branch during 1985 than during 1984 (Figure 5-37). The differences in the swamp and creek mouth ichthyoplankton densities probably were due to differences in the level of the Savannah River during those years. During the spring of 1984, the swamp was intermit tently flooded by cool river water (Paller 1985). Most of the larvae taken from the swamp during 1984 were collected when the swamp was flooded. Conversely, the Savannah River generally remained below flood stage during 1985, and relatively few ichthyoplankton were collected from Fourmile Branch.

Effects of Temperature on Ichthyoplankton

Introduction

Aho et al. (1986) studied larval fish assemblages at nine sampling locations in the Savan nah River swamp system during 1985. Five stations spanned a wide range of temperatures (temperatures elevated 2-19'C [7.2-34°F] above ambient) along the plumes from Fourmile Branch and Pen Branch. These were numbered 1 through 5, from hottest to coolest. The other four stations near the mouth of Steel Creek were at ambient temperatures (Aho [1986] has a description of station locations). Each sampling station was an area roughly 50 m (164 ft) in diameter and included both channels and adjacent shallows (except two stations that had no distinct channels).

Effect of Reactor Operation

At stations within the thermal plumes, water temperatures changed suddenly when reactors stopped and started. During these reactor cycles, recording thermometers showed tempera tures fell to ambient levels (and subsequently rose) at about $1^{\circ}C$ (1.8°F)/hr at the Fourmile Branch stations. Stations 4 and 5 had more gradual temperature changes due to their dis tances from the main flow from Pen Branch. The most extreme fluctuations among the nonthermal sites occurred at the station in a shallow disturbed area that lacked canopy veg etation, where temperatures changed 20'C (36°F) over two weeks in February.

Table 5-92. Number and Percent Composition of Ichthyoplankton Collected from Fourmile Branch, February-July **1985**

	Number	Percent
blueback herring	3	1.7
minnows	14	8.0
carp	7	4.0
mosquitofish	2	1.2
brook silverside	2	$1.2\,$
sunfish and/or bass	14	8.0
crappie	3	1.7
darters	6	3.4
unid. ichthyoplankton ^a	123	70.7
total	174	99.9

Source: Paller et al. 1986. a Primarily eggs.

Figure **5-36.** Monthly Mean Water Temperature and Ichthyoplankton Density at Each Sampling Site in Fourmile Branch, February-July 1985

Figure 5-37. Total Ichthyoplankton Densities at Fourmile Branch Locations during the 1984 and 1985 Sampling Program (Source: Paller et al. 1986)

No larvae were collected in the first week of January at any location, although water tem peratures exceeded 15'C (59°F) in the hottest thermal areas. Densities began to rise at ther mal Stations 1 through 3 by the end of January. In the ambient temperature areas, density peaks occurred later in the year.

Densities

The timing of peak ichthyoplankton densities differed by as much as eight weeks between the hottest thermal area and the ambient temperature stations (Figure 5-38). Seasonal pat terns were advanced, even at Station 5, where temperatures were only about two degrees warmer than in the natural or previously disturbed habitats (Figure 5-38).

Maximum densities occurred two to four weeks earlier in the previously disturbed areas, where canopy trees had been killed, than in the natural areas; although water temperatures were similar (Figure 5-38).

Reproductive Phenology

In the thermal areas, reproductive phenology was disrupted in several respects. Most pro nounced was a disruption in the sequence of appearance of larvae of different taxa. Larval cyprinids, golden shiners, brook silversides, sunfish, and darters all were taken at Stations 1 and 2 by the first week of February, suggesting that many of the normally spring- or summer-spawning species began to reproduce in the thermal areas in early January. Some of these species may have been capable of spawning during the fall and early winter in thermal areas. Water quality values other than temperature appeared similar at all stations. Thus, shifts in reproduction or changes in species composition probably result directly or indirectly from differences in temperature, light intensity, or trophic base of the commu nity.

Figure 5-38. Seasonal Changes of Total Ichthyoplankton Density at each Station (These station numbers do not corre spond to those on the map in Figure 5-33)

F- and H-Area Impacts

Introduction

Paller and Storey (1990) electrofished Fourmile Branch during June 26-July 2, 1990, to assess the impacts of outcropping groundwater from the F- and H-Area seepage basins on fish abundance and distribution. Effluents formerly discharged to the basins contained sodium hydroxide, nitric acid, low levels of radionuclides (mostly tritiated water), and dis solved metals (Looney et al. 1988, Haselow et al. 1990). These effluents seeped into the ground, migrated through the subsurface strata, and outcropped into Fourmile Branch and adjacent wetlands. Seepage basin discharge appeared to cause elevated conductivity, total dissolved solids, nitrate, phosphate, sodium, potassium, and, possibly, cadmium levels in Fourmile Branch downstream from the seepage basins (Looney et al. 1988). In addition, gross beta and tritium levels were above either the proposed or established drinking water standards at one or more points in Fourmile Branch (Haselow et al. 1990).

Effects of Seepage Basin Constituents on Fish

The occurrence of elevated levels of several seepage basin constituents in Fourmile Branch water raised concerns about possible impacts to aquatic organisms. Gladden et al. (1985) reviewed the historical data from the upper reaches of Fourmile Branch to determine if out cropping effluent from the seepage basins was adversely affecting the instream communi ties. He concluded that there was no clear evidence of adverse impacts due to the seepage basin effluent, but said the data were insufficient to evaluate possible local effects. To pro vide more information, a sampling program was designed to assess the abundance, distribu tion, and tissue contaminant levels of Fourmile Branch fish upstream and downstream from the seepage basins. This program included seven sample stations: six in Fourmile Branch, including locations upstream from (Stations 1 and 2) and downstream from (Stations 3 - 6) the seepage basins, and one in Pen Branch (Figure 5-3 1). The sample station in Pen Branch was included primarily to serve as a source of uncontaminated fish to establish background levels of potential tissue contaminants. Three 100-m (328 ft) stream segments were electro fished at each sample station.

Sampling Program

Fish assemblage structure differed among sample stations. Pirate perch, redbreast sunfish, and creek chubsuckers *(Erimyzon oblongus)* dominated Station 1. Several types of shiners *(Notropis* spp.) (dusky, yellowfin, or taillight) and sunfishes *(Lepomis* spp.) (dollar, spotted, or redbreast) dominated Stations 2-4. Mosquitofish, redbreast sunfish, spotted sunfish, and yellow bullhead dominated Stations 5 and 6. The number of fish collected per 100 m (catch per unit effort or CPUE) also differed among stations (Table 5-93). The lowest CPUE in Fourmile Branch occurred at Station 2 (44.7 fish/100 m) and the highest occurred at station 4 (149.7 fish/100 m). However, a comparison of collections at Stations 1 and 2 (above the outcropping zone) with Stations 3 and 4 (below the outcropping zone) indicated that species number and total (i.e., all species summed) CPUE were higher below the outcropping zone than above (Figure 5-39; Table 5-94). On an individual species basis, four species decreased below the outcropping zone, while nine increased. These differences were not indicative of adverse impacts due to seepage basin outcropping and were more likely from habitat differ ences among stations.

Table 5-93. Relative Abundance (Percent Composition) of Fish Collected at Six Sample Stations in Fourmile Branch and One Sample Station in the Headwaters of Pen Branch, June 1990

Key to sample station locations:

1 Fourmile Branch 1.1 km (0.7 mi) upstream from Road 4.

2 Fourmile Branch at Road 4.

3 Fourmile Branch at Road C.

4 Fourmile Branch at Road A-7.

5 Fourmile Branch downstream at Road A (not on map in Figure 5-33).

6 Fourmile Branch at Cassels Pond (just upstream of delta/swamp).

Figure 5-39. Mean Electrofishing Catch per Unit Effort at Sample Stations in Fourmile Branch (Number of species collected is in parentheses)

Table 5-94. Mean (Standard Error) Electrofishing Catch Per Unit Effort (CPUE Expressed as Number Fish Per 100 m) at Six Sample Stations in Fourmile Branch and One Sample Station in the Headwaters of Pen Branch, June 1990

Key to sample station locations:

1 Fourmile Branch 1. **1** km (0.7 mi) upstream from Road 4.

2 Fourmile Branch at Road 4.

3 Fourmile Branch at Road C.

4.Fourmile Branch at Road A-7.

5 Fourmile Branch downstream of Road A (not on map in Figure 5-33).

6 Fourmile Branch at Cassels Pond (just upstream of delta/swamp).

Section 5.3-Fourmile Branch Fish

Table 5-94. Mean (Standard Error) Electrofishing Catch Per Unit Effort (CPUE Expressed as Number Fish Per **100** m) at Six Sample Stations in Fourmile Branch and One Sample Station in the Headwaters of Pen Branch, June 1990

Key to sample station locations:

I Fourmile Branch 1.1 km (0.7 mi) upstream from Road 4.

2 Fourmile Branch at Road 4.

3 Fourmile Branch at Road C.

4.Fourmile Branch at Road A-7.

5 Fourmile Branch downstream of Road A (not on map in Figure 5-33).

6 Fourmile Branch at Cassels Pond (just upstream of delta/swamp).

Changes in Fish Assemblage with Habitat Alterations

While not a consequence of seepage basin operation, species number and total CPUE decreased downstream from Station 4 (Table 5-94). Mosquitofish, a species commonly associated with thermal and post-thermal sites on the SRS (Aho et al. 1986), increased in abundance below Station 4. These changes are likely from habitat alterations associated with past thermal discharge from C Reactor. Decreases in species number and CPUE at these stations indicate that recovery from C-Reactor operation was not yet complete at the time of this study.

Fish Communities Comparison with Unimpacted Streams

The fish community below the outcropping zone, but above the former point of C Reactor cooling water entry, also was evaluated by comparing it to the fish communities in nearby unimpacted streams of similar size and generally similar habitat. Three unimpacted stream reaches (upper Meyers Branch, upper Steel Creek, and upper Pen Branch [Station 7]) were used in this comparison. Species number at Stations 3 and 4 in Fourmile Branch was com parable to species number in the other streams. Relative abundance at Stations 3 and 4 was generally similar to that in the other streams, except that sunfishes constituted a slightly higher percentage of the community and minnows a slightly lower percentage.

Fish Analysis for RCRA Trace Metals

Fish collected from Fourmile Branch and Pen Branch were analyzed for Resource Conser vation and Recovery Act (RCRA) trace metals using standard U.S. Environmental Protec tion Agency methods. Loehle and Paller (1990) report the results of the analyses. Silver, arsenic, beryllium, cadmium, copper, nickel, lead, antimony, and thallium were all either undetectable or had only a few detectable values. Mercury values were all well below 1 µg/l. For the total group of fish analyzed, there were no differences among sampling sites for aluminum, chromium, or zinc. Selenium concentrations differed among sites; fish col lected near the H-Area and two control sites had the highest concentrations. When the anal ysis was restricted to sunfishes only, the seepage basin site was shown to be slightly elevated. Among species, yellowfin shiners had higher aluminum and zinc concentrations than sunfishes and bottom fish.

Summary of Studies

These studies provided no indication that outcropping groundwater from the F- and H-Area seepage basins adversely affected the fish community in Fourmile Branch. Because there was an absence of adverse changes in community structure below the outcropping zone, community structure in this region generally was comparable to community structure in other relatively unimpacted SRS streams, and there was no indication of significant accu mulation of heavy metals in the fish. However, they do indicate that habitat alterations from past C-Reactor operations still were influencing fish community structure in Fourmile Branch at the time of the study.

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5.4 Pen Branch

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Drainage Description and Surface Hydrology

General Description

Pen Branch drains an area of about 55 $km^2(21 mi^2)$ and is approximately 24 km (15 mi) long. The creek flows southwesterly, from its headwaters about 3.2 km (2 mi) east of K Area to the Savannah River swamp (Figure 5-40). After entering the swamp, the creek flows parallel to the Savannah River for about 8 km (5 mi) before it enters and mixes with the waters of Steel Creek about 0.4 km (0.2 mi) from the mouth of Steel Creek on the Savan nah River. In its headwaters, Pen Branch is a largely unperturbed blackwater stream, similar to the headwater reaches of Fourmile Branch. Indian Grave Branch is the principal tributary of Pen Branch.

Pen Branch discharges into the Savannah River floodplain swamp rather than flowing directly into the Savannah River. The discharge of Pen Branch into the swamp formed a delta where water temperatures typically ranged from 25 to 40'C (45 to 72°F) above ambi ent during reactor operations. The flow from Pen Branch spreads over the delta and contin ues through the swamp as shallow sheet flow until entering the lower reaches of Steel Creek; from there it travels to the Savannah River. When the Savannah River inundates the flood plain swamp, Pen Branch flows along the northern border of the swamp and crosses the Steel Creek delta. When the Savannah River is not flooding, the Pen Branch flow enters the Steel Creek channel downstream from the swamp. By the time Pen Branch discharged into Steel Creek during reactor operation, its temperature was near ambient, due to dilution and cooling in the swamp.

Figure 5-40. Location of Pen Branch on SRS

Effluents Contribution

Until K Reactor shut down in 1988, Indian Grave Branch received thermal effluent from K Reactor. With K-Reactor discharge, the natural flow of about 0.3 m³/s (10 ft³/s) was increased to about 11.3 m³/s (400 ft³/s). K-Reactor cooling water discharges, which origi nated from the Savannah River, changed the water quality and temperature and flow regimes in Pen Branch (Firth et al. 1986). Currently, the Pen Branch system receives nonthermal effluents (i.e., nonprocess cooling water, ash basin effluent waters, powerhouse waste water, and sanitary waste water) from K Area and sanitary effluent from the Central Shops Area. K Reactor has been placed on permanent shutdown; therefore, flow and temperature will no longer affect the stream.

Pen Branch, via Indian Grave Branch, receives the following National Pollutant Discharge Elimination System (NPDES) permitted discharges: cooling water, blowdown, powerhouse wastewater, stormwater, 186 basin overflow, and sanitary wastewater.

Flow Measurements

The U.S. Geological Survey measures flow at several locations on Pen Branch. Table 5-95 summarizes flow statistics for Pen Branch at SRS Road B and at SRS Road A-13.2 (Figure 5-41). Records at SRS Road A-13.2 date back to November 1976 (no records from February 1983 through April 1983). In water-year 1995, the mean flow of Pen Branch at Road A-13.2 was 1.6 m³/s (55.8 ft³/s). Over the period of record (water-years 1977-1995) at Road A 13.2, the mean flow was 5.9 m³/s (210 ft³/s), the 7-day low flow was 0.25 m³/s (8.8 ft³/s) and the 7Q10 was 0.15 m^3 /s (5.46 ft³/s). The maximum, minimum, and mean daily flows from December 1983 to September 1995 at SRS Road B are shown in Figure 5-42. The maximum, minimum and mean daily flows from October 1982 to September 1995 at Road A-13.2 are shown in Figure 5-43.

Figure 5-41. Flow Measurement Sampling Stations for Pen Branch

Table 5-95. Flow Summary for Pen Branch

cms **=** cubic meters per second.

cfs = cubic feet per second.

Figure 5-42.Maximum, Mean, and Minimum Flow for Pen Branch at SRS Road B, December 1988-September 1995

Water Chemistry and Quality

Studies and Monitoring

Water-Quality Monitoring

The WSRC Environmental Monitoring Section (EMS) has conducted routine water quality monitoring of the Pen Branch system since 1973. EMS monitors one location on Pen Branch near Road A-17 (Figure 5-44; location 04) monthly for physical and biological water quality indicators and quarterly for metals. EMS also collects an additional sample annually and analyzes it for pesticides, herbicides, and PCBs. All routine water quality mon itoring data reported in the following sections can be found in the annual SRS Environmen tal Reports.

Comprehensive Cooling Water Study

Five locations on the Pen Branch system were studied from 1983 to 1985 as part of the Comprehensive Cooling Water Study (CCWS). This study was designed to assess present and proposed SRS activities on water quality. The Pen Branch sampling locations (Figure 5 44) include the following:

- Pen Branch at Road B (01) measured the effects of the effluent from the Central Shops Area
- * Indian Grave Branch downstream of K-Reactor Effluent (02) measured the effects of the thermal effluent
- Pen Branch at Road A-13 (03) downstream of confluence of Indian Grave Branch and Pen Branch

Figure 5-44. Water Chemistry and Quality Sampling Stations for Pen Branch

• Pen Branch at Road A-17 (04) - measured concentrations after release to the Savannah River Swamp " Pen Branch upstream of Steel Creek confluence (05) - measured thermal effluents and upriver waters during periods of high river water, which overflows into the Savannah River Swamp The data collected downstream of K-Reactor effluent during the CCWS reflect impacts associated with reactor operation and are not relevant to the current conditions of the stream. However, because limited monitoring data are available for the Pen Branch system, all data will be presented. Comprehensive results and discussion of CCWS data can be found in Newman et al. (1986) and Lower (1987). Priority Pollutants Survey In 1984, a special instream survey of priority pollutants was conducted to determine the lev els of volatile, acid, and base/neutral organic compounds in the Pen Branch system. Three stations - near Road C, Road B, and Road A - were established in Pen Branch. Newman et al. (1986) and Lower (1987) documented these results. Lower (1987) also reported the results of analyses for pesticides and PCBs in the Pen Branch system. Field Data Water Temperature During reactor operation, mean temperatures in thermal portions of the Pen Branch water shed (33.5 - 48.1°C [95.9 - 118.5°F]) ranged from 18 to 33°C (32 to 59°F) above those of the upstream nonthermal waters (17.4°C [63.3°F]) (Table 5-96). The temperatures at the thermal sites fluctuated more widely than those of the nonthermal site due to the reactor cycle. The shutdown of K Reactor in 1987 decreased temperatures to an average of 22°C (71.6°F) in the Pen Branch system. pH Measurements The pH values for the thermal sites (mean 7.46) were higher than those of the nonthermal site (mean 6.89), reflecting Savannah River source-water pH levels. Over the past 9 years, the pH near Road A-17 has ranged from 5.7-8.6 (Table 5-96). Physical Characteristics and General Chemistry Dissolved Oxygen The dissolved oxygen concentration is inversely related to water temperature, reflected in

the data the CCWS generated. The mean dissolved oxygen concentrations in the thermal waters were much lower (5.3-7.5 mg/I or 87-90% saturation) than those at the nonthermal site. Mean dissolved oxygen concentration was 8.12 mg/1 at the Pen Branch nonthermal site. Because there has been no thermal input to the Pen Branch system since 1987, the mean dis solved oxygen concentrations (8.5 mg/1 between 1987 and 1991 and 9.1 mg/l between 1992

Table **5-96.** Pen Branch Field Data

NA = Not analyzed.

Blank spaces = Mean not calculated due to insufficient data in report.
^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

 b 1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

 c 1992-1995 = Data taken from Arnett 1993, 1994, 1995, and 1996.

and 1995) at Road A-17 have been similar to the concentrations measured at the nonthermal site during the CCWS (Table 5-97).

Suspended Solids and Turbidity

The CCWS measured a wide range of total suspended solids concentrations (mean 3.14 13.9 mg/1). At locations with slower water velocities, the total suspended solids concentra tions were lower than at locations with higher water velocities. Data collected from 1987 to 1991 at Road A- 17 indicate slightly higher total suspended solids concentrations (mean 7.7 mg1l). Between 1992 and 1995, total suspended solids concentrations were even higher (8.25 mg/l). Turbidity between 1992 and 1995 was about half of what it was during the CCWS (7.08 NTU).

Conductivity

Mean specific conductivity increased from $45.6 \mu S/cm$ at the nonthermal sites to 73.4 pS/cm at the thermal sites (Table 5-97). This change was linked to the specific conduc tance of Savannah River source water (Lower 1987). Routine water quality monitoring since the CCWS has measured a wide range of specific conductance (13-171 pS/cm), with a mean of 77 uS/cm.

Major Anions and Cations

Alkalinity, Chloride, and Sulfate

During the CCWS, mean concentrations of total alkalinity, chloride, and sulfate ranged from 13.8 to 17.9 mg/I, 2.48 to 6.02 mg/l, and 2.53 to 5.26 mg/l, respectively. The maximum total alkalinity, chloride, and sulfate concentrations were measured at the thermal sites and likely reflected the chemistry of the Savannah River water used to cool the reactor. Mean concen trations of total alkalinity (18.2 mg/I), chloride (7.1 mg/I), and sulfate (7.4 mg/I) from 1987 to 1991 have been slightly higher than the ranges measured during the CCWS (Table 5-98). Between 1992 and 1995, mean values of chloride, alkalinity, and sulfate were lower still (Table 5-98).

Calcium, Magnesium, Sodium, and Potassium

Concentrations of total calcium, magnesium, and sodium measured from 1987 to 1991 near Road A-17 are similar to those concentrations measured during the CCWS (Table 5-99). The CCWS determined that nearly all of the calcium, magnesium, sodium, and potassium in Pen Branch waters were in the dissolved fraction (Newman et al. 1986). The mean concen tration of calcium was slightly higher at the nonthermal site than at the thermal sites. Mag nesium, sodium and potassium, however, were higher at the thermal sites than at the nonthermal sites. SRS does not measure potassium during routine water quality monitoring. Calcium, magnesium, and sodium values between 1992 and 1995 were similar to those measured during the CCWS (Table 5-99).

Aluminum, Iron, and Manganese

Concentrations of total and dissolved aluminum measured during the CCWS were higher in the waters of the thermal sites than in those of the nonthermal site (Table 5-99). Approxi mately 92% of the aluminum was associated with the solid phase at the nonthermal site and

Table **5-97.** Pen Branch Physical Characteristics and General Chemistry

NA = Not analyzed.

NTU = Nephelometric Turbidity Units.

Blank Spaces = Mean not calculated due to insufficient data in report.
^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b,987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

c1992-1995 **=** Data taken from Arnett 1993, 1994, 1995, and 1996.

Table 5-98. Pen Branch Major Anions

NA = Not analyzed.

Blank Spaces = Mean not calculated due to insufficient data in report.

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

 $c_{1992-1995}$ = Data taken from Arnett 1993, 1994, 1995, and 1996.

Table **5-99.** Pen Branch Major Cations (Total)

NA = Not analyzed.

Blank Spaces = Mean not calculated due to insufficient data in report.

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

 $°1992-1995 = Data taken from Arnett 1993, 1994, 1995, and 1996.$

92-93% was associated with the solid phase at the thermal sites. Total and dissolved iron concentrations were higher at the nonthermal site than those at the thermal sites. Approxi mately 83-86% of the iron was in the solid phase. Mean total manganese concentrations ranged from 0.071 to 0.104 mg/l. Roughly 20-28% of the manganese was in the solid phase (Newman et al. 1986). From 1987 to 1995, concentrations of total aluminum, iron, and man ganese were slightly lower than those measured during the CCWS (Table 5-99). The higher concentrations during the CCWS reflected Savannah River source water concentrations.

Nutrients

Phosphorus

All measured forms of phosphorus were higher in the thermal portions of this system than at the nonthermal site during the CCWS. Whereas 13% of the phosphorus was present as dis solved orthophosphate at the nonthermal site, 49-59% at the thermal site was dissolved orthophosphate. At the nonthermal site, 62% of the phosphorus was not orthophosphate, and 27-38% of the phosphorus at the thermal sites was not orthophosphate. The speciation and concentrations of phosphorus in the thermal sites reflected, at least partially, the phosphorus speciation and concentrations in the Savannah River (Newman et al. 1986). Only total phos phorus is measured during the routine monitoring, but concentrations of total phosphorus near Road A-17 were similar to those measured at Road A-17 during the CCWS (Table **⁵** 100).

Nitrogen

Concentrations of nitrogen species are in Table 5-100. Although the mean concentrations of organic nitrogen did not differ greatly between the thermal and nonthermal sites, the per centage of nitrogen in the organic form was 73% in the nonthermal waters and only 23-26% in the thermal waters (Newman et al. 1986). Routine monitoring does not measure for organic nitrogen in Pen Branch. During the CCWS, mean ammonia, nitrite, and nitrate con centrations were higher at the thermal sites than at the nonthermal site, reflecting the nitro gen species and concentrations in the Savannah River. Since 1987, concentrations of ammonia and nitrate have remained similar to those concentrations measured during the CCWS. Nitrite has not been measured since the CCWS.

Trace Elements

Newman et al. (1986) measured low levels of trace elements during the CCWS. Routine monitoring detection limits are higher than the concentrations measured during the CCWS; therefore, only CCWS data are discussed in this section (Table 5-101). Mean total arsenic concentrations were similar in the nonthermal waters (1.6 μ g/l) and the thermal waters (1.3 to 2.8 µg/l) of Pen Branch. Mean total cadmium concentrations ranged from 0.29 µg/l upstream of the Steel Creek confluence to 0.99 **gig/l** at Road A- 13. Although mean total chromium concentrations were as high as 10.4 µg/l, mean dissolved chromium concentrations were below the detection limit at all sites (Newman et al. 1986). Mean total copper concentrations were approximately 2.1 -3.2 μ g/l. The highest mean concentration of total nickel (4.3 μ g/l) was found near Road A-13 (4.3 μ g/l). Mean total zinc concentrations ranged from 4.7-6.2 µg/l. Mercury and uranium and concentrations were below the detection limit in Pen Branch waters (Newman et al. 1986).

Table 5-100. Pen Branch Nutrients

NA = Not analyzed.

ND = None detected.

Blank Spaces = Mean not calculated due to insufficient data in report.

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 199 **1;** Davis et al. 1989; Mikol et al. 1988.

'1992-1995 = Data taken from Arnett 1993, 1994, 1995, and 1996.

^dNondetectable quantities were excluded from the calculation of the mean.

eNitrite **+** nitrate; 1992-1993.

 f 1994-1995.

Table **5-101.** Pen Branch Trace Elements (Total)

 $NA = Not analyzed.$

ND = Not detected.

Blank Spaces = Mean not calculated due to insufficient data in report.

^aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

 c 1992-1995 = Data taken from Arnett 1993, 1994, 1995, and 1996.

dAll nondetected quantities were excluded from the calculation of the mean.

Organic Carbon

Total organic carbon concentrations were similar for thermal and nonthermal sites in Pen Branch (Table 5-102). However, the nonthermal site had slightly more organic carbon in the dissolved phase (77%) relative to those of the thermal sites (68-69%) (Newman et al. 1986).

Priority Pollutants

The Pen Branch system had below detectable concentrations of all 88 tested volatile, acid, and base/neutral organics. The 1984 study also confirmed the lack of variability between thermal and nonthermal waters (Lower 1987).

Pesticides, Herbicides, and PCBs

Water samples collected annually during routine monitoring are analyzed for pesticides, herbicides, and PCBs. None of the analytes has been detected in the Pen Branch system.

Lower (1987) reports the results of analyses for pesticides, herbicides, and PCBs from 1982 to 1985; results from 1967 to 1981 can be found in Gladden et al. (1985). During these peri ods, concentrations also were near or below detection limits at all locations.

Chemical, Including Radionuclide, and Toxicity Assessment Studies

In 1994, a study was done to determine if the macroinvertebrate *Ceriodaphnia* is adversely affected by Pen Branch water that does not receive National Pollutant Discharge Elimina tion System (NPDES) discharges; if *Ceriodaphnia* can be cultured for extended times in Pen Branch water; and if *Ceriodaphnia* cultured in Pen Branch water are sensitive to a reference toxicant. SRS surface waters are extremely soft, with hardness ranging from approximately 2 to 30 mg/I. Waters this soft may not have adequate calcium or trace minerals to support long-term survival of *Ceriodaphnia.* Detailed results of this study are in Specht (1994a) and in Chapter 7 of this document.

Pen Branch water was never acutely toxic (measured as percent survival) to *Ceriodaphnia dubia* in 11 monthly tests. However, it was chronically toxic (measured as reproductive suc cess) in 5 of 11 monthly tests. Results of the acute reference toxicant (sodium chloride) tests on Pen Branch water indicated that the test organisms in that water were slightly more sensi tive to the reference toxicant than other organisms cultured in standard dilute mineral water. Reproduction in Pen Branch waters in the presence of the reference toxicant was more suc cessful than reproduction in the control population in the presence of the reference toxicant. These results suggest that water from Pen Branch may be superior to dilute mineral water as dilution and control water for chronic toxicity testing of SRS waters (Specht 1994a).

No studies on chemical or radionuclide contaminants have been done in Pen Branch.

Table 5-102. Pen Branch Carbon Concentrations

NA = Not analyzed.

Blank Spaces = Mean not calculated due to insufficient data in report.

aCCWS = Comprehensive Cooling Water Study (Newman et al. 1986).

b1987-1991 = Data taken from Arnett et al. 1992; Cummins et al. 1990, 1991; Davis et al. 1989; Mikol et al. 1988.

Algae

Phytoplankton Phytoplankton has not been studied in the Pen Branch drainage system. The potential for activities to impact the phytoplankton is low. Phytoplankton contribute insignificantly to the food chain base in this and other shallow streams. Periphyton Studies of the ecology of the thermal and nonthermal streams of SRS that were conducted between 1983 and 1985 (Firth et al. 1986; Specht 1987) included the periphyton community of Pen Branch. The analyses of the periphyton communities in these studies included taxo nomic identifications, chlorophyll analyses, and measurement of ash-free dry weight. Reactor operations undoubtedly affected periphyton species composition and abundance. Periphyton differed in type between thermal and nonthermal sites. Green algae and diatoms characterized the nonthermal sites, while thick mats of blue-green algae were unevenly dis tributed at the thermal sites (Specht 1987). Also, variations in periphyton biomass values were higher at the thermal stream sites than at the nonthermal sites. The densely canopied Pen Branch swamp station had a significantly lower periphyton biomass than the Pen

by the status of K Reactor (Firth et al. 1986).

Branch delta station. Biomass at the Pen Branch swamp station did not appear to be affected