Near-Field and Altered-Zone Environment Report Volume I: Technical Bases for EBS Design Revision 1

Dale G. Wilder



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Prepared by Yucca Mountain Site Characterization Project (YMP) participants as part of the Civilian Radioactive Waste Management Program. The YMP is managed by the Yucca Mountain Site Characterization Project Office of the U.S. Department of Energy, Las Vegas, Nevada.

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Acronyms and Abbreviations

	•
ACD	advanced conceptual design
AML	areal mass loading
APD	areal power density
AZ	altered zone
BWR	boiling water reactor
СН	Calico Hills Formation
DHLW	defense high-level waste
DOE	U.S. Department of Energy
EBS	engineered barrier system
ECM	equivalent continuum model
ESF	Exploratory Studies Facility
FFE	far-field environment
LA	license application
LAD	license application design
LLNL	Lawrence Livermore National Laboratory
LML	lineal mass loading
LPL	lineal power loading
MTU	metric tons of uranium
NF	near field
NFE	near-field environment
NFER	Near-Field Environment Report
NTS	Nevada Test Site
PPw	Prow Pass welded unit
PTn	Paintbrush Tuff vitric nonwelded unit
PWR	pressurized water reactor
QA	quality assurance
QARD	Quality Assurance Requirements and Description (DOE, 1997)
RIB	Reference Information Base (DOE, 1995)
SCP	Site Characterization Plan (DOE, 1988)
SNF	spent nuclear fuel
SHT	Single Heater Test
SZ	saturated zone
TBM	tunnel boring machine
TDIF	Technical Data Information Form
T-H	thermal-hydrological
T-H-M-C	thermal-hydrological-geomechanical-geochemical
TCw	Tiva Canyon welded unit
Tpt	Topopah Spring Tuff of the Paintbrush Group (stratigraphic unit)
ISPA	Total System Performance Analysis
ISw1, 2, and 3	Topopah Spring Tuff (thermal/mechanical units, per Fig. 2-3)
UNE	underground nuclear explosion
UZ	unsaturated zone
WP .	waste package
NPP	Waste Package Plan (Harrison-Giesler, 1991)
(M	Yucca Mountain
(MP	Yucca Mountain Site Characterization Project
(MSCO	Yucca Mountain Site Characterization Project Office

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Glossary

ACD rev 00 design

Altered Zone (AZ)

Areal mass loading

Axial WP spacing

Cladding

Drift diameter

∆Tdrift

EBS performance

Emissivity of WP

Lineal mass loading

Line-load design

This report uses the Advanced Conceptual Design (ACD) from TRW, 1996, as the Reference Design for the repository. The geometry of the ACD has a spacing between drift centerlines roughly the same as axial spacing between WPs. The drift spacing is 22.5 m, and the areal mass loading AML is 83.4 MTU/acre.

In general, the region extending from tens to hundreds of meters outward from the repository, where temperatures are sufficiently low to allow liquid water to exist in pores and fractures. In this region, fundamental changes to hydrologic, mineralogic, or chemical conditions may take place, but these conditions do not interact directly with the waste packages. Rather, they interact with the near-field environment.

AML is expressed as metric tons of uranium emplaced per acre of drift (MTU/acre).

The spacing of waste packages along the emplacement drift axis, expressed as center-to-center spacing.

The tubes in which nuclear fuel pellets are incased.

Diameter of drift excavation, but does not account for ground support or other engineered structures.

Temperature difference between the WP and drift wall.

The extent to which the engineered barrier system contributes to isolation of radionuclides, either by containment or limits to release from the waste package. Although the EBS may also contribute to performance by physical/chemical inclusion of radionuclides into its components, this is not normally considered as part of isolation performance.

Determines how efficiently heat is transferred by thermal radiation between a WP and the drift floor, drift wall, or adjacent WPs.

LML is expressed as metric tons of uranium emplaced per meter of drift (MTU/m).

One of several possible alternative WP arrangements in a repository at YM, which lineally concentrates decay heat by placing WPs nearly end to end along drifts, i.e., tight axial spacing. For this report, the line-load design has the same AML (83.4 MTU/acre) as the ACD rev 00 design, although the AML could vary depending on drift spacing. Two line-load (LL) designs are considered:

 The 0.1-m spacing, 1.11-MTU/m line-load design, which assumes a 0.1-m gap between WPs, resulting in an LML of 1.11 MTU/m and drift spacing of 53.8 m. Near-field environment (NFE)

Percolation flux

Point-load design

Reference Design

Thermal conductivity of invert or backfill

Thermal conductivity of rock

Thermal conductivity of WP

Waste container

Waste form

Waste package

WP layout

 The 1.0-m spacing, 0.94-MTU/m line-load design, which assumes a 1.0-m gap between WPs, resulting in an LML of 0.94 MTU/m and drift spacing of 46.1 m.

Encompasses the region of rock and drift that is proximal to the waste packages and interacts with them; sometimes called the waste package environment (WPE). In general, the NFE or WPE is where temperatures reach above the boiling point and where resulting changes in environmental conditions dominate. These changes may be reversible.

Generally refers to the volume of water moving downward through unit cross-sectional area per unit time within the rock mass below the zone where evapotranspiration returns moisture to the atmosphere. For this report, the term refers specifically to downward water flux through the potential repository host rock (the TSw2 thermomechanical unit).

A repository design wherein waste is placed with sufficient spacing that environmental conditions of temperature and moisture do not reflect the influence of adjacent waste packages. To a large extent, the ACD rev 00 design is a point-load design.

See ACD rev 00 design.

The measure of how effectively heat is conducted through introduced invert or backfill materials.

The measure of how effectively heat is conducted through the repository rock.

The measure of how effectively heat is conducted through the waste package.

The container that encloses the waste form, any handling or shipping containers, fillers, support structures, etc. The term "container" refers to the container itself, not to the entire package including waste.

The type of radioactive waste material, e.g., spent fuel or defense wastes in reprocessed glass waste logs.

The complete package of waste and containers that includes the waste, any internal structural and criticality-control items, fillers, and the container.

The geometric configuration of waste emplacement. For this study, the layout consists of axial spacing along a drift between waste package (WP) centers and lateral spacing between drift centers (called drift spacing). Although other design options exist (e.g., two parallel rows of WPs in a drift or WPs placed off-center in a drift), this report is limited to layouts of WPs centered within drifts.

Near-Field and Altered Zone Environment Report Volume I: Technical Bases for EBS Design

Executive Summary

This report presents an updated summary of results for the waste package (WP) and engineered barrier system (EBS) evaluations, including materials testing, waste-form characterization, EBS performance assessments, and near-field environment (NFE) characterization. Materials testing, design criteria and concept development, and waste-form characterization all require an understanding of the environmental conditions that will interact with the WP and EBS. The *Near-Field Environment Report* (NFER) was identified in the *Waste Package Plan* (WPP) (Harrison-Giesler, 1991) as the formal means for transmitting and documenting this information.

The purpose of the new Near-Field and Altered-Zone Environment Report is to update the Preliminary NFER, to convey current understanding of environmental conditions, and to provide technical information needed for repository design decisions and performance analyses. Volume I provides the technical bases for design of the EBS, which is derived from our current knowledge of the waste-package and altered-zone environment. Volume I also discusses the altered zone (AZ) as it might impact performance analyses; however, Volume II is the principal source of information regarding the AZ. Volume II consists of 11 sections providing much more detailed discussion of the topics in Volume I as they pertain to particular fields of study.

The design and performance of the WP and EBS will depend on the hydrological, geochemical, and geomechanical conditions that develop over time in the rock extending for considerable distances into the mountain. For that reason, emphasis in this report is placed not only on conditions that will result and their associated parameter values, but also on processes that will change the environmental conditions. That is, it is necessary to understand not just the statistical distribution of environment conditions, but also the physical processes themselves.

The emphasis of the safety argument has changed over the years, both in response to changing understanding of the regulations as well as of the repository and natural system. When the *Site Characterization Plan* (SCP) was prepared, it was understood that the containers must contribute to satisfy requirements of the regulations by substantially complete containment of the waste for 300 to 1000 years. The waste form was required to release no more than 1 in 10⁵ of the 1000-year radionuclide specific inventory for up to 10,000 years. Because the container was assumed to perform for 1,000 years--as opposed to the waste form for 10,000 years--the stress on the NFE was to provide an environment that was compatible with waste from performance by keeping the pH close to neutral and avoiding high pH.

The current strategy has placed more reliance on the Engineered Barrier System, specifically the container. Current information indicates that geochemical processes that result from the coupling of heat and fluids generated from waste emplacement will form large volumes of sorptive minerals within the AZ. Furthermore, there will likely be a significant change in the hydrologic system. However, there are large uncertainties in our understanding of the natural system, and recent evidence suggests much higher percolation fluxes than originally anticipated. As a result, there has been a tendency to place greater emphasis on very robust WPs (e.g., ceramics) that will be able to last for times that exceed 10,000 years regardless of the water and water chemistry to which they are exposed.

The impact of the NFE on the evolving waste isolation strategy depends, to a large degree, on the portion of the waste isolation system being considered. The three main aspects of the system that will contribute to isolate waste from the accessible environment are:

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- The waste form itself.
- The container in which the waste is placed.
- The natural system (rock mass).

The waste form contributes to isolation of radionuclides mainly by limiting the rate at which the waste can be released. The NFE factors that influence the rate of release are the amount, temperature, and chemical content of water that is able to contact the waste.

Container performance depends on the container design and material selected, and on their interaction with the NFE, specifically the quantity, temperature, and chemical content of both liquid water and water vapor as well as drift gaseous makeup (i.e., presence or absence of oxygen and CO_2). The mode of contact of water (dripping, continuous contact, condensation) also is an important factor in the performance of the container, or more specifically, on corrosion kinetics.

The natural system contributes to isolation in three ways. First, it limits the amount of water that can contact the waste and WPs. Factors that limit the amount of water include the unsaturated rock mass and the tight rock matrix. Second, the natural system can prevent radionuclide migration through chemical exchange or sorptive processes with the rock minerals and pore water. Third, the natural system contributes to isolation by delay in transport, which will either allow shorter-lived radionuclides to decay or will allow for delayed releases and dilution processes to lower the dose rates. Of course, there are also processes within the natural system than can concentrate radionuclides by allowing them to accumulate over time.

The rate of radionuclide release (and dose rate) at the accessible environment depends on the time that release from the EBS begins, the magnitude and duration of the release, and decrease in magnitude (attenuation) of radionuclide concentration resulting from radioactive decay and dilution during the time required for radionuclides to move through the natural system from the repository to the accessible environment. A conceptual model for the release of radionuclides from the WPs and the EBS is summarized below.

Radionuclide release from the EBS requires two conditions:

A breach, or failure, of the metal barriers of the WP, including the outer portion of the container that is designed to consist of a thicker corrosion-allowance material, and the inner portion of the container designed to consist of a thinner corrosion-resistance material, as well as the cladding around the spent fuel pellets or pour ? canisters around the Defense High Level glass waste forms.

Contact of the waste form by liquid water, which can dissolve or suspend radionuclides and transport them through the failed WP and other elements of the EBS to the unexcavated rock host.

Postemplacement cladding failure will be controlled primarily by the temperature within the WP. Breach of the outer barrier is determined by it's corrosion rate, particularly the localized rate of attack or "pitting rate." Breach of the inner corrosion-resistant barrier will be controlled almost entirely by the internal environment until the outer barrier is breached. Even if a container fails, negligible amounts of radioactive material will be released from the repository unless liquid water is advected into the drift, into the breached WP, into the waste form, out of the WP, and into the repository horizon rock. Hence the timing of advected water entry, flux of liquid water, and it's chemistry will be the most important NFE variables for performance assessment after failure of the WPs.

The Performance Assessment implications of NFE conditions suggest that design goals should include the development of an EBS that, in concert with the natural system, will provide a long delay between emplacement and the first contact of waste by water. The overall design goal is to prevent water from contacting the waste form as long as possible and to limit the amount of waste contacted by water after contact begins. Survival of the WP can be greatly increased if the relative humidity at the WP surface is kept low. An overriding characteristic of the interaction between the WPs/repository and the NFE is that it results from coupled processes. The significance and magnitude of the coupling varies both spatially and temporally. To assist in the prioritization, the dominant

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processes were considered.

As stated, interactions between the EBS components and the NFE are, to large extent, dominated by the possible presence of liquid water and elevated humidity. Thus, an assessment of EBS performance can focus on hydrological aspects and associated geochemistry. Fracture flow is a key issue. We are now focusing almost entirely on the following processes as priorities:

- 1. Understanding how and when water will contact the WPs.
- 2. Determining the chemical composition of fluids and gases that contact the container or waste forms.
- 3. Understanding what happens in the AZ at distances that can extend more than hundreds of meters away from the repository.
- 4. Determining how geomechanics couples to the hydrology and geochemistry.

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Near-Field and Altered-Zone Environment Report Volume I: Technical Bases for EBS Design

Rev. 1

1.0 Overview

The United States Department of Energy (DOE) is investigating the suitability of Yucca Mountain as a potential site for the nation's first high-level nuclear waste repository. As shown in Fig. 1-1, the site is located about 120 km northwest of Las Vegas, Nevada, in an area of uninhabited desert.

Figure 1-1. Location of the Yucca Mountain site.

Favorable aspects of Yucca Mountain (YM) as a potential repository site include its arid nature and the sorptive properties of the rock materials. The arid environment results in unsaturated conditions at the potential emplacement horizon, which is the Topopah Spring Tuff (Tpt) of the Paintbrush Group. The major advantages of unsaturated conditions are that container corrosion, waste-form leaching, and radionuclide transport mechanisms are minimized because of the lack of contact between liquid water and the waste package (WP). Although recent studies indicate that percolation fluxes may be higher than originally anticipated, the major advantages should still apply, but their significance must be evaluated.

Lawrence Livermore National Laboratory (LLNL) is a Yucca Mountain Site Characterization Project (YMP) participant and is responsible for WP and engineered barrier system (EBS) evaluations, including materials testing, waste-form characterization, EBS performance assessments, and near-field environment (NFE) characterization. Materials testing, design criteria and concept development, and waste-form characterization all require an understanding of the environmental conditions that will interact with the WP and EBS. The *Near-Field Environment Report* (NFER) was identified in the *Waste Package Plan* (WPP) (Harrison-Giesler, 1991) as the formal means for transmitting and documenting this information.

1.1 Purpose and Scope of This Report

The first NFER was published in April, 1993 as a preliminary report (Wilder, 1993) because the design and characterization studies were not mature. Since publication of the Preliminary NFER, information regarding the environment of the potential repository site has progressed considerably. This new report is Revision 1 of the NFER. The term "Preliminary" has been dropped from the title to reflect the greater maturity of both design and characterization efforts. In addition, at the time the Preliminary NFER was published, a companion report was intended that would focus on the altered zone (AZ) environment—that is, the environment that develops away from the WPs and impacts the waste and its containers only in terms of its influences on the WP environment. Subsequently, a decision was made to combine both WP and AZ reports into the NFER. This decision recognizes the interdependence of the WP and AZ environments and the fact that the studies documented in the companion volume (Volume II of this report) apply to both of these zones of the environment.

The purpose of this new Near-Field and Altered-Zone Environment Report is to update the Preliminary NFER, to convey current understanding of environmental conditions, and to provide • وحدل



Figure 1-1.

technical information needed for repository design decisions and performance analyses. Volume I provides the technical bases for design of the EBS, which is derived from our current knowledge of the waste-package and altered-zone environment. Volume I also discusses the AZ as it might impact performance analyses; however, Volume II is the principal source of information regarding the AZ.

Volume II consists of 11 sections providing much more detailed discussion of the topics in Volume I as they pertain to particular fields of study. Volume II includes information on many different design options, property values, and sensitivity analyses. Volume I focuses on the ACD (reference) design and on project reference property values. To allow scientists and engineers to determine the validity of information in Volume I, Volume II documents the scientific and engineering data, models, and approaches used to gather information. Of particular relevance in Volume II are an extended and updated description of hydrothermal modeling (Sec. 1) and new discussions of the AZ environment (Sec. 10) and thermodynamic data determination of radionuclides (Sec. 11). Volume II documents the state of knowledge of relevant processes, conditions, and properties as of August 23, 1996, and, unless otherwise noted, it constitutes the basis for the material presented in this Revision 1 of Volume I.

1.1.1 Limitations

24 Evaluations of the expected environment are based on currently available understanding of properties, parameters, analyses, and assumptions, as well as reference designs, and on prototype 7.12 or scoping tests and analyses. In some cases, information from similar rock type or formations as well as from specific mineral or chemical conditions is used. Field tests needed to characterize the responses of YM and to test the model conceptualizations have not been completed at the time of this report. Although the information in this report is considered accurate vis-a-vis modeling approximations, the validation exercises have not been completed. Future assessments will be required to determine if the environment as described here is appropriate for conditions at YM.

1.1.2 Emphasis on Processes

The design and performance of the WP and EBS will depend on the hydrological, geochemical, and geomechanical conditions that develop over time in the rock extending for considerable distances into the mountain. For that reason, emphasis in this report is placed not only on conditions that will result and their associated parameter values, but also on processes that will change the environmental conditions. That is, it is necessary to understand not just the statistical distribution of environment conditions, but also the physical processes themselves.

A mechanistic understanding of processes and events as they unfold over time is essential for several reasons. For example, the degree of rock saturation is one important aspect of the repository environment. However, what the initial saturation happens to be is insufficient information for design and decisions. It is the amount, timing, and temporal distribution (steadystate or episodic) of water contacting the waste that are pertinent to design. To estimate the amount of water expected to contact waste at or over a given time, we must integrate saturation E. I makes data with other factors, including, (1) percolation flux, (2) fracture-matrix flow models, (3) fracture distribution, (4) emplacement configuration, (5) areal mass loading, (6) projected WP #2-ET Control thermal characteristics, and (7) repository design. Similarly, events as wide ranging as El. Also. ohento proprod prentime the impact of compart protons construction activities, microbial processes, climate change, and many others may modify repository conditions. For these reasons, the report focuses not only on detailed technical information but also on the interaction of processes and the integration of values.

1.1.3 Revisions

This Near-Field and Altered-Zone Environment Report is the first of the scheduled NFER revisions identified in the WPP. However, the WPP has been superseded by the Annual Planning Whith Y? Suf-

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process. Although the process does not have the same revision points as those specified in the WPP, it includes updating the NFER at major milestones, such as at completion of field testing. and at times needed for major decisions or documents, such as viability assessment and license application. This revision (Rev 1 of the NFER) will support the viability assessment and associated design and performance assessment. A second revision of the NFER will be based on the Large Block Test at Fran Ridge and the single heater test (SHT) in the Exploratory Studies Facility (ESF). The second revision will support the license application design (LAD) phase. A third revision based on ESF test results will support defense of the license application (LA) and the LA amendment to emplace waste. Additional reports are likely, and they will cover Tebe use for subsequent work related to the performance confirmation phase. Ared the spe before 6/92-6

1.1.4 Quality Assurance (QA) Controls

At the time of the Preliminary NFER, a distinction was made between quality-affecting work and work that was not quality-affecting. Subsequently, changes have been made in the QA program. All of the work or activities reported in this Near-Field and Altered-Zone Environment Report comes under the control of the Quality Assurance Requirements and Description (QARD; DOE, 1997). All data included in this report, except for referenced information from other sources . 12-1 not collected under YMP controls, were collected under controls of the QARD and can therefore be gualified for use. The NFER is not an activity that generates data itself, rather, the material presented in this report is a compilation/synthesis of data and other information that have been collected under other activities and reported elsewhere. The QA status of the data is determined by the activities and reports from which they were synthesized with the specific controls for those activities noted in Scientific Notebooks associated with those activities. For the most part, the data reported in Volume I was taken from Volume II, which identified the QA status of the data by sections in an appendix. For convience, this information is included in this volume (see Appendixes A and B), even though it is a duplication. No new data are transmitted in this report that were not already transmitted to YMP by Technical Data Information Form (TDIF).

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The NFER is a synthesis of information, where "synthesis" is used in the dictionary sense of being a combination of thesis (data, etc.) and logical deduction. It was also defined to be a synthesis report in the YMP sense, that is, a compilation of all relevant data regarding the NFE and AZ environments. It is not, however, a synthesis in the sense that is sometimes used in the YMP to mean a final, one-time compilation of all the data that will be used. Rather, as pointed out in Section 1.1.3, the NFER is revised periodically to reflect the information that will be developing. This revision (Revision 1) is based on a mixture of qualified and nongualified information that must be used as a whole to make the conclusions and assessments. Therefore, Revision 1 of the NFER must be considered to be <u>ponqualified</u> at this point. Subsequent revisions will be based on qualified data and will be qualified for use during licensing. If it is judged necessary, Revision 1 of the NFER will be qualified as a whole by the process of corroborating the data based on laboratory and field data, as feasible, when data are available and/or by peer review in accordance with Section 2.0, Quality Assurance Program.

1.1.5 Importance of NFE/AZ to Waste Isolation

The emphasis of the safety argument has changed over the years, both in response to changing understanding of the regulations as well as of the repository and natural system. When the Site Characterization Plan (SCP) was prepared, it was understood that the containers must contribute to satisfy requirements of the regulations by substantially complete containment of the waste for 300 to 1000 years. The waste form was required to release no more than 1 in 10^5 of the 1000-year radionuclide specific inventory for up to 10,000 years. Because the container was assumed to perform for 1,000 years-as opposed to the waste form for 10,000 years-the stress on the NFE was to provide an environment that was compatible with waste form performance by keeping the pH close to neutral and avoiding higher pH. This was the reason for the requirement

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in the SCP to limit the use of concrete or cementitious materials in the repository. There were also SCP requirements to prevent water contact by maintaining a capillary barrier (gap between WP and borehole rock) and limiting rock sloughing into waste-emplacement boreholes. The natural system was then relied on to provide a ground water travel time that exceeded 1000 years and also to retard the transport of the radionuclides by sorption and similar processes. The natural system was assumed to provide this function for times that extended up to 1,000,000 years under some regulatory interpretations. Because of the emphasis on long times and on retardation of radionuclides, it was deemed important to keep from destroying the sorptive minerals (zeolites). To protect the sorptive minerals as well as other components of both natural and engineered systems a series of constraints were placed on temperatures. In the case of sorptive minerals, these temperature limits are reflected in the SCP.

At the time of the Preliminary NFER, the greatest emphasis for isolation continued to be on the unperturbed natural system, although credit for lifetimes of the container beyond 1000 years was being considered, and thus concepts to maintain a very dry NFE (extended dryout) were being stressed. Because of the possibly conflicting goals of high temperatures to keep packages dry and lower temperatures to prevent dehydration or alteration of sorptive minerals (as well as thermal performance issues of other components of the repository/waste disposal system), a series of thermal goals was formalized. During this time, the requirement for pH control or limits on use of concrete was relaxed.

The current strategy has placed more reliance on the Engineered Barrier System, specifically the container. This emphasis is illustrated by several draft Waste Isolation Strategies. Although none have been formally accepted by the DOE, they place increased emphasis on the EBS with either greatly reduced emphasis on the natural system or at least equal emphasis on the EBS. The strategy is in transition, and the thermal goals have remained largely as they were during the time of the Preliminary NFER. Current information indicates that geochemical processes that result from the coupling of heat and fluids generated from waste emplacement will form large volumes of sorptive minerals within the AZ. Furthermore, there will likely be a significant change in the hydrologic system. However, there are large uncertainties in our understanding of the natural system, and recent evidence suggests much higher percolation fluxes than originally anticipated. As a result, there has been a tendency to place greater emphasis on very robust WPs (e.g., ceramics) that will be able to last for times that exceed 10,000 years regardless of the water and water chemistry to which they are exposed. The much greater emphasis on very robust containers occurred since publication of Volume II of this report. This evolving strategy has not fully addressed performance issues beyond the 10,000 year time frame. Because the strategy has not been officially adopted and is evolving, this report will not focus on it but recognizes and addresses the implications.

The impact of the NFE on the evolving waste isolation strategy depends, to a large degree, on the portion of the waste isolation system being considered. The three main aspects of the system that will contribute to isolate waste from the accessible environment are:

- The waste form itself.
- The container in which the waste is placed.
- The natural system (rock mass).

In addition, many engineering options (such as drip shield or backfill) can enhance the performance of these three components. However, the actual isolation of waste as required by regulations (e.g., 10 CFR 60) depends only on the functioning of these three components. For instance, a drip shield will not contain the waste or control the rate of release, except as it impacts water contact. Thus, a drip shield is more a determinant of the NFE that a component of the isolation system. Whereas other components can contribute to isolation, it is the author's

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contention that they can be covered as part of the NFE rather than the isolation system. Perhaps the distinction is not as clear for backfill because such material can provide performance similar to that of the natural system, although at a much smaller geometric scale. As discussed below, the NFE impacts the three major components listed above in different ways, but is a very important factor in the performance of each of them.

The waste form contributes to isolation of radionuclides mainly by limiting the rate at which the waste can be released. The NFE factors that influence the rate of release are the amount, temperature, and chemical content of water that is able to contact the waste. The properties of the waste that determine this rate are discussed in the *Waste Form Characteristics Report* (Stout and Leider, 1997). Note that this does not address the gaseous radionuclide releases (e.g., ¹⁴C), which are not expected to be the major releases. The NFE will be a major factor in the release rates from the waste form because the quantity, temperature, and chemistry of water contacting the waste are all parts of the NFE. Water chemistry contacting the waste will likely be dominated by introduced materials (including potentially significant amounts of corrosion products), and the quantity of water will certainly be influenced by the mobilization of water due to waste emplacement. Microbial activities as well as the presence of colloids from corrosion products and microbes can have a major impact on the mobilization of radionuclides from the waste form.

Container performance depends on the container design and materials selected, and on their interaction with the NFE, specifically the quantity, temperature, and chemical content of both liquid water and water vapor as well as drift gaseous makeup (i.e., presence or absence of oxygen and CO₂). The mode of contact of water (dripping, continuous contact, condensation) also is an important factor in the performance of the container, or more specifically, on corrosion kinetics. Once again, these factors are all components of the NFE, and the processes noted for the waste form will be important to the container performance, although the influence of corrosion products and colloids will not be as significant. Even if a ceramic waste container is selected, it will be impossible to guarantee 100% performance of the ceramic because of undetected manufacturing flaws, damage, and other issues. Thus, the NFE becomes critical, and given the trend toward increased reliance on the container, it will be the most critical portion of the environment.

The natural system contributes to isolation in three ways. First, it limits the amount of water that can contact the waste and WPs. Factors that limit the amount of water include the unsaturated rock mass and the tight rock matrix. Second, the natural system can prevent radionuclide migration through chemical exchange or sorptive processes with the rock minerals and pore water. Third, the natural system contributes to isolation by delay in transport, which will either allow shorter-lived radionuclides to decay or will allow for delayed releases and dilution processes to lower the dose rates. Of course, there are also processes within the natural system that can concentrate radionuclides by allowing them to accumulate over time.

Processes within the NFE/AZ can impact the water distribution, specifically the redistribution of water due to waste heat. Processes that occur in the AZ can also contribute to significant changes in porosity and permeability, which can have dramatic impacts on the water contact issues. These processes can also result in significant modifications of the mineralogy of the system, resulting in vastly different sorptive capacities of the system (not all of which are detrimental). In addition, if dryout occurs within the AZ, there is increased capacity for imbibition so that radionuclides in any water that is imbibed from the fractures into the pores would tend to be physically held by capillary forces for very long times. Thus, the AZ can have potentially positive impacts on performance.

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1.1.6 Importance to Performance Assessment

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The rate of radionuclide release (and dose rate) at the accessible environment depends on the time that release from the EBS begins, the magnitude and duration of the release, and decrease in magnitude (attenuation) of radionuclide concentration resulting from radioactive decay and dilution during the time required for radionuclides to move through the natural system from the repository to the accessible environment. A conceptual model for the release of radionuclides from the WPs and the EBS is summarized below.

Radionuclide release from the EBS requires two conditions:

- A breach, or failure, of the metal barriers of the WP, including the outer portion of the container that is designed to consist of a thicker corrosion-allowance material, and the inner portion of the container designed to consist of a thinner corrosion-resistance material, as well as the cladding around the spent fuel pellets or pour canisters around the Defense High Level glass waste forms.
- Contact of the waste form by liquid water, which can dissolve or suspend radionuclides and transport them through the failed WP and other elements of the EBS to the unexcavated rock, host. [Once again, this does not address release of gaseous radionuclides, such as ¹⁴C. The will be However, these radionuclides are not identified by the Total System Performance Analysis) draphword (TSPA) or Waste Isolation Strategies as major concerns and are not included in NFE discussions.]

Postemplacement cladding failure will be controlled primarily by the temperature within the WP. An additional factor is the amount of oxygen that is able to contact the cladding. If temperatures are high at the time O_2 contacts cladding, the cladding will oxidize rapidly. In addition, if temperatures are high and O_2 can contact the waste (penetrate the cladding) it will oxidize with significant volume increase thereby bursting the cladding. Furthermore, oxidized fuel is much more soluble than non-oxidized fuel. WP temperature depends on the rate of heat transfer between the WP and its surroundings. Heat transfer rate, in turn, depends on whether or not backfill is used, the spacing between drifts and between WPs, the thermal output of waste, and other factors. The heat transfer rate from WPs is not expected to be particularly sensitive to the presence within the repository of liquid water or its chemistry. The mode of heat transfer may be impacted, as well as distributed along the drift, but overall transfer will be fairly insensitive.

Breach of the outer barrier is determined by its corrosion rate, particularly the localized rate of attack or "pitting rate." TSPA95 analysis suggests that the depth of pits in the corrosionallowance material can be modeled statistically by a random pitting factor times the general corrosion depth. The general corrosion rate is a function of NFE conditions (humid air or aqueous surroundings). For humid air conditions, the corrosion rate is much more sensitive to relative humidity than to temperature. Some of the rate parameters also depend strongly on the chemical composition of solids on the WP surface, because the composition of salts can control the humidity at which a liquid film can form. These solids could be manufacturing residues or salts left by the evaporation of water.

Breach of the inner corrosion-resistant barrier will be controlled almost entirely by the internal environment until the outer barrier is breached. At that time, the NFE could affect corrosion. The materials being considered have extremely slow general corrosion rates under anticipated repository conditions. The highly localized forms of corrosion depend on the chemistry of water entering the drift and on microbiological attack. The main considerations are halide ions entering the drift, which can substantially increase pitting, and nutrients for microbes, which can accelerate corrosion by creating a more aggressive microenvironment.

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Even if a container fails, negligible amounts of radioactive material will be released from the repository unless liquid water is advected into the drift, into the breached WP, into the waste form, out of the WP, and into the repository horizon rock. Hence, the timing of advected water entry, flux of liquid water, and its chemistry will be the most important NFE variables for performance assessment after failure of WPs. All of these issues may be dominated by changes within the NFE from processes involving:

- Introduced materials, such as corrosion products, backfill, and altered drift liner materials.
- Mineralogical changes that can influence the hydrologic properties and water chemistry.
- Coupled thermal-mechanical-hydrological processes within the AZ, which may completely change the flow fields above and below the repository.

Such processes could alter the physical and geochemical processes in the NFE that control WP performance as well as the transport of radionuclides in both the unsaturated zone and in the saturated zone below the repository.

1.1.7 Importance of NFE Conditions to Design

The Performance Assessment implications of NFE conditions suggest that design goals should include the development of an EBS that, in concert with the natural system, will provide a long delay between emplacement and the first contact of waste by water. The overall design goal is to prevent water from contacting the waste form as long as possible and to limit the amount of waste contacted by water after contact begins. The NFE that develops above the repository may change the amount, chemistry, and mechanisms (location, timing, and duration) of water seeping into drifts and ultimately contacting the waste form. Calculated fluxes around the drifts during the first few hundred years, driven by decay heat, are in the range of hundreds of millimeters per year. Thus, depending on the design, percolation flux into the drifts from condensed water may far exceed that of meteoric water.

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Survival of the WP can be greatly increased if the relative humidity at the WP surface is kept low. Current calculations suggest that a thermal loading strategy can be used to reduce humid air corrosion during the thermal pulse to rates below those under ambient conditions. The duration and magnitudes of this reduction depend on many factors, including percolation flux.

The importance of NFE conditions to design can be seen by current discussions of design modifications. Recent results of site investigations indicate that the ambient infiltration flux could average as high as 5 to 10 mm per year, with large temporal and spatial variations above and below those averages. If these ambient fluxes were allowed to contact a significant fraction waste inventory, such contact could lead to releases larger than some of "standards. Hence, conceptual designs are being development of the second development of lifetimes, respectively.

1.2 Changes Since Publication of the Preliminary NFER

1.2.1 Definitions of Near-Field versus Altered-Zone Environments

The Preliminary NFER focused on environmental conditions that will have a direct impact on the WP container materials and on the waste form. From work that supported the Preliminary NFER, it was recognized that interactions between the environment and waste could potentially

cause important property changes that could extend for considerable distances into the rock mass or natural system. Although the same suite of processes exist in all zones of the perturbed environment, those that dominate the NFE or even the far-field environment (FFE) (essentially, the unperturbed environment) are not necessarily the same as those that dominate in the region between them. Therefore, an altered zone (AZ) was defined and is now the focus of considerable attention. As noted earlier, information on the AZ is included in this report to provide a complete description of all aspects of the perturbed environment.

This report defines the near-field environment (NFE) as encompassing the region of rock and drift that is proximal to the WPs and interacts with them. The NFE has sometimes been called the WP environment (WPE). In general, the NFE or WPE is where temperatures reach above the boiling point. This environment evolves over time in response to waste emplacement. Some of the resulting changes in environmental conditions may be reversible.

The AZ refers to those regions where the rock mass undergoes changes due to coupled processes again related to emplaced waste. These changes may also be reversible over long times but are typically long-lived or permanent. In general, the AZ is a region where temperatures are sufficiently low to allow liquid water to exist in pores and fractures, although it may include vapor-dominated regions as well. The AZ is the region where rates of geochemical interactions with the environmental parameters are a maximum. Conditions in the AZ do not interact directly with the waste packages but, rather, interact with the NFE. Furthermore, changes in the AZ are more significant than those in the FFE, where ambient conditions prevail. The distinction between the NFE and FFE is based on processes and resultant changes rather than on a single geometric scale. The AZ may extend considerably farther for some processes than for others, ranging from tens to hundreds of meters outward from the repository.

1.2.2 Emplacement and Repository Concepts

An important development since publication of the Preliminary NFER is that major changes in the disposal concept have taken place. Emplacement is no longer planned in open boreholes and thin-walled containers. Rather, emplacement is currently envisioned in large, multi-barrier WPs within emplacement drifts (horizontal tunnels).

Three repository design options are evaluated in this Near-Field and Altered-Zone Environment Report. The three designs are all based on an areal mass loading (AML) of 83.4 MTU/acre. The designs are:

- The Advanced Conceptual Rev 00 Design (ACD) (CRWMS M&O, 1996), as shown in Fig. 1-2(a). In the ACD design, the axial center-to-center spacing of WPs varies from 10.3 to 14.6 m, and the drift-to-drift spacing is 22.5 m. Because WP spacing is large in comparison to their length and is from 46 to 65% as large as the drift spacing, the WP heat tends to be point (or isolated) sources. The waste types are indicated in the figure.
- An alternative design has been suggested by LLNL in which the WPs are placed nearly endto-end, as shown in Fig. 1-2(b). Two versions of WP spacings are analyzed: one with 1.0-m, end-to-end WP spacing; and one with 0.1-m, end-to-end WP spacing (or gaps). This results in center-to-center spacings of 5.68 to 6.68 m. The drift spacing to maintain an AML of 83.4 MTU/acre is 53.8 m, Thus, the WP spacing is roughly 10 to 12% of the drift spacing, and the WP heat tends to be a line source (conduction between packages of different waste types evens out the differences in heat output). This alternative design is referred to as the line-load (LL) design.
- The 1997 modified design, as shown in Fig. 1-3 wherein a slightly larger drift spacing (28 m versus 22.5 m for the ACD design) is used, together with closer WP spacing (6.6 to 14.23 m versus 10.3 to 14.6 m for the ACD design). The end-to-end gap or spacing for this design is 1 to 8.5 m versus 5 to 10 m for the ACD design, so that there is somewhat less isolation of the

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heat output. In addition, the WP spacing is a slightly smaller percent of the drift spacing (29 to 51% versus 46 to 65% for ACD). Thus, the difference between drift and pillar thermal conditions is greater than that for the ACD design.

Figure 1-2. Plan view of the WP layout for (a) the ACD point-load design and (b) the LLNLproposed line-load (LL) design. Both designs result in an AML of 83.4 MTU/acre. (The waste types are described in more detail in Section 4.1.1 of this report.)

Figure 1-3. Plan view of the WP layout for the 1997 modified repository design. The AML is 85 MTU/acre, and drift spacing = 28 m.

The first two designs are the basis for analysis in Volume II. Because the 1997 modified design falls somewhere between the ACD and line-load cases in terms of WP spacing and ratio of WP to drift spacing, the ACD and line-load designs can be viewed as the bounding cases for designs currently under consideration. Where the modified design includes 1- to 2-m end-to-end spacing, the line-load analysis is more pertinent; where that spacing is closer to 8.5 m, the ACD analysis is more germane. Nevertheless, the ACD and line-load designs bound the conditions.

Whereas the ACD and line-load designs bound the average conditions, the actual distribution of waste types can impact specific conditions associated with individual WPs. The ACD and lineload designs use the same distribution of waste types; the only difference is the spacing of WPs and drifts. However, the 1997 modified design not only varies the spacing, but also the way the waste types are distributed. Therefore, the environmental conditions given in this report for individual waste types should not be used to necessarily bound the conditions that could be expected for individual waste containers under the modified design. The analysis of these environmental conditions will be included in future versions of this report, if the modified design remains pertinent.

1.2.3 Waste Age

At the time the Preliminary NFER was published, many of the analyses were based on the hottest or youngest waste that was considered likely to be emplaced. The age for that waste was approximately 10 years out of core. No detailed information on what the actual emplacement age will be is available (some young waste could be emplaced first, depending on how waste is delivered from utilities). However, the age of the overall inventory has increased, and 10-yr-old waste is probably an unrealistic value.

The ACD design specifies a mix of 10-, 26-, and 40-yr-old SNF, with 26-yr-old as the most prevalent. This represents the typical waste age discussed in Rev 1. As was the case at the time of the Preliminary NFER, several issues continue to make the age of SNF subject to great uncertainty. Current discussions are considering interim storage, which could result in considerably older waste at the time of emplacement. Waste age has the greatest impact on peak WP and rock temperatures and on the time that the peak occurs. In general, older waste results in a lower peak, which occurs at later times.

1.2.4 Prioritization of Variables and Parameters

Another distinction between the Preliminary NFER and this report is in the organization of information according to environmental priorities. The Preliminary NFER was organized by following a temporal sequence, from preemplacement activities and conditions to postemplacement conditions and performance implications. A primary purpose of that report was to inform the reader about the complexity of changing conditions and interactions over time, and it remains a valuable reference tool for that reason. However, the earlier report made no

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(a) ACD design

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attempt to prioritize processes that may change the environment or the most relevant environmental parameters themselves.

This report includes an attempt to rank order, or at least focus on, the factors or conditions that are important for design decisions and Performance Assessment. The purpose of prioritizing the most germane factors and focusing on a few design options is to facilitate design decisions and to center attention on which of the myriad variables may be the most important for overall repository design considerations and performance of the EBS.

A key consideration in the ranking is the interactions between the WPs/repository and the NFE. These interactions are not separable; that is, the WP/repository impacts the NFE, but the NFE also impacts performance of the WP/repository. For example, we know that areal mass loading (AML) defines the repository environment to a larger extent than does a host of other factors, ranging from drilling and ventilation to radiation field from WPs, although they are also relevant. Furthermore, we now know that the way WPs are configured in the repository has a far greater role in determining whether near-field thermal-hydrological (T-H) behavior is beneficial (versus deleterious) than the areal mass loading of the repository itself. Indeed, WP layout may be more consequential than variability or uncertainty in the natural system parameters. However, the processes within the NFE that are triggered by the waste loading can modify the thermal regime and locally impact the heat distribution so that it is more typical of other AMLs.

An overriding characteristic of the interaction between the WPs/repository and the NFE is that it results from coupled processes. The significance and magnitude of the coupling varies both spatially and temporally. To assist in the prioritization, the dominant processes were considered.

The forward and back coupling of processes makes the NFE an environment typified by interrelated interactions. This idea can be shown conceptually by using an interaction matrix of the type proposed by Hudson(1989). Figure 1-4 is an interaction matrix for the NFE, where the key components of the NFE are shown along the diagonal, and the interaction of processes is shown off diagonal. Rather than showing all components and all coupling, which can be very complex, it is the objective of this matrix to focus on the key parameters or components and on the key processes. The matrix reflects, in general, the organization of this report.

Figure 1-4. Simplified matrix showing coupled processes in the NFE.

Figure 1-5 shows two examples of forward and back-coupling that are typical of both NFE and AZ environments. The first example is the coupling between temperatures and moisture conditions. Temperature regimes are directly coupled to moisture conditions through processes of evaporation/dryout. Many more processes are discussed in greater detail in the hydrology section (Sec. 4 of this report), but in a simplistic view, the moisture conditions are determined by the forward coupling shown on the matrix as the short arrow (A) to the right. However, the changes to moisture conditions will back-couple to the temperatures by processes involving moisture/heat transfer. Vaporization of water and its condensation and return to the vaporization zone (heat pipe) can strongly alter the heat regime. This is shown by the short arrow (a) to the left of moisture, indicating back-coupling. Another example is the interaction of temperature and moisture conditions through rock-water interactions to change hydrologic properties (see the B-b loops in Fig. 1-5). This is a multiply coupled process: both temperature and moisture conditions couple to change hydrologic properties, which, in turn, back-couple to the temperature field in modification to convection, water flow pathways, and potential coefficient of conductivity changes, and also potentially to the moisture conditions by modifications of flow pathways.

The boxes shown here on the diagonal are generalized topics that identify the major subject areas discussed in this report. As Jiao and Hudson (1995) point out, a matrix such as this actually



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Temperature	• Vaporization • Mobilization	 Rock-water interactions Manmade material interactions (kinetics) 	Thermodyn. of vapor	 Dehydration Rock-water Interactions (kinetics) 	 Dehydration Rock-water Interactions Material (kinetics) 	Stress changes Coefficient impacts	 Microbe survival Metabolic processes
• Latent heat • Heat pipes	Moisture conditions	Volumetrics •Water • Rock • Materials			 Volumetrics -water -rock Contact modes 	Water contribution to crack growth	Survival
		Water, chemistry.	Vapor partiai pressure		 Reaction thermodyn. Solubility Precipitation Carbonation 		Biomass generation Nutrient supply Metabolic processes
			Relative humidity conditions	Vapor/rock interactions	Vapor/rock/ solid interactions		
Attered flow paths	Altered flow paths		Changed flow conditions	Hydrologic properties	Water contact	• Healing of fractures • Changes water avail. for cracking	
Thermal conductivity changes	Water release/ consumpt. in reactions	• Altered minerals • Rock-water interactions	Partial press. (salts)	 Fracture/pore plugging Fracture healing Colloid transport 	Mineral and solid material (changes)	Volume changes (phase transform and mineral changes)	
		· · ·		Closing/ opening fractures Microcracks Coalescence flow path	Stress Impacts Mineral phase Microcracks (new surfaces)	Mechanical Conditions (changes)	
N/A	Biomass water consumpt/ release	Biomass micro- chemistry	Biomass control of T/RH relations	Biomass plugging	Biomass dissolution/ mineralization	• Biomass microcracking • Volume changes	Microblotic conditions

Figure 1-4



represents one part of a hierarchy of interaction matrices. In this report, the general matrix is either expanded as details are required, or a lower-level (more refined) matrix is used to explain the concepts and processes germane to a specific topic.

Figure 1-5. Forward and back coupling processes in the NFE.

1.2.4.1 Dominant Processes in the NFE. In the NFE, the dominant processes can be grouped into two categories: those occurring during the heating phase and those occurring after the waste cools.

During the heating phase, the processes will include:

- Relative humidity reductions due to strong thermal gradients.
- Removal of water from the system, slowing or shutting off many of the processes and resulting in weaker coupling between the geochemistry of system components.
- Thermal expansion under spatially varying thermal gradients, which will change rock and rock-support-system stresses.
- Potential microcracking of both rock and rock-support and invert systems.
- Potential precipitation of salts by water removal from vaporization.
- Dormancy or sporation of microbial colonies.
- Potential carbonation, depending on CO₂ availability, of concrete portions of the EBS.
- Displacement of O₂ by water vapor.

The rates and significance of coupling will depend on the balance between two factors that have opposite effects. At elevated temperature, kinetics are increased, which will increase the rates and significance of coupling with introduced materials and microbial communities. However, when temperatures approach or reach the boiling point, evaporation and boiling will remove water from the system. Without water, microbial processes become minimal or are eliminated, many geochemical processes are minimized or eliminated (except vapor processes, which can be increased), and aqueous corrosion or interactions with man-made materials are shut down. If water removal is significant, many of the coupled thermal-hydrologicalgeomechanical-geochemical processes are eliminated.

After cooldown, water will return to the NFE through the fractured, densely welded host rock mainly by relatively rapid fracture flow and by very slow movement of moisture and resaturation through the matrix. The amount of water that can return through fractures will depend on how much of the condensate from water driven out of the NFE during the thermal phase can be stored in fractures and on percolation flux at the time.

Once water returns, the NFE will be dominated by strongly coupled processes. Depending on how effectively the thermal field prevents water return during heating and on how far the water removal extends, the coupling will be re-established, but with lower kinetics because of lower temperatures. The NFE will likely be strongly influenced by the effects of processes that took place during the thermal phase. Water chemistry will mainly depend on interactions with introduced materials, precipitated salts left by the water that was removed, and, to some extent, on changes in mineralogy of the NFE rock. Interactions with introduced materials, particularly concretes, will depend on modifications to the concretes that resulted during the thermal pulse. Some effects, such as precipitated salts, may decrease with time. Because the time frames considered can be lengthy (hundreds to thousands of years), the significance of these processes must be based on the specific time of concern.

1.2.4.2 Dominant Processes in the AZ Environment. In the AZ, hydrological processes will be dominated by increased water availability and saturation associated with vapor condensation at elevated temperatures. Geochemical processes will be dominated by rock-fluid interactions,



A Temperature	• Vaporization • Mobilization	 Rock-water interactions Manmade material interactions (kinetics) 	B Thermodyn. of vapor	Dehydration Rock-water Interactions (kinetics)	 Dehydration Rock-water Interactions Material (kinetics) 	• Stress changes • Coefficient Impacts	 Microbe survival Metabolic processes
• Latent heat • Heat pipes	Molsture conditions	Volumetrics • Water • Rock • Materials	B		Volumetrics -water -rock Contact modes	Water contribution to crack growth	Survival
		Water chemistry	Vapor partial pressure		 Reaction thermodyn. Solubility Precipitation Carbonation 		Biomass generation Nutrient supply Metabolic processes
b	F D		Relative humidity conditions	Vapor/rock Interactions	Vapor/rock/ solid Interactions		
Altered flow paths	Attered flow paths		Changed flow conditions	-Hydrologic properties	Water contact	 Healing of fractures Changes water avail. for cracking 	
Thermal conductivity changes	Water release/ consumpt. in reactions	Altered minerals Rock-water interactions	Partial press. (salts)	 Fracture/pore plugging Fracture healing Colloid transport 	Mineral and solid material (changes)	Volume changes (phase transform and mineral changes)	
				 Closing/ opening fractures Microcracks Coalescence flow path 	Stress impacts Mineral phase Microcracks (new surfaces)	Mechanical Conditions (changes)	
N/A	Biomass water consumpt/ release	Biomass micro- chemistry	Biomass control of T/RH relations	Biomass plugging	Biomass dissolution/ mineralization	• Biomass microcracking • Volume changes	Microbiotic conditions

Figure 1-5

6

vapor contact at elevated temperatures, and reactive transport. The AZ will be less transient or dynamic than the WPE in that residence times for water will be much longer. Because of the durations, the processes will tend to go toward completion, and the changes will tend to be more dramatic.

Many of the geochemical processes in the regions of the NFE and AZ will be the same (for example, recrystallization, dissolution and precipitation, cation exchange, sorption, etc.). However, their impact on rock properties will be different because of contrasts in heating and cooling rate and in the abundance and temperature of liquid water. The difference will be in the magnitude and nature of chemical and mineralogical changes, and in the duration of specific processes. As a result, performance of the overall repository system will be affected differently in the two regions because of mineral-water interactions. In addition to mineral-water interactions, there will be a significant difference due to introduced material interactions. This is true not only because the introduced materials are located mainly in the drifts or NFE but also because of the higher temperatures, dynamic or transient water conditions, and the very large volume percent of introduced materials that occurs in the NFE as opposed to the AZ. Thus, introduced materials will be a significant, if not dominant, factor in determining the NFE, whereas mineral-water interactions will be one of the principal factors that determines the AZ environment.

1.2.4.3 Summary of Prioritization. Although the dominant processes vary for the NFE and AZ, evaluations have shown the overriding importance of hydrology and geochemistry, and the strong influence of heat on those processes. Thus, we can begin to prioritize the most essential issues related to EBS design and performance. It is logical to break down the parameters and processes into two lists. Table 1-1 identifies the most important NFE and AZ parameters, that is, the environmental parameters shown to have the greatest effects on the EBS and WPs. Table 1-2 identifies, in approximate order, the factors shown to have the greatest influence on the NFE and AZ. Subsequent sections of this report discuss these parameters in more detail and some of the complex processes and interactions among them.

Table 1-1. Near-field and altered-zone environmental parameters.

I. Parameters shown to have the greatest effects on the EBS and WP Amount of water contacting the WPs Water chemistry

Relative humidity

II. Parameters expected to have moderate or variable effects on the EBS and WP Native microbial species

III. Parameters expected to have minimal (little or controllable) effects on the EBS and WP. (These factors could have secondary effects through coupling with more significant parameters.)

Rock fall physically damaging metal waste containers (different priority if ceramic) Geomechanics Table 1-2. Factors that influence the near-field and altered-zone environment.

M/D lourest	greatest effects on the INTE and AZ
WP layout	•
WP heat output	
Areal mass loading (AMI	L)
Backfill	
Introduced (man-made) r	materials
Percolation flux	· ·
Microbiological effects	
Canada a stinition in	cluding
 • drilling • excavation • ventilation Radiation field from intaction 	t WPs

1.2.5 Single versus Multiple Repository Design Efforts

Repository design must not only consider what the NFE is, but once design decisions are made, it must also consider how those decisions will affect the NFE. This is an iterative process. It involves a determination of the design that will function within the given environment, a consideration of the environment that will then develop as a result of that design, and revision of the design, as necessary, for the newly developed environment.

An alternative to such an iterative approach is to assume different designs and then to assess the NFE for those design alternatives. This is akin to a sensitivity study—which has the advantage of a shorter schedule because the second assessment is eliminated—and it is the approach taken in this report. However, a disadvantage is that many different cases must be assessed and reported. The Preliminary NFER included sensitivity analyses; however, it was more focused on the iterative approach. In contrast, Rev. 1 evaluates multiple designs and sensitivity. To minimize the cumbersome nature of reporting all cases that have been considered, Volume I focuses on a few design assumptions that represent the most likely ones, or reference designs (ones that bound the 1997 modification). We refer the reader to Volume II for information on the many other design options that were evaluated.

1.3 Modeling Advances

Whereas considerable progress has been made since publication of the Preliminary NFER in the characterization of YM, particularly in the availability of the underground drifts and ESF facilities for sampling and mapping, perhaps even more progress has been made in the conceptual understanding and models of coupling and in the computer codes available to assist in the analyses of those conceptualizations. Of course, there have also been significant improvements in the capabilities of computers to handle the computationally intensive requirements of these models and codes. This report is not focused on discussions of those advances. The reader can find more detailed discussion in Hardin and Chesnut (1997) and in Glassley (1997).

Of course it needs to be understood that although there have been advances in models, they are nonetheless simplifications of the real world. The material in this report relies heavily on numerical models of an idealized system with simplified processes and properties and should not be construed as actual, detailed NFE and AZ conditions unless verified by testing etc. This does not mean that the analyses are useless, nor does it mean that they should not be used in making decisions. As discussed in the LBT status report (Wilder et al., 1997), it will be impossible to measure these processes and properties directly for the time scales required, and therefore models with their inherent assumptions will always be the basis for understanding the evolving environment and for making decisions. It will be imperative that careful programs be conducted that are design to test these models. While some of the uncertainties will be reduced with time, the same caveat—that the models are idealized and simplified abstractions of the real world will always apply.

1.4 Focus of Current Work

As stated, interactions between the EBS components and the NFE are, to a large extent, dominated by the possible presence of liquid water and elevated humidity. For example, leaching of most of the radionuclides from the waste forms is almost exclusively aqueous, and corrosion of most candidate container materials is either dominated by liquid water contact or by relative humidity that is sufficiently high to allow water films to form on the containers. The transport of radionuclides through the rock mass depends exclusively on aqueous mechanisms (except for gaseous radionuclides associated with spent fuel and those generated slowly by radioactive decay from the glass waste form). Thus, an assessment of EBS performance can focus on hydrological aspects and associated geochemistry. Fracture flow is a key issue. As noted in the Preliminary NFER, "if radionuclide transport does occur, it will most likely involve fracture flow" (Wilder, 1993).

We are now focusing almost entirely on the following processes as priorities:

- 1. Understanding how and when water will contact the WPs, especially during the first several thousand years following waste emplacement. This work includes determining the amount of fluids, as well as vapor or gases, that contact the container or waste forms, and optimizing ways to keep water away from the WPs.
- 2. Determining the chemical composition of fluids and gases that contact the container or waste forms. Introduced man-made materials, such as the containers, invert material, cement grout, concrete, and rock-support system, will be a dominant factor in water chemistry. The chemistry of water contacting the WPs will also depend on whether water drains quickly down fractures or migrates slowly in the matrix and on process such as refluxing.
- 3. Understanding what happens in the AZ at distances that can extend more than hundreds of meters away from the repository. This focus arises because conditions in the AZ set the initial water chemistry as well as changes in hydrological properties.
- 4. Determining how geomechanics couples to the hydrology and geochemistry. Issues include:
 - Microfracturing to create new flow paths and reactive surfaces.
 - Fracture closing or changed hydrologic pathways.
 - Altered pore spaces.

2.0 Ambient Yucca Mountain Environment

Yucca Mountain (YM) consists of a layered sequence of variably fractured, nonwelded to densely welded tuff units with an eastward tilt of about 5 to 30° (Montazer and Wilson, 1984). At YM, the thickness of the unsaturated zone varies from 500 to 750 m. The potential repository is located within this unsaturated zone, lying approximately 350 m below the ground surface and 225 m above the water table (Klavetter and Peters, 1988). Montazer and Wilson (1984) also report the absence of perennial streams at YM. Therefore, recharge due to rainfall occurs episodically. Flint (1991) reports that the mean annual precipitation at YM varies (areally) from 150 to 240 mm/yr.

At the time of preparation of the Preliminary NFER in 1993, it was reported that overall infiltration was expected to be very low. Percolation flux estimates based on equivalent continuum model (ECM) analysis for the repository horizon ranged from 0.05 to 1.0 mm/yr, possibly upward due to vapor transport from the underlying saturated zone (see Vol. II, Sec. 1.0) (Montazer and Wilson, 1984). More recent estimates based on infiltration studies, ³²Cl, calcite, and other evidence, have been revised upward to a range from 1.0 to 15 mm/yr (Bodvarsson and Bandurraga, 1996). It is judged that 5 mm/yr is the most likely value. However this revision of Volume I, focuses on 0.3 mm/yr to be consistent with values used in Volume II.

The geologic unit being considered as the host rock is in the Topopah Spring Tuff (Tpt) of the Paintbrush Group. The host rock is a densely welded, fractured, devitrified ash-flow tuff. The Preliminary NFER reported estimates of fracture density ranging from 20 to 42 fractures per cubic meter based on information available at that time, which was obtained from surface outcrops and from boreholes (MacIntyre et al., 1990; Scott and Castellanos, 1984; Dudley et al., 1990; Wilder, 1990). These data did not distinguish fracture spacing along any direction, although attempts to estimate fracture spacing in terms of fractures per meter were representative (see Volume 1, Rev 0). Since the Preliminary NFER, the main drift of the Exploratory Studies Facility (ESF) has been excavated and mapped. Results of mapping (Albin et al., 1997) more than 10,000 fractures within the middle nonlithophysal and lower lithophysal zones (upper repository horizon strata) indicate that there are four sets of fractures. The first set is oriented at 120°/80° (strike/dip). The second set is oriented 220°/80°. A third set is oriented approximately 310°/30°. A fourth set was identified that has strike similar to sets 1 and 3, ranging between 270° and 330° (90° to 150°), and dips intermediate between these sets at 40° to 60°.

The relative prominence of the fracture sets depends on location within the drift; however Set 1 is always, by far, the most prominent (from 47 to 74% of total fractures). Set 2 is the next most dominant of those sets that are associated with a particular orientation (as opposed to random) and varies from 5 to 22% of the fracture population. These two sets comprise 61 to 79% of the fractures observed in the drift, as is clearly shown by the azimuth distribution histogram (Fig. 2-1). A correction for orientation bias from drift orientation was made to evaluate Set 2. However, in the author's view, the drift and fracture Set 2 orientations were too similar (only 17° from the edge of the distribution that starts at 200°), and the correction can be misleading (Yow and Wilder, 1983). Albin et al. (1997) recognized this problem and used an averaging method to moderate the effect. Albin et. al. (1997) interpreted four fracture domains within the drift. These domains occur in separate intervals of the drift, as shown in Fig. 2-2. As can be observed, the fracture frequency average—except in Domain 3, which was defined as a fracture zone—is about 2.5 fractures per meter.

Figure 2-1. Distribution of fractures by azimuth in the main drift. Modified from Albin et al., (1997).







Figure 2-2. Fracture-density histogram plots of the number of fractures, cooling joints, and vapor-phase partings per meter in each 2-meter interval along the main drift. Modified from Albin et al., (1997).

One of the critical NFE aspects of the fractures is the way the fractures are distributed. Set 1 is dominant, and its spacing becomes critical, as will be discussed in Section 4 of this report. The average spacing of Set 1 is 0.89 fractures per meter for domain 1, 1.1 fractures per meter for domain 2, 3.1 fractures per meter for domain 3, and 1.2 fractures per meter for domain 4. Fracture aperture estimates range from 43 to 127 μ m (Buscheck, 1990). As discussed in greater detail in Volume II, this fracturing has a major impact on the NFE.

2.1 Geologic Framework

The geologic framework was originally described in Chapter 1, Volume 1 of the Site Characterization Plan (SCP; DOE, 1988). The SCP pointed out that Yucca Mountain is in the southern part of the Great Basin, the northernmost subprovince of the Basin and Range physiographic province. The overall structural features that characterize the Basin and Range province are more or less regularly spaced, subparallel mountain ranges and intervening intermontane sedimentary basins that are formed by extensional faulting.

As noted in ISM2.0 (CRWMS M&O, 1997), YM consists of several outflow volcanic sheets, the thickest of which are units within the Paintbrush Canyon Group, which includes the Topopah Spring Tuff (Tpt), which is the repository horizon unit. Faulting at YM is typical of Basin and Range physiographic province. The repository block as identified in the SCP is roughly defined by Solitario Canyon and Ghost Dance faults along the east and west boundaries, respectively, with possible fault bounding on the north. Subsequent investigations have refined this understanding, but these refinements do not make significant changes in the impacts of the overall geologic structure on the NFE.

The stratigraphy of the YM area, as defined in the SCP, has been refined over the years and subdivided into units that are useful in performance analyses. The Tpt is a formal stratigraphic unit and has been subdivided into more specific thermal/mechanical units, which are particularly useful for hydrologic analyses. Both the formal and thermal/mechanical stratigraphic columns are shown in Fig. 2-3. A complete listing of geologic units as reported in the ISM2.0 (CRWMS M&O, 1997) is shown in Table 2-1. For the remainder of this report, the thermal/mechanical unit designation is used. Figure 2-4 shows an east-west cross section through the repository site. The potential repository horizon is in the thermal/mechanical TSw2 unit (Ortiz et al., 1985). This unit is a devitrified, welded, rhyolitic tuff that is moderately to densely welded tuff. It appears to consist of a mass of intact blocks separated by ubiquitous planar cooling fractures, which tend to be strata-bound, and by discrete tectonic fractures, which may not be planar but tend to be more extensive. Price, Connolly, and Keil (1987) report that the majority of the rock is a fine-grained matrix containing gray-colored regions of vapor-phasealtered material that vary in size and are quite common. In addition to these main components, the rock contains small (open and closed) lithophysal and "healed" fractures filled with quartz or calcite.

A detailed, 3D, site-scale hydrological model has also been prepared by Bodvardsson and Bandurraga (1996). This model provides detailed descriptions of the hydrologic units and their properties.

Figure 2-3. General stratigraphic column at Yucca Mountain. Modified from Elayer (1997).

Figure 2-4. East-west cross section through the repository siting volume. Modified from Elayer (1997).



Figure 2-2.














		1
Modeled units	Stratigraphy (Buesch et al., 1996; Sawyer et	Thermal / mechanical units
(unit names are in column 2	al., 1994)	(Ortiz et al., 1985)
alluvium	NA	UO
Rainier Mesa Tuff	Trm	UO
rbylite of Comb Peak	Tpk	N/A
Tiva Canyon Tuff undifferentiated	Tpc .	TCw
crystal-poor densely welded vitric subzone of Tiva Canyon Tuff	Трсру3	TCw
crystal-poor nonpartly welded vitric subzone of Tiva Canyon Tuff	Tpcpv1-2	PTn
Pre-Tiva Canyon Tuff bedded tuff	Tpbt4	PTn
Yucca Mountain Tuff	Тру	PTn
pre-Yucca Mountain Tuff bedded tuff	Tpbt3	PTn
Pab Canyon Tuff	Трр	PTn
nre-Pab Canyon Tuff bedded tuff	Tpbt2	PTn
Toponah Spring Tuff upper nonpartly-welded vitric subzones	Tptrv2-3	PTn
Topopal Spring Tuff upper densely welded vitric subzone	Tptrv1	TSw1
Topopan Spring Tuff xl-rich nonlithophysal zone	Tptrn	TSw1
Topopat Spring Tuff xI-tich lithophysal zone	Tptrl	TSw1
Topopan Spring Tuff lithic rich member	Tptf	TSw1
Toponah Spring Tuff upper lithophysal zone	Tptpul	TSw1
Toponah Spring Tuff middle nonlithophysal zone	Tptpmn	TSw2*
Toponah Spring Tuff lower lithophysal zone	Tptll	TSw2
Topopah Spring Tuff lower nonlithophysal zone	Tptpln	TSw2
Topopal Spring Tuff lower densely welded vitric subzone	Tptpv3	TSw3
Topopah Spring Tuff poppartly welded vitric subzones	Tptpv1-2	CHn1
nre-Toponab Spring Tuff bedded tuff	Tpbt1	CHn1
Calico Hills Formation undifferentiated	Tac	CHn1
pre-Calico Hills Formation bedded tuff	Tacbt	CHn2
Prow Pass Tuff upper nonwelded zone	Tcp [unw]	CHn3
Prow Pass Tuff welded zone	Tcp [w]	PPw
Prow Pass Tuff lower nonwelded zone	Tcp [lnw]	CFUn
pre-Prow Pass Tuff bedded tuff	[bt]	CFUn
Bullfrog Tuff upper nonwelded zone	Tcb [unw]	CFUn
Bullfrog Tuff welded zone	Tcb [w]	BFw
Bullfrog Tuff lower nonwelded zone	Tcb [Inw]	CFMn1
pre-Bullfrog Tuff bedded tuff	[bt]	CFMn2
Tram Tuff undifferentiated	Tct	CfFMn2, TRW
Pre-Tram Tuff bedded tuff	[bt]	N/A
lower Tertiary unit undifferentiated	N/A	N/A
Paleozoic and older units	N/A [pz]	N/A

Table 2-1. Geologic units as reported in ISM2.0 (CRWMS M&O, 1997).

* Top of TSw2 as used in repository design is above the top of Tptpmn (Elayer, 1995).

2.2 Geochemistry

Petrographic analysis shows that the repository horizon tuff consists of primary minerals, such as sanidine, plagioclase, quartz, biotite, iron-titanium oxides, allanite and zircon, that formed at temperatures in excess of 600°C in a magma chamber prior to eruption of the tuff. It also contains secondary minerals, such as cristobalite, quartz, alkali feldspars, and smectite clays, that formed during cooling and later alteration of the tuff at temperatures less than 500°C. The chemical composition of the rock is shown in Table 2-2. Because these values were determined from cores, they may not be representative of the composition of mineral coatings along fractures.

Distinguished from the primary and secondary minerals are the minerals that formed on fracture surfaces [e.g., calcite, smectite, quartz, cristobalite, alkali feldspar, and the zeolites mordenite, chabazite, clinoptilolite, and heulandite (Carlos, 1985, 1989; Lin and Daily, 1989)]. These minerals form layers that possess physical properties distinct from those minerals in the rock matrix and are estimated to be a maximum of 5% of the total rock mass. The mineral assemblage along a fracture surface varies from one location to another, and appears to reflect, at least in part, the previous location of the saturated zone (Carlos, 1989). Based on x-ray diffraction (XRD) analyses, the mineral assemblage of a typical fracture surface within the repository horizon might consist of 7% smectite, 7% cristobalite, 12% quartz, 29% clinoptilolite, and 45% alkali feldspar (Carlos, 1985).

See Glassley (1986) for a more detailed discussion and a description of the mineralogic and chemical properties of the preemplacement environment.

Constituent	60	61	62	Average	Standard deviation
SiO ₂	78.4	78.9	78.9	78.73	0.24
Al ₂ O ₃	12.0	12.3	12.2	12.17	0.12
Fe ₂ O ₃	1.016	0.973	1.000	0.996	0.018
CaO	0.492	0.451	0.480	0.474	0.017
MgO	0.1217	0.1281	0.1126 .	0.123	0.007
TiO ₂	0.1108	0.0927	0.0984	0.101	0.008
Na ₂ O	4.07	3.92	4.25	4.08	0.13
K ₂ O	3.71	3.18	2.94	3.28	0.32
P ₂ O ₅	0.01	0.01	0.03	0.02	0.01
MnO	0.0624	0.0455	0.0488	0.052	0.007

Table 2-2. Percentages of major constituents in Topopah Spring Tuff (core samples 60, 61, and 62 from drill hole USW G-3). Fe₂O₃ represents total iron. Data from Schuraytz et al. (1986).

2.3 Hydrology

Because of the small pore size and the unsaturated conditions, a significant suction potential (matrix potential, capillary force, or suction) develops in the rock. Recent measurements of saturation indicate higher saturations than originally estimated. Samples from the ESF Single Heater Test indicate saturation ranges from 80.5 to 99.0%, which compare with saturation values of $85 \pm 12\%$ for the same interval from surface boreholes (CRWMS M&O, 1996). LLNL evaluation of the Lawrence Berkeley Laboratory site scale model (Bodvarsson and Bandurraga, 1996) yields saturations ranging from 88.6 to 96%. Thus, the suction potential may be less than previously anticipated.

As noted in the Preliminary NFER, LLNL performed suction potential tests at room temperature and at elevated temperatures. The results are shown in Figs. 2-5(a) and 2-6. Tests

performed more recently have filled in the data gaps. Based on these recent results, shown in Fig. 2-5(b), the matric potential at the best current estimate for saturation (i.e., about 90%) is not greatly different than that for the assumed 65% saturation; both values are less than 100 bars (10 MPa).

Figure 2-5. (a) Suction potential as a function of water saturation at 20°C for Topopah Spring Tuff. (b) Matric potential as a function of saturation from the Single Heater Test. Drying refers to the portion of the test cycle when moisture was being removed. Wetting refers to the portion of the test cycle when moisture was being added. Because suction potential cannot be measured at 100% saturation, this value is designated as a theoretical calculation.

Figure 2-6. Suction potential as a function of water saturation at 70°C for Topopah Spring Tuff. Because suction potential cannot be measured at 100% saturation, this value is designated as a theoretical calculation.

Current data are significantly different from data reported in the Preliminary NFER. Current data show the suction potential >500 bar up to 20 to 25% saturation compared to about 10% saturation in the Preliminary NFER. This result probably reflects variability of rock properties. The current tests were performed on samples from the ESF, whereas earlier tests relied on samples from surface boreholes. It was noted in the Preliminary NFER that considerable variability was shown among these samples, implying a heterogeneity of suction properties in the Topopah Spring Tuff. The earlier data showed approximately a factor of 3 difference in suction potential between samples 1 and 3 at 15% saturation. Nevertheless, the room-temperature drying data of the earlier measurements were within the spread of corresponding psychrometer results of Klavetter and Peters (1987). The current data have not been analyzed sufficiently to determine how they compare with results from other tests.

The saturated water permeability of an intact Topopah Spring Tuff sample at room temperature measured by LLNL is about 0.3 μ D, although measurements by others (Moore et al., 1986) at room temperature ranged from 0.85 to 64 μ D. The saturated water permeability of fractured Topopah Spring Tuff is at least 3 orders of magnitude greater than that of intact tuff. At room temperature, cylindrical tuff samples containing a single natural hairline through-fracture have apparent water permeabilities ranging from 0.85 to 13 mD (Daily et al., 1987). The water permeability of an intact Topopah Spring Tuff sample is virtually independent of temperature, time, and the dehydration and rehydration processes.

The matrix properties of the hydrostratigraphic units at Yucca Mountain are summarized in Table 2-3 (Klavetter and Peters, 1988). Matrix property values are continuing to evolve (Bodvarsson and Bandurraga, 1996; Flint, 1997). The units generally fall into two categories: (1) welded tuffs of very low permeability and low porosity (TCw, TSw1, TSw2, and TSw3), and (2) nonwelded vitric tuffs of high permeability and high porosity (PTn and CHnv). The zeolitized nonwelded CHnz has low permeability and intermediate porosity, and the welded PPw has intermediate permeability and porosity. The permeabilities of the nonwelded vitric tuffs are 4 to 5 orders of magnitude greater than those of the welded tuffs and the zeolitized nonwelded CHnz.

Physical properties including porosity, grain density, bulk density at in situ saturation, and dry bulk density have been determined from core samples, and values of these parameters are summarized in Table 2-4. Additional discussion can be found in Vol. II, Sec. 4.0.

On the basis of samples taken from surface coring, the TSw2 was expected to have 65% saturation (±19%) (although, as noted, measurements on ESF samples show saturations of approximately 90%) and a porosity of 11% (DOE, 1990; Klavetter and Peters, 1986). Air in the unfilled voids is expected to be moist, although not always at 100% humidity. For further discussion, see Vol. II, Sec. 1.0.



Figure 2-5(a & b)

Describe | |









Tinita	Sample	Porosity	Permeability	<u>s</u> b	0 ^c (10 ⁻² m ⁻¹)	Rc
TCw	G4-1	0.08	9.7 × 10-19	0.002	0.821	1.558
PTn	GU3-7	0.40	3.9 × 10-14	0.100	1.500	6.872
TSw1	G4-6	0.11	1.9 × 10 ⁻¹⁸	0.080	0.567	1.798
TSw2	G4-6	0.11	1.9 × 10 ⁻¹⁸	0.080	0.567	1.798
TSw3	GU3-11	0.07	1.5×10 ⁻¹⁹	0.080	0.411	2.058
CHnv	GU3-14	0.46	2.7×10^{-14} .	0.041	. 1.60	3.872
CHnz	G4-11	0.28	2.0×10^{-18}	0.110	0.308	1.602
PPw	G4-18	0.24	4.5×10^{-16}	0.066	1.41	2.639

Table 2-3. Matrix properties of Yucca Mountain tuff (Klavetter and Peters, 1988).

*Unit designations: TC = Tiva Canyon, PT = Paintbrush, TS = Topopah Spring, CH = Calico Hills, PP = Prow Pass, w = welded, n = nonwelded, v = vitric, z = zeolitized.

^bS_r is the residual liquid saturation (not in situ saturation).

 α and β are fitting parameters for capillary pressure and relative permeability curves, respectively.

Table 2-4. Physical and thermal properties of rock in the potential repository horizon.

Parameter	Value		
Physical properties			
Porosity (%)	12±4ª		
Grain density (g/cm ³)	2.55 ± 0.03^{2}		
Bulk density at in situ saturation (g/cm^3)	2.30 ± 0.09^{a}		
Dry bulk density (g/cm ³)	2.22 ± 0.10 ^a		
Thermal properties			
Dry matrix thermal conductivity (W/m°K)	2.51 ± 0.17 ^b		
Dry in situ thermal conductivity (W/m°K)	2.1 ± 0.2 ^b		
Saturated in situ thermal conductivity (W/m°K)	2.1 ± 0.2 ^{b,c}		

*DOE (1995).

^bNimick (1990).

 9.65 ± 0.19 in situ saturation with lithophysical cavities and fractures assumed dry.

2.4 Geomechanics

A substantial number of laboratory measurements have been made to determine the mechanical strength of intact samples from the proposed repository horizon. These data indicate that the intact rock is quite strong with a uniaxial strength of 155 MPa (±59 MPa) and a high Young's modulus. Uncracked samples have stress-strain curves that show nearly linearly elastic behavior until failure. Samples with cracks exhibit nonlinear stress-strain behavior as expected when stress is above 50% of the failure stress. Typical stress-strain curves for 50.8-mm-diam saturated samples tested under drained conditions are shown in Fig. 2-7. Most of the tests for compressive strength have been conducted on samples that were saturated with water and tested under drained conditions. This represents a minimum value as rocks are generally weaker when saturated with water. Furthermore, as noted by Biotnott (1997) the uniaxial stress path commonly used to infer elastic constants is strongly influenced by nonlinearities. This becomes more critical when coupled to thermal and other processes, as will be discussed in the Section 4.

The dynamic properties of the rock are also important, especially in the area of seismic design. These properties include the compressional-wave (*P*-wave) velocity (C_p), the shear-wave (*S*-wave) velocity (C_s), the dynamic deformation (Young's) modulus (E_d), and the dynamic Poisson's ratio (V_d). Recommended values of these parameters are given in item 2.2.2 of the *Reference Information Base* (RIB) (DOE, 1995) as $C_p = 3400 \text{ m/s}, C_s = 2040 \text{ m/s}, E_d = 23.5 \text{ GPa}$, and $v_d = 0.22$.

Price (1986) studied the effect of sample size on mechanical properties of Topopah Spring Tuff and found that both the ultimate strength and axial strain at failure are inversely related to sample diameter, whereas Young's modulus and Poisson's ratio are independent of sample diameter. The effect of sample size on ultimate strength is illustrated in Fig. 2-8.

Estimates of rock-mass strength are based on (1) the known behavior of intact rock, (2) the known joint characteristics, and (3) the presence of applied or confining stresses. A rock-mass strength criterion of

 $(\sigma_1)_{\text{ultimate}} = 16.0 + 10.2 (\sigma_3)^{0.602},$

(1)

where $0 < \sigma_3 < 25$ MPa, is presented in item 1.2.6 of the RIB. Equation (1) is based on an assumed rock-mass rating (RMR) of 61 and an unconfined compressive strength of 16.0 MPa.

In situ stress values for the potential repository horizon have been determined from measurements in drill holes USW G-1, USW G-2, and USW G-3 (Stock et al., 1984, 1985). Table 2-5 shows the average mean value and range for vertical stress, which is the maximum principal stress and is due to the overburden rock at the site, in addition to the ratios of minimum and maximum horizontal stresses to vertical stress, and the bearings of minimum and maximum horizontal stresses.

A stress profile for the in situ stress near the ESF has been estimated using two-dimensional finite-element analysis similar to those presented in Bauer et al. (1985). The stress profile can be found in Vol. II, Sec. 4.0.

Figure 2-7. Plots of axial stress vs axial strain for uniaxial measurements on 50.8-mm saturated Topopah Spring Tuff samples at 22°C. Measurements were taken at a strain rate of 10^{-5} s⁻¹. Adapted from Price (1986).

Figure 2-8. Plot of ultimate strength vs sample diameter for uniaxial measurements on saturated Topopah Spring Tuff samples at 22°C. Measurements were taken at a strain rate of 10^{-5} s⁻¹. Adapted from Price (1986).

Table 2-5. Values and ranges of principal stresses in the potential repository horizon at Yucca Mountain (Stock et al., 1984, 1985).

Parameter	Average value ^a	Range
Maximum principal stress (vertical)	7.0 MPa (1015 psi)	5.0 to 10
Ratio of minimum horizontal stress to vertical stress	0.5	0.3 to 0.8
Ratio of maximum horizontal stress to vertical stress	0.6	0.3 to 1.0
Bearing of minimum horizontal stress	N57°W	N50°W to N65°W
Bearing of maximum horizontal stress	N32°E	N25°E to N40°E

*Average value for a depth of about 0.3 km (1000 ft).

















3.0 Processes That Form the NFE

As discussed in Section 1, waste packages will interact with a near-field environment that will be altered from the original Yucca Mountain conditions. The NFE changes arise from several factors, including construction of the repository, emplacement of waste, and the possible backfilling of drifts. All of these activities can cause chemical changes, possible mineralogical or basic rock-mass property changes, as well as hydrologic and thermal changes. These alterations may be significant relative to the EBS design and performance. Because NFE development is driven by coupled T–H–M–C processes that are initiated by repository-related activities, the processes leading to the NFE related to those activities are discussed in this section. The specific NFE conditions that may result are summarized in Section 4. The intention of this report is not to give all possible scenarios but, rather, to indicate the types of processes that can alter the environment so that the reader can consider the impacts on design options.

3.1 Construction

The following discussion includes all activities involved in construction of the facilities that take place prior to emplacement of waste. It does not include construction activities that will take place subsequent to waste emplacement (including closure activities).

3.1.1 Excavation of Rock

The most significant impacts assessed for excavation are the alteration of the stress field in the rock near drifts and changes in moisture due to exposure of surfaces to air. Moisture changes are discussed in Section 3.1.3, below. The changes in stress will result in an overall increase in hoop stresses, which will tend to close fractures, particularly during the period of increased temperature. For further discussion, refer to the Preliminary NFER.

3.1.2 Fluids Introduced during Drilling and Mining

As explained in the Preliminary NFER, the introduction of mining and drilling fluids may increase the matrix saturation surrounding the openings. However, no evidence was found of significant moisture increases during construction of the 5-mile-long access drift, which should be representative of construction and emplacement drifts. The Preliminary NFER contains discussions of possible impacts of drilling and mining, if they should occur.

3.1.3 Mine Ventilation

The Preliminary NFER noted that mine ventilation may lead to drying of the matrix surrounding emplacement drifts, where ventilation of the shafts, ramps, and drifts will remove moisture from the rock via evaporation. This appears to have been the case within the access drift. No water has been observed to drip into the tunnels, except when the ventilation is shut off or when rock is isolated from ventilation. Ions dissolved in the affected pore water will leave behind a small residue of soluble and insoluble salts as the pore water evaporates. No samples of such residue have been taken.

3.1.4 Introduction of Materials by Man

Materials will be introduced, both intentionally and unintentionally, during construction of the facilities. These materials may include such things as material added in the form of crushed rock, granular minerals, engine exhaust, rubber products and other polymers, metals, and

concretes. The major potential for perturbations of the environment due to man-made materials will exist after emplacement.

3.2 Emplacement of Waste

When waste is emplaced, heat from the radioactive decay of the waste will cause changes in the environment, which will likely dominate the geochemical-hydrological-mechanical conditions. Heat has the potential to drive water away from the WPs, to control the relative humidity, and to change the relative hydraulic conductivity, gas permeability, imbibition characteristics, and other properties of the rock mass.

3.2.1 Thermal Loading

3.2.1.1 Dry Out. As reported in the Preliminary NFER, after waste emplacement, thermal loading will drive moisture away from waste emplacement openings. Because of the large increase in vapor pressure, nearly all of the air will be driven away from the boiling zone (leaving the gas phase with 100% water vapor). Tests and analyses of matrix drying show that vapor transport preferentially occurs into openings and along fracture faces. Upon reaching the fractures, most of the vapor is driven away from the boiling zone toward a condensation zone where it condenses and drains down fractures until it is either (1) entirely imbibed by the matrix, (2) enters the saturated zone, or (3) flows back into the boiling zone where it boils again forming a "heat pipe" by refluxing. Low wetting sorptivity of TSw2 appears to facilitate drainage to great depths below the repository. The extent to which condensate drains away from the emplacement zone depends on the competition between gravity drainage and imbibition by the rock matrix. Results of field tests at G-Tunnel (Ramirez, et al., 1991; Nitao and Buscheck, 1995) and at the ESF Single Heater Test (Finley et al., 1997) indicate that condensate drainage can be significant.

3.2.1.2 Rewetting. As the rock cools below boiling conditions, vapor and liquid will flow back into the dried region toward the drifts. The flow will be dominantly along fractures. Saturation gradients will also cause water redistribution in the matrix; however, the very low permeabilities will cause matrix flow to be slow. Thus, flow along fractures will dominate rewetting of the dried-out zone.

Studies show that rewetting occurs at a much slower rate than drying. The details of emplacement configurations, fuel age, AMLs, and percolation flux will have profound effects on the time of dryout and return of moisture to the WP areas. Design development must take these phenomena into account.

3.2.1.3 Dissolution and Precipitation of Minerals. Strong coupling will take place between thermo-hydrologic and geochemical processes. Within the NFE/AZ in the regions where rock is dried, the processes will be dominated by vapor flow and chemical exchanges associated with vapor flow. In the portions of this region where temperatures are very near boiling, the rock will undergo drying with dissolved minerals left in the pores of the rock. This zone may be fairly limited in extent, but the zone will move outward with time so that much of the rock mass may be in this zone at some time. As water is removed, dissolved minerals will be deposited in the throats of pores, particularly the smaller ones. This may result in fundamental changes of the matrix hydrologic properties. Studies have shown that, in addition to salt deposition, considerable mineral changes can result from rock-water interactions at elevated temperatures. This topic is discussed in detail in Sec. 10.4.4, Volume II.

The zone of rock where temperatures are elevated, but are below the boiling point, will be very active both hydrologically and geochemically with potential for very active dissolution and precipitation of minerals. Numerous experiments have demonstrated that the chemical composition of the water is likely controlled by the solubility of mineral phases present in the rock (Knauss, 1987; Knauss and Peifer, 1986; Knauss et al., 1986). See Sec. 10.4.3, Volume II.

3.2.1.4 Increased Stress. The largest effect on the geomechanical behavior of the NFE is expected to be due to the thermal cycle in which the rock will be heated and then allowed to slowly cool. Thermal loading of the NFE will alter the stress in the rock near the emplacement drift as a function of time. During the period of temperature increase, stress in the near-field rock will increase as the rock tries to expand. However, due to geometry of the excavated drifts, the stress fields in the NFE will be complex, and some zones may even be put in tension for an extended period of time. As the temperature decreases, the overall stress levels will decrease, and the entire stress will again be compressive.

Deformations within the rock mass can impact the hydrologic flow system. Many laboratory studies (Bandis et al., 1983; Barton et al., 1985) have documented that greater impact often results from shear rather than normal deformations. This is especially true once stresses are removed, as would be the case for the repository once the thermal pulse has decayed. Permanent changes in hydrologic properties or flow paths can result from stresses normal to fractures, but they tend to be smaller than those caused by shear displacements. This has also been observed in field studies (Wilder, 1987). A "qualitative evaluation of the coupling complexity between mechanical behavior and void space evolution between the joint walls during shear displacement..." has been suggested by Archambault et al. (1997), but has not been applied to NFE evaluations.

3.2.2 Introduced Materials

In addition to material introduced as part of construction, significant volumes of material will be brought in at the time of emplacement, including the container materials and EBS packing materials. The significance of such materials (except for dust, coatings, lubricants, and so forth) is long-term once they interact (corrode) with the NFE. As noted, the major potential for perturbations of the environment due to man-made materials will exist after emplacement, when heat and microbial activity will increase the chances for impact on geochemistry of the water. It is anticipated that introduced material will dominate the WP or NFE chemistry, particularly if microbial activity becomes significant.

3.3 Backfilling

Environmental conditions most likely to be impacted by backfilling include the hydrologic characteristics and an early, transient thermal effect of different heat conduction in the backfill compared with that of the rock mass. Backfilling will also influence opening stability, but that is a design-related issue, not an environmental characteristic. As is discussed in Section 4, backfill can have a significant effect on WP temperatures, on humidity, and on the potential for water to contact the WPs.

As discussed in the Preliminary NFER, it is likely that backfill materials will be the tuff removed during excavation. Excavated materials will either be temporarily stored in drifts, where it is assumed there would be some sort of dust control by ventilation, or on the surface. The moisture in the excavated materials would be reduced because ventilation air and surface conditions are expected to be considerably drier than the in situ rock saturation conditions. With handling and breaking of rock, this drying process would be facilitated. Thus, backfill will likely have lower saturation than the original rock-mass saturation. However, backfill will possibly have greater saturation than the dried-out rock mass that develops as a function of waste heat, unless the backfill materials have been stored in areas of drifts that were above the boiling point of the rock. Once emplaced, the backfill will act as a porous medium that, depending on the material chosen, could disperse any episodic fracture-dominated flow. Although crushed tuff is the most likely backfill material (it is readily available), significant advantages are associated with using backfill consisting of impervious grains (Wilder, 1990; Sec. 1.10.14, Vol. II) rather than crushed tuff. If such a granular backfill is used, there will be longterm flow diversion without wicking. The thermal conduction of granular backfill will be lower than that of the rock mass because of its granular nature. This will not only modify thermal conditions, but because of that modification, will also have dramatic impacts on relative humidity as well as evaporation and deposition of salt and its location.

3.4 Closure

Major activities associated with closure that might impact the environmental conditions include (1) ceasing of ventilation, (2) emplacement of seals in openings, and (3) the introduction of construction materials. Because the details of closure are yet to be determined, and further field studies of hydrothermal-chemical-mechanical coupling will be conducted prior to those details being developed, closure effects will be discussed in subsequent revisions of this report.

3.5 Climate or Site Environment Changes with Geologic Time

Hydrologic discussions for the NFE included in this report assume that conditions will remain unsaturated for the 10,000-yr performance period. The discussions do not address changes (e.g., climatic) that might take place during the 10,000-yr time frame. However, preliminary information is included that addresses how sensitive the hydrologic response might be to differing flux conditions. The flux at the depths of the NFE depends on the response of the overlying units.

Current estimates suggest that the flux is considerably greater than values used at the time of the Preliminary NFER, and they are higher than those values addressed in Volume II, Rev. 1. The information available at the time of preparing this report has not been updated to address the higher flux estimates—namely, in the range of 10 mm/yr. The highest flux values evaluated in this report are in the range of 0.3 to 5 mm/yr. We are in the process of evaluating the hydrologic properties used in the higher flux estimates so that these analyses can be performed using internally consistent parameters. In addition, analyses of whether the higher flux (10 mm/yr) is representative of actual conditions are in progress. Field measurements are likely to be available prior to the next update of the this report, and these values will allow an assessment to be made of the appropriate upper value of flux. Regardless of what value is deemed appropriate for upper ranges, the analysis contained in this report should be representative of the lower flux ranges. If it is determined that the upper ranges are higher than those reported here, that information will be incorporated in subsequent revisions.

4.0 Near-Field Environment

Because the NFE is the environment that interacts with emplaced waste, it is, by definition, the post-emplacement environment that develops through process interactions with the waste. The interactions that develop the NFE (as well as the AZ) are highly coupled processes. To facilitate discussion in this report, the coupled processes are broken down either into singly coupled processes (see, for example, loop A-a in Fig. 1-5) or into multiply coupled processes (see, for example, loop B-b in Fig. 1-5), but with a limit on the number of processes analyzed separately.

Figure 4-1 is a simplified model of the coupled-process interactions that will develop the NFE. Recall from Section 1 that entries along the diagonal of this matrix are the principal components—in this case, of the NFE. Other entries in the matrix represent the processes. Those interactions located to the right of the diagonal are the ones that are forward coupled; those to the left are back coupled.

Figure 4-1. Matrix of coupled processes in the NFE.

For example, waste emplacement generates heat that will mobilize water by boiling or evaporation. Thus, to the right of the environmental parameter of temperature is the process of boiling, as indicated by the arrow. Below this process is the changed moisture conditions of the NFE, shown on the diagonal. The forward coupling indicates that the heat drives a change in moisture. However, the changed moisture conditions may also result in changes in thermal conductivity and in removal of heat by convection. These changes would result in a changed thermal regime. Therefore, back-coupling from moisture conditions to temperature is shown by the arrow to the left from moisture conditions to the process box of latent heat and then up to the environmental condition of heat (temperature). The coupling matix shown in Fig. 4-1 is neither comprehensive nor definitive of all of the interactions that must be considered for the NFE. It is simply provided to help understand the organization of technical discussions in this report and the coupling, which can be quite complex.

Volume I of this report focuses only on those issues that are deemed to have direct impact on the design of the repository, or on the Total System Performance Analysis (TSPA). Many additional issues associated with the NFE and coupled processes must be considered when assessing the NFE, but they are not deemed critical to the design or to the PA abstraction. These issues are discussed in Volume II of this report. For example, one of the issues important to PA is changes in hydrologic properties. Such changes are considered in the results reported in this volume, but the details of how those changes occur are of less direct interest and therefore are not specifically discussed in Volume I. Volume I, Rev 0, of the Preliminary NFER discussed changes that can occur in rock properties (see Sec. 4.2.1 of that report). Volume I, Rev 0 also discussed changes that can occur in hydrologic and geochemical properties (see Sec. 4.2.2). Volume II, Rev. 1 (Sec. 1.7 and 2.0) discussed these effects in detail. No attempt is made here to repeat that information, only to summarize results for design and TSPA purposes.

To aid the reader in locating a variety of issues discussed in greater detail in Volume II, Rev. 1, see the tables in Section 4.4, below. In addition, the discussions in this section are based on numerical models of an idealized system and should not be construed to reflect actual, detailed NFE and AZ conditions. However, models with their inherent assumptions will always be the basis for understanding the evolving environment and for making decisions. Even though some of the uncertainties will be reduced with time, the conclusions need to be applied with care, realizing that the models are idealized and simplified abstractions of the real world.



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Temperature	• Vaporization • Mobilization	Rock-water Interactions Manmade material Interactions (kinetics)	Thermodyn. of vapor	• Dehydration • Rock-water Interactions (Idnetics)	 Dehydration Rock-water interactions Material (kinetics) 	• Stress changes • Coefficient Impacts	 Microbe survival Metabolic processes
• Latent heat • Heat pipes	Moisture conditions	Volumetrics •Water • Rock • Materials			Volumetrics -water -rock Contact modes	Water contribution to crack growth	Survival
		Water chemistry	Vapor partial pressure		 Reaction thermodyn. Solubility Precipitation Carbonation 		Biomass generation Nutrient supply Metabolic processes
			Relative humidity conditions	Vapor/rock Interactions	Vapor/rock/ solid interactions		
Altered flow paths	Attered flow paths		Changed fiow conditions	Hydrologic . properties	Water contact	 Healing of fractures Changes water avail. for cracking 	
Thermal conductivity changes	Water release/ consumpt. In reactions	Altered minerals Rock-water interactions	Partial press. (salts)	Fracture/pore plugging Fracture healing Colloid transport	Mineral and solid material (changes)	Volume changes (phase transform and mineral changes)	
				Closing/ opening fractures Microcracks Coalescence flow path	Stress impacts Mineral phase Microcracks (new surfaces)	Mechanical Conditions (changes)	
N/A	Biomass water consumpt/ release	Biomass micro- chemistry	Blomass control of T/RH relations	Biomass plugging	Biomass dissolution/ mineralization	• Biomass microcracking • Volume changes	Microbiotic conditions

Figure 4-1

4.1 Temperatures of the NFE

For this revision of the NFER, no effort has been made to back-couple geochemistry etc. to the temperature fields, although it is recognized that such back-coupling could alter the temperature fields, possibly in fundamental ways. This will require an iteration of the alteration impacts back into the thermal conductivities. Such an iteration will be included in subsequent revisions of the NFER, once the design is more mature and laboratory and field test data are available on impact of mineral changes and microfracturing on thermal conductivities and hydrologic properties.

Temperatures of the NFE are mainly determined by:

- Thermal output of the waste.
- Configuration of the waste.
- Thermal conductivity of the rock.
- Heat redistribution by moisture/vapor.

First, each of these parameters or processes is discussed, then the results of analyses that consider these factors is presented.

4.1.1 Thermal Output of the Waste

The WPs being considered for emplacement have, for design analyses, been divided into six different classes based on heat outputs. Heat output is a function of waste type, waste age, and amount of waste in any given package. The WP types, in descending order of heat output, are:

- 1. WPs with 21 PWR assemblies of 10-yr-old spent nuclear fuel (SNF), 17.9-kW heat output.
- 2. WPs with 21 assemblies of 26-yr-old PWR SNF with a total heat output of 9.2kW.
- 3. WPs with 44 assemblies of 26-yr-old BWR SNF with 6.1-kW heat output.
- 4. WPs with 12 PWR assemblies of 40-yr-old SNF with 3.2-kW heat output.
- 5. WPs with Defense High Level Waste (DHLW) from Savannah River, 2.8-kW heat output.
- 6. WPs with DHLW from Hanford with 1.6-kW heat output.

Approximately 38% of the waste inventory consists of the 21 assembly PWRs. The heat output will depend on the age of the SNF (as noted by comparing output of type 1 and 2 listed above). The 26-yr-old waste is considered to be more typical of the likely age of emplaced PWR waste than 10-yr-old SNF (see waste age discussion in Sec. 1.2.3 of this report). Whereas the DHLW packages are the next most prevalent waste type, their heat output does not contribute significant amounts of heat. The BWR SNF is the next most numerous WP type (about 26%). The 12 PWR assemblies only represent about 6% of the WPs.

Even though the U.S. has other waste types, those listed are the ones that are currently being considered for disposal at YM if it were to become a repository. If other waste types are later considered, the analyses that follow would need to take such wastes into account.

Analyses of temperatures in the NFE/AZ have taken into account the heat outputs for the WP types, including their percentages and possible locations or distributions. However, the

details of WP distribution are of greater interest for individual WP temperatures and local NFE temperatures. The AZ and average NFE temperatures are influenced more by the total AML and drift spacing than by details of emplacement distribution by WP type. This is important because the distribution of the WPs shown on the design assumptions (see Figs. 1-2 and 1-3) may not be the actual distribution of emplaced waste. Final distributions will depend on several operational constraints.

4.1.2 Configuration (Layout) of the Waste

This report only discusses the results in Volume II that are focused on two design cases, ACD and LL, both of which were performed for approximately 84 MTU/acre loading. These two cases should bound the conditions for those designs options currently being considered.

One of the options common to all of the design cases is the incorporation (or lack thereof) of backfill. From an NFE perspective, properly chosen backfill can have very positive performance aspects that derive from:

- 1. Extending the duration of elevated WP temperatures.
- 2. Establishment of strong thermal gradients that result in significant RH reductions (near the WPs) and lengthened duration of those reductions.
- 3. Diversion of seepage flow that might enter drifts, thus moderating impacts of higher flux.
- 4. Evaporation/boiling of any water that enters drifts during the thermal regime within the backfill rather than on WPs. This has the significant benefit of preventing salts on WPs, which can lower the magnitude of the RH at which corrosion becomes active.

Backfill is addressed in Volume I for the two reference design cases, the ACD and LL designs.

Volume II considers many other design options, including variations in drift diameter, various AMLs, and various waste ages. Refer to Volume II for information on designs that may fall beyond the scope of those considered in the bounding calculations. Tables 4-5 and 4-6 later in this section are provided to assist in locating the appropriate sections in Volume II for the various design options.

4.1.3 Thermal Conductivity of the Rock

Theoretically, the conductivity of the rock should be a strong function of its moisture conditions, although other factors such as microcracking and mineralogic changes can contribute to changes in conductivity. As reported in the Preliminary NFER, the dry matrix thermal conductivity is 2.51 ± 0.17 W/m.°K. The in situ measurements of thermal conductivity, both dry $(2.1 \pm 0.2 \text{ w/m.°K})$ and at a saturation of 0.65 $(2.1 \pm 0.2 \text{ W/m.°K})$, showed that the thermal conductivity may not vary significantly. That the dry and 65% saturation conductivities measured in situ are the same is not consistent with understanding based on intact sample measurements and may indicate that the presence of fractures has a more significant or masking influence on conductivity than does saturation. Recent measurements on intact TSw2 samples show that the thermal conductivity for dry samples is 1.50 ± 0.44 W/m.°K and for saturated samples is 2.29 ± 0.42 W/m.°K. The calculations presented in this report use the Reference Information Base (RIB, Version 4) value of 2.1 W/m.°K.

4.1.4 Heat Redistribution by Moisture and Vapor

Hydrological conditions are the strongest coupled-process determinants of temperatures that develop within the system, particularly where temperatures reach the boiling point. Therefore, the coupling between heat output and hydrology is used to determine the temperature field of the NFE. The calculations that will be presented consider these changes as well as the changes

from convection/conduction to merely conduction as moisture is removed from rock zones. Other processes (e.g., geochemical and mineralogical) will then be coupled to the thermal and hydrological conditions. Thus, coupling is discussed, first, based on an analysis of single coupling of hydrological conditions to the thermal output of the waste considering the design layout, configuration, and characteristics of the waste to determine the temperature field of the system. Following that discussion, we describe the coupling of geochemistry and geomechanics to the coupled hydrothermal results.

4.1.5 Results: Temperatures in the NFE

Temperatures that are of greatest interest for the NFE are those on the WP surfaces and on the surface of the rock. For the AZ, the temperature distributions within the entire rock mass—at least to the point that they approach ambient temperatures—are important. Because much of the information presented for the two regions comes from the same calculations, there is some overlap between the information in this section and in Section 5, which considers the AZ.

Several different analyses have been performed for temperatures in the NFE. For consistency, the analyses reported in Volume II are the focus of this report. More recent calculations are included in a *Materials Research Society* paper (Buscheck, 1996), which used a finer gridding as well as other model refinements. Those results show peak temperatures that are 5 to 10% higher than the ones reported in Volume II. Because this difference is well within the variation expected due to uncertainties in assumptions, it was decided that consistency between Volumes I and II is of greater value than incorporating more up-to-date information that was not as complete (less sensitivity analyses of other design options and of assumptions).

Another set of calculations sometimes referred to are those of TSPA95. These analyses assume that heat generated by WPs could be modeled as a line load. As noted earlier, this is only appropriate for the LL design option. For the point loading typified by ACD and for the modified design, this is not as appropriate as the analyses reported in Volume II that account for a large number of WP distributions/layouts. Finally, several analyses have been performed using models that only account for heat flow by conduction. These analyses are appropriate in a limited number of applications and are particularly useful for details within the drifts or WPs; however, they are not generally appropriate to the NFE or AZE because they do not account for the potentially significant coupling with water. For this reason, the analyses reported here will again be limited to those in Volume II, which include this coupling.

Figures 4-2 and 4-3 show the calculated temperatures around the drifts and on the WP surface to illustrate the difference between the ACD and line-load approaches, respectively. Two line-load options were considered: one in which WPs are placed with only 0.1-m spacing between them, and one in which WPs are placed with 1.0-m spacing. In each scenario considered, the drift spacing was determined by equating the spacing of waste and drifts to the AML and then calculating the drift spacing for the given waste spacing that would yield 83.4 MTU/acre.

Figure 4-2. Calculated temperatures around the drifts for the ACD design, the LL design with 1.0-m spacing, and the LL design with 0.1-m spacing. The calculations shown here are specific to WP placement and use WPs with 21 assemblies of 26-yr-old PWR SNF.

Figure 4-3. Calculated temperatures on the WP surface for the ACD design, the LL design with 1.0-m spacing, and the LL design with 0.1-m spacing. The calculations shown here are specific to WP placement and use WPs with 21 assemblies of 26-yr-old PWR SNF.

Drift-wall and waste package surface temperatures are critical parameters of the NFE. The peak temperatures are shown in Table 4-1, whereas the temperatures over time extending to 10,000 years are shown in Figs. 4-2 and 4-3. The very early time responses are not well defined in





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these plots, and those beyond 10,000 years are not shown. The temperature data for these times can be found in Volume II (see Figs. 1.10.5.1.3, 1.10.5.1.6, 1.10.5.1.9, and 1.10.5.1.13).

	AC	D Rev 00	Line-load design, 0.1m g		
Waste Package	Without backfill	With backfill	Without backfill	With backfill	
$T_{\text{peak}} (t < 100 \text{ yr})$	107-192°C	same, no backfill yet ^a	172-203°C	same, no backfill yet*	
T_{neal} (t > 100 vr)	104-144°C	117-353°C	147-158°C	231-314°C	
Wall Rock					
T _{peak} (t < 100 yr)	100-165°C	same, no backfill yet ^a	161–184°C	same, no backfill yet ^a	
T_{mark} (t > 100 vT)	100-129°C	104-144°C	142-148°C	135-137°C	

Table 4-1. Peak temperatures in rock at the upper drift wall and on the WP surface for ACD and line-load designs; percolation flux = 0.3 mm/yr.

*Backfill is emplaced at 100 yr; therefore, there is no difference at t < 100 yr between the backfilled and nobackfilled cases.

The temperature responses of the ACD and LL designs are markedly different. The axial WP spacing in the ACD design is large enough to thermally isolate WPs from each other, which results in a broader range of temperatures. For the ACD case without backfill, peak drift-wall temperatures range from 100° to 165°C. (As shown in Table 4-1, the peak WP temperatures range from 107° to 192°C). The average peak wall rock temperature (at t < 100 yr) is 132.5°C, and the average WP temperature is 149.5°C. There is a range of 65° in wall rock temperatures and 85° in WP temperatures. This represents a range of ±25% and ±28% of the average wall rock and WP temperatures, respectively. In contrast, the wall rock temperatures for the LL range from 161° to 184°C, and for the WPs the temperatures range from 172° to 203°C. This range represents ±6.7% and ±8.3% of the average temperature of 172.5° and 187.5°C for wall rock and WPs respectively. As is apparent, the temperatures are much more uniform for the LL than for ACD case.

The design option of backfilling the drifts was evaluated for the ACD and LL cases. It was assumed that backfilling was not emplaced until 100 yr after waste emplacement. The impact on WP and wall rock temperatures is also shown in Table 4-1. These analyses assumed complete backfilling around the WPs, including the gap between WPs. In the LL case, leaving the gap open facilitates thermal communication between WPs and therefore would have performance advantages. However, additional design features would be required to prevent backfill materials from filling (at least partially) the gap, and for that reason, the analyses reported are for the gap being filled. A more complete analysis of impacts of gap filling is discussed in Volume II, Sec. 1.10.5.2, and temperature information for the open gap can be found in Table 1.10.5.2.2 of Volume II. As can be seen, the peak temperature of the wall rock for backfilled versus nonbackfilled drifts are not significantly different (no change for ACD and 7 to 11°C greater temperatures for backfilled LL case). The greatest difference is for WP temperatures, where backfill increases temperature by 13 to 209°C for ACD and 84 to 156°C for the LL design. As a result, when backfilled, the uniformity of the LL case compared to the ACD remains pronounced. The average WP peak temperature after backfilling is 235° C with a range of $\pm 50\%$ of the average for the ACD case, and is 272.5°C with a range of \pm 15% of the average for the LL design.

Analyses of the current or modified design have been based on a modified ACD with 1.0-m spacing for significant portions of the drifts. Therefore, Figs. 4-2 and 4-3 showing the line-load design with 1.0-m spacing are more representative of portions of the current design than is the ACD design.

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As seen in Figs. 4-2 and 4-3, the distinction between temperatures resulting from different designs is reduced with time. This is also true for the distinction between wall rock and WP temperatures. As can be seen in Fig. 4-4, for times that extend beyond about 3000 years, there are only minor differences (about 5°C) between the temperature on the WP surface and in the rock at the drift wall. This plot was based on the TSPA95 reference case with AML = 83.4 MTU/acre. The plot is included merely to illustrate that, at longer times, the temperature differences between WP and rock decrease. Thus, the impacts of temperature differences between WPs and drift walls will be greatest during the first few thousand years after emplacement. The TSPA95 case was essentially a line load, but it was judged sufficiently representative of the average response of the design cases to illustrate this point. Therefore, the calculations were not redone to be more directly comparable with the ACD or LL cases.

Figure 4-4. Temperature on the WP surface and in rock at the drift wall for the TSPA95 reference case with AML = 83.4 MTU/acre, LML = 0.47 MTU/m, and drift spacing = 22.5 m.

The preceding discussions were for average conditions. Table 4-2 shows the temperature conditions for individual WPs by waste type based on the ACD and LL cases. This table shows that there is considerably more variability in peak WP temperatures for the ACD design than for the LL design. For the first 100 years, the average WP temperature for the ACD design is 134°C, and the range of temperatures is equivalent to 64% of the average. In contrast, the LL design has a higher average temperature of 180°C and much less variability, with a range of temperatures equivalent to 18% of the average. The difference is even more pronounced for times beyond 100 years. For t > 100 yr, the average ACD WP temperature is 212°C, and the range is 111% of that value. The average LL WP temperature is 256°C, and the range is only 6% of the average value.

Table 4-2 Temperature on the upper WP surface by waste type for the ACD versus LL design with 0.1-m spacing. Full sand backfill is used with $K_{\text{th}} = 0.6 \text{ W/m}^{\circ}\text{C}$, and percolation flux = 0.3 mm/yr.

Parameter	Hanford DHLW	Savannah DHLW	40-yr-old PWR MPC	26-yr-old BWR MPC	26-yr-old PWR MPC	10-yr-old PWR MPC
ACD				· · · · · · · · · · · · · · · · · · ·		
$\frac{T_{\text{peak}}}{(t < 100 \text{ yr})}$	106°C	121℃	112°C	127℃	145℃	192°C
$\frac{T_{\text{peak}}}{(t > 100 \text{ yr})}$	117°C	136°C	172°C	222°C	272°C	353°C
Line load						
T _{peak} (t < 100 vr)	171°C	184°C	171°C	174°C	178°C	203°C
$\frac{T_{\text{peak}}}{(t > 100 \text{ yr})}$	252℃	258°C	250°C	251°C	256°C	266°C

Figure 4-5 shows the three-dimensional temperature distribution for the same amount of waste per acre (83.4 MTU/acre) at 100 and 500 yr for the ACD and the line-load designs with 0.1-m spacing and full sand backfill at 100 yr. Temperatures less than 112°C are transparent to emphasize the relatively hot regions in the NFE. (For an ambient percolation flux of 0.3 mm/yr, rock temperatures above 105°C correspond to a reduction in relative humidity. Therefore, the colored regions in the illustration are zones of superheated conditions where temperatures are hotter than the nominal boiling point (96°C), and moisture conditions are drier than ambient. Condensate drainage will generally occur around the colored regions.)

The large axial spacing in the ACD design thermally isolates WPs from one another, causing rock dryout and condensate shedding to occur in a spherical fashion. The size of any spherically





Figure 4-4.



shaped superheated zone is directly correlated with the heat output of the WP in that zone. Only half of the WPs generate enough decay heat to develop a superheated zone at 100 yr. In addition, only a very small portion of the repository can be considered to be in a marginally superheated state at 500 yr for the ACD design.

In contrast, for the line-load design, condensate shedding occurs in a cylindrical fashion, and condensate flux at the repository horizon only occurs in the rock pillars separating the emplacement drifts. Consequently, the line-load design is associated with a much larger superheated volume that the ACD design, even though the two designs share the same AML.

Figure 4-5. Temperature distribution for (a and c) ACD versus (b and d) line-load design with 0.1-m spacing between WPs. Temperatures are shown at (a and b) 100 yr and at (c and d) 500 yr. For all cases shown, AML = 83.4 MTU/acre, fill sand backfill is present, drift diameter = 5.5 m, and ambient percolation flux is 0.3 mm/yr. Emplacement drifts are parallel to the axial distance axis.

Refer to the tables in Section 4.4, below, for a list of cross references to Volume II, which considered many other parameters that can affect temperatures at the WP and drift wall.

4.2 Water (Liquid and Vapor) Fluxes in the NFE

From the list of priorities identified in Section 1, a key consideration for nuclear waste isolation is whether (more precisely, when and how) water contacts a WP. The manner in which decay heat influences the distribution of liquid saturation and liquid-phase flux in the NFE is the focus of this section.

Water may contact a WP either by liquid-phase (advective or diffusive) flow or by condensation of water vapor that forms a liquid film on the WP (function of RH). Water contact may arise in the following three ways:

- Drift seepage. Advective liquid-phase flow of water that enters the drift (and flows through the backfill if present) as a result of ambient percolation or decay-heat-driven condensate flow. This seepage can include episodic, nonequilibrium fracture flow or steady weep.
- Wicking. Transport of moisture driven by matric potential gradients (i.e., capillary pressure gradients). This is considered to be primarily an advective liquid-phase transport process called imbibition; however, binary gas-phase diffusion can also play a role. Because wicking can occur as two-phase flow, it does not necessarily require a continuous liquid phase.
- Cold-trap effect. Axial water-vapor flow and condensation within the drift is driven by axial variations in temperature (T) and vapor pressure (P_V) along the drift. Water vapor is transported by gas-phase advection and diffusion from areas of higher T and P_V to areas of lower T and P_V where it condenses, which causes RH to increase (to as high as 100%). Large condensation rates can arise in the cooler areas if three conditions are met: (1) high RH in the rock at the drift wall, (2) WP heat output varies substantially from WP to WP, and (3) WPs are thermally isolated from one another. It is important to note that this mechanism does not require liquid water to enter the drift.

For all three of these effects, the most important thermal-management design factor is the WP layout rather than the overall AML of the repository itself (i.e., high versus low AML).

Liquid-phase flux that may enter a drift is the result of ambient percolation flux and decayheat-mobilized vapor and condensate flow. One of the issues that will be discussed in this report is whether (or under what conditions) decay heat decreases or increases the magnitude (or probability) of liquid-phase flux entering a drift. Depending on the AML, decay heat is capable of generating a zone of counter-current, two-phase flow above the repository, which is sometimes



Figure 4-5.



called the refluxing zone or the heat-pipe zone (Nitao, 1988; Pruess et al., 1990; Buscheck and Nitao, 1994). This zone is associated with liquid-phase fluxes that are much larger than ambient. If the AML is large enough, this zone can be displaced well above the repository horizon by the superheated (above-boiling-temperature) zone. If the AML is not large enough or the ambient percolation flux is too large, it is possible for the heat-pipe zone to remain at the repository horizon. A heterogeneous permeability distribution can also cause the heat-pipe zone to remain at the repository horizon in regions where condensate flow is focused (see Volume II, Sec. 1.10.2.5).

Whether liquid-phase flux, arising from ambient percolation flux and decay-heat-mobilized condensate flux, may enter emplacement drifts depends on the magnitudes of the fluxes. As noted earlier, this report considers ambient fluxes in the range between 0 to 5 mm/yr, with 0.3 mm/yr as the most likely. Current estimates, although not universally accepted, are much higher, in the range of 1 to 15 mm/yr. The overall magnitude of the condensate flux depends on AML, not layout configuration. As shown in Fig. 4-6, the total flux at the repository level that results from 83.4 MTU/acre for percolation fluxes of 0.3 mm/yr and 0.05 mm/yr is less than 1 mm/yr, except between 3,000 to 6,000 yr, when fluxes peak at about 1.6 mm/yr (at approximately 4,000 years). Moreover, liquid-phase flux in the vicinity of an 83.4-MTU/acre repository is dominated by condensate flux for 2000 to 5000 yr (see Buscheck, 1996, in Sec. 1.10, Vol. II). Thus, there is a potential total percolation/condensate flux of about 1 mm/yr that theoretically could enter the drifts.

Figure 4-6. Liquid-phase flux at the repository horizon at a location midway between the repository center and the outer perimeter. AML = 83.4 MTU/acre, and repository area = 3.06 km².

Whether this amount of flux can actually enter the drifts depends on several factors. The first is how much of the condensate flux will be available to flow into the drifts and how much will be diverted away from drifts. As noted by Wilder (1990), condensate will be driven away from the drifts in all directions (radially) and will tend to be deposited in the greatest volumes over the pillars between drifts rather than above the drifts. (Although this analysis was made for the borehole emplacement case, the geometric conditions apply to current designs, so that this conclusion remains valid.) Thus, unless flow in pillars is restricted by processes such as fracture healing (see Section 5), it is not appropriate to combine both the average percolation and condensate fluxes to determine the potential flux into the drifts. How large a percentage of the condensate would be deposited over the drifts will depend on the repository design or configuration—specifically drift spacing. The assumption of uniform condensate buildup in the rock above the drifts gives an upper bound, since the condensate flux would be independent of location, and all of the focusing mechanisms would tend to focus condensate away from drifts rather than over them.

A second factor is whether condensate will leave the system before it can return to the drifts. This depends on whether the condensate will be thermally perched so that it can reside in fractures/matrix in the condensate zone rather than flowing through the fractures and away from the condensate zone. Laboratory studies by Roberts and Lin (1997) indicate that for a 25-µm fracture, thermal perching is overcome once the condensate builds up a head exceeding 29 m. At that point, the condensate would drain through fractures. Of course, how far it can flow along the fracture is a function of length of flow path in the superheated zone, the heat flux, and the quantity of condensate available to continue to feed water into the fracture (Buscheck and Nitao, 1994). Because the heat flux is much higher near the drifts, the condensate would tend to drain preferentially through pillar areas. If the condensate is thermally perched, the fractures have a very small storage volume for water to reside in. Analyses of the relative volumes of available storage versus volumes of condensate generated indicate that more than 90% of the condensate generated will have to either flow laterally away from the repository drifts and/or down







fractures to the saturated zone, or fill up the matrix pores, which would result in a thick zone of nearly saturated rock, or leave the system entirely.

A third factor is whether the local heat flux, q_k is large enough to evaporate the local incoming liquid-phase mass flux q_{liq} . Spatial variability can result in local regions where q_{liq} will prevail. Because q_k increases with proximity to the drift, it is more likely for q_{liq} to prevail in the rock farther away from the drifts. Near the drifts, q_k will be 2.4 times greater, on average, for the line-load design than for the point-load design.

Superheated conditions in the repository rock will occur as long as the local heat flux q_h is enough to evaporate the local incoming liquid-phase mass flux q_{hiq} . The extent or rapidity at which the water that is within the pores of the matrix will be vaporized depends on the extent that $q_h > q_{hiq}$. If q_{hiq} is too large, then q_h is insufficient to generate superheated conditions. Spatial variability can result in local regions where q_{hiq} will prevail. Because q_h increases with proximity to the drift, it is more likely for q_{hiq} to prevail in the rock farther away from the drifts. Near the drifts, q_h will be 2.4 times greater, on average, for the line-load design than for the point-load design.

Fourth, studies by Nitao (1997) indicate that for total percolation fluxes below 10 mm/yr, the water will not enter the drifts. If the ambient percolation flux approaches 10 mm/yr, then during the thermal regime the issues of q_h versus q_{liq} , as well as condensate drainage become very important. After the waste and surrounding rock cools down (2,000 to 6,000 yr), then the condensate will not be added to the ambient percolation flux. It will have either drained away or have been imbibed into the matrix. Thus, for the longer time frames, the ambient percolation flux can be used to determine the water entry into drifts. During this time, unless the percolation flux exceeds the 10-mm/yr value that is currently envisioned as the maximum value, then water would not enter the drifts.

The above discussion assumes that there are no drift liners or other ground support systems to hinder water flow into the drifts. The current design calls for emplacement of pre-cast concrete liner as ground support. The liner will not be sealed at the joints but will interlock segments. This will provide some additional resistance to flow into the drifts but will not be entirely tight. Where fractures intersect the drift opening, the water would tend to be diverted around the liner segments, and if open fractures were available, much of the water would flow into those fractures. This would not divert all of the water from the drift, but it would result in reductions so that assessments based on the unrestricted flow into the drifts would be a conservative estimate. Recent estimates of potential for flow into drifts from the ambient percolation flux have concluded that no flow will enter the drifts (Bodvarsson and Bandurraga, 1996).

4.3 Relative Humidity Conditions in the Drift and NFE

One of the performance issues related to the NFE is the relation between relative humidity (*RH*) and corrosion rates of candidate metals for the WP containers. It has been observed that corrosion rates initiate, and kinetics become increased, at *RH* values that are higher than about 70% (McCright, 1997; see Van et al., 1994). Buscheck (1996) noted that one of the results of the heat output of the WPs is the lowering of the *RH* on the surface of the WPs. A temperature difference or gradient in the drift will result in lowering of the *RH* in the drift, or more specifically, on the WP surface, to values below what they would be if the *RH* was entirely determined by equilibrium with the humidity or water/moisture conditions in the rock. Without a difference in temperature between the rock wall surface and the drift air or WP surfaces, the *RH* in the drift would be in equilibrium with humidity conditions in the rock, which are a function of saturation of the rock. Except for very low rock saturations, this would mean that the *RH* within the drifts would be very high. Analyses have indicated that *RH* within the drifts will exceed 70% for rock saturation above 20%. Thus, the rock would need to be reduced in saturation from its

ambient state, currently estimated to be around 90% saturation, to below 20% saturation before there would be a sufficient reduction in *RH* to be out of the high-corrosion-potential regime of *RH* equal to or exceeding 70%. In contrast, the thermal gradient is highly effective at lowering the *RH*.

Table 4-3 shows the calculated RH conditions on the WP surface and on the emplacement drift wall (for average repository conditions) that would result from the two repository designs/layouts considered, i.e., the ACD and LL 0.1-m-gap cases. Because backfill has a strong influence on temperature gradients, the information is shown for both backfilled and open drifts.

Table 4-3. Relative humidity in rock at the upper drift wall and on the WP surface for ACD and line-load designs, assuming percolation flux = 0.3 mm/yr and backfilling occurs at 100 yr.

	AC	D Rev 00	Line-load design		
Time of RH	Without backfill	With backfill	Without backfill	With backfill	
WP surface					
RH(t < 100 vr)	25-86%	Same ^a	16-21%	Same	
RH(t = 120 vr)	29-85%	1-57%	18-23%	1-4%	
RH(t = 2000 yr)	86-97%	43-98%	72-76%	44-57%	
RH(t = 10.000 vr)	92-99.9%	60-99.7%	94-98%	68-81%	
Rock Surface					
RH(t < 100 yr)	36-98%	Same ^a	21-25%	Same	
RH(t = 120 yr)	36-97%	45-99%	23-27%	31%	
RH(t = 2000 yr)	90-97%	92-99%	76-77%	78%	
RH(t = 10.000 vr)	99.8-99.8%	99.8-99.8%	99.1-99.1%	99.1%	

*Backfill is emplaced at 100 yr; therefore, there is no difference at t < 100 yr between backfilled and nonbackfilled cases.

Comparison of the values in Table 4-3 shows that there is a reduction in RH that results from emplacement of waste for both the WP and the drift wall. The reductions are greater for the WP surfaces because the temperatures are higher there than on the drift walls. Regardless of the design option considered, the RHs are considerably lower than the values of nearly 98 to 99% that are representative of the ambient conditions (the values for t = 10,000 yr are essentially ambient).

A significant difference between the RH reductions resulting from the ACD and the LL designs arises because the axial WP spacing in the ACD design is large enough to thermally isolate WPs from each other. However, the reductions are much more dramatic for the WPs when backfill is emplaced and the LL design option is used. In this case, the humidities remain below the critical values for several thousand years. Even after 10,000 yr, they are in the 70 to 80% range. The difference between the ACD and LL case is not merely in the magnitude of the reduction, but also in the variability or range of the reduction. This can be seen by comparing average values. In the ACD case at 120 yr, the average RH = 29% and the range of values is $\pm 28\%$ of the average; whereas for the LL case at 120 yr, the average RH = 2.5% and the range is only $\pm 1.5\%$ of the average. Likewise at 2,000 yr, the ACD RH = 70.5% and the range is \pm 27.5% of the average; whereas the LL case RH = 50.5 % and the range is only ± 6.5 % of the average. The range in RH for the ACD case represents ±97% and ±39% of the average values at 120 and 2000 yr, respectively. In contrast, the range of RHs for the LL case represents $\pm 60\%$ and $\pm 13\%$ of the average values at 120 and 2000 yr, respectively. The reductions in RH for the wall rock are not as significant for the backfill case as for the nonbackfilled case. This reflects the temperature gradient that results when backfill is emplaced, which causes a much reduced WP relative humidity, but a less reduced wall-rock relative humidity.

The preceding discussions were for average conditions. Table 4-4 shows the RH conditions for individual WPs by waste type based on the ACD and LL cases. This table shows that repository design can cause significant differences in the RHs that develop on the WP surfaces, particularly when the drifts are backfilled. One way to consider this impact is to look at the range of RH conditions that develop. Prior to backfilling, the range of RHs for the ACD case is from 25 to 86%. The average RH is 55%, and the range is equivalent to 111% of the average value. In contrast, the range of RHs for the LL case is from 16 to 21%, the average is 20%, and the range is equivalent to 25% of the average value. Thus, the RH environment is much more uniform for the LL case prior to backfill.

At 120 yr (20 yr after backfill is emplaced), the RH range is from 1 to 57% for the ACD case, the average RH is 19%, and the magnitude of the range of RHs is 300% of the average value. For the LL case, the RH range is from 2 to 3%, the average RH is 3%, and the range is equivalent to 35% of the average. Similar differences are seen at 2,000 yr, where the ACD average value is 70%, and the range is equivalent to 78% of the average value. For the LL case, the average RH is 50%, and the range is equivalent to only 6% of the average. Similar differences are seen for the time at which the RH returns to a value of 65%.

$flux = 0.3 \mathrm{mm/yr}$.								
Time of RH	Hanford DHLW	Savannah DHLW	40-yr-old PWR MPC	26-yr-old BWR MPC	26-yr-old PWR MPC	10-yr-old PWR MPC		
ACD								
RH_{max} (t < 100 yr)	86%	61%	65%	50%	42%	25%		
RH_{max} (t = 120 yr)	57%	34%	13%	5%	2%	1%		
RH(t = 2000 yr)	98%	97%	71%	62%	50%	43%		
t(RH = 65%) ^b	160 yt ⁴	210 yr*	1590 yr	2250 ут	8480 yt	13,660 ут		
t(RH = 90%) ^b	950 yr	860 yt	25,870 yr	35,690 ут	54,680	62,670 yr		
Line load								
RH_{max} (t < 100 yr)	21%	21%	21%	21%	19%	16%		
RH_{max} (t = 120 yr)	3%	3%	3%	3%	3%	2%		
RH(t = 2000 yr)	51%	52%	50%	51%	49%	49%		
$t(RH = 65\%)^{b}$	3930 ут	3730 ут	4250 yr	3960 ут	5430 ут	4600 yr		
$t(RH = 90\%)^{b}$	33,320 yr	30,480 ут	36,800 yr	34,890 yr	39,920 yr	39,820 ут		

Table 4-4. Relative humidity on the upper WP surface by waste type for the ACD versus LL design with 0.1-m spacing. Full sand backfill is used with $K_{\text{th}} = 0.6$ W/m°C, and percolation flux = 0.3 mm/yr.

*RH is never less than 65% prior to backfill.

^bTime at which relative humidity reaches this value.

Figures 4-7 and 4-8 show the relative humidities over time extending to 10,000 years. The very early time responses are not well defined in these plots, and those beyond 10,000 years are not shown. The humidity data for these times can be found in Volume II (see Figs. 1.10.5.1.3, 1.10.5.1.6, 1.10.5.1.9, and 1.10.5.1.13). Beyond about 2000 years, the WP RH is maintained at about 95%, whereas the rock RH returns to 99 to 100%. The reason for this difference is that the decay heat flux is nearly constant, so that there is a sufficient difference in temperature between WP and rock wall to depress the WP RH by about 5%.

Figure 4-7. Calculated relative humidities on upper drift wall for the ACD design, LL design with 1.0-m spacing, and LL design with 0.1-m spacing (no backfill). The calculations shown here are specific to WP placement and use WPs with 21 assemblies of 26-yr-old PWR SNF.









Figure 4-8. Calculated relative humidities on the WP surface for the ACD design, LL design with 1.0-m spacing, and LL design with 0.1-m spacing (no backfill). The calculations shown here are specific to WP placement and use WPs with 21 assemblies of 26-yr-old PWR SNF.

The forgoing discussion of RH primarily considered a percolation flux of 0.3 mm/yr, consistent with Volume II. However, the estimates of flux have been more recently revised to greater values, with 5 mm/yr as the value that is recommended (Bodvarsson and Bandurraga, 1996). Values as high as 10 to 15 mm/yr have been suggested. The effects of higher percolation were analyzed in Section 1.10.6 of Volume 2 and are merely summarized here. The relationship between relative humidity on the WP (RH_{wp}) and percolation flux (q_{perc}) was found to fall into three ranges. For $q_{perc} < 0.3 \text{ mm/yr}$, the duration of reduced RH_{wp} increases strongly with decreasing q_{perc} . For $0.3 < q_{perc} < 1.0 \text{ mm/yr}$, the duration of reduced RH_{wp} is relatively insensitive to q_{perc} . For $q_{perc} > 1.0 \text{ mm/yr}$, reduces the period during which RH_{wp} is less than 65% from 1660 to 1200 yr. Because the decay-heat-driven liquid-phase flux is much greater than q_{perc} for the first 1000 yr, it is not surprising that RH_{wp} was found to be only moderately sensitive to q_{perc} of the first 1000 yr. Even if q_{perc} is as large as 10 to 20 mm/yr, the decay-heat-driven liquid-phase flux is much greater than q_{perc} for 0.3 < q_{perc} during the first 500 yr.

Other studies (e.g., Bodvarsson and Bandurraga, 1996) suggest that much less reduction of RH_{wp} would occur at higher q_{perc} (such as 5 mm/yr). However, those studies used hydrological properties that result in much stronger capillary-driven rewetting of the dryout zone (both the fracture and matrix continua) than are predicted using the hydrological properties in Volume II of this report. An important distinction between the percolation-flux sensitivity study conducted in Volume II and the other sensitivity studies is whether the intrinsic hydrological properties of the fracture and matrix continua were held constant while percolation flux was varied. Other studies (e.g., Bodvarsson et al. 1996) simultaneously varied percolation flux and the intrinsic hydrological properties, which makes it difficult to discern the influence of individual factors.

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Post-test analyses of the Single-Heater Test (Buscheck et al., 1997) show that the use of recent hydrological property sets calibrated to percolation fluxes of 3.6 and 6.2 mm/yr predicted much lower temperatures than those measured in the superheated zone surrounding the heater. The 3.6- and 6.2-mm/yr property sets resulted in unreasonably high rates of capillary-driven rewetting of the dryout zone around the heater, preventing the temperatures at many of the thermocouple locations around the heater from rising above the nominal boiling point. This calculated temperature behavior differed markedly from the observed behavior, which rose sharply past the nominal boiling temperatures. An important finding was that the 3.6- and 6.2-mm/yr property sets hampered the predicted dryout for the Single-Heater Test more as a result of the extremely high rates of capillary-driven rewetting than as a result of high percolation flux. Similarly, the greatly hampered dryout around emplacement drifts that has been attributed to high percolation flux is more the result of extremely high (and probably unrealistic) capillary-driven rewetting than high percolation flux itself. This finding illustrates the importance of using thermal tests to calibrate hydrological properties that are required for drift-scale thermal-hydrological modeling.







4.4 Guide to Temperature and RH Data in Volume II

Volume II considers many design details that can affect temperature and relative humidity, in addition to the ones discussed in this report. The variations assessed include different values for WP spacing, drift diameter, percolation flux, emissivity, and backfill thermal conductivity. As an aid to the reader, the following tables contain cross references, by topic, to the appropriate sections, tables, and illustrations in Volume II.

Table 4-5. Guide to temperature and relative humidity data in Volume II, Rev. 1 by section number.

Abbreviations:

WP = effects on waste package upper surface dw = effects at drift wall T = temperature RH = relative humidity AML = areal mass loading expressed in MTU/acre percolation flux expressed in mm/year ACD = Advanced Conceptual Design Rev 00 LL = line-load design (LLNL comparison case)

Design type	TWP	Tdrift wall	RHWP	RHdrift wal	AML
ACD no backfill					
Flux = 0.3	1.10.5.1	1.10.5.1	1.10.5.1	1.10.5.1	83.4
	1.10.5.3	1.10.5.3	1.10.5.3	1.10.5.3	
	1.10.5.7	1.10.5.7	1.10.5.7	1.10.5.7	
Fiux = 0	1.10.6	1.10.6	1.10.6	1.10.6	83.4
LL no backfill					
Flux = 0.3 to 5	1.10.6	1.10.6	1.10.6	1.10.6	83.A
Flux = 0.3	1.10.5.1	1.10.5.1	1.10.5.1	1.10.5.1	83.4
	1.10.5.3	1.10.5.3	1.10.5.3	1.10.5.3	
	1.10.5.6	1.10.5.6	1.10.5.6	1.10.5.6	
	1.10.5.7	1.10.5.7	1.10.5.7	1.10.5.7	
	1.10.6		1.10.6		
Flux = 0	1.10.6	1.10.6	1.10.6	1.10.6	83.4
ACD with backfill					į
Flux = 0.3	1.10.5.2	1.10.5.2	1.10.5.2	1.10.5.2	83.4
	1.10.5.3	1.10.5.3	1.10.5.3	1.10.5.3 ·	
	1.10.5.4	1.10.5.7	1.10.5.4	1.10.5.7	
	1.10.5.5		1.10.5.5		
	1.10.5.7	1	1.10.5.7		
Flux = 0	1.10.6	1.10.6	1.10.6	1.10.6	83.4
LL with backfill					
Flux = 0.3 to 5	1.10.6	1.10.6	1.10.6	1.10.6	83.4
Flux = 0.3	1.10.5.2	1.10.5.2	1.10.5.2	1.10.5.2	83.4
	1.10.5.3	1.10.5.3	1.10.5.3	1.10.5.3	
,	1.10.5.4	1.10.5.6	1.10.5.4	1.10.5.6	
	1.10.5.5	1.10.5.7	1.10.5.5	1.10.5.7	
	1.10.5.6		1.10.5.6		
	1.10.5.7	•	1.10.5.7		
	1.10.6		1.10.6		
Flux = 0	1.10.6	1.10.6	1.10.6	1.10.6	83.4
AML variations	1.10.5.7	1.10.5.7	1.10.5.7	1.10.5.7	25 to 120
Abbreviations:

WP = effects on waste package upper surface dw = effects at drift wall T = temperatureRH = relative humidity

 $\Delta T_{drift} = temp difference between WP and dw$ RHwp/dw = relative humidity difference

for waste package and drift wall

AML = areal mass loading expressed in MTU/acre

LML = lineal mass loading expressed in MTU/m (WP spacing) percolation flux expressed in mm/year thermal conductivity Kth expressed in W/m°C

TSPA95 = Total Systems Performance Assessment (Andrews, 95) ACD = Advanced Conceptual Design rev 00 (Reference case)

LL = line-load design (LLNL comparison case)

			TTTL	Decign		•	Perc.	Emiss.	Backfill	
Figure or	Variable		when	Design	AMT.	LML	finx	(or expla	inatory n	otes)
Table No.	studied	Where	(years)	type	NULL					
1.10.2 Loca	ized and exi	enaca aryo	MI TEguiles	warries	0 to 100	0.63, 1.25		_	no	
F1.10.2.1	Dryout, T	for variou	S WP ages	veries	varies	0.63, 1.25			no	
F1.10.2.2	AML	for variou	S WP Lges	Varies	40	0.26-1.25			no	
F1.10.2.3	T, RH	WP	10,000	Varies	40	0.20				
1.10.3 Perfo	1.10.3 Performance of engineered backfill				04.00	0.62	0	03	ves	$K_{\rm rel} = 0.3$
T1.10.3.1	T, RH	WP, dw	100,000	varies	24,80	0.03		03	Ves VS DO	
F1.10.3.1	T, RH	WP, dw	10,000	varies	24,80	0.85		0.0	- <u>-</u>	
1.10.4 Engi	neered back	ill scenari	<i>os</i>						adal arrun	mions
T1.104.1	RH calc			TSPA95	ļ	0.47-0.53		Various n		er drift
F1.10.4.1	T. ATANA	WP	10,000	TSPA95	83.4	0.47		recungu	AF VS CHOM	4.16 4:00
F1 1042	RH		10,000	TSPA95	83. 4	varies		varies ba		02.004
E1 1043	T PH AT		10,000	•	80	0.63	0 ю 0.3		yes, K _{th} =	. 0.5 10 0.4
F1.10.45	T DU AT.	nn	10,000		24	0.63	0 to 0.3		yes, $K_{\rm th}$ =	= 0.3 to 0.4
F1.10.4A	1, Kn. 61	m WP/dw	10,000		80	0.63	0 to 0.3	•	yes, $K_{\rm th}$ =	= 0.3 to 0.4
F1.10.4.5	RI wp/dw	WP/dw	100.000		24	0.63	0 to 0.3		yes, K _{th} =	: 0.3 to 0.4
F1.10.4.6	Kriwp/dw	-111/10	aina and d	rife spacing	· · · · · · · · · · · · · · · · · · ·					
1.10.5.1 Inj	iuence of ax	Lai WP Spc	ung uni u	ACD IL	83.4	varies	0.3	0.8	no	
T1.10.5.1.1	T, RH	drin wall	Valies	ACD LL	83.4	varies	0.3	0.8	TIO	
T1.10.5.1.2	T, KH	WP		11.	83.4	1.11	0	0.8	BO	
T1.10.5.1.3	T.RH	different	willie types	11	83.4	1.11	0.3	0.8	no	
T1.10.5.1.4	T, RH	different	waste types		83.4	0.94	0.3	0.8	no	
T1.10.5.1.5	T. KH	anierent	Vasie types	ACD U.	83.4	varies	0.3	0.8	no	
F1.10.5.1.1	T	3D View	1000000	ACD	83.4	0,46	0.3	0.8	no	
F1.10.5.1.2	Liq. flux	dw	100,000	ACD IL	83.4	A6. 1.11	0.3	0.8	no	
F1.10.5.1.3	<u>T, RH</u>	dw	2000	ACD IL	83.4	A6. 1.11	0.3	0.8	no	
F1.10.5.1.4	T, RH	dw	10,000	ACD II	83.4	46.1.11	0.3	0.8	no	
F1.10.5.1.5	T, RH	dw	10,000	ACD II	83.4	46.1.11	0.3	0.8	no	
F1.10.5.1.6	T, RH	WP	100	ACD II	83.4	46.1.11	0.3	0.8	no	
F1.10.5.1.8	T, RH	WP	200	ACD II	83.4	46 1.11	0.3	0.8	no	
F1.10.5.1.9	T, RH	WP	10,000	ACD II	83.4 83.4	46 1.11	0.3	0.8	no	1
1.10.5.1.10	T, RH	WP	100,000	TACD, LL	01/	0.04	0.3	0.8	no	1
1.10.5.1.11	T, RH	WP, dw	2000		03.4	0.04	0.3	0.8	BO	1
1.10.5.1.12	T, RH	WP. dw	10,000		02.4	1 0.04	03	0.8	100	1
1.10.5.1.13	T. RH	WP, dw	100,000		63.4	0.74				

Abbreviations:

WP = effects on waste package upper surface dw = effects at drift wall T = temperature RH = relative humidity AML = areal mass loading expressed in MTU/acre

LML = lineal mass loading expressed in MTU/m (WP spacing) percolation flux expressed in mm/year

thermal conductivity Kth expressed in W/m°C

ACD = Advanced Conceptual Design rev 00 (Reference case) LL = line-load design (LLNL comparison case)

Figure or	Variable		When	Design			Perc.	·		
Table No.	studied	Where	(years)	type	AML	LML	flux	Emiss.	Backfill	Notes
		L								
1.105.2 Inf	luence of en	gineered L	ackjul		02.4		0.2	0.0	aver	R06
T1.10.5.2.1	T, RH	dw	Varies	ACD, LL	63.4	VEILES	03	0.0	yes	K. = 0.6
T1.10.5.2.2	T, RH	WP	Varies	ACD, LL	63.4	Varies	0.3	0.0	<u>ycs</u>	N: -06
T1.10.5.2.3	<u>T, RH</u>	different	waste types	ACD	83.A	UA0	0.3	0.6	yes	Nth = 0.0
T1.10.5.2.4	T, RH	different	waste types		83.4	1.11	0.3	0.8	yes	Ath E U.U
T1.10.5.2.5	T, RH	different	waste types	L	83.4	1.11	0.3	0.8	yes, illso i	uis wr gap
F1.10.5.2.1	Т	3D view	100, 500	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	
F1.10.5.2.2	T, RH	WP, dw	10,000	ш	83.4	1.11	0.3	0.8	no vs yes	varies Kth
F1.10.5.2.3	T, RH	dw	2000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	varies WP
F1.10.5.2.4	T, RH	dw	10,000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	varies WP
F1.10.5.2.5	T, RH	WP	2000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	varies WP
F1.10.5.2.6	T, RH	WP	10,000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	varies WP
F1.10.5.2.7	T, RH	WP	100,000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	varies WP
F1.10.5.2.8	T, RH	WP, dw	2000	L	83.4	1.11	0.3	0.8	yes	varies WP
F1.10.5.2.9	T, RH	WP, dw	10,000	LL	83.4	1.11	0.3	0.8	yes	varies WP
1.10.5.2.10	T, RH	WP, dw	100,000	Ш	83.4	1.11	0.3	0.8	yes	varies WP
1.10.5.2.11	T, RH	WP, dw	2000	L	83.4	0.94	0.3	0.8	yes	varies WP
1.10.5.2.12	T, RH	WP, dw	10,000	L	83.4	0.94	0.3	0.8	yes	varies WP
1.10.5.2.13	T, RH	WP, dw	100,000	LL	83.4	0.94	0.3	0.8	yes	varies WP
1.10.5.3 Inf	uence of wa	ste packas	e emissivit	y				•		
T1.10.5.3.1	T. RH	đw	varies	ACD, LL	83.4	3 values	0.3	0.3	no	
T1.10.5.3.2	T. RH	WP	varies	ACD, LL	83.4	3 values	0.3	0.3	no	
T1.10.5.3.3	T. RH	WP	varies	ACD	83.4	0.46	0.3	0.3	TIO	varies WP
T1.10.5.3.4	T. RH	WP	varies	LL	83.4	1.11	0.3	0.3	no	varies WP
T1 10535	T. RH	WP	varies	LL	83.4	0.94	0.3	0.3	tio	varies WP
T1 10.5.3.6	T. RH	dw	varies	ACD. LL	83.4	3 values	0.3	0.3	yes	
T1 10537	T. RH	WP	varies	ACD. LL	83.4	3 values	0.3	0.3	yes	
T1 10538	TRH	WP	varies	ACD	83.4	0.46	0.3	0.3	yes	varies WP
T1 10530	TRH	WP	varies	L	83.4	1.11	0.3	0.3	ves	varies WP
1 10 5 3 10	TRH	WP	varies	LL.	83.4	0.94	0.3	0.3	yes	varies WP
E1 10 5 3 1	T	WP dw	1 87 100	varies	83.4	0.4 to 1.2	0.3		10	
1 - 1.10.J.J.I	14	1 111, UW	· ·	101100	1	1			1	

Abbreviations:

WP = effects on waste package upper surface dw = effects at drift wall T = temperatureRH = relative humidity emissivity expressed in mm/year thermal conductivity Kth expressed in W/m°C AML = areal mass loading expressed in MTU/acre

LML = lineal mass loading expressed in MTU/m (WP spacing) percolation flux expressed in mm/year

ACD = Advanced Conceptual Design rev 00 (Reference case) LL = line-load design (LLNL comparison case)

Figure or	Variable		When	Design			Perc.			
Table No.	studied	Where	(years)	type	AML	LML	flux	Emiss.	Backfill	Notes
1.10.5.4 Infl	uence of wa	e and back	fill							
T1.10.5.4.1	% drift occu	pied		varies	83.4				varies	
T1.10.5.4.2	T, RH	WP	varies	ACD	83.4	0 <i>A</i> 7	0.3	0.3	varies	
T1.10.5.4.3	T, RH	WP	varies	Ц	83.4	0.94	0.3	0.3	varies	
T1.10.5AA	T, RH	WP	varies	ACD	83.4	0.46	0.3	8.0	yes	varies WP
T1.10.5.4.5	T. RH	WP	varies	Ц	83.4	1.11	0.3	8.0	yes	varies WP
F1.10.5.4.1	2D RH distr	ibution	120	ACD	83.4	0.46	0.3	0.3	varies	
F1.10.5.4.2	2D RH distr	ibution	120	L	83.4	0.94	0.3	0.3	varies	
F1.10.5A.3	T, RH	WP	10,000	ш	83.4	0.94	0.3	0.3	varies	off-center
F1.10.5AA	T, RH	WP	10,000	Ц	83.4	0.94	0.3	0.3	. yes	center
F1.10.5A.5	RH	WP	100,000	ACD	83.4	0.47	0.3	0.3	yes	
F1.10.5.4.6	RH	WP	100,000	ACD	83.4	0.47	0.3	0.3	partial sloping backfil	
F1.10.5.4.7	RH	WP	100,000	ACD	83.4	0.47	0.3	0.3	no backfill	
1.10.5.5 Influence of engineered backfill thermal conductivity										
T1.10.5.5.1	T, RH	dw	varies	ACD, LL	83.4	3 values	0.3	0.8	yes	$K_{\rm th} = 0.3$
T1.10.5.5.2	T. RH	WP	varies	ACD, LL	83.4	3 values	0.3	0.8	yes	$K_{\rm th} = 0.3$
T1.10.5.5.3	T, RH	different	waste types	ACD	83.4	0.46	0.3	8.0	yes	$K_{\rm th} = 0.3$
T1.10.5.5.4	T, RH	different	waste types		83.4	1.11	0.3	0.8	yes	$K_{\rm th} = 0.3$
T1.10.5.5.5	T, RH	different	waste types	Ц	83.4	0.94	0.3	8.0	yes	$K_{\rm th} = 0.3$
F1.10.5.5.1	T, RH	WP	· 2000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	$K_{\rm th} = 0.3$
F1.10.5.5.2	T, RH	WP	10,000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	$K_{\rm th} = 0.3$
F1.10.5.5.3	T, RH	WP	100,000	ACD, LL	83.4	.46, 1.11	0.3	0.8	yes	$K_{\rm th} = 0.3$
1.10.5.6 Infl	uence of drij	ft diamete	r							
T1.10.5.6.1	T, RH	đw		Ц	83.4	1.11	0.3	.0.8	no	diam 3-6.5
T1.10.5.6.2	T, RH	WP		Ц	83.4	1.11	0.3	0.8	no	diam 3-6.5
T1.10.5.6.3	T, RH	dw		Ц	83.4	1.11	0.3	8.0	yes	diam 3-6.5
T1.10.5.6.4	T, RH	WP		L	83.4	1.11	0.3	0.8	yes	diam 3-6.5
T1.10.5.6.5	T, RH	different	waste types	L	83.4	1.11	0.3	0.8	yes	diam 3 m
T1.10.5.6.6	T, RH	different	waste types	L	83.4	1.11	0.3	0.8	yes	diam 4 m
T1.10.5.6.7	T, RH	different	waste types	L	83.4	1.11	0.3	0.8	yes	diam 6.5 m
F1.10.5.6.1	peak T	WP, dw		L	83.4	1.11	0.3	0.8	no	diam 3-6.5

Abbreviations:

WP = effects on waste package upper surface dw = effects at drift wall T = temperatureRH = relative humidity

AML = areal mass loading expressed in MTU/acre LML = lineal mass loading expressed in MTU/m (WP spacing) percolation flux expressed in mm/year emissivity expressed in mm/year

 ΔT_{drift} = temp difference between WP and dw

ACD = Advanced Conceptual Design rev 00 (Reference case) ΔT_{pillar} = temp difference, dw vs rock centerline LL = line-load design (LLNL comparison case).

Figure or Table No.	Variable studied	Where	When (years)	Design type	AML	LML	Perc. flux	Emiss.	Backfill	Notes
1.10.5.7 Infl	uence of are	al mass lo	ading							
T1.10.5.7.1	T, RH	dw	varies	ACD	25-83.A	0.46	0.3	0.8	no	diam 5.5
T1.10.5.7.2	T, RH	WP	varies	ACD	25-83.A	0.46	0.3	0.8	no	diam 5.5
T1.10.5.7.3	T, RH	WP	varies	ACD	25	0.46	0.3	8.0	no	diam 5.5
T1.10.5.7A	T, RH	WP	varies	ACD	45	0.46	0.3	0.8	no	diam 5.5
T1.10.5.7.5	T, RH	dw	varies	LL_	25-120	1.11	0.3	0.8	no	diam 5.5
T1.10.5.7.6	T, RH	WP	varies	L	25-120	1.11	0.3	0.8	no	diam 5.5
T1.10.5.7.7	T, RH	WP	varies	Ц	25	1.11	0.3	8.0	no	diam 5.5
T1.10.5.7.8	T, RH	WP	varies	Ц	45	1.11	0.3	8.0	no	diam 5.5
T1.10.5.7.9	T, RH	WP	varies	ц	100	1.11	0.3	0.8	no	diam 5.5
1.10.5.7.10	T, RH	WP	varies	Ц	120	1.11	0.3	0.8	no	diam 5.5
1.10.5.7.11	T, RH	dw	varies	ACD	25-83.4	0.46	0.3	8.0	yes	$K_{\rm th} = 0.6$
1.10.5.7.12	T, RH	WP	varies	ACD	25-83.4	0.46	0.3	0.8	yes	$K_{\rm th} = 0.6$
1.10.5.7.13	T, RH	WP	varies	ACD	25	0.46	0.3	0.8	yes	$K_{\rm th} = 0.6$
1,10.5.7.14	T, RH	WP	varies	ACD	45	0.46	03	0.8.	yes	$K_{\rm th} = 0.6$
1.10.5.7.15	T, RH	dw	varies	IL .	25-120	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
1.10.5.7.16	T, RH	WP	varies	Ш	25-120	1.11	03	0.8	yes	$K_{\rm th} = 0.6$
1.10.5.7.17	T, RH	WP	varies	Ц	25	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
1.10.5.7.18	T, RH	WP	varies	Ц	45	1.11	0.3	8.0	yes	$K_{\rm th} = 0.6$
1.10.5.7.19	T, RH	WP	varies	Ц	100	1.11	0.3	8.0	ye s	$K_{\rm th} = 0.6$
1.10.5.7.20	T, RH	WP	varies	L	120	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
F1.10.5.7.1	Liquid flux	different v	waste types	ACD, LL	83.4	.46, 1.11	0.3	8.0	no	
F1.10.5.7.2	Liquid flux	different v	waste types	ACD, LL	25	46, 1.11	0.3	0.8	no	
F1.10.5.7.3	∆T _{drift}	different v	waste types	L	25 - 120	1.11	0.3	0.8	по	
F1.10.5.7.A	ΔT _{drift}	different v	waste types	LL .	25, 83 4	1.11	0.3	0.8	no	
F1.10.5.7.5	RH	WP	varies	LL	25 - 120	1.11	0.3	0.8	ло	
F1.10.5.7.6	RH	WP	varies	LL	25 - 120	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
F1.10.5.7.7	ΔT _{pillar}	Pwr WP	varies	Ц	25 - 120	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
F1.10.5.7.8	ΔT _{niller}	DHLW	varies	LL	25 - 120	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
F1.10.5.7.9	ΔT _{drift}	WP, dw	varies	LL	25 - 120	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
1.10.5.7.10	∆T _{drift}	different v	waste types	Ш	25, 83 4	1.11	0.3	0.8	yes	$K_{th} = 0.6$

Abbreviations:

WP = effects on waste package upper surface dw = effects at drift wall T = temperature RH = relative humidity AML = areal mass loading expressed in MTU/acre LML = lineal mass loading expressed in MTU/m (WP spacing) percolation flux expressed in mm/year ACD = Advanced Conceptual Design rev 00 (Reference case) LL = line-load design (LLNL comparison case)

Figure or	Variable		When	Design			Perc.			
Table No.	studied	Where	(years)	type	AML	LML	flux	Emiss.	Backfill	Notes
1 10 6 1 6			and adopted	Facts on T	Hbehov	ior				
1.10.6 Infia	ence of amo	ieni perco			92 4	A6 1 11	03	0.8	70	
T1.10.6.1	Liquid flux	đw	100,000	ACD, LL	63.4	40, 1.11	0.05	0.0		
T1.10.6.2	Liquid flux	dw	100,000	ACD, LL	63.4	A0, 1.11	0.05	0.0	10	
T1.10.6.3	T, RH	dw	varies	ACD	<u>83.4</u>	UAD	0-03	0.8	no	
T1.10.6A	T, RH	WP	Varies	ACD	83.4	0.45	0-0.3	8.0	no	
T1.10.6.5	T, RH	different	waste types	ACD	83.4	0.45	0	0.8	no	
T1.10.6.6	T, RH	different	waste types	ACD	83.4	0.46	0.05	0.8	no	
T1.10.6.7	T. RH	different	waste types	ACD	<u>83.4</u>	0.46	03	8.0	TIO .	
T1.10.6.8	T, RH	đw	varies	L	83.4	1.11	0-5	0.8	no	
T1.10.6.9	T, RH	WP	varies	Ц	83.4	1.11	0-5	0.8	no	
T1.10.6.10	T, RH	different	waste types	Ц	83.4	1.11	0	0.8	no	
T1.10.6.11	T, RH	different	waste types	L	83.4	1.11	0.05	0.8	no	
T1.10.6.12	T, RH	different	different waste types		83.4	1.11	0.3	0.8	no	
T1.10.6.13	T, RH	different	waste types	Ц	83.4	1.11	1.0	0.8	no	
T1.10.6.14	T. RH	different	waste types	L	83.4	1.11	5.0	0.8	no	
T1.10.6.15	T. RH	đw	varies	ACD	83.4	0.46	0-0.3	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.16	T. RH	WP	varies	ACD	83.4	0.46	0-0.3	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.17	T, RH	different	waste types	ACD	83.4	0.46	0	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.18	T, RH	different	waste types	ACD	83.4	0.46	0.05	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.19	T, RH	different	waste types	ACD	83.4	0.46	0.3	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.20	T. RH	dw	varies	Ц	83.4	1.11	0-5	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.21	T. RH	WP	varies	Ц	83.4	1.11	0-5-	0.8	yes	$K_{\rm 1h} = 0.6$
T1.10.6.22	T, RH	different	waste types	Ц	83.4	1.11	0	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.23	T, RH	different	waste types	Ш	83.4	1.11	0.05	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.24	T, RH	different	waste types	u	83.4	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
Ť1.10.6.25	T, RH	different	waste types	Ц	83.4	1.11	1.0	0.8	yes	$K_{th} = 0.6$
T1.10.6.26	T, RH	different	waste types	Ц	83.4	1.11	5.0	0.8	yes	$K_{\rm th} = 0.6$
T1.10.6.27	T, RH	different	waste types	ш	83.4	1.11	0.3	0.8	no	near center
T1.10.6.28	T, RH	different	waste types	u	83.4	1.11	0.3	0.8	no	outer edge
T1.10.6.29	T, RH	different	waste types	L	83.4	1.11	0.3	0.8	yes	near center
T1 10 6 30	TRH	different	waste types	LL	83.4	1.11	0.3	0.8	ves	outer edge

ADDIEVIALO

emissivity expressed in mm/year

thermal conductivity Kth expressed in W/m°C

Abbreviations:

WP = effects on waste package upper surface dw = effects at drift wall T = temperatureRH = relative humidity percolation flux expressed in mm/year

AML = areal mass loading expressed in MTU/acre

LML = lineal mass loading expressed in MTU/m (WP line spacing) percolation flux expressed in mm/year

ACD = Advanced Conceptual Design rev 00 (Reference case) LL = line-load design (LLNL comparison case)

bulk permeability k_b expressed in millidarcy

thermal conductivity Kth expressed in W/m°C

Figure or Table No.	Variable studied	Where	When (years)	Design type	AML	LML	Perc. flux	Emiss.	Backfill	Notes
1.10.6 Influ	ence of ambi	ient perco	lation flux	and edge ef	fects on T	-H behavi	ог			
F1.10.6.1	Liquid flux	51 m	above drift	ACD, LL	83.4	046, 1.11	varies	0.8	no	
F1.10.6.2	flux, RH	51 m	above drift	Ц	83. A	1.11	0-5	0.8	no	
F1.10.6.3	Liquid flux	varies	100,000	min scale	83.4		0.3-0.5			
F1.10.6A	Liquid flux	varies	2000	min scale	83.4		0.3-0.5			
F1.10.6.5	RH	varies	10,000	min scale	83.4		0-0.3			
F1.10.6.6	RH	varies	100,000	min scale	83.4		0-0.3			
F1.10.6.7	RH	WP	100,000	ACD	83. 4	0.46	0-0.3	0.8	no	
F1.10.6.8	RH	WP	100,000	LL .	83.4	1.11	0-5	0.8	no	
F1.10.6.9	RH	WP	100,000	ACD	83.4	0.46	0-0.3	0.8	yes	$K_{\rm th} = 0.6$
F1.10.6.10	RH	WP	100,000	Ц	83.4	1.11	0-5	8.0	yes	$K_{\rm th} = 0.6$
F1.10.6.11	T. RH	WP	100	L	83.4	1.11	0.3	8.0	no	
F1.10.6.12	T.RH	WP	2000	LL	83.4	1.11	0.3	0.8	no	
F1.10.6.13	T. RH	WP	10,000	L	83.4	1.11	0.3	0.8	no	
F1.10.6.14	T, RH	WP	100,000	ш	83.4	1.11	0.3	0.8 .	no	
F1.10.6.15	T, RH	WP	2000	ш	83.4	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
F1.10.6.16	T, RH	WP	10,000	L	83.4	1.11	0.3	0.8	yes	$K_{th} = 0.6$
F1.10.6.17	T, RH	WP	100,000	Ц	83.4	1.11	0.3	0.8	yes	$K_{\rm th} = 0.6$
1.10.7 Com	parison of di	iscrete-fra	cture mode	l with equiv	valent con	tinuum ma	odel			
F1.10.7.1	Lio, satur	ation at 50	and 500 yrs		48					L ₆ = 280
F1 10 7.2	Lig. satura	tion at 100	to 1000 yrs		48					k ₀ = 280



4.5 Geochemistry of the NFE

The chemistry of water entering drifts and contacting WPs will depend not only on the chemistry of water as it moves through the rock mass and interacts with the rock, but also on the chemical interactions of that water with introduced materials and microbial communities that exist within the drifts. These topics are discussed separately. In addition, performance of the system will depend on the mineralogy of zones within the natural system that are perturbed as a result of the emplacement of waste, specifically the thermal pulse that results from emplacement. Mineralogy is discussed in Section 5 of this report.

Figure 4-9 is a matrix showing the main components of the NFE associated with water chemistry on the diagonal and the coupled processes off-diagonal.

Figure 4-9. Matrix of coupled processes in the NFE related to water chemistry.

To determine the composition of water that might contact the WPs, the evolution of water compositions over a range of scenarios must be determined. The chemistry of the initial water is probably only of concern during the early time prior to the thermal pulse within the rock exceeding the boiling point of water. After boiling, the water is mobilized, and any water that returns into the drifts subsequently will be a combination of meteoric and condensate water. The condensate would be very dilute, essentially distilled, water but would have strong interaction with the rock because it would be condensing at elevated temperatures. The extent of rock-water interaction will be a function of the temperatures, the duration of contact, whether the rock is unaltered, or whether it had already interacted with other water (typical of the more stable condensate zone of later time frames). The meteoroic waters may have been exposed to elevated temperatures, but such would only be the case for the first few thousand years after emplacement. The mixture of meteoric and condensate waters would then interact strongly with the materials within the drift that had been introduced into the system. Such materials will include significant volumes of cementitious materials, irons or steel, and various constructionrelated materials. An exhaustive list of these materials is available in Volume II, Sec. 6.

Water contacting the EBS would also interact with any minerals or salts left by the water that had been removed from the NFE during the thermal pulse. The likely result will be that between the introduced material interactions, dissolution of precipitated minerals, as well as the microbial interactions. The water chemistry will be largely determined by interactions in the very near field regardless of what the water chemistry might have been in the condensate or meteoric waters. Therefore, water chemistry within the AZ is not discussed here (see Volume II for more information), and is only discussed in Section 5.0 (AZ) as it determines the chemistry of reactions.

As noted, water might drip into the drifts and contact the WPs in the period prior to water removal by boiling. Such contact is less likely because this is also the period during which the ground-support system should be essentially intact and would tend to inhibit water influx into the drifts. This is especially the case if continuous concrete liners are used as ground support. However, the liner segments are not intended to be sealed, and water can enter through the joints between segments. During early elevated temperatures prior to boiling, evaporation of water is an important consideration in determining the water chemistry as well as in establishing the nature of secondary precipitates that may be deposited along fracture surfaces or rock pores. In such regions, secondary precipitates could be dissolved later by the percolation of meteoric and condensate waters. It is also important to determine the nature of precipitates deposited on the WPs directly as water is evaporated or boiled, thus leaving behind the mineral salts. The deposition on WPs is important not only in determining future water chemistries but also in relative humidity corrosion interactions because salt depositions can facilitate corrosion at much lower *RH*. The specific precipitates that may develop will reflect the initial water composition and that which evolves as evaporation takes place.

Temperature 5:	• Vaporization • Mobilization	Rock-water interactions Manmade material interactions (kinetics)	- Thermodyn of vapor	Dehydration Rock-water Interactions (kinetics)	 Dehydration Rock-water interactions Material (kinetics) 	Stress changes Coefficient impacts
• Latent heat • Heat pipes	Moisture conditions	Volumetrics • Water • Rock • Materials			• Volumetrics -water -rock • Contact modes	Water contribution to crack growth
		Water chemistry	vapor partial pressure		 Reaction thermodyn. Solubility Precipitation Carbonation 	
			Relative humidity conditions	Vapor/rock interactions	Vapor/rock/ solid Interactions	
Attered flow paths	Altered flow paths		Changed flow conditions	Hydrologic properties	Water contact	Healing of fractures Changes water avail. for cracking
• Thermal conductivity changes • Heat consumption release in chemical	Consumpt/ release H ₂ O (chemical reactions)	• Altered minerals • Rock-water Interactions	Partial press. (saits)	Fracture/pore plugging Fracture healing Colloid transport	Mineral and solid material (changes)	Volume changes (phase transform and mineral changes)
mactions				Closing/ opening fractures Microcracks Coalescence flow path	Stress impacts Mineral phase Microcracks (new surfaces)	Mechanical Conditions (changes)

Figure 5-15



zone may remain in place for thousands of years, therefore this change in hydrologic properties may be quite significant.

These results are expected to be the foundation upon which refined estimates of mineral facies would be developed for specific repository scenarios of waste emplacement strategies, infiltration fluxes, and rock properties. These scenarios have yet to be developed. Future work will focus on developing these mineralogical and chemical scenarios, and on reducing the uncertainties in key thermodynamic and kinetic data to better bound the behavior of the repository.

5.4.2 Hydrothermal Alteration of Vitric Tuff

Alteration of the basal vitrophyre of the Topopah Spring Tuff (TSw3 unit) represents a potentially profound modification of the lithologic units that compose the potential repository block at Yucca Mountain. The results of earlier hydrothermal alteration experiments, in which the devitrified (repository-equivalent) TSw was reacted with well J-13 water at [150] and 250°C, showed that the primary alteration products were the Ca-rich, mordenite-type zeolite, dachiardite, and mordenite itself (Knauss et al., 1987). Aqueous SiO₂ activity was controlled by cristobalite saturation. Fluid chemistry was in accord with the production of these zeolites. In contrast, results of earlier hydrothermal alteration experiments, in which the vitric airfall/ashflow at the base of the TSw was reacted with well J-13 water at 250°C, showed production of the zeolite clinoptilolite, exclusively (Knauss, 1987).

Knauss continued the series of runs using vitric airfall/ashflow from the base of the TSw and found that even after 6 months at 150°C, the SiO₂ activity remained below amorphous silica saturation, and no zeolites precipitated. However, at 350°C, amorphous silica saturation was quickly reached, and the glass was completely consumed by the end of the run. Because the temperatures generated by waste disposal^{*} are well below 250°C, this would imply that zeolites and clays would be formed at the base of the TSw, but their exact compositions remain to be established.

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Hydrothermal alteration of TSw vitrophyre in well J-13 water was also studied(Knauss, 1987). Although compositionally identical to the vitric basal airfall/ashflow material, the vitrophyre has a much smaller specific surface area. At 250°C, it is clear from SEM/EMP analyses that early in the run the dominant product was a clay, but as the run progressed, the zeolite clinoptilolite became the major product. Given sufficient time and reaction progress, one would expect the vitrophyre to result in mostly zeolite alteration.

Knauss also studied hydrothermal alteration of naturally zeolitized TSw in well J-13 water at 90° and 150°. No new run products were identified in either run, and there was no obvious evidence of dissolution. T Close here real. 5

Preliminary results from the three experiments to study the hydrothermal alteration of vitric Calico Hills Tuff (Tchnv) at 90° (150°, and 250°C contrast with those from the other experiments. Whereas the basal airfall/ashfiow and vitrophyric TSw samples were largely fresh (unaltered) glass, the vitric Tch material contained a significant amount of smectite. The smectite was observed to coat the matrix shards and pumice lapilli. These results demonstrate that the presence of secondary minerals in the vitric tuffs may accelerate, relative to rates of formation in "pristine" vitric materials, the formation of a new generation of minerals under repository conditions. By implication, this leads to the conclusion that alteration of the underlying vitric material will not be rate limited by nucleation effects or alteration minerals.

5.4.3 Amorphous Silica Precipitation

To ascertain how quickly alteration of the near-field and altered zone will occur, it is necessary to rigorously determine the dissolution and precipitation rates of minerals that are present in the mountain in its ambient state, and which may form at future times. Some laboratory measurements have been

conducted using pure mineral phases and simple, controlled solutions. However, it has often been noted that inferences of rates from field measurements are not consistent with laboratory measurements. Therefore, experiments were conducted in the field to measure the precipitation rate of amorphous silica, a mineral that may play an important role in the evolution of the mountain system.

Dissolution of silica in response to water movement at elevated temperatures and subsequent precipitation of silica on cooling may affect local porosity and permeability. Changes in porosity and permeability will, in turn, affect flow pathways, imbibition, and heat transfer. A study was conducted to measure rates of precipitation in natural and laboratory settings as the first stage of an effort to place bounds on kinetic effects over a range of humidities (Carroll et al., 1996, in Sec. 10.4.2, Vol. II; Johnson and Glassley, 1996, in Sec. 10.4.4, Vol. II).

This study confirmed the need of complementary laboratory and field studies to determine the factors that control mineral-water interactions in tuff. Amorphous silica precipitation rates appear to be controlled by distinct mechanisms in laboratory versus field studies. Rates from a field study are 10 to 300 times higher than those predicted from laboratory experiments. The deviation in rates increases as temperature decreases and solution composition increases, and may be influenced by the presence of other chemical species in solution. The precipitation mechanism in the field may be controlled by surface defects, such as dislocation or impurities.

These results indicate that laboratory measurements of precipitation rates may significantly underestimate the rates at which alteration minerals will form in the natural system when perturbed by repository heat. Mineral precipitation rates may be considered instantaneous for modeling purposes, and the primary rate-limiting step in alteration processes will be the dissolution kinetics of mineral phases.

5.4.4 Bounds on Water Composition and Secondary Mineral Development in the AZ

Heat-driven evaporation, boiling, and condensation processes will lead to movement of water vapor away from the repository, resulting in increased saturation of some rocks within some parts of the repository block (Nitao, 1988; Buscheck and Nitao, 1992, 1993). The saturation may approach 100% under some operation scenarios after waste is emplaced. During cool down, or as a result of fracture flow or dehydration of hydrous minerals, it is possible that liquid water may return to or enter the NFE, or may affect rock units bounding the repository horizon. Most of the minerals currently present in the rock making up the potential repository horizon are not in thermodynamic equilibrium with water at elevated temperatures, nor is the glassy material preserved within the PTn or the TSw3 units. As a result, water will interact with existing minerals and glass, causing new minerals to form, existing minerals to change their compositions or dissolve, and modifying the water chemistry in the process. Therefore, a range of simulations were conducted to establish how the mineralogy and water chemistry will evolve as a function of time. To account for uncertainty in the rates at which reactions may procede, as described in Section 5.4.3, simulations were conducted assuming that the rate constants of dissolution and precipitation reactions could have a plus or minus error of an order of magnitude.

The extent to which reactions will progress in the near-field and AZ systems may be limited by water availability. In such cases, the natural systems will only progress part way along the reaction paths suggested by the modeling described here. In these studies, the initial water composition reacting with the rock materials was assumed to be that of J-13 water. The composition used was the average composition reported in Harrar et al. (1990).

5.4.4.1 Results. The water compositions and secondary mineralogies obtained at steady-state conditions, at a given temperature, for the tuffs are constrained by the fixed composition of the starting material, and the solubilities of the starting phases. For vitric material, however, dissolution is treated as though it is congruent; hence, the amount of material dissolved (or, in other words, the extent of reaction progress) will determine the mineralogical products and water chemistry. In this case, therefore, a steady-state composition will not be achieved, except under extreme conditions in which very high absolute amounts of glass are dissolved.

The mineral suites described here must be considered representative "facies" of the assemblage of minerals that may form, and not absolute descriptions of the specific identities of minerals that will be present. In particular, the occurrence of zeolites, such as mesolite, clinoptilolite, and stilbite, or the occurrence of muscovite in simulations must be taken to indicate that a zeolite(s) and a sheet silicate are expected to form, but the specific composition of those phases will depend on local effects that cannot be accounted for *a priori*.

Throughout reaction progress, saponite (a clay), carbonate, and fluorapatite are present at low abundances at all temperatures. These phases are joined by clays (celadonite and smectite), authigenic feldspar (microcline), and zeolites (clinoptilolite at low temperature, stilbite at 50° and 75°C, with mesolite present throughout the temperature range), relatively early in the reaction progress period, and they persist to steady-state conditions. The actual time of appearance of specific phases depends on temperature. At 40°C, important secondary alumino-silicate phases do not appear until after approximately 100 days of reaction progress. At temperatures greater than 50°C, important aluminosilicates appear within a few days. However, in all cases, steady-state conditions with relatively high abundances of secondary minerals are not approached until thousands of days have elapsed. In all cases, one of the last phases to appear is a silica polymorph (in this case, chalcedony), but it quickly becomes the most abundant secondary phase. At high temperatures (90°C), garnet and tremolite appear in the secondary mineral assemblage. These phases probably would not be the stable phases that would form under these conditions, given evidence from metamorphic petrology in low-temperature systems. Their presence in the simulations attests to the limitations of the thermodynamic data currently available.

The initial water composition reflects equilibration of J-13 water at the temperature of the simulation. At all temperatures, this initial composition is moderately alkaline (pH of ~8.5), oxidizing (Eh between 580 and 680 millivolts), and has a low ionic strength (~0.0028 molal). Concentrations of the dissolved elements are low.

Water composition remains relatively constant at a given temperature, until reaction progress reaches approximately 100 days. At that point, the composition changes significantly due to formation of secondary mineral phases, and dissolution of significant quantities of the rock components. Calcium and magnesium are both largely removed from solution due to precipitation of carbonates and, to a lesser extent, a variety of silicates. Thus, as long as the system is changing and dynamic, we would not expect significant changes in secondary mineral phases; however, once conditions become stable (e.g., 50 yr after emplacement), it will only take a few years to experience significant changes.

Na, Si, and Al increase significantly, because the amount added to solution by dissolution of the solids is much greater than the amount removed by precipitation. Nevertheless, the final concentration reached for these elements is controlled by the solubilities of the major secondary phases. Silica, for example, is buffered at chalcedony saturation, at steady-state conditions. Therefore, in the AZ, where temperature and saturation remain relatively stable for long times, it is expected that silica will be in equilibrium with chalcedony saturation.

The redox state is effectively determined by the assumption that the solution is in equilibrium with atmospheric gases. Although this may not be the case within the drifts at early times, it is most likely to be the case in the AZ. Hence, total oxygen is assumed to be maintained at approximately 20% of the coexisting gas phase.

Variation in pH is strongly controlled by the fact that CO₂ fugacity is fixed in these simulations. These results differ somewhat from those described in Sec. 3.4.1 of Volume II, where water evaporation was modeled. In the latter study, the effect of including rock interaction with the water was not considered. The pH remains relatively alkaline when CO₂ is controlled by atmospheric gases. Results of the single heater test (SHT) indicate that within the active boiling regime, excess CO₂ is generated by rock-water interactions. This would change the pH conclusions. However, it is unknown whether excess CO₂ will be generated within the AZ. That generated in the boiling zone could flow into the AZ, but after the initial thermal perturbation, the excess CO₂ should decrease.

The conservative elements Cl, F, Li, and P showed no significant variation during reaction progress because they are not incorporated into any secondary phases that form in significant quantities.

For dissolution of relatively small volumes of glass, the secondary phases that form during reaction progress are similar to those formed in the tuffs (pyrolusite, clays, and zeolites), except for the respective amounts of secondary phases. These results are qualitatively consistent with the description of glass alteration (Levy, 1984a,b) in the vicinity of Yucca Mountain. The main exception is calcite, which persists to the end point of the calculations for all temperatures except 40°C.

Dissolution of large volumes of glass rather quickly achieves a state in which the only secondary phases that are present are pyrolusite, chalcedony, and hematite. In none of the simulations with vitric material was muscovite or microcline present as reaction products at any temperature.

Water composition in simulations involving small degrees of glass dissolution reached end points that were nearly identical to those for the tuff-water system. The simulations involving large volumes of glass ultimately achieved much higher total concentrations of all elements except Si, which dropped to very low values controlled by large volumes of precipitated chalcedony. In addition, at high degrees of glass dissolution, pH drops, reaching acidic values (-3.8) in the low temperature simulations and slightly alkaline (-7.8) at the 90°C simulation. This pH behavior primarily reflects the effects of temperature, the glass chemistry components, and the solubility of C-bearing aqueous species (e.g., HCO3⁻) and solids on pH. In these simulations, total dissolved elemental carbon decreased by three orders of magnitude, which will profoundly influence alkalinity.

To evaluate the effects of uncertainty in reaction rate constants on mineral development, simulations were conducted in which sanidine dissolution rate was varied by an order of magnitude, plus or minus. This was done to document the effect of this uncertainty on time predictions of when specific minerals will appear or vanish during the reaction progress. The results shown in Fig. 5-16 for sanidine abundance demonstrate that the propagated error increases with time, until sanidine vanishes. The error bars represent uncertainty regarding the amount of the mineral present at any given time. The error in predicting when a phase appears or disappears from the model can be substantially greater than the prediction of the amount made. For example, the maximum uncertainty in the amount of sanidine produced is about 50%, but the uncertainty of when it is consumed is from about -60% to +2100%.

Figure 5-16. Estimates of sanidine abundance over time.

5.4.4.2 Conclusions. These simulations were designed to be preliminary for bounding water compositions that could evolve within the AZ and migrate into the NFE, and for bounding the mineralogy that could evolve in the AZ. The key constraints upon which the simulations were based and details of the uncertainties are discussed in Sec. 10.4.3.4 of Volume II. The results provide bounds on secondary mineralogy and water chemistry and are a baseline from which further refinements in simulation strategies can be developed.

The results document that the timing of mineral development is most sensitive to dissolution kinetics. Precipitation kinetics are sufficiently fast that they do not significantly influence mineral development for most mineral phases. This knowledge allows uncertainty limits to be quantified for projections of mineral development based mainly on dissolution kinetics.

These results also complement studies of the thermodynamic properties of mineral phases (Bruton, 1996, in Sec. 3.4.2, Vol. II) documenting that specific types of minerals (facies) can be accurately predicted during evolution of the repository. However, the development of individual mineral compositions is less accurately predicted. The studies also provide quantitative descriptions of the uncertainty expected in simulations of the evolution of the AZ.

Retardation of radionuclides passing through the AZ is directly related to the identity and abundance of minerals along the flow path. These results show that an analysis of radionuclide transport must consider the effects of water chemistry as well as the impact of dissolution kinetic and thermodynamic properties on the models.



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5.4.5 Formation of Flow and Transport Barriers in the AZ

Thermal gradients and resultant advective mass transport within the AZ will invariably lead to porosity evolution as a consequence of reaction-controlled mineral dissolution and precipitation both along fractures and within bounding matrix blocks (Fig. 5-17). Preliminary reactive transport modeling conducted in FY 1995 suggested that these processes might result in significant modification of flow porosity both above and below the repository (Glassley, 1995).

Figure 5-17. Matrix of coupled processes.

Reactive transport simulations can be viewed as an iterative, five-stage process, in which the steps are:

- 1. Precise specification of the initial chemical and hydrologic state of the fluid-rock system.
- 2. Translation (which often includes approximation) of the specifications into the analogous model parameters of the software.
- 3. Specification of certain system-independent numerical parameters.
- 4. Successful execution of the software.
- 5. Interpretation of calculated results.

Using this five-stage process, the approach taken in this study was to first establish the validity of the software-encoded numerical model by successfully predicting the outcome of two well-characterized, experimentally verifiable "benchmark" problems. Next, increasingly complex, yet still experimentally verifiable problems having some relevance to porosity evolution in the backfill zone were addressed. Finally, a critical simulation that cannot be addressed directly by physical experiment was carried out: modeling of long-term porosity evolution in the AZ. These benchmark comparisons are reported in more detail in Sec. 10.4.4.1 of Volume II and provide some confidence that the results of the verification are realistic, if not exact. Only those simulations that apply to the AZ are reported here.

5.4.5.1 Compositional Specifications and Parameterization of the AZ. The composition and relative abundance of mineral phases in the potential repository host rock, the Topopah Spring Tuff of the Paintbrush Group (TSw2), have been determined experimentally by Warren et al. (1984), and translated into model compositions/abundances by Delany (1985) for use in EQ3/6 calculations. Because GIMRT, unlike EQ3/6, is not yet configured to account for solid solutions, it is necessary to further approximate Delany's (1985) model TSw2 mineralogy by resolving its alkali feldspar and plagioclase compositions into the corresponding end-member K-feldspar, albite, and anorthite components. Mg-beidellite and biotite compositions compose only 1.1 volume% of the TSw2, and they represent the only Mg- and Fe-bearing phases in the system. In these preliminary simulations, Mg-beidellite and biotite were eliminated from the model TSw2 composition.

The fluid phase anticipated to form within, and react with, the TSw2 in the AZ is derived from condensation of a vapor phase moving radially outward from the repository (Buscheck et al., 1993). This condensate is likely to be a highly dilute, neutral-pH fluid. The fluid was modeled to initially contain very low concentrations of aqueous ions and complexes augmented by the additional component CO₂, which is necessary owing to the presence of atmospheric gases. Again, a cautionary note regarding CO₂ beyond atmospheric is made. The initial fluid composition, specified as the log molal concentration of basis species and given in Table 5-1, is much more dilute than J-13, reflecting its condensation from a vapor phase.



Temperature	• Vaporization • Mobilization	Rock-water interactions Manmade material interactions (kinetics)	Thermodyn.	• Dehydration • Rock-water interactions (kinetics)	Dehydration Rock-water Interactions Material (kinetics)	• Stress changes • Coefficient Impacts
• Latent heat • Heat pipes	Moisture conditions	Volumetrics • Water • Rock • Materials			Volumetrics -water -rock Contact modes	Water contribution to crack growth
		Water chemistry	Vapor partial pressure		Reaction thermodyn. Solubility Precipitation Carbonation	
			Relative humidiby conditions	Vapor/rock Interactions	Vapor/rock/ solid interactions	
Altered flow paths	Altered flow paths		Changed flow conditions	Hydrologic properties	Water contact	 Healing of fractures Changes water avall. for cracking
Thermal conductivity changes	Water release/ consumpt. in reactions	 Altered minerals Rock-water interactions 	Partial press. (salts)	Fracture/pore plugging Fracture healing Cotiold transport	Mineral and solid material (Changes)	Volume changes (phase transform and mineral changes)
				Closing/ opening fractures Microcracks Coalescence flow path	Stress Impacts Mineral phase Microcracks (new surfaces)	Mechanical Conditions (changes)

Figure 5-17.

Table 5-1. Summary of infiltrating fluid composition used in this study. In the simulations, Ca^{+2} and HCO_3^- represent the mass- and charge-balancing species, respectively.

Basis species	Log activity
Н .	-7
K*	-10
Na*	-10
Ca*2	-5
Al ⁺³	-20
SiO ₂ (aq)	-10
HCO3	5

We employed the thermodynamic data base typically used with both the original and our modified version of GIMRT, mastertemp.data, which is a reformatted version of the "composite" data base distributed with the EQ3/6 software package (Wolery, 1992). Two distinct AZ environments were addressed with GIMRT. The first was the localized, interior backfill environment. The second was the much larger-scale, relatively outlying condensation zone (Buscheck et al., 1993), where gravity-driven flow of vapor-exsolved fluid represents the most likely catalyst for significant, long-term porosity evolution in the AZ.

5.4.5.2 Backfill Simulations. We used crushed TSw material having an initial porosity of 40%, where a constant outlet flux of 266.7 $m^3/m^2/yr$ was maintained over six months. Because the porosity of actual backfill material and especially the *in situ* fluid flux are likely to be lower than the adopted values, the simulations provide an upper bound to the potential magnitude of porosity evolution.

In the first TSw2-backfill simulation, the system was maintained in isothermal equilibrium at 90°C. Steady-state saturation indices for all primary and potential secondary minerals were achieved within one hour. As shown in Fig. 5-18, the fluid remains undersaturated with respect to all primary and potential secondary minerals at this inlet node throughout the simulation. Hence, at this location, all primary phases will dissolve to varying degrees controlled by the encoded kinetic rate law, and no secondary minerals will precipitate. At the outlet node (Fig. 5-19), mineral saturation indices reach steady-state values in about 9 hours. Here again the fluid remains undersaturated with respect to all primary minerals, but in contrast to the initial node, several secondary phases (kaolinite, gibbsite, diaspore, boehmite, and muscovite) achieve steady-state supersaturation with the fluid.

Figure 5-18. Mineral saturation indices (log Q/K) vs time at the inlet node of a GIMRT simulation at 90°C, where initial composition of the 31-cm, 1-D, fluid-rock system is given in Table 5-1. Solid symbols denote primary minerals; open symbols denote secondary minerals.

Figure 5-19. Mineral saturation indices (log Q/K) vs time at the outlet node of a GIMRT simulation at 90°C.

At 90°C, kinetic dissolution rates for the undersaturated primary phases are on the order of -10^{-1} volume%/yr for cristobalite and albite, -10^{-2} volume%/yr for quartz, and $<-10^{-3}$ volume%/yr for K-feldspar and anorthite. As a result, over six months, the aggregate dissolution of primary minerals will be <0.25 volume%. Some of this minute volume loss will be filled by precipitation of secondary phases following their sequential supersaturation along the flowpath. Specifically, gibbsite will precipitate first (at node 2; nodes were located at each cm along the flowpath), subsequently joined by diaspore (node 4), kaolinite (node 9), boehmite (node 14), and finally muscovite (node 29). However, because maximum





Figure 5-18.

K

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precipitation rates for these phases are so slow at 90°C (ranging from 10^{-11} volume%/yr for kaolinite to 10^{-14} volume%/yr for muscovite and boehmite), the aggregate precipitation of secondary minerals is <2.5(10^{-3}) volume/yr throughout the domain, i.e., negligible. As a consequence, six months of reactive flow through crushed TSw at 90°C results in a relatively uniform but minute increase in porosity (from 40 to ~40.25%) as a function of minor dissolution of primary phases and trace precipitation of alteration minerals (Figs. 5-20 and 5-21). Although the changes are very small, when considering the times of concern for repository performance (10,000 yr or longer), the changes can be significant. At about 3000 yr, the aggregate is 10%; at 100,000 yr, it could be significant. It is unknown whether these rates continue or if they decrease with time.

Figure 5-20. Porosity and volume fractions of primary and secondary minerals as a function of distance at six months for a GIMRT simulation at 90°C. Solid symbols denote primary minerals; open symbols denote secondary minerals.

Figure 5-21. Porosity and volume fractions of trace secondary minerals as a function of distance after six months for a GIMRT simulation at 90°C.

A second GIMRT simulation was performed identical to the first except for the temperature, which was increased to 250°C. Although this temperature exceeds the maximum likely to be attained in the near-field rock, depending on emplacement configuration, it may be realistic and is within the plausible temperature ranges in the backfill close to drifts. These temperatures are probably unrealistic for the AZ, which will be in the temperature range closer to 100° to 150°C. Thermodynamic and kinetic considerations suggest that 250°C will lead to readily observable porosity evolution and alteration-phase precipitation. We know that similar changes to those observed at 250°C would apply to lower temperatures (especially at long time frames), but in a qualitative sense. We do not have sufficient information at present to be more quantitative.

Although the time scale for attainment of steady-state saturation indices is similar for the 90°C and 250°C simulations, the enhanced solubilities and faster reaction rates of primary TSw2 minerals at the higher temperature and their effect on dissolution (and secondary phase precipitation) are readily apparent. Despite concomitant precipitation of kaolinite, whose reaction rate and volume fraction ultimately attain 50 volume%/yr and 14 volume%, respectively, this near-complete consumption of cristobalite, albite, and quartz causes porosity to increase from 40 to 70% at the initial node (Fig. 5-22), and, of course, results in the late marked decline in saturation indices for these and all other silica-bearing phases (Fig. 5-23). Because the differences between the 90°C and 250°C cases are impacted by reaction kinetics, except for those products that are in thermodynamic equilibrium, the long time frames of elevated temperatures expected in the AZ will probably result in mineralogy/chemistry similar to the 250°C case by the time the thermal pulse decays.

Figure 5-22. Porosity and volume fractions of primary and secondary minerals as a function of time at the inlet node of a GIMRT simulation at 250°C.

Figure 5-23. Mineral saturation indices (log Q/K) vs time at the inlet node (1 cm) of a GIMRT simulation at 250°C, where initial composition of the 31-cm, 1-D, fluid-rock system is given in Table 5-1. Solid symbols denote primary minerals; open symbols denote secondary minerals.

As shown in Fig. 5-24, the fluid is supersaturated with not only the same six secondary minerals that characterize the inlet node, but also with albite and quartz. Primary albite ceases to dissolve, and secondary albite begins to precipitate within 20 minutes, quartz follows suit about 40 minutes later, and within nine hours the cristobalite dissolution rate has dropped to and essentially leveled off at about 20%



Figure 5-20.

















Figure. 5-23.



of its original magnitude. As a result, after six months, total quartz and albite volume fractions have actually increased (Fig. 5-25), and the net volume fraction of cristobalite, albite, and quartz has only declined from 45.9% to 37.2% (versus the inlet node). Hence, despite slightly reduced precipitation of secondary kaolinite relative to the inlet (10.2% versus 15.2%), the six-month outlet porosity has actually decreased from 40 to 38.8% (Fig. 5-25).

Figure 5-24. Mineral saturation indices (log Q/K) vs time at the outlet node of a GIMRT simulation at 250°C.

Figure 5-25. Porosity and volume fractions of primary and secondary minerals as a function of time at the outlet node of a GIMRT simulation at 250°C.

One would expect the gradient in dissolution/precipitation behavior to be sharpest across the first few nodes, where the fluid concentration gradient is steep. This is exactly the case, as illustrated by the time-integrated spatial distribution of saturation indices and porosity/volume fractions (Figs. 5-26 and 5-27). Thus, gradients that develop during initial AZ development close to the WP environment may not experience complete chemical reactions. However, the rapidity of growth in the dryout zone and its associated condensate zone will slow as a function of decreasing rate and increased diameter as the spherical volumes increase. In addition, the kinetics are sufficiently high that reactions can be essentially fully developed in a relatively short time. Therefore, the products of the 250°C case should be representative of the majority of the AZ. Note that the region of enhanced porosity is restricted to the first few nodes and is characterized by very steep gradients, whereas the region of reduced porosity accounts for roughly 90% of the reaction domain and is characterized by much subtler variations.

Figure 5-26. Mineral saturation indices (log Q/K) as a function of distance after six months for a GIMRT simulation at 250°C.

Figure 5-27. Porosity and volume fractions of primary and secondary minerals as a function of distance after six months for a GIMRT simulation at 250°C.

Significant porosity evolution in the 90°C backfill zone will not be realized for at least 100 yr. Moreover, recalling the likely overestimation of initial porosity, this time frame must be considered extremely conservative. Because accurate values for the initial porosities and flow rates relevant to the potential crushed-TSw2 backfill environment are currently unavailable, a more detailed, longer time-scale simulation addressing porosity evolution here cannot be pursued at this time. Current information indicates that if pre-cast or cast-in-place concrete liners are used, the water chemistries (pH) would be affected, and the fluid would be more reactive. Therefore, the extent of alteration of the backfill would be increased. However, no quantitative studies of this topic have been performed to date.

The potential repository horizon lies well above the present-day water table within the welded, devitrified section, which is characterized by low porosity (11%) and extremely low, fracture-controlled permeability $(10^{-18} m^2)$ (Klavetter and Peters, 1988). As described by Buscheck et al. (1993), hydrology of the post-emplacement NFE may be characterized by a boiling zone, where vapor moves radially outward from the EBS/WP and an outer but partially overlapping condensation zone, where temperature-dependent condensation of the vapor leads to primarily gravity-driven flow of a fluid phase. In the vicinity of emplacement drifts, average calculated trajectories indicate that the condensate may flow along path lines characterized by an initial isothermal segment (at 90° to 100°C) of several meters and subsequent cooling segment (–1° to –3°C/m), also of several meters.

The temporal evolution of mineral saturation indices at the inlet (0.25 m), center (5 m), and outlet (10 m) nodes are shown in Figs. 5-28 through 5-30. Initial steady-state mineral saturation indices are attained





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Figure 5-26.



Figure 5-27.

after about 1 month, during which time the fluid obtains supersaturation with quartz and the secondary phases kaolinite and pyrophyllite, but remains undersaturated with respect to cristobalite (and the primary feldspars) throughout the reaction domain. As a result, over the entire reactive flowpath, cristobalite (and the feldspars) will dissolve, whereas quartz, kaolinite, and pyrophyllite will precipitate. Given the relatively sluggish dissolution rates of the feldspars at 90° to 95°C, dissolution of these primary aluminous phases and dependent precipitation of kaolinite and pyrophyllite will be very minor. Hence, the fundamental mechanism of porosity evolution in this environment is expected to be metasomatic replacement of cristobalite by quartz.

Figure 5-28. Mineral saturation indices (log Q/K) as a function of time at the inlet node (0.25 m) of a GIMRT simulation at 95°C, where initial composition of the 10-m, 1-D, fluid-rock system is given in Table 5-1. Solid symbols denote primary minerals; open symbols denote secondary minerals.

Figure 5-29. Mineral saturation indices (log Q/K) as a function of time at the center node (5 m) of a GIMRT simulation at 95°C.

Figure 5-30. Mineral saturation indices (log Q/K) as a function of time at the outlet node (10 m) of a GIMRT simulation at 95°C.

As shown in Fig. 5-31, at the inlet (pure dissolution) node, behavior analogous to the center and outlet of the domain is observed until cristobalite has been completely dissolved at about 650 yr. Subsequently, the dilute inlet fluid begins to dissolve quartz, which disappears completely after 3,100 yr, and then albite.

Figure 5-31. Porosity and volume fractions of primary and secondary minerals as a function of time at the inlet node (0.25 m) of a GIMRT simulation at 95°C.

Figures 5-32 through 5-35 show time-integrated spatial variations in porosity and mineral volume fractions. The complete replacement of cristobalite (initial volume fraction = 39.6%) by quartz throughout the reaction domain leads directly to a background increase in porosity from 11% to about 16% after roughly 3,100 yr (Figs. 5-34 and 5-35). Superimposed is a secondary porosity enhancement, resulting from quartz dissolution (initiated following complete cristobalite consumption) at the first few inlet nodes, beginning with the inlet node at 650 yr. Because the 90° to 95°C dissolution rate of quartz is only about 40% that of cristobalite, it takes dilute inlet fluid roughly 2,500 years to completely dissolve the inlet-node quartz from its maximum concentration (29 volume%), which coincided with cristobalite disappearance at 650 yr. This secondary quartz dissolution front advances quite slowly, and the advection of quartz-derived SiO₂(aq) maintains nearly uniform levels of slight quartz supersaturation and resultant minor quartz precipitation over the entire downstream reaction domain. As a result, this secondary porosity enhancement is characterized by steep gradients that advance downstream quite slowly (Figs. 5-34 and 5-35).

Figure 5-32. Porosity and volume fractions of primary and secondary minerals as a function of distance after 100 yr for a GIMRT simulation. Temperature for the first 5 m is 95°C, then decreases by 1°C/m over the last 5 m.

Figure 5-33. Porosity and volume fractions of primary and secondary minerals as a function of distance after 1000 yr for a GIMRT simulation. Temperature for the first 5 m is 95°C, then decreases by 1°C/m over the last 5 m.







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Figure 5-29.







volume fraction (%)

Figure 5-31.







volume %



Figure 5-33.

Figure 5-34. Porosity and volume fractions of primary and secondary minerals as a function of distance after 5000 yr for a GIMRT simulation. Temperature for the first 5 m is 95°C, then decreases by 1°C/m over the last 5 m.

Figure 5-35. Porosity and volume fractions of primary and secondary minerals as a function of distance after 10,000 yr for a GIMRT simulation. Temperature for the first 5 m is 95°C, then decreases by 1°C/m over the last 5 m.

Important implications for the post-emplacement evolution of porosity in the AZ can be drawn from this simulation. First, thermodynamic, kinetic, and volumetric constraints strongly suggest that long-term (> 3,000 yr), gravity-driven flow of vapor-derived dilute condensate at 90° to 95°C will result in nearly 50% background porosity enhancement (from 11% to 16%) as a consequence of mole-for-mole, quartz-forcristobalite replacement. Should this local "condensation" zone persist for longer time periods, the *in situ* porosity may further increase dramatically, albeit slowly, to more than 65% as quartz dissolution occurs. It is conceivable, if not likely, that the post-emplacement AZ will evolve toward extreme porosity heterogeneity, characterized by relatively large porosities in the condensation-zone "cap" above the repository and relatively low porosities (which nevertheless exceed pre-emplacement porosity) in the outlying margins of this cap. In addition, these extremes may be joined by a narrow, steep-gradient interface. All of the dissolved silica will, of course, eventually precipitate farther downstream, presuming gravity-driven flow continues down thermal gradients. Hence, it is anticipated that a zone of very low (less-than-ambient) porosity may develop at depth, perhaps below and radially outward from the repository. Above the condensate cap, there will also be porosity decreases.

5.4.5.3 Conclusions. Recognizing the abundance of compositional and hydrologic approximations, assumptions, and simplifications that are necessarily embodied in these still-preliminary models, they nevertheless serve to elucidate the fundamental time scales (thousands of years) and geochemical processes (e.g., cristobalite-for-quartz replacement) likely to govern active evolution of the AZ. They can also be used to imply the distant-future porosity distributions that are so critically important. Our initial modeling efforts suggest that post-emplacement reactive transport will lead to porosity enhancement in the AZ, most dramatically in the condensation "cap" overlying the repository. Further modeling of fluid transport to regions below the emplacement drifts will evaluate the extent to which flow barriers may form due to porosity reduction.

Our results to date strongly suggest that significant porosity changes will occur in the postemplacement, repository-block environment. Long-term performance assessment of this environment must explicitly account for such porosity evolution and its effect on dependent geochemical and hydrologic processes. Porosity changes in the backfill would also be significant because the purpose of the backfill is to act as a hydrological barrier. Buscheck has pointed out that backfill may need to be quartz sand, not crushed tuff, to provide diversion of percolation without wicking. A further consideration is the geochemical alteration of backfill materials, particularly crushed tuff. If porosity, hence permeability and matric potential, significantly change, then the backfill may not serve its intended function.






Figure 5-35.

6.0 Performance Implications

The near-field and altered-zone environments impact the overall performance of the repository system in three general areas:

- Performance of the waste container materials, particularly metallic container materials.
- Dissolution/mobilization and transport of radionuclides from the waste.
- Flow fields and retardation characteristics of the natural system.

6.1 Impacts on Waste Container Performance

The current repository concept includes the use of metal containers that consist of outer corrosionallowance materials in concert with inner corrosion-resistant materials. Combined, the materials are intended to prevent contact of water with the waste materials and to provide for containment, thus isolation, of the waste. The length of time that these containers will function and how well they function very much depends on how much water can contact the WPs, the chemistry of the water, and the relative humidity of air surrounding the WPs.

Estimates of water volumes that can potentially contact the WPs have been based on the Project's understanding of ambient percolation fluxes and the thermohydrologic response of the system to the repository heat load as envisioned in the current design for the repository. In terms of WP water contact, this revision of the NFER has placed more emphasis on flux that is created by displaced water combining with ambient percolation flux than on the saturation conditions and timing for those conditions. This change of emphasis from that of the Preliminary NFER is due to two factors.

First, as the YMP has continued to study water flow through the mountain, it has become apparent that the flow is dominated by fracture flow more than by matrix hydrologic conditions. This was recognized at the time the Preliminary NFER was written and was the basis for the statement "...Thus, fracture flow is the only credible mechanism to bring liquid water into contact with the WPs."

The second change arises from the recent recognition that ambient percolation fluxes may be much higher than values originally considered possible. At the time the Preliminary NFER was written, the general feeling was that the total percolation flux was very low or zero. Current estimates for flux based, in part, on neutron logging measurements and isotopes released by atmospheric nuclear weapons tests—range as high as 10 to 15 mm/yr, with 5 mm/yr as the most accepted value (Bodvarsson and Bandurraga, 1996). However, such estimates are based largely on analyses that used assumed hydrologic property values, the representativeness of which has not been demonstrated and is not entirely consistent with the findings of others (Buscheck, 1997).

Although it was previously recognized that periodic or episodically higher fluxes could occur, data were not available to quantify the significance of these transient fluxes. Hence, greater emphasis in the Preliminary NFER was placed on the conditions of overall saturation as an indicator of potential water contact with the waste. The current revision places much less emphasis on saturation changes and their duration as an indicator of WP performance (it is still important in AZ processes, as discussed in Section 5). Instead, more emphasis is placed on actual flux estimates, which will be dominated by fracture flow. It is difficult to relate saturation to flux directly because saturation that exists in the matrix is relatively immobile and does not contribute to flux to the same extent that the same saturation in fractures would. Because the saturation profiles shown in Figure 5-6 are based on assumptions of equilibrium between the moisture in matrix and pores, they do not relate directly to flux of water that could enter the drifts.

One possible way to assess the flux entering drifts is to account for the current or ambient percolation flux in the mountain—which is an indication of flux within fractures because the matrix will contribute such a small percentage to the overall flow—and then to assess the quantity of additional water that can

be added to this flux by thermal mobilization or redistribution of water. In this regard, it is important to note that the total condensate flux reported in Section 4.3 represents the total amount of water that can be vaporized by the heat available, which consists of both meteoric percolation and thermally mobilized water. Thus, if the meteoric percolation flux is greater, then the amount of water that can be vaporized will be reduced. The total flux will remain constant for any given heat flux.

The influence of percolation flux on drift-scale thermal-hydrological behavior was investigated in Section 1.10.6 of Volume II for the flux range of 0 to 5 mm/yr. It is important to realize that the decayheat-mobilized, liquid-phase flux in the condensate zone overlying the repository is much greater than this range for about 1000 yr. For the 83.4-MTU/acre repository design, the maximum decay-heatmobilized liquid-phase flux can be as large as 1000 mm/yr during the first 100 yr. The maximum liquidphase flux declines to about 200 mm/yr at 200 yr, 100 mm/yr at 500 yr, and 20 mm/yr from 1000 to 3000 yr. These flux values are for homogeneous conditions; spatial variability in heating conditions or fracture properties can cause fluctuations about these averages. For example, greater fluxes can occur in regions of focused condensate flow or where the decay-heat flux is less than average. On the basis of test results from air permeability, it is likely that the fracture network within the TSw unit can easily accommodate fluxes of up to hundreds of millimeters per year.

Two significant cautions apply to this assessment. First, the same problems in defining flux based on saturation apply to estimating the ambient flux; that is, assumptions must be made to evaluate the current flux. These assumptions deal with the equilibrium between fracture and matrix flow, as well as the rock mass properties versus those determined from intact, small-scale samples. Current estimates of flux are as high as 10 to 15 mm/yr, but these values are not confirmed by direct measurements within the TSw2. Shallow infiltration flux is estimated by periodic neutron logging of shallow boreholes. Analyses of these measurements provide a lower bound estimate of the flux from the surface to the depth of measurement. This measurement will not detect water moving rapidly though the fractures because large fluxes can be handled with no measurable change in water content. Only the water actually imbibed into the matrix will be measured by the neutron log. These relatively shallow measurements do not account for diversion and other processes, and thus may not be representative of the percolation flux at the repository horizon. (See Section 4.3 for further information on flux.)

As explained in Section 4.3, the amount of water that can contact waste depends on both the meteoric water percolation flux and on the flux of water that is mobilized by being driven out of the pores of the rock. Furthermore, depending on the repository design, there can be local areas where water is more readily able to enter the drifts than others because the thermal output of waste will drive water away from the waste for many years. For the percolation flux cases analyzed in this report, up to 5 mm/yr, the estimate of how much water could contact the WPs is less than 1 mm/yr. In addition, studies by both Nitao (1997) and Bodvardsson and Bandurraga (1996) indicate that there will be very little or no flow into the drifts at the percolation fluxes that are believed to exist at YM. Even after the thermal pulse no longer restricts inflow, after thousands of years, Nitao found that unless percolation flux exceeded 10 mm/yr, there would not be influx into the drifts. Further, as noted in Section 4.5, the use of a concrete liner will moderate or inhibit any inflow that might occur during the early times when the liner remains intact. It is also recognized that inflow will not be uniform but, rather, restricted mostly to major inflow zones. The inflowing water will interact with emplaced materials that have been altered in response to repository conditions. This alteration is expected to evolve over time. It is anticipated that the result of these interactions will be a solution of moderately high pH (9 to 11), with an ionic strength greater than that of the present ambient pore waters. Depending on how inflow is distributed both spatially and temporally, inflow could accelerate corrosion, but inflow is not seen as a major determinant of WP lifetime. The corrosion-allowance material can be designed with thickness necessary to provide for long-lived containers if general corrosion, not pitting or microbial-induced corrosion, is the concern.

Relative humidity, on the other hand, is perhaps the most significant environmental variable affecting corrosion rate of the outer barrier, therefore survival of the WPs. As discussed in Section 4.3, both the magnitude and time distribution of *RH* depend on the repository design/layout. The design can prevent conditions in which corrosion is most aggressive, namely when temperatures and relative humidities are high, for many thousands of years, or it can create conditions that will be adverse to container survival.

This is especially true for individual waste containers having low heat output or those located near the edge of the repository. The ACD and LL designs have different RH conditions associated with them, and the ACD design has great variability in RH conditions from package to package (see Section 4.3 and Table 4-4).

The period of time during which design influences RH also depends on percolation flux. The relationship between relative humidity on the WP (RH_{wp}) and percolation flux falls into three flux ranges. For a percolation flux of less than 0.3 mm/yr, the duration of reduced RH_{wp} increases strongly with decreasing flux. For the range 0.3 to 1.0 mm/yr, the duration of reduced RH_{wp} is relatively insensitive to percolation flux. For percolation flux greater than 1.0 mm/yr, the duration of reduced RH_{wp} decreases with increasing flux. Increasing the percolation flux from 1 to 5 mm/yr, reduces the period during which $RH_{wp} < 65\%$ from 1660 yr to 1200 yr. For the first 1000 yr, RH_{wp} is only moderately sensitive to flux between 0.3 and 5 mm/yr. Even if the flux is as large as 10 to 20 mm/yr, the decay-heat-driven liquid-phase flux is much larger than the percolation flux during the first 500 yr. Therefore, percolation flux will only have a moderate impact on RH after 500 to 1000 years. If the percolation flux is as high as 10 mm/yr, the RH will essentially be ambient (close to 100%).

Preliminary unpublished analyses of thermal variability have been made for the ACD and the 1997 modified designs (Blink and Chesnut, 1977). Figure 6-1 shows the package arrangement for the modified design that had been described earlier in Fig. 1-3, but with both the axial position and the WP heat output shown. Figure 6-2 shows the initial heat output as a function of axial position along an emplacement drift. As can be seen, there will be a very significant peak of thermal output around the 21 PWR assembly, which will result in much higher local temperatures around these packages with associated lowering of *RH*. In contrast, much cooler temperatures prevail around the 44 BWR assemblies. This design will result in much higher *RH* in the areas of lower heat output and possibly even condensate collection and dripping in the cooler areas.

Figure 6-1. Location of individual waste packages for the modified design.

Figure 6-2. Heat output (kW/m) as a function of the location (axial position) of individual waste packages for various averaging lengths.

Figure 6-3 shows the effect of the time-dependent (calculated) RH and T environment on general corrosion rate for an ACD layout but with only 24 MTU/acre, compared with corrosion predicted for ambient (fixed) RH and T. This was considered to be a conservative estimate of the impact of waste-heat generation on corrosion because it represents the smallest MTU/acre value that has been considered to be feasible. Even for this conservative case, the difference between ambient and thermally perturbed conditions is dramatic. The corrosion rate does not change much, compared to ambient conditions, until a few years after emplacement, then drops rapidly at closure as the RH drops to very low values. For several hundred years after closure, the calculated corrosion rate is less than 10⁻¹⁰ microns per year. As humidity increases from the minimum value reached shortly after closure, the corrosion rate increases rapidly but does not reach the ambient corrosion rate until almost 100,000 years after closure.

Figure 6-3. Corrosion rate of the outer WP barrier at ambient conditions compared with the rate predicted for the 24-MTU/acre repository.

Figure 6-4 shows the cumulative depth of corrosion of the outer WP barrier for both an ambient (fixed) RH and T environment and for the predicted RH and T repository conditions representative of the conservative 24-MTU/acre case. The predictions show that almost 100,000 years would be required to remove half the original outer barrier thickness by general corrosion for the repository RH and T, as contrasted with approximately 10,000 years for the ambient conditions. These predictions do not take into account localized attack. Analyses using a Monte Carlo approach, as well as analyses for the 83.4-













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Figure 6-3.

MTU/acre design, will be reported in the future. Such simulations of the propagation of millions of pits per WP suggest failure of the outer barrier in only a few thousand years.

Figure 6-4. Cumulative corrosion depth of the outer WP barrier at ambient conditions compared with the values predicted for the 24-MTU/acre repository.

In addition, the preliminary analyses of Blink and Chesnut (1977) have also identified the periods of time during which WPs in the ACD, 83.4-MTU/acre design will be exposed to especially adverse environments. The simultaneous occurrence of high RH and T promotes rapid corrosion, whereas lowering the RH retards corrosion at all temperatures. Figures 6-5 and 6-6 show the conditions of corrosion susceptibility for two design options with an assumed 6-mm/yr ambient flux. Figure 6-5 shows results for the ACD design case without backfill. Fig. 6-6 shows results for the LL case with 0.1-m gaps between WPs and with sand backfill. Recall that these two designs represent the bounding cases for corrosion. For the ACD design case, both "hot" and "cool" waste packages enter the window of susceptibility—that is, the combinations of RH and T that cause rapid corrosion—within a few hundred years and do not exit this region until after 70,000 years. In contrast, model calculations for the LL design (Fig. 6-6) show that the "hot" packages never enter the aggressive corrosion region. The "cool" packages enter the aggressive region at about 10,000 years and exit at about 18,000 years. For the LL design with backfill, water evaporates within the backfill and is not deposited on the WPs. Therefore, the RH at which corrosion starts is not lowered by the presence of salts on the WP surfaces. Other engineering options, such as drip shields, limit contact by liquid water, but they do not influence RH. Even continuous drift liners would only prevent such exposure if they could be designed to have lifetimes that exceed 10,000 years. Thus, the most effective engineering option that offers protection from corrosion is the simple selection of WP placement and backfill.

Figure 6-5. WP relative humidity versus surface temperature for the ACD design with an AML of 83 MTU/acre, no backfill, and an ambient flux of 6.2 mm/yr (Blink and Chesnut, 1997).

Figure 6-6. WP relative humidity versus surface temperature for the 0.1-m LL design with an AML of 83 MTU/acre, sand backfill, and an ambient flux of 6.2 mm/yr (Blink and Chesnut, 1997).

At present, the program is considering ceramic coatings or other options to make the WP lifetime less sensitive to environmental variables. However, unless 100% defect-free manufacturing and inspection can be guaranteed, the environment will still be of concern.

6.2 Impacts on the Waste Form

Several issues regarding waste form dissolution and radionuclide release/mobilization are related to the NFE. First, the amount and timing of dissolution or radionuclide release is directly related to the amount of water contacting the waste, its chemistry at the time of contact, and the temperature of the waste at the time of contact (which influences the kinetics or rapidity of reactions). The amount of water that contacts the waste provides an upper bound on the amount of water that contacts the WPs. This value would tend to be conservative in that the WP container would have to first fail before water could contact the waste itself. It is possible that the WP could be penetrated by humid air corrosion, exacerbated by microbes and crevice processes, before any actual flow of liquid water contacts the WPs. Another source of conservativism is introduced by assuming that all water contacting a failed WP also contacts the waste itself. Much of the water will never contact waste, especially that fraction of waste protected by cladding remaining intact after the decay of the thermal pulse. Part of the water contacting WPs will also be consumed in corrosion processes and products. Thus, depending on the design, the waste would not likely be contacted by water until well after the thermal pulse and its impact on overall flux had decayed.





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Figure 6-4.

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Water chemistry will be very much influenced by interactions with introduced materials and possibly microbes, and by the temperatures of the system at the time of those interactions. Studies are not mature in this area, but it is likely that the pH will be elevated by the significant amount of concrete in the system. If crushed backfill is used, the water will probably be elevated in silica. How much influence on water chemistry the corrosion products will have is not currently known. The amount of impact would depend on whether failures had occurred by general corrosion—in which case there would be large volumes of corrosion product that would both hinder water flow and also influence its chemistry—or whether the failures resulted from localized failures, such as pitting. Insufficient information is available on the design or the interaction with materials to allow definitive discussion at this time.

6.3 Impacts on Flow Fields and Transport

The impacts on system performance that involve flow and transport are essentially due to the AZ as it develops with time. The major conclusion that can be reached at this time is that there will be a significant zone that will be exposed to both elevated temperatures and moisture (saturations). As shown in Fig. 5-6, a zone of elevated temperatures between 60° and 96°C and saturations between 60 and 100% moves away from the repository horizon and increases in size. Below the repository is a zone from 425 to 500 m below surface that remains at elevated temperatures and saturations for at least 9,000 years. A similar zone is located above the repository around 250 m below the ground surface; however, this zone is not as long-lived and extensive.

The conditions in these zones are conducive to chemical/mineralogical reactions. These zones are likely to experience equilibrium conditions. Estimates of the alteration potential show that during early times, the TSw3 unit may develop alteration products of zeolites and clays. However, the kinetics of the zones close to the repository are such that these zones may not have water present for sufficiently long periods to allow the chemical reactions to go to completion. Farther away from the repository, nearly complete chemical reactions are expected so that the mineralogy and water chemistry will be in equilibrium.

As described in Sec. 5.4.5, long-term condensate zones will allow considerable porosity enhancement (from 11 to 16%), and the values could be even larger if the zones persist for long time periods. This would certainly be the case in the zone below the repository that is stable for many thousands of years, and for the condensate cap if condensate is able to develop above the repository. Dissolved silica would be deposited further out. Thus, a zone of enhanced porosity would develop around the repository, and a zone with reduced porosity would develop farther out. Studies of fracture healing indicate that the zones where water flows through will potentially have healing of fractures. Therefore, there is the potential to have increased porosity (and possibly permeability) around the repository with a zone of decreased porosity and permeability further out. If these zones developed, they would tend to divert percolating meteoric water away from the repository area, would restrict the downward flow of water that might have flowed through the repository (thus reducing waste transport), and create an enlarged porosity around the repository that would be able to store water so that it would not flow through the fractures. On the other hand, the zone wherein the temperatures and saturations are high is much larger and more stable (in duration) below the repository than above it. This might imply that the restrictions to flow above the repository would not develop to the same extent as below the repository. In this case, environmental conditions might develop wherein water could build up (bathtub scenario) over time. However, the development of heat-pipes in this region could cause greater rock-water interactions that would potentially cause greater changes to occur above the repository.

The AZ is less sensitive to repository design than is RH, water contact, and other parameters within the drifts and on the WPs. Therefore, it is the WP and waste form performance issues that need to be considered in design of the repository. The AZ is more sensitive to total mass of waste emplaced and on its density. Thus, for any given AML, the general AZ development will be similar, especially at longer time frames. Nevertheless, a zone that is part of both the AZ and NFE can be influenced by design, namely, the one that develops in the pillars between waste emplacement drifts. Condensate will tend to

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drain down the pillars during early emplacement times. Depending on the spacing of drifts, the pillars may be in the environment of elevated temperatures and saturations (between 60 and 97°C, and greater than 60% saturation). Because condensate will be elevated in temperature, it may not require that the conduction of heat cause elevated temperatures to create conditions conducive to porosity/permeability changes. However, the larger the pillar, the longer the time for drainage to occur before the temperatures are elevated to the extent that kinetics will result in these hydrologic changes. Therefore, design options that include wider drift spacing would result in a larger percentage of the condensate draining away from the zone above the repository. This would reduce the amount of flux to re-enter the drifts and would also reduce the heat pipes that may form above the repository horizon.

An additional performance implication of the AZ is the significant amounts of zeolites and secondary minerals that are likely to form below the repository. The extent of changed sorptive capacity of the overall system is not known at this time. However, preliminary indications suggest fundamental increases in both amounts of and lateral extent of these minerals.

Finally, if the thermal conditions that extend into the water table create the types of convection cells that are postulated, they might contribute to increased mixing/dilution of any radionuclides. Of course, increases in temperatures within the saturated zone may also cause the rocks to be out of thermodynamic equilibrium with the water, which could cause mineralogical changes that are unknown at present.

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Appendix A Identification of Q and NonQ Data

(Appendix B of Volume II)

The NFER and the LLNL conducted work reported in the NFER have all been performed in accordance with QARD or predecessor QA program procedures and therefore the material in the NFER is considered "Q"- This determination is in accordance with DOE guidance dated 8/19/96 states "All reports, analyses, or models must be done under the QA program and are considered "Q", if they describe, predict, or defend how the natural barriers will or are performing their function to isolate waste (their safety function). These Q reports/analyses/models can use non-Q data that is clearly identified as such-the report is still Q...."

The planning for the activities reported in the NFER has long recognized that studies to understand mechanistic processes, lab tests on non YM materials or on non-qualified samples, or scoping calculations and lab tests may not be used directly in license activities related to Site Characterization, but will provide a basis for design, planning, or judging the ranges of conditions within which the conditions of Yucca Mountain will fall. As such the strategy incorporated in planning of these activities included a series of studies that were performed under the LLNL approved QA program that implements the QARD or predecessor YMP QA programs. The strategy provided for qualification of any data that had been determined to be non-Q, but later was deemed necessary for licensing etc., through any of a number of possible approaches including: peer review; comparison of the results with studies in which the stringent controls were applied; reperforming of the specific studies that were found applicable, or review of the procedures applied to determine if they met all requirements of qualified data. In keeping with the DOE guidance discussed above, the Q status of data used or reported in the NFER are reflected in Appendix C following the TDIF number. This Q status is listed to identify which data would need to be qualified or verified for use in quality affecting work.

Model results (viewed by some as data, but not rigorously defined as data) fall into another category which needs to be considered. First, for the results to be relied on, any codes used in the analyses must be properly qualified. Qualification of the codes is only a small part of the process of qualifying results of models. Other portions relate to whether the parameters used were appropriate, whether the model itself (which includes the conceptualization and the abstraction process) is appropriate to, or valid for, Yucca Mountain. The codes used in work reported in this Rev 1 of the NFER are as follows:

Hydrologic Codes:	
V-Tough	Fully QA qualified
NUFT	LLNL developed outside of YMP. Not
	qualified, but comparisons to qualified codes
	have been performed
Geochemistry Codes:	
EQ3/6 .	Fully QA qualified
GEMBOCHS (data base)	Some data qualified, some QA status
	indeterminate
GEMBOCHS suite of codes	not qualified-qualification left to user by
	comparison with analytical solutions of the
	appropriate sets.
SUPCRT92 (code)	commercial/academic code-not gualified
(data base)	Published (peer reviewed) but not qualified
OS3D/GIMRT	Commercial Code-
REACT	Commercial Code

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Geomechanics Codes	
FLAC ABAQUS FRACROCK	commercial code-not qualified commercial code-not qualified LLNL developed outside of YMP. Not qualified.
Man-Made Materials Codes	
EQ3/6 GEMBOCHS (data base) GEMBOCHS suite of codes	Fully QA qualified Some data qualified, some QA status indeterminate. Microbial will be added. not qualified—qualification left to user by comparison with analytical solutions of the appropriate sets.

As was the case for data qualification, the codes that are found to be of importance in characterization of the Near-Field Environment, if not already qualified, will need to be qualified. The QA procedures for qualifying codes will be followed which will start with an Individual Software Plan to determine the appropriate way to qualify the code.



Appendix B NFER Vol. II Chapter Data Status

(Appendix C of Volume II)

Overall NFER-LL960807504241.012 (non-Q)

Chapter 1: Hydrothermal Modeling

This chapter does not report data, it reports results of modeling only, using data from RIB or referenced data from other sources. Because the codes used in the models varies, model information by section is listed.

Section 1.1 Introduction-no data

Section 1.2 No new data-refers to published data or data from other sources Section 1.3 No data-model analyses only using V-TOUGH (This section is

Rev 0 and therefore was based on V-TOUGH prior to its qualification) Section 1.4 No data, refers to work of others

Section 1.5 No new data, refers to work of others

Utilizes Log-Normal Model which is unqualified but used and reported in outside literature

Section 1.6 No new data, analyses based on V-TOUGH prior to qualification Section 1.7 No data, analyses/models only based on V-TOUGH

Section 1.8 No data, model results based on Qualified V-TOUGH Section 1.9 No data, model results based on Qualified V-TOUGH

Section 1.10.1-1.10.3 No data, model results based on both V-TOUGH and NUFT as well as hybrid model.

Section 1.10.4 No data, model results based on V-TOUGH and hybrid. Sections 1.10.5.1 through 1.10.5.3 used qualified V-Tough code. Sections 1.10.5.4 through 1.10.7 are model results (no data) using NUFT

Chapter 2: Laboratory Determined Hydrologic Properties and Processes

LL950404604242.012 (non-Q) LL950812704242.017 (non-Q) LL960201404244.011 (non-Q) LL960100604244.007 (non-Q) LL960201304244.010 (non-Q) LL960100704244.008 (non-Q) LL950916504242.018 (non-Q) LL950406104242.016 (non-Q) LL940800704242.002 (non-Q)

Chapter 3: Geochemistry

None. Modeling only based on use of qualified EQ3/6 as well as unqualified GMIRT and REACT.



Chapter 4: Geomechanics LL960101004243.005 (non-Q) LL960401704243.008 (non-Q) LL960201104243.006 (non-Q) Modeling results based on use of non-qualified codes.

Chapter 5: Radiation Effects No new data.

Chapter 6: Man-Made Materials LL950813104245.004 (Q) LL940803904245.001 (non-Q) LL960807404245.008 (non-Q)

Chapter 7: Integrated Testing LL960401051051.007 (non-Q) LL960400951051.006 (non-Q) LL940803851051.002 (non-Q)

Chapter 8: Electrical Potentials No new data.

Chapter 9: Field Thermal Tests LL950103004244.004 (non-Q)

Chapter 10: Altered Zone, Geochemistry (10.4) Section 10.1, and 10.2 contain no data, model results only based on qualified V-TOUGH code. Section 10.3 No data Section 10.4 LL960201004241.011 (non-Q)

Chapter 11: Introduction to the Thermodynamic Data Determination Data from NEA sources.